## Compilation of 2013 pink ling Slope RAG documents from ISL

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11 November 2013

This document is a compilation of all the documents supplied by ISL to the Slope RAG during 2013 for the pink ling assessment.

Documents are generally in reverse date order (most recent documents first), except the three CPUE documents are grouped together.

The first document contains a summary of the stock assessment results from the models which the RAG agreed to use as the base models for providing management advice.

# Pink Ling assessment results 

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11 November 2013

## Summary

This report summarises the results of ISL's pink ling assessment that was presented to the Slope RAG in November 2013. It contains the final assessment results from the models which were accepted by the RAG as the base models for the eastern and western stocks. These results differ slightly from those presented to the meeting due to some corrections and additions.

For both stocks, the base models (and some sensitivities) were taken through to MCMC to provide posterior distributions for virgin biomass, current stock status (depletion), and RBC (as estimated from the ling control rule). Also, for the eastern stock, a brief risk analysis was performed.

For the eastern stock, current stock status was estimated at $25 \% B_{0}$ with a $95 \%$ CI of $17-38 \%$ $B_{0}$. The imprecision in estimated stock status was amplified in the estimation of RBC which had a $95 \%$ CI of 0-550 t. However, when a full Bayesian assessment is available to perform risk analysis, the use of a generic control rule is not needed to provide management advice on TACs.

Stochastic projections were performed for the eastern stock for a range of constant catch strategies. The model projections suggest that the stock can be rebuilt to $48 \% B_{0}$ within one mean generation time ( 8.8 years) when total removals are 250 t per year. If two mean generation times are allowed for the rebuild then total removals can be up to 400 t per year. Long-term yield is estimated at 540-640 $\mathrm{t}(95 \% \mathrm{CI})$. That is not to say that constant catches should be taken for more than a period of 2-3 years. The results merely indicate that it is safe to take constant catches in the range of 250-400 t per year until a new assessment becomes available.

For the western stock there is considerable uncertainty with regard to the strength of recent recruitment. This uncertainty flows through into the estimates of stock status and RBC (with the upper limits of the $95 \%$ CIs corresponding to very strong recent recruitment). Current stock status is estimated at $58 \% B_{0}$ with a $95 \%$ CI of $41-86 \% B_{0}$. RBC is estimated at 807 t with a $95 \%$ CI of 430-1710 t . Long-term yield is estimated at 470-960 t ( $95 \%$ CI).

## Detailed results

The following tables provide further results, some of which may wish to be presented in the AFMA documentation of the 2013 assessment. Details of the methods and run specifications can be found below (in "2013 Pink ling stock assessment: final ISL model results").

Table 1: MPD estimates: virgin biomass $\left(B_{0}\right)$, stock status ( $\boldsymbol{B}_{\text {current }} / B_{0}$ ), natural mortality $(M)$, and RBC estimates for the base models.

| Stock | $\boldsymbol{B}_{\boldsymbol{0}}(\mathbf{t})$ | Stock status <br> $\left(\boldsymbol{\%} \boldsymbol{B}_{\boldsymbol{0}}\right)$ | $\boldsymbol{M}\left(\mathbf{y r}^{\mathbf{- 1}}\right)$ | $\mathbf{R B C}(\mathbf{t})$ |
| :--- | ---: | ---: | ---: | ---: |
|  |  |  |  |  |
| East | 5550 | 23 | 0.24 | 75 |
| West | 4590 | 52 | 0.23 | 619 |

Table 2: Eastern model, MCMC: median and $\mathbf{9 5 \%}$ credibility intervals for natural mortality, virgin biomass, and current stock status for the base model and two sensitivities.

| Model | $\boldsymbol{M}\left(\mathbf{y r}^{-1}\right)$ |  |  | $\boldsymbol{B}_{\boldsymbol{0}}(\mathbf{t})$ |  | Stock stat $\left(\% \boldsymbol{B}_{\boldsymbol{0}}\right)$ |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | :---: |
|  | Median | $\mathbf{9 5 \%} \mathbf{C I}$ | Median | $\mathbf{9 5 \%} \mathbf{C I}$ | Median | $\mathbf{9 5 \%} \mathbf{C I}$ |  |
|  |  |  |  |  |  |  |  |
| Base | 0.24 | $0.22-0.26$ | 5620 | $4990-6460$ | 25 | $17-38$ |  |
| Fixed $\boldsymbol{M}$ | 0.24 |  | 5510 | $5090-6060$ | 25 | $18-33$ |  |
| Split CPUE | 0.24 | $0.22-0.26$ | 5720 | $5030-6610$ | 25 | $16-40$ |  |

Table 3: Eastern model, MCMC: estimated probabilities of current SSB exceeding reference points for the base model and two sensitivities.

| Model | $\mathbf{P}\left(\boldsymbol{B}_{2013} \geq \mathbf{2 0 \%} \boldsymbol{B}_{\mathbf{0}}\right)$ | $\mathbf{P}\left(\boldsymbol{B}_{2013} \geq \mathbf{3 5 \%} \boldsymbol{B}_{\boldsymbol{0}}\right)$ | $\mathbf{P}\left(\boldsymbol{B}_{\mathbf{2 0 1 3}} \geq \mathbf{4 8 \%} \boldsymbol{B}_{\mathbf{0}}\right)$ |
| :--- | ---: | ---: | ---: |
|  |  |  |  |
| Base | 0.85 | 0.07 | 0.00 |
| Fixed $\boldsymbol{M}$ | 0.91 | 0.01 | 0.00 |
| Split CPUE | 0.83 | 0.08 | 0.00 |

Table 4: Eastern model, MCMC: median and 95\% credibility intervals for the RBC calculated from the ling control rule and the yield when the biomass is in deterministic equilibrium at $48 \% B_{0}$.

|  | Median (t) | $\mathbf{9 5 \%}$ CI (t) |
| :--- | ---: | ---: |
|  |  |  |
| RBC | 122 | $0-550$ |
| Long-term yield | 582 | $540-640$ |

Table 5: Eastern model, MCMC, stochastic projections: performance indicators for constant catch strategies ranging from zero catch up to 500 t annual catch. "E" denotes expected value (average). " P " denotes probability. $B_{i}$ is the mid-year female spawning biomass in year $i$. The rebuild year is the first year in which at least $50 \%$ of the projections are at or above $48 \% B_{0}$.

| Annual <br> catch (t) | $\mathbf{E}\left(\mathbf{B}_{\mathbf{2 0 1 5}} / \mathbf{B}_{\mathbf{0}}\right)$ | $\mathbf{E}\left(\mathbf{B}_{\mathbf{2 0 2 0}} / \mathbf{B}_{\mathbf{0}}\right)$ | $\mathbf{P}\left(\mathbf{B}_{\mathbf{2 0 1 5}}>\mathbf{B}_{\mathbf{2 0 1 3}}\right)$ | $\mathbf{P}\left(\mathbf{B}_{\mathbf{2 0 2 0}}>\mathbf{B}_{\mathbf{2 0 1 3}}\right)$ | $\mathbf{P}\left(\mathbf{B}_{\mathbf{2 0 1 5}}<\mathbf{0 . 2}\right)$ | $\mathbf{P}\left(\mathbf{B}_{\mathbf{2 0 2 0}}<\mathbf{0 . 2}\right)$ | $\mathbf{r}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |
| $\mathbf{\text { Rear }}$ |  |  |  |  |  |  |  |
| 0 | 0.33 | 0.56 | 1.00 | 1.00 | 0.01 | 0.00 | 2019 |
| 250 | 0.30 | 0.44 | 0.98 | 0.99 | 0.04 | 0.00 | 2022 |
| 300 | 0.30 | 0.42 | 0.96 | 0.99 | 0.05 | 0.01 | 2024 |
| 350 | 0.29 | 0.39 | 0.93 | 0.97 | 0.07 | 0.02 | 2026 |
| 400 | 0.28 | 0.37 | 0.88 | 0.93 | 0.09 | 0.04 | 2029 |
| 450 | 0.28 | 0.35 | 0.82 | 0.90 | 0.11 | 0.07 | 2034 |
| 500 | 0.27 | 0.32 | 0.75 | 0.82 | 0.14 |  | 0.11 |

Table 6: Western model, MCMC: median and 95\% credibility intervals for natural mortality, virgin biomass, and current stock status for the base model and one sensitivity.

| Model | $\boldsymbol{M}\left(\mathbf{y r}^{-1}\right)$ |  |  | $\boldsymbol{B}_{\boldsymbol{0}}(\mathbf{t})$ |  | Stock status (\% $\left.\boldsymbol{B}_{0}\right)$ |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | :---: |
|  | Median | $\mathbf{9 5 \%} \mathbf{C I}$ | Median | $\mathbf{9 5 \%} \mathbf{C I}$ | Median | $\mathbf{9 5 \%} \mathbf{C I}$ |  |
|  |  |  |  |  |  |  |  |
| Base | 0.23 | $0.20-0.27$ | 5130 | $4030-6730$ | 58 | $41-86$ |  |
| Fixed $\boldsymbol{M}$ | 0.24 |  | 4830 | $4150-5940$ | 56 | $46-67$ |  |

Table 7: Western model, MCMC: estimated probabilities of current SSB exceeding reference points for the base model and two sensitivities.

| Model | $\mathbf{P}\left(\boldsymbol{B}_{2013} \geq \mathbf{2 0 \%} \boldsymbol{B}_{\mathbf{0}}\right)$ | $\mathbf{P}\left(\boldsymbol{B}_{2013} \geq \mathbf{3 5 \%} \boldsymbol{B}_{\boldsymbol{0}}\right)$ | $\mathbf{P}\left(\boldsymbol{B}_{2013} \geq \mathbf{4 8 \%} \boldsymbol{B}_{\boldsymbol{0}}\right)$ |
| :--- | ---: | ---: | ---: |
|  |  |  |  |
| Base | 1.00 | 1.00 | 0.86 |
| Fixed $\boldsymbol{M}$ | 1.00 | 1.00 | 0.95 |

Table 8: Western model, MCMC: median and $95 \%$ credibility intervals for the RBC calculated from the ling control rule and the yield when the biomass is in deterministic equilibrium at $48 \% B_{0}$.

|  | Median (t) | $\mathbf{9 5 \%}$ CI (t) |
| :--- | ---: | ---: |
|  |  |  |
| RBC | 807 | $430-1710$ |
| Long-term yield | 661 | $470-960$ |

# 2013 Pink ling stock assessment: final ISL model results 

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## Executive Summary

This document presents the final assessment results for ISL's assessment of pink ling in 2013. Methods and previous work on data preparation, including CPUE analysis for the trawl fishery, are summarised.

The base models presented conform with the base-model specifications agreed at the September RAG meeting. Final details of the base model specifications were agreed by email and include the use of time-blocking for the eastern trawl CPUE series (with the top three linking vessels used) and the exclusion of the non-trawl CPUE indices (for both stocks).

For both stocks, the base models (and some sensitivities) were taken through to MCMC to provide posterior distributions for virgin biomass, current stock status (depletion), and RBC (as estimated from the ling control rule). Also, for the eastern stock, a brief risk analysis was performed.

For the eastern stock, current stock status was estimated at $25 \% B_{0}$ with a $95 \%$ CI of $17-38 \%$ $B_{0}$. The imprecision in estimated stock status was amplified in the estimation of RBC which had a $95 \%$ CI of 0-550 t. However, when a full Bayesian assessment is available to perform risk analysis, the use of a generic control rule is not needed to provide management advice on TACs.

Stochastic projections were performed for the eastern stock for a range of constant catch strategies from 250 t to 500 t per year. The model projections suggest that the stock can be rebuilt to $48 \% B_{0}$ within one mean generation time ( 8.8 years) when total removals are 250 t per year. If two mean generation times are allowed for the rebuild then total removals can be up to 400 t per year. Long-term yield is estimated at 540-640 t ( $95 \%$ CI).

For the western stock there is considerable uncertainty with regard to the strength of recent recruitment. This uncertainty flows through into the estimates of stock status and RBC (with the upper limits of the $95 \%$ CIs corresponding to very strong recent recruitment). Current stock status is estimated at $58 \% B_{0}$ with a $95 \%$ CI of $41-86 \% B_{0}$. RBC is estimated at 807 t with a $95 \%$ CI of 430-1710 t . Long-term yield is estimated at 470-960 t ( $95 \%$ CI).

## Introduction

This document describes the model results for the pink ling assessment undertaken by ISL in 2013. The stock assessment methods, including data preparation and the derivation of CPUE indices, are summarised (details are given in earlier Slope RAG documents: Cordue, drafts 16 , Cordue \& Punt, draft). The ISL base models for the eastern and western stocks conform with the base-model specifications agreed at the September Slope RAG meeting. The base models and some sensitivities are taken through to MCMC.

Some additional work done during the November RAG meeting is given in Appendix 5.

## Methods

## Data preparation

For the eastern and western stocks, catch histories (trawl, non-trawl) and length frequency data by method (trawl, line), zone, sampling type (port, onboard), and depth stratum (0$300 \mathrm{~m}, 300-500 \mathrm{~m}, 500+\mathrm{m}$ )(samples scaled to individual catches) were supplied by CSIRO. Raw age-length data were also supplied (for which standard length measurements were converted to total length where appropriate).

The length frequency data were stratified and scaled following earlier recommendations (Cordue, draft 1). In the east, stratification was by depth and zone for the trawl data (which precluded the use of port samples). For the west, and for line in the east, there was no stratification (but the samples for some years were omitted where there were very few operations and/or fish).

The age-length data were also stratified and scaled for eastern trawl as recommended earlier (Cordue, drafts 2, 4). Non-sexed age-length data (almost all from Zone 20) were used to construct age-length keys which were applied to the corresponding length frequencies to produce age frequencies for the eastern assessment.

The tables below give the years for which composition data were used in the base models (years with port-sampled length-data are in red).

|  | Length frequencies |  | Conditional age-length |  |  | Age frequencies |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :---: |
|  | Trawl | Line | Trawl | Line | Trawl | Line |  |
| East |  |  | 1979 |  |  |  |  |
|  |  |  |  |  |  |  |  |
|  |  |  | 1994 |  |  |  |  |
|  |  |  | 1995 |  |  |  |  |
|  |  |  | 1996 |  |  |  |  |
|  |  |  |  |  |  |  |  |


|  | 1998 |  | 1998 |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 1999 |  | 1999 |  |  |  |
|  | 2000 |  |  |  |  |  |
|  | 2001 |  | 2001 |  |  |  |
|  | 2002 | 2002,2002 |  |  |  |  |
|  | 2003 | 2003 |  | 2003 |  |  |
|  |  | 2004,2004 |  | 2004 |  |  |
|  | 2005 | 2005 | 2005 |  |  |  |
|  | 2006 | 2006,2006 |  |  |  |  |
|  |  | 2007 |  |  |  |  |
|  |  |  |  |  |  | 2008 |
|  |  |  |  | 2009 |  | 2009 |
|  |  |  | 2010 |  | 2010 | 2010 |
|  |  |  | 2011 | 2011 | 2011 | 2011 |
|  |  | 2012 |  | 2012 | 2012 |  |


|  | Length frequencies |  | Conditional age-length |  |
| :--- | :--- | :--- | :--- | :--- |
|  | Trawl | Line | Trawl | Line |
| West |  |  | 1987 |  |
|  | 1992 |  |  |  |
|  | 1993 |  |  |  |
|  | 1994 |  |  |  |
|  | 1995,1995 |  | 1995 |  |
|  | 1996,1996 |  |  |  |
|  | 1997,1997 |  | 1997 |  |
|  | 1998,1998 |  | 1998 |  |
|  | 1999,1999 |  | 1999 |  |
|  | 2000,2000 |  |  |  |
|  | 2001,2001 | 2001 |  | 2001 |
|  | 2002,2002 | 2002 | 2002 |  |
|  | 2003,2003 | 2003 | 2003 | 2003 |
|  | 2004,2004 | 2004 | 2004 | 2004 |
|  | 2005,2005 | 2005 | 2005 | 2005 |
|  | 2006,2006 | 2006 | 2006 | 2006 |
|  | 2007 | 2007 |  | 2007 |
|  |  |  |  |  |
|  | 2009 |  |  |  |
|  | 2010 | 2010 | 2010 | 2010 |
|  | 2011 |  | 2011 |  |
|  | 2012 | 2012 | 2012 | 2012 |

## CPUE indices

The results of ISL's CPUE analysis of trawl data were presented at the September RAG meeting (Cordue, draft 3). The analysis filtered catch records based on depth and catches of species (other than ling) in an attempt to define a consistent "fishery" in which ling was a
bycatch species (although ling targeting was not explicitly excluded). Also, the issue of potential changes in ling catchability from 1999 to 2000 (in the east, due to the sale of ling quota from trawl operators to line operators) and from 2006 to 2007 (structural adjustment) was addressed by time-blocking of vessel effects. The indices produced by ISL were almost identical to those produced by CSIRO for the western fishery, but they were quite different in the east.

After the September RAG meeting, further analysis was done for the eastern fishery (Cordue, draft 5). It was shown that the results were sensitive to the filters that were used and to the choice of 'linking vessels" (vessels assumed to have constant vessel effects across two consecutive time blocks). Nevertheless, it was agreed (by email) to use an eastern CPUE time series, with time-blocking and linking vessels, in the base stock assessment model. The chosen indices had minimal filtering and used three linking vessels chosen objectively (Cordue, draft 6). The base indices are intermediate between indices produced by using two or four objectively chosen linking vessels. Also, the base indices are not very different from the indices produced by CSIRO (see Cordue, draft 6).

In the western assessment, the CSIRO indices were used in ISL's base model.

It was agreed, by email, that the CPUE indices for the line fisheries, produced by CSIRO, would not be used in the base models because they had limited spatial coverage (when minimal requirements were placed on an acceptable number of records in each block, e.g., see Table 5 in Whitten et al. draft).

## Model structure

A single-area model with a single time-step was used for both stocks. Ages (1-30+), sex, and maturity were in the partition (although the latter is irrelevant as there were no fisheries that preferentially selected mature fish). The two fisheries (trawl and non-trawl) were assumed to be year-round and mortality was modelled using the Pope approximation to Baranov (CASAL's standard option). Further details of the models are:

| Model years | $1970-2013$ | Stock status assessed mid- <br> year 2013 |
| :--- | :--- | :--- |
| Biomass parameterisation | $B_{0}$ | Estimated parameter. $R_{0}$ is <br> derived. |
| Recruitment parameterisation | Haist, lognormal prior, <br> sigmaR $=0.7$ | Also, a moderate penalty on <br> year class strengths (YCS) <br> averaging to 1. |
| YCS estimated (i.e., <br> recruitment deviations) | East: 1969-1977, 1983-2009 <br> West: 1975-2009 initially, <br> but 2009 dropped as its very <br> high estimate was based on <br> little data. | Required to have at least 8 <br> observations in the sexed <br> age-length data used |


| Steepness | 0.75 | As used in 2012. A conservative value - it could be higher. Fixed. |
| :---: | :---: | :---: |
| Maturity | Logistic at age: $\mathrm{a}_{50}=5 \mathrm{yr}, \text { ato } 95=2 \mathrm{yr}$ | Approximates the lengthbased curve used in the 2012 assessment. Fixed. |
| Trawl selectivities | Three blocks in the east: 1970-99, 2000-2006, 20072013. Two in the west: 19702006, 2007-2013. Double normal at age, same for males and females. | Estimated in the model. Timing of blocks indicated by events and confirmed by data analysis. <br> Separate male and female selectivities in a sensitivity. |
| Non-trawl selectivities | Logistic at age, same for males and females. | Estimated in the model. Separate male and female selectivities in a sensitivity. |
| Growth | Separate male and female von Bertalanffy | Estimated in the model. |
| Length-weight relationship | $\begin{array}{\|l\|} \hline \text { a } 2.93 \mathrm{e}-9 \\ \text { b } 3.139 \\ \hline \end{array}$ | Fixed at 2012 assessment values. (cm to tonnes) |

## MPD methods

## Model runs

The base models had the model structure described above and estimated $M$. For the western model the prior on $M$ was $\mathrm{N}($ mean $=0.2, \mathrm{CV}=0.2)$. For the east, the prior on $M$ was taken from an early portion of the western MCMC chains and was $\mathrm{N}($ mean $=0.22, \mathrm{CV}=0.06$ ). See "Estimation of natural mortality" below.

For both assessments, MPD sensitivities were done at fixed $M$ (low $=0.2$, medium $=0.24$, high $=0.28$ ), alternative maturity ogives (shifted up or down one year), and SSB defined as beginning of year biomass (rather than mid-year). Retrospective models for the 2012 assessment were also done for both stocks as were models which included the non-trawl CPUE indices ( $\mathrm{CV}=0.2$ ). For the east there were additional sensitivities: include Kapala biomass indices and length frequencies; use CSIRO trawl CPUE indices; use two-linkingvessel trawl CPUE; and split the fully-linked (no time blocks) trawl CPUE into two series 1986-1999, 2000-2012 with a prior on the q-ratio, $\mathrm{LN}($ mean $=1, \mathrm{CV}=0.2$ ).

## Data weighting

The CVs for the trawl CPUE indices were determined following the recommendation of Francis (2011) to fit a smoother outside the model and estimate the CV from the residuals (see Appendix 1).

Length data were modelled with multinomial distributions with effective sample sizes initially assumed to be equal to the number of operations (onboard sampling) or landings (port sampling). Age frequencies had a starting sample size equal to the number of otoliths
divided by 10 . The age-length data had initial sample sizes equal to the number of otoliths. The effective sample sizes were tuned following the iterative reweighting approach of Francis (2011) and as extended to conditional age data in Appendix 1. However, the age-length data were not fully tuned as this would have created problems estimating growth within the model (i.e., effective sample sizes would have been so low that growth would have been misspecified). The age-length data were only thinned by a factor of 4 and, for the line data, the top $15 \%$ of fish, by length, were retained (to help with growth estimation).

The ranges for initial sample sizes and the final sample sizes (after tuning) are given below.

| Data set | Initial range | Final range |
| :--- | ---: | ---: |
|  |  |  |
| East |  |  |
| LF trawl | $33-82$ | $55-118$ |
| LF line | $8-78$ | $3-27$ |
| LF line (port) | $2-10$ | $7-33$ |
| AF trawl | $45-53$ | $5-6$ |
| AF line | $33-59$ | $3-7$ |
| Age-length trawl | $72-707$ | $16-170$ |
| Age-length line | $57-309$ | $13-76$ |
|  |  |  |
| West |  |  |
| LF trawl | $9-49$ | $8-42$ |
| LF trawl (port) | $4-29$ | $4-28$ |
| LF line | $5-40$ | $4-33$ |
| Age-length trawl | $92-528$ | $23-132$ |
| Age-length line | $40-370$ | $9-92$ |

The biggest change from initial effective sample sizes was for the age frequencies which, in terms of mean age, were not fitted well by the model. The Francis method essentially determines effective sample sizes on the basis of the fit to annual mean lengths or ages (the fits for each data set are given in Appendix 2).

## Estimation of natural mortality ( $M$ )

The initial MPD estimate of $M$ for the east was rather high ( 0.27 ) and the likelihood profile suggested it was sensitive to the weights used for the age-length data and the trawl length frequencies (Appendix 3). It was concluded that $M$ would not be well estimated in the eastern model. However, for the western model the MPD estimate of $M$ seemed sensible for ling (0.23) and from the likelihood profile it looked like it was relatively insensitive to the data weights (Appendix 3). Therefore, it was decided to first estimate $M$ in the western assessment and then use the western posterior distribution as a prior for $M$ in the eastern assessment. As it happened, there was insufficient time to wait for the western MCMC to finish and the prior for $M$ was chosen from the early part of the chains (after a burn-in period). The prior used for the eastern base model was $M \sim \mathrm{~N}$ (mean $=0.22, \mathrm{CV}=0.06)$.

## Estimation of sex-specific selectivities

Initial eastern models used sex-specific fishing selectivities. However, the eastern trawl selectivities varied greatly across time blocks (Appendix 3). Runs were done using sexindependent selectivities (i.e., the same selectivities for males and females) and the likelihoods and age-length residuals were examined. There was no apparent residual pattern by sex in the age-length data and the improvement in likelihood when using sex-specific selectivities did not justify the use of the extra parameters (Appendix 3). Therefore, sexindependent selectivities were used for the eastern and western base models.

## Estimation of RBC and long-term yield

The pink ling control rule was applied using the MPD estimate of parameters and assuming that the exploitation rates of the trawl and non-trawl fisheries were maintained in the same proportion as was estimated for 2013. $F_{48}$ was determined by scaling the current exploitation rates (up or down) until the corresponding long-term equilibrium SSB, under deterministic recruitment, was equal to $48 \% B_{0}$ (i.e., $B_{48}$ ). The estimate of RBC was taken to be the catch in 2014 associated with $F_{\text {target }}$ where:

$$
\begin{aligned}
F_{\text {target }} & = & F_{48} \\
& = & F_{48}\left(B_{\text {current }}-0.2\right) / 0.15 \\
& = & 0
\end{aligned}
$$

$$
\begin{aligned}
& \text { for } B_{\text {current }} \geq B_{35} \\
& \text { for } B_{20}<B_{\text {current }}<B_{35} \\
& \text { for } B_{\text {current }} \leq B_{20} \text {. }
\end{aligned}
$$

The long-term yield was estimated as the catch associated with $F_{48}$ when the stock reached deterministic equilibrium.

## MCMC methods

For the east, three models were taken through to MCMC: base, fixed $M=0.24$, and split trawl-CPUE. For the west, there were two MCMC models: base, fixed $M=0.24$.

Three independent chains of at least 1 million were run for each model (much longer chains were needed for the western base ( 3 million) and split trawl-CPUE ( 5 million)). One in every one thousand samples were retained. A burn-in length was determined for each model using plots of the objective function for the three chains, looking for when the chains fluctuated over a consistent range. For each model, histograms of the three chains (with the burn-in removed) showed good agreement between the main variables of interest (virgin biomass, stock status and natural mortality) except for the western base and the split-trawl CPUE (these were not well converged for current stock status - but they were "adequately" converged in that much longer chains would be unlikely to change the median estimated stock status). A single chain for each model was formed by combining the three chains with the burn-in removed. The chains for the western base model and the split trawl-CPUE were thinned to retain 3000 samples.

## Estimation of RBC and long-term yield

The methods applied to the MPD estimates were applied to each sample from the final chain. This was done by running deterministic projections at different levels of scaled exploitation rates (assuming the trawl and non-trawl proportions at the MPD) and determining $F_{48}$ and the associated catches, in 2014, and at equilibrium (by linear interpolation). The control rule was then applied to determine the RBC estimate for each member of the chain (and hence to an RBC posterior distribution).

## Projections and risk assessment

For the eastern base model, short and long-term projections (to 2050) were performed for constant catch strategies at zero catch and from 250 t to 500 t (split between trawl and nontrawl in proportion to the MPD exploitation rates in 2013). These were full stochastic projections where year class strengths from 2010 onwards were randomly sampled from the last 10 years that were estimated (2000-2009) - to reflect recent recruitment, which may be a little less than long-term average recruitment.

Six performance indicators were calculated for each constant catch strategy:

- $\mathrm{E}\left(\boldsymbol{B}_{2015} / \boldsymbol{B}_{0}\right)$ : mean stock status in 2015
- $\mathrm{E}\left(\boldsymbol{B}_{2020} / \boldsymbol{B}_{0}\right)$ : mean stock status in 2020
- $\mathrm{P}\left(B_{2015}>B_{2013}\right): \quad$ probability that SSB in 2015 is greater than that in 2013
- $\mathrm{P}\left(B_{2020}>B_{2013}\right): \quad$ probability that SSB in 2020 is greater than that in 2013
- $\mathrm{P}\left(B_{2015}>0.2\right): \quad$ probability that SSB in 2015 is greater than $20 \% B_{0}$
- $\mathrm{P}\left(B_{2020}>0.2\right): \quad$ probability that SSB in 2020 is greater than $20 \% B_{0}$

Also, the year in which there was at least a $50 \%$ probability of SSB exceeding $48 \% B_{0}$ was determined for each constant catch strategy. To put the timeframe required to rebuild to the target in context, the mean generation time for ling was calculated: the average age of a mature fish in an unexploited population (assuming $M=0.24$ and the base-model maturity assumption) - which was 8.8 years.

## Results

## Eastern stock

## MPD results

The model provided an excellent fit to the trawl CPUE time series (Figure 1). The estimated growth curves were consistent with the sexed age-length data from the line fishery (Figure 2). The trawl selectivities were estimated to be highly domed with the mode shifting higher over each time period ( 2.8 years, 3.1 years, 4.5 years)(Figure 3). The trawl-selectivity modes are at younger ages than the assumed $50 \%$ maturity at 5 years and the estimated $50 \%$ selection for the non-trawl fishery of 5.9 years. The selection for non-trawl port-sampled fish showed somewhat younger fish were sampled at port compared to those sampled at sea (Figure 4).

The estimated year class strengths (YCS) showed moderate variability (standard deviation of $\log \mathrm{YCS}=0.54$ ) with weak recent YCS in 2005 and 2009 (Figure 5).

Estimated stock status (depletion) for the MPD base model is just above 20\% $B_{0}$ in 2013 (Table 1, Figure 6). The MPD estimates of virgin biomass and stock status are sensitive to $M$ (as expected) and also to the use of sex-independent selectivities as opposed to sex-specific selectivities. The latter result is because of the paucity of data with which to estimate sexspecific selectivities for the trawl fishery in three different time blocks. The high estimated stock status for the sex-specific selectivities model is because a very domed trawl selectivity for females is estimated for 2007-2013 (see Appendix 3). This means that females are relatively invulnerable to trawl fishing during this period and therefore SSB (which is female only) increases quickly (Figure 6).

The estimate of $M$ is very stable (Table 1) due to the very strong prior placed on it. It cannot be considered a reliable estimate of $M$ because the strong prior was constructed from the early chains in the western base model before they had converged (although they were consistent at the time). The base model is almost a fixed- $M$ model, but for the MCMC estimation it is preferable to use it instead of a fixed- $M$ model as it allows $M$ to move (at least a little bit).

Table 1: Eastern stock, MPD estimates: virgin biomass $\left(\boldsymbol{B}_{0}\right)$, stock status ( $\left.\boldsymbol{B}_{\text {current }} / \boldsymbol{B}_{0}\right)$, and natural mortality $(M)$ estimates for the base model and sensitivity runs. *This differs from the result in Whitten et al. draft (which was in error).

| Model run | $\boldsymbol{B}_{\boldsymbol{0}}(\mathbf{t})$ | Stock status <br> $\left(\% \boldsymbol{B}_{\boldsymbol{0}}\right)$ | $\boldsymbol{M}\left(\mathbf{y r}^{\mathbf{- 1}}\right)$ |
| :--- | ---: | ---: | ---: |
|  |  |  |  |
| Base | 5550 | 23 | 0.24 |
| Sex-specific selectivities | 5820 | 31 | 0.23 |
| Fixed and lower $M$ | 6830 | 17 | 0.20 |
| Fixed and higher $M$ | 4670 | 33 | 0.28 |
| Maturity 1 year younger | 6010 | 27 | 0.23 |
| Maturity 1 year older | 5090 | 20 | 0.24 |
| Including line CPUE | 5490 | 18 | 0.24 |
| Including Kapala data | 5510 | 23 | 0.23 |
| CSIRO trawl CPUE | 5420 | 22 | 0.24 |
| Two-linking-vessel CPUE | 5660 | 27 | 0.24 |
| Split fully-linked CPUE | 5550 | 23 | 0.24 |
| Split fully-linked CPUE (no prior) | 5570 | 24 | 0.24 |
| Beginning-of-year SSB | 6140 | 26 | 0.24 |
|  |  |  |  |
| Retrospective 2012 | 5570 | 23 | 0.24 |
| SS3 2012 assessment | 6930 | 19 | 0.24 |

The MPD estimate of RBC for the eastern base model is just 75 t . This is despite the current exploitation rates corresponding to $F_{40}$ and the catch at $F_{48}$ being 337 t . The penalty in the control rule is very severe because the MPD estimate of stock status is close to $20 \% B_{0}$.

## MCMC results

For the base model, the non-trawl fishing selectivity and the corresponding port-sampling selection pattern are reasonably tightly estimated (Figure 8). The same is true for the trawlfishing selectivity pattern in the second time block, but the selectivities in the other time blocks are less well estimated (Figure 8). In particular the selectivity pattern in the most recent time block is very imprecise. The pattern of MCMC estimated year class strengths is similar to that in the MPDs (Figure 9).

In the base model, virgin biomass (mid-year, female only) is estimated to be about 5500 t (Table 2, Figure 10). Current stock status (depletion) is estimated at $25 \% B_{0}$ with a fairly broad credibility interval (Table 2, Figure 11).

Table 2: Eastern model, MCMC: median and $\mathbf{9 5 \%}$ credibility intervals for natural mortality, virgin biomass, and current stock status for the base model and two sensitivities.

| Model | $\boldsymbol{M}\left(\mathbf{y r}^{-1}\right)$ |  |  | $\boldsymbol{B}_{\boldsymbol{0}}(\mathbf{t})$ |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Stock stat $\left(\% \boldsymbol{B}_{\boldsymbol{0}}\right)$ |  |  |  |  |  |  |
|  | Median | $\mathbf{9 5 \%} \mathbf{C I}$ | Median | $\mathbf{9 5 \%} \mathbf{C I}$ | Median | $\mathbf{9 5 \%} \mathbf{C I}$ |
|  |  |  |  |  |  |  |
| Base | 0.24 | $0.22-0.26$ | 5620 | $4990-6460$ | 25 | $17-38$ |
| Fixed $\boldsymbol{M}$ | 0.24 |  | 5510 | $5090-6060$ | 25 | $18-33$ |
| Split CPUE | 0.24 | $0.22-0.26$ | 5720 | $5030-6610$ | 25 | $16-40$ |

The base-model SSB trajectory shows median estimated SSB at a bit above $20 \% B_{0}$ since the mid 2000s with an increase in the last two years (Figure 12). The distribution of SSB (as a percentage of $B_{0}$ ) in each year of the trajectory is quite broad. According to the base model and sensitivities, current biomass is very likely to be above the limit reference point of $20 \%$ $B_{0}$ and very unlikely to be above the target of $48 \% B_{0}$ (Table 3). There is more than a $2.5 \%$ chance of the current biomass being below $20 \% B_{0}$ and of being above $35 \% B_{0}-$ which is of relevance to the imprecision to be expected in the estimation of RBC.

Table 3: Eastern model, MCMC: estimated probabilities of current SSB exceeding reference points for the base model and two sensitivities.

| Model | $\mathbf{P}\left(\boldsymbol{B}_{2013} \geq \mathbf{2 0 \%} \boldsymbol{B}_{0}\right)$ | $\mathbf{P}\left(\boldsymbol{B}_{2013} \geq \mathbf{3 5 \%} \boldsymbol{B}_{0}\right)$ | $\mathbf{P}\left(\boldsymbol{B}_{2013} \geq \mathbf{4 8 \%} \boldsymbol{B}_{0}\right)$ |
| :--- | ---: | ---: | ---: |
|  |  |  |  |
| Base | 0.85 | 0.07 | 0.00 |
| Fixed $\boldsymbol{M}$ | 0.91 | 0.01 | 0.00 |
| Split CPUE | 0.83 | 0.08 | 0.00 |

The RBC calculated from the ling control rule is very poorly estimated with the $95 \%$ CI extending from $0 t$ to over $500 t$ (Table 4, Figure 13). This is primarily because of the uncertainty in current stock status which has a $95 \%$ CI extending from below $20 \% B_{0}$ to over $35 \% B_{0}$ (which are defining limits in the control rule).

Long-term yield appears well-estimated and is in the region of about 600 t (Table 4).
Table 4: Eastern base model, MCMC: median and 95\% credibility intervals for the RBC calculated from the ling control rule and the yield when the biomass is in deterministic equilibrium at $48 \% B_{0}$.

|  | Median (t) | $\mathbf{9 5 \%}$ CI (t) |
| :--- | ---: | ---: |
|  |  |  |
| RBC | 122 | $0-550$ |
| Long-term yield | 582 | $540-640$ |

The median RBC is just 122 t , but it is only of academic interest as the RBC is so poorly estimated. In any case, it is inappropriate to base management advice on the output from a generic control rule when a full Bayesian assessment is available to perform a risk assessment.

It is very likely that the eastern stock is below the management target of $48 \% B_{0}$ and hence future total removals need to be below the long-term yield to allow SSB to rebuild towards the target. According to the base model, this can be done in a suitable timeframe (e.g., no more than 2 times the mean generation time of 8.8 years) with a constant annual catch of up to 400 t (Table 5). If total removals are 250 t per year then the rebuild would be expected in about one mean generation time (Table 5).

Table 5: Eastern base model, MCMC: performance indicators for stochastic projections using constant-catch rebuilding strategies (the "rebuild year" corresponds to a $\mathbf{5 0 \%}$ probability of being at or above $48 \% B_{0}$ ).

| Annual <br> catch $(\mathbf{t})$ | $\mathbf{E}\left(\mathbf{B}_{2015} / \mathbf{B}_{0}\right)$ | $\mathbf{E}\left(\mathbf{B}_{2020} / \mathbf{B}_{0}\right)$ | $\mathbf{P}\left(\mathbf{B}_{2015}>\mathbf{B}_{2013}\right)$ | $\mathbf{P}\left(\mathbf{B}_{2020}>\mathbf{B}_{2013}\right)$ | $\mathbf{P}\left(\mathbf{B}_{2015}<\mathbf{0 . 2}\right)$ | $\mathbf{P}\left(\mathbf{B}_{2020}<\mathbf{0 . 2}\right)$ | Rebuild <br> year |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |
| 0 | 0.33 | 0.56 | 1.00 | 1.00 | 0.01 | 0.00 | 2019 |
| 250 | 0.30 | 0.44 | 0.98 | 0.99 | 0.04 | 0.00 | 2022 |
| 300 | 0.30 | 0.42 | 0.96 | 0.99 | 0.05 | 0.01 | 2024 |
| 350 | 0.29 | 0.39 | 0.93 | 0.97 | 0.07 | 0.02 | 2026 |
| 400 | 0.28 | 0.37 | 0.88 | 0.93 | 0.09 | 0.04 | 2029 |
| 450 | 0.28 | 0.35 | 0.82 | 0.90 | 0.11 | 0.07 | 2034 |
| 500 | 0.27 | 0.32 | 0.75 | 0.82 | 0.14 | 0.11 | 2047 |

## Western stock

## MPD results

The model provided a poor fit to the trawl CPUE time series. The predictions follow the same trend but fail to reach the height of the peaks or the depths of the troughs (Figure 15). The estimated growth curves are consistent with the sexed age-length data from the line fishery (Figure 16). The trawl selectivities are estimated to be slightly domed with the mode shifting higher between the two time periods ( 3.9 years, 4.9 years)(Figure 17). The trawl-selectivity modes are at younger ages than the assumed $50 \%$ maturity at 5 years and the estimated $50 \%$ selection for the non-trawl fishery of 6.0 years. The selection for trawl-caught port-sampled fish is highly domed in comparison to the at-sea sampling (Figure 17).

The estimated year class strengths (YCS) showed low to moderate variability (standard deviation of $\log \mathrm{YCS}=0.41$ ) with some relatively recent good recruitment from the 2003 and 2004 cohorts (Figure 18).

Estimated stock status (depletion) for the MPD base model is just above $50 \% B_{0}$ in 2013 (Table 6, Figure 19). The MPD estimates of virgin biomass and stock status show some sensitivity to $M$ (as expected) but are fairly robust to the other scenarios (Table 6, Figure 20). The estimate of $M$ is very robust to the scenarios investigated (Table 6).

Table 6: Western stock, MPD estimates: virgin biomass ( $\boldsymbol{B}_{0}$ ), stock status ( $\boldsymbol{B}_{\text {current }} / \boldsymbol{B}_{0}$ ), and natural mortality $(M)$ estimates for the base model and sensitivity runs.

| Model run | $\boldsymbol{B}_{\boldsymbol{0}}(\mathbf{t})$ | Stock status <br> $\left(\% \boldsymbol{B}_{\boldsymbol{0}}\right)$ | $\boldsymbol{M}\left(\mathbf{y r}^{\mathbf{- 1}}\right)$ |
| :--- | ---: | ---: | ---: |
|  |  |  |  |
| Base | 4590 | 52 | 0.23 |
| Sex-specific selectivities | 4600 | 53 | 0.23 |
| Fixed and lower $M$ | 4450 | 45 | 0.20 |
| Fixed and higher $M$ | 6070 | 65 | 0.28 |
| Maturity 1 year younger | 5130 | 56 | 0.23 |
| Maturity 1 year older | 4040 | 49 | 0.23 |
| Including line CPUE | 4550 | 51 | 0.23 |
| Beginning-of-year SSB | 5080 | 55 | 0.23 |
| Trawl CPUE CV 10\% | 4550 | 55 | 0.23 |
|  |  |  |  |
| Retrospective 2012 | 4320 | 49 | 0.23 |
| SS3 2012 assessment | 4710 | 43 | 0.21 |

The poor fit the trawl CPUE indices was investigated to see if it could be improved by specifying a constant CV of $10 \%$ in a sensitivity run (in the base model the CV is $15 \%$ ). This improved the fit to the CPUE indices somewhat but not dramatically (Figure 21). It had little effect on the estimates of virgin biomass and stock status (Table 6).

The MPD estimate of current exploitation rates corresponds to $F_{58}$ and the estimate of RBC is 619 t (being the 2014 catch at 1.4 times the current exploitation rates). The MPD estimate of long-term yield is 578 t .

## MCMC results

For the base model, the non-trawl fishing selectivity is reasonably tightly estimated (Figure 22). The same is not true for the trawl-fishing selectivity patterns or the port sampling of trawl-caught fish (Figure 22). The trawl selectivities may or may not be domed, although it seems likely that the port-sampling selectivity is more domed than the at-sea fishing selection pattern.

The pattern of MCMC estimated year class strengths is similar to that in the MPDs except that the 2008 YCS is very imprecise - it could be very strong or well below average (Figure 23). The uncertainty in the 2008 YCS is the major cause of uncertainty with regard to current stock status and RBC.

Natural mortality appears to be well estimated, with the median at 0.23 and a fairly tight credibility interval (Table 7, Figure 24). The prior used for $M$ in the eastern base model was N (mean $=0.22, \mathrm{CV}=0.06$ ); had there been time to wait for the western model to finish, the
prior would have been $\mathrm{N}($ mean $=0.23, \mathrm{CV}=0.07)$, which would have given slightly more optimistic results for the eastern stock.

In the base model, virgin biomass (mid-year, female only) is estimated to be about 5000 t (Table 7, Figure 25). Current stock status (depletion) is estimated at $58 \% B_{0}$ with a very broad credibility interval (Table 7, Figure 26).

Table 7: Western model, MCMC: median and $95 \%$ credibility intervals for natural mortality, virgin biomass, and current stock status for the base model and one sensitivity.

| Model | $\boldsymbol{M}\left(\mathbf{y r}^{\mathbf{- 1}}\right)$ |  |  | $\boldsymbol{B}_{\boldsymbol{0}}(\mathbf{t})$ |  | Stock status $\left(\% \boldsymbol{B}_{\boldsymbol{0}}\right)$ |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | :---: |
|  | Median | $\mathbf{9 5 \%} \mathbf{C I}$ | Median | $\mathbf{9 5 \%} \mathbf{C I}$ | Median | $\mathbf{9 5 \%} \mathbf{C I}$ |  |
|  |  |  |  |  |  |  |  |
| Base | 0.23 | $0.20-0.27$ | 5130 | $4030-6730$ | 58 | $41-86$ |  |
| Fixed $\boldsymbol{M}$ | 0.24 |  | 4830 | $4150-5940$ | 56 | $46-67$ |  |

The base-model SSB trajectory shows median estimated SSB at about $60 \% B_{0}$ for the last five years (Figure 27). According to the base model and the sensitivity, current biomass is very likely to be above the target reference point of $48 \% B_{0}$ (Table 8). There is a zero probability estimated for the biomass to be below $35 \% B_{0}$.

Table 8: Western model, MCMC: estimated probabilities of current SSB exceeding reference points for the base model and two sensitivities.

| Model | $\mathbf{P}\left(\boldsymbol{B}_{2013} \geq \mathbf{2 0 \%} \boldsymbol{B}_{\mathbf{0}}\right)$ | $\mathbf{P}\left(\boldsymbol{B}_{2013} \geq \mathbf{3 5 \%} \boldsymbol{B}_{\boldsymbol{0}}\right)$ | $\mathbf{P}\left(\boldsymbol{B}_{2013} \geq \mathbf{4 8 \%} \boldsymbol{B}_{\mathbf{0}}\right)$ |
| :--- | ---: | ---: | ---: |
|  |  |  |  |
| Base | 1.00 | 1.00 | 0.86 |
| Fixed $\boldsymbol{M}$ | 1.00 | 1.00 | 0.95 |

The RBC calculated from the ling control rule is imprecisely estimated with the $95 \%$ CI extending from 430 t to over 1700 t (Table 9, Figure 28). This is mainly due to the large uncertainty in current stock status (because of uncertainty in recent recruitment). Long-term yield is also fairly imprecise with the $95 \%$ CI from about 500-1000 t (Table 9, Figure 29).

Table 9: Western base model, MCMC: median and $\mathbf{9 5 \%}$ credibility intervals for the RBC calculated from the ling control rule and the yield when the biomass is in deterministic equilibrium at $48 \%$ B $_{0}$.

|  | Median (t) | $\mathbf{9 5 \%}$ CI (t) |
| :--- | ---: | ---: |
|  |  |  |
| RBC | 807 | $430-1710$ |
| Long-term yield | 661 | $470-960$ |

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Figure 1: Eastern base model MPD-fit to trawl CPUE indices. Dashed lines are 95\% CIs.


Figure 2: Eastern base model MPD: "fit" to age-length data from the line fishery. Model growth curves shown as solid lines. Lowess fits to data shown as dashed lines.


Figure 3: Eastern base model MPD: estimated fishery selectivities (age-based, sex independent, three time blocks for trawl).


Figure 4: Eastern base model MPD: estimated fishery selectivity for the non-trawl fishery and the corresponding ogive for the port sampled non-trawl fish.


Figure 5: Eastern base model MPD: estimated "true year class strengths" (annual recruitment divided by virgin recruitment).


Figure 6: Eastern base model MPD: estimated stock status (annual SSB divided by virgin SSB; female biomass only; mid-year). Lines at $0.2,0.35$, and 0.48 .


Figure 7: Eastern base model MPD: estimated stock status for the base model and sensitivity runs. Lines at $0.2,0.35$, and 0.48 .


Figure 8: Eastern base model, MCMC: estimated fishing selectivities. Each box contains the middle $50 \%$ of the distribution. Whiskers extend over the full range.


Figure 9: Eastern base model, MCMC: estimated true year class strengths $\left(\boldsymbol{R}_{y} / \boldsymbol{R}_{0}\right)$. Each box contains the middle $50 \%$ of the distribution. Whiskers extend over the full range.


Figure 10: Eastern base model, MCMC: marginal posterior distribution for $\boldsymbol{B}_{0}$. The black dot marks the median of the posterior and the red dot the MPD estimate.


Figure 11: Eastern base model, MCMC: marginal posterior distribution for stock status $\left(B_{2013} / B_{0}\right)$. The black dot marks the median of the posterior and the red dot the MPD estimate.


Figure 12: Eastern base model, MCMC: SSB trajectory. Each box contains the middle $\mathbf{5 0 \%}$ of the distribution. Whiskers extend over the full range. Lines at $\mathbf{2 0 \%}, \mathbf{3 5 \%}$, and $48 \%$.


Figure 13: Eastern base model, MCMC: marginal posterior distribution for RBC. The black dot marks the median of the posterior and the red dot the MPD estimate.


Figure 14: Eastern base model, MCMC: marginal posterior distribution for long-term yield. The black dot marks the median of the posterior and the red dot the MPD estimate.


Figure 15: Western base model MPD-fit to trawl CPUE indices. Dashed lines are 95\% CIs.


Figure 16: Western base model MPD: "fit" to age-length data from the line fishery. Model growth curves shown as solid lines. Lowess fits to data shown as dashed lines.


Figure 17: Western base model MPD: estimated fishery selectivities and the ogive for port sampling of trawl-caught fish (age-based, sex independent, two time blocks for trawl).


Figure 18: Western base model MPD: estimated "true year class strengths" (annual recruitment divided by virgin recruitment).


Figure 19: Western base model MPD: estimated stock status (annual SSB divided by virgin SSB; female biomass only; mid-year). Lines at $\mathbf{0 . 2 , 0 . 3 5}$, and 0.48 .


Figure 20: Western base model MPD: estimated stock status for the base model and sensitivity runs. Lines at $0.2,0.35$, and 0.48 .


Figure 21: Western base model MPD-fit to trawl CPUE indices ( $\mathrm{CV}=\mathbf{1 5 \%}$ ) compared to the fit when the CPUE CV = $\mathbf{1 0 \%}$. Dashed lines are $\mathbf{9 5 \%}$ CIs (for the CV = $\mathbf{1 5 \%}$ ).


Figure 22: Western base model, MCMC: estimated fishing selectivities. Each box contains the middle $50 \%$ of the distribution. Whiskers extend over the full range.


Figure 23: Western base model, MCMC: estimated true year class strengths $\left(R_{y} / R_{0}\right)$. Each box contains the middle $50 \%$ of the distribution. Whiskers extend over full range.


Figure 24: Western base model, MCMC: marginal posterior distribution for natural mortality status ( $M$ ). The black dot marks the median of the posterior and the red dot the MPD estimate.


Figure 25: Western base model, MCMC: marginal posterior distribution for $\boldsymbol{B}_{0}$. The black dot marks the median of the posterior and the red dot the MPD estimate.


Figure 26: Western base model, MCMC: marginal posterior distribution for stock status $\left(B_{2013} / B_{0}\right)$. The black dot marks the median of the posterior and the red dot the MPD estimate.


Figure 27: Western base model, MCMC: SSB trajectory. Each box contains the middle $\mathbf{5 0 \%}$ of the distribution. Whiskers extend over the full range. Lines at $\mathbf{2 0 \%}, \mathbf{3 5 \%}$, and 48\%.


Figure 28: Western base model, MCMC: marginal posterior distribution for RBC. The black dot marks the median of the posterior and the red dot the MPD estimate.


Figure 29: Western base model, MCMC: marginal posterior distribution for long term yield. The black dot marks the median of the posterior and the red dot the MPD estimate.

## Appendix 1: Data weighting following Francis

Francis (2011) proposed methods to determine a constant CV for a biomass time series and to calculate the effective sample sizes for age and length frequencies assumed to have multinomial distributions. This appendix provides explicit equations for the calculation of a CV for a CPUE time series and also extends the method Francis applied to age and length frequencies to conditional age-at-length data.

## CPUE indices

Francis (2011) proposed that a biomass time series should be fitted by a smoother outside of the model and that a constant CV be estimated for the time series. The idea is to choose a level of smoothness which corresponds to what would be considered a good fit within the stock assessment model (thus allowing for assumptions with regard to natural mortality and how variable a biomass trajectory should be).

For each pink ling stock, the trawl CPUE indices ( $X_{i}$ ) were assumed to be median-unbiased relative biomass indices with a constant CV:

$$
\ln \left(X_{i}\right)=\ln \left(q B_{i}\right)+\ln \left(\varepsilon_{i}\right)
$$

where $q$ is the proportionality constant, $B_{i}$ is the selected biomass in year $i$, and $\ln \left(\varepsilon_{i}\right) \sim \mathrm{N}\left(0, \sigma^{2}\right)$. The maximum likelihood estimator of the CV was used for given $q B_{i}$ (which are obtained from the fit given by the smoother):

$$
\widehat{\sigma^{2}}=\frac{1}{n} \sum_{i}\left[\ln \left(\frac{x_{i}}{q B_{i}}\right)\right]^{2}
$$

and

$$
\widehat{C V}=\sqrt{e^{\widehat{\sigma^{2}}}-1}
$$

For pink ling: The lowess smoother in R was used and a CV of $15 \%$ was chosen by eye for both the eastern and western trawl CPUE indices (Figures A1.1 and A1.2). A CV of 10\% could be justified on the basis of the fits that have been seen in preliminary stock assessment models. A CV of $20 \%$ is too high.

## Conditional age-length data

Francis (2011) developed equations for his method TA1.8 for age and length frequencies which were assumed to have multinomial distributions. The method can be extended to conditional age-at-length data in several ways. However, following the principle of having
only a single standardised residual each year (i.e., mean age in our case) a natural approach for conditional age-at-length data is to standardize "observed mean age - expected mean age (given the lengths and sexes of the aged fish)".

We develop the equations assuming we have independent samples within each length class and across length classes. In reality this is far from the truth, but we hope that once we have an appropriate vector of effective sample sizes that the derived distribution is approximately "correct".

For a given year, assume that there are $n_{j}$ independent samples of age for the $j$ th length class and let $A_{j i}$ be the $i$ th sample. Suppose there are $m$ length classes (note, the classes can be defined by sex and length) and $N$ is the total number of age samples. Then the observed mean age is,

$$
\bar{A}=\frac{1}{N} \sum_{j=1}^{m} \sum_{i=1}^{n_{j}} A_{j i}=\frac{1}{N} \sum_{j=1}^{m} n_{j} \bar{A}_{j}
$$

We need an equation for the variance of $\bar{A}$ so that we can work out how to standardise the variance of the annual residual, $\bar{A}-\bar{S}$, where $\bar{S}=\mathrm{E}(\bar{A})$ is the mean age (given the lengths and sexes of the aged fish) for the catch from the specified fishery (i.e., calculated from the age-length structure of the catch).

Within a given length class the age samples are independent and identically distributed, so it is easy to calculate $\bar{S}$ and work out the variance of each $\bar{A}_{j}$ (and hence to the variance of $\bar{A}-\bar{S})$.

Let,

$$
\mu_{j}=\mathrm{E}\left(A_{j i}\right)=\sum_{a=\min (a g e)}^{a=\max (a g e)} p_{j a} a
$$

where $p_{j a}$ is the proportion of fish at age $a$ in the $j$ th length class for the catch after perturbation by ageing error (if any). Then we have:

$$
\mathrm{E}(\bar{A})=\frac{1}{N} \sum_{j=1}^{m} n_{j} \mathrm{E}\left(\bar{A}_{j}\right)=\frac{1}{N} \sum_{j=1}^{m} n_{j} \mu_{j}
$$

Also, let

$$
\mu_{j 2}=\mathrm{E}\left(A_{j i}^{2}\right)=\sum_{a=\min (a g e)}^{a=\max (a g e)} p_{j a} a^{2}
$$

Then,

$$
\operatorname{Var}\left(A_{j i}\right)=\mathrm{E}\left(A_{j i}^{2}\right)-\left(\mathrm{E}\left[A_{j i}\right]\right)^{2}=\mu_{j 2}-\mu_{j}^{2}
$$

and

$$
\operatorname{Var}\left(\bar{A}_{j}\right)=\frac{1}{n_{j}}\left(\mu_{j 2}-\mu_{j}^{2}\right)
$$

Again, from independence assumptions (in this theoretical setting where we have independent samples and a "correct" effective sample size),

$$
\operatorname{Var}(\bar{A})=\frac{1}{N^{2}} \sum_{j=1}^{m} n_{j}\left(\mu_{j 2}-\mu_{j}^{2}\right)
$$

We can rejig this equation to get to a form that Francis used for TA1.8:
We have,

$$
N=\sum_{j=1}^{m} n_{j} \text { and let } r_{j}=\frac{n_{j}}{N}
$$

then,

$$
\operatorname{Var}(\bar{A})=\frac{1}{N} \sum_{j=1}^{m} r_{j}\left(\mu_{j 2}-\mu_{j}^{2}\right)=\frac{v}{N}
$$

We can now generalise to the residual in any year $y$ :

$$
\operatorname{Var}\left(\bar{A}_{y}-\bar{S}_{y}\right)=\operatorname{Var}\left(\bar{A}_{y}\right)=\frac{v_{y}}{N_{y}}
$$

and adopt the notation and iterative reweighting scheme of Francis (2011) whereby we want to find $w$ to scale our original vector of sample sizes $\tilde{N}_{y}: \quad N_{y}=w \tilde{N}_{y}$

We note that,

$$
\operatorname{Var}\left(\bar{A}_{y}-\bar{S}_{y}\right)=\frac{v_{y}}{N_{y}}=\frac{v_{y}}{w \tilde{N}_{y}}
$$

and therefore,

$$
\operatorname{Var}\left(\frac{\bar{A}_{y}-\bar{S}_{y}}{\left(v_{y} / \tilde{N}_{y}\right)^{0.5}}\right)=\frac{v_{y}}{w \tilde{N}_{y}} \frac{\tilde{N}_{y}}{v_{y}}=\frac{1}{w}
$$

It follows that if we can iteratively reweight until we get $w=1$ then all of our standardised residuals will have a variance of 1 (which is what is expected of standardised residuals). The scalar at each iteration can be chosen by guesswork, aiming to get the "standard deviation of the standardised residuals equal to 1 ". Alternatively, we can use the formula that Francis (2011) obtained by equating the variance of the "vector of annual standardised residuals" (a vector of numerical values available at each iteration) to the formula for the variance of each standardised residual (note the subscript $y$ on the right-hand "Var" operator):

$$
\begin{aligned}
& \frac{1}{w}=\operatorname{Var}\left(\frac{\bar{A}_{y}-\bar{S}_{y}}{\left(v_{y} / \tilde{N}_{y}\right)^{0.5}}\right)=\operatorname{Var}_{y}\left(\frac{\bar{A}_{y}-\bar{S}_{y}}{\left(v_{y} / \tilde{N}_{y}\right)^{0.5}}\right) \\
& \rightarrow w=1 / \operatorname{Var}_{y}\left(\frac{\bar{A}_{y}-\bar{S}_{y}}{\left(v_{y} / \tilde{N}_{y}\right)^{0.5}}\right)
\end{aligned}
$$



Figure A1.1: Lowess fits to the original ISL-base eastern-trawl CPUE indices at different levels of smoothness with the corresponding estimated CV.


Figure A1.2: Lowess fits to the original ISL-base western-trawl CPUE indices at different levels of smoothness with the corresponding estimated CV.

Appendix 2: MPD fits to annual mean age, mean length, and mean conditional age (for Francis weighting)


Eastern base model: fits to annual mean age and length. Dashed lines are $\mathbf{9 5 \%}$ confidence intervals for the final effective sample sizes.


Eastern base model: fits to mean conditional age. Vertical lines are $\mathbf{9 5 \%}$ confidence intervals for the final effective sample sizes.


Western base model: fits to annual mean length. Dashed lines are 95\% confidence intervals for the final effective sample sizes.


Western base model: fits to mean conditional age. Vertical lines are $\mathbf{9 5 \%}$ confidence intervals for the final effective sample sizes.

Appendix 3: Likelihood profiles and some miscellaneous MPD diagnostics


Eastern base model: likelihood profile for $\boldsymbol{B}_{\boldsymbol{0}}$.


Eastern base model: likelihood profile for $M$.


Western base model: likelihood profile for $\boldsymbol{B}_{\boldsymbol{0}}$.


Western base model: likelihood profile for $M$.


Eastern model: sensitivity with sex-specific fishing selectivities (which vary enormously across the three time blocks).

Table: Total objective function value for the models with sex-independent and sexspecific fishing selectivities. AIC (for example) requires at least 2 log-likelihood units for each extra parameter to justify the use of the extra parameters.

| Stock | No sex | Sex-specific | Difference | Extra <br> parameters |
| :--- | ---: | ---: | ---: | ---: |
| East |  |  |  |  |
| West | 1670 | 1662 | 8 | 13 |



Eastern base model: boxplot of standardised residuals for conditional age data, by sex and age, for trawl-caught fish, (dominant pattern is by age - related to growth and selectivity; similar pattern for both sexes).


Eastern base model: boxplot of standardised residuals for conditional age data, by sex and age, for line-caught fish, (slightly different pattern for males and females).


Western base model: boxplot of standardised residuals for conditional age data, by sex and age, for trawl-caught fish, (dominant pattern is by age - related to growth and selectivity; similar pattern for both sexes).


Eastern base model: boxplot of standardised residuals for conditional age data, by sex and age, for line-caught fish, (similar pattern for both sexes).

## Appendix 4: MPD fits to age and length frequencies



Eastern base model: non trawl length frequencies (1 of 2)


Eastern base model: non trawl length frequencies (2 of 2)


Eastern base model: port sampled non-trawl length frequencies


Eastern base model: trawl length frequencies (1 of 2)


Eastern base model: trawl length frequencies (2 of 2)


Eastern base model: non trawl age frequencies


Eastern base model: trawl age frequencies


Western base model: non-trawl length frequencies (1 of 3)


Western base model: non-trawl length frequencies (2 of 3)


Western base model: non-trawl length frequencies (3 of 3)


Western base model: port-sampled trawl length frequencies (1 of 4)


Western base model: port-sampled trawl length frequencies (2 of 4)


Western base model: port-sampled trawl length frequencies (3 of 4)


Western base model: port-sampled trawl length frequencies (4 of 4)


Western base model: trawl length frequencies (1 of 5)


Western base model: trawl length frequencies (2 of 5)


Western base model: trawl length frequencies (3 of 5)


Western base model: trawl length frequencies (4 of 5)


## Appendix 5: Comparison of eastern base model with "SS3-type" runs

During the November RAG meeting two additional MPD sensitivity runs were done for the eastern base model to see if the difference between the ISL model and CSIRO's model using length-based selection (which was contrary to the base model specification) was due to the length-based selection. According to the results below (Table A5.1, Figure A5.1), it appears that much of the difference was due to the difference in fishing selectivities and the use of beginning-of-year SSB (the CSIRO model estimated stock status at $19 \% B_{0}$ ).

Table A5.1: Eastern base model compared with "SS3-type" alternative models: using length-based selection for the fishing selectivities; additionally using beginning-of-year SSB.

| Model | $\boldsymbol{B}_{\boldsymbol{0}}$ | Stock status <br> (depletion) | $\boldsymbol{M}$ | Total objective | Neg. Log. <br> Like. Trawl <br> CPUE |
| :--- | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |
| Base | 5549 | 0.23 | 0.24 | 1670.45 | -46.1 |
| Length selectivities | 6326 | 0.17 | 0.23 | 1684.42 | -46.4 |
| + Beginning of year | 7059 | 0.19 | 0.24 | 1688.00 | -44.3 |



Figure A5.1: Eastern ISL-base model compared with an "SS3-type" model: lengthbased selectivities and beginning-of-year SSB (ISL-base model in black).

# Final eastern CPUE analysis from ISL in 2013 

P.L. Cordue, ISL

16 October 2013

## Introduction

This is a short document, the final in the series, on CPUE analysis from ISL in 2013 with regard to pink ling. The previous document, dated 30 September 2013, and entitled, " $A$ further look at filtering and vessel-linking for pink ling CPUE" contains details of the potential linking vessels to use in the eastern trawl CPUE analysis.

## Methods

## Filtering

The species-based filtering for the analyses in this document was very minimal:

$$
\operatorname{lig}>0 \& \operatorname{gem}<250 .
$$

This filter just deals with the gemfish targeting problem in 1986-1989. Further work is needed on filtering for the eastern trawl analysis but there is no more time available in 2013. The other filters used in the previous analyses (associated with depth, tow duration, latitude, longitude, and missing values) were still used; as were the same explanatory variables.

## Vessel linking

The top four candidate linking vessels for each of the two time-block pairs were progressively added as linking vessels to the CPUE standardisation:

- top 2 vessels for each pair of time blocks
- top 3 vessels for each pair of time blocks
- top 4 vessels for each pair of time blocks.

The CPUE indices in each case were calculated, compared with each other, and with CSIRO's eastern trawl indices.

## Results and Conclusions

The use of just the two top linking vessels for each pair of time blocks leaves a little kink in the CPUE indices from 1999 to 2000, with the indices from 2000 onwards all raised relative to the indices when there are no time blocks (Figure 1). Adding the third linking vessel to each time block moves the indices, from 2000 onwards, down a bit (Figure 1). When the fourth vessel is added there are only subtle difference between the resulting indices and the no-time-block indices (Figure 1).

In comparison to the CSIRO indices the biggest difference is from 1986 to 1989 because of the gemfish filtering (Figure 2). As expected, the time-blocked indices (for the two or three top linking vessels) from 2000 onwards are above the CSIRO indices because of the time blocking (Figure 2).

The CPUE indices with the top three linking vessels is a natural choice for the eastern base stock-assessment model as it is intermediate for the two time-blocked series using two or four linking vessels (Figure 1). It is also intermediate between the CSIRO indices and the timeblocked indices using two linking vessels (Figure 2). The natural sensitivities to run in the stock assessment are the CSIRO indices and the time-blocked indices using two linking vessels.


Figure 1: Standardised CPUE for eastern trawl when all vessels are linked (no time blocks), or the top 2,3 , or 4 candidate vessels are linked across their respective time blocks.


Figure 2: Standardised CPUE for eastern trawl when the top 3 or 4 candidate vessels are linked across their respective time blocks; compared to the CSIRO indices (no time blocks and no gemfish filter).

# A further look at filtering and vessel-linking for pink ling CPUE 

P.L. Cordue

30 September 2013

## Introduction

At the September Slope RAG meeting it was suggested that ISL conduct further CPUE analysis to look at how robust the indices were to the choice of vessels used in the linking of time blocks and also how filtering was affecting the indices. The main concern was for the eastern stock and therefore the additional analysis only deals with that stock.

A recommendation is made on the methods to use to construct the CPUE indices for use in the base eastern stock assessment model.

## Methods

## Filtering

Logbook catch records for the eastern zones ( $10,20,30,60$ ) were filtered according to various criteria and standardised CPUE indices were calculated without time-blocking of vessel effects. Many filters were tried, but results are presented in this document for the following filters:

1: $\quad$ lig $>0$ and (gre $>0$ or trs $>0$ or oth $>0$ or red $>0$ )
2: $\quad$ lig $>0$ and gem $<250$ and (gre $>0$ or trs $>0$ or oth $>0$ or red $>0$ )
3: $\quad$ lig $>0$ and (gre $>0$ or trs $>0$ )

For filter 2, various models with second order interactions were run: depth*latitude, depth*month, month*latitude, and all three together.

## Time-blocking of vessel effects

Formal statistics were calculated for each vessel which potentially could be used as a linkingvessel between time blocks for the eastern-stock CPUE standardisation. This allows fully objective methods to be developed for the choice of linking vessels. Standardised CPUE indices were calculated for one such method to compare with the existing indices proposed as ISL's base eastern model. The indices were also compared with those from no time blocking (i.e., all vessels linking across all time blocks) and those from no vessel linkage at all (simply for comparative purposes - not a viable option because of the confounding between year and vessel effects).
The filter used in these models was the one originally proposed by ISL:

4: $\quad$ lig $>0$ and $($ gre $>100$ or trs $>100$ or oth $>100$ or red $>100)$

The acceptance rule for a vessel to be a candidate as a linking vessel between two time blocks was determined before filtering: at least 30 records in at least 4 years in each block; and a CV of less than $50 \%$ for the annual proportions of positive ling tows in each time block. The time blocks were [1986-1999], [2000-2006], and [2007-2012].

The statistics calculated for each candidate vessel for each linkage were:

- mean of the annual proportions of positive ling trawls within each block
- ratio of the (above) block means
- absolute value of the $\log$ of the ratio
- the difference between the proportion of positive ling tows in each block
- $95 \%$ CI for the above difference

Note the distinction between the mean of the annual proportions within a block (which is where the proportion of positive ling tows is calculated for each year and then the mean of the proportions is calculated) and the proportion of positive ling tows within a block (which is where the number of positive tows in the whole time block is divided by the number of tows in the time block).

## Results

## Filtering

With regard to filtering of catch records for the eastern stock, it was noted at the meeting that the differences in the first time block (1986-1999) between the ISL and CSIRO CPUE indices were due to the use of different records - i.e., the filtering procedure used. Two of the early points in the ISL series are higher than the CSIRO points, and the points from 1991-1999 are lower (Figure 1). This indicates that the choice of filter can be important. That is, it matters whether one uses all positive ling records in the main depth range or whether one excludes some records in an attempt to obtain data from a "consistent fishery".

The reason why it is desirable to define a fishery to some extent (through filtering) is to avoid strong targeting behaviour (on any species) confounding year effects. The problem is that in some years there may be a lot of targeting on some particular species or group of species. The model may not be able to account for the change in targeting with the variables available to it (e.g., depth, month, latitude).


Figure 1: A comparison of the base ISL eastern trawl CPUE indices and the CSIRO indices.

In the eastern data, a classic example of this problem is the targeting of gemfish in 1986-1988 when the catches were huge compared to later years. All other things being equal (e.g., depth, month, latitude), when vessels were targeting gemfish, it seems likely that the average ling catch was different from when they were not targeting gemfish. Similarly, when vessels were "targeting" ling (to the extent that they could), then average ling catches were higher than when vessels were not "targeting" ling. It is easy to deal with the problem of gemfish targeting because records with "large" gemfish catches can simply be removed. It is much more difficult to deal with ling "targeting" because we must not use a filter that depends on the amount of ling that was caught in each trawl.

The exclusion of "large" gemfish catches (by using filter 2) causes the 1986-1988 CPUE indices to increase relative to those produced by the filter which includes them (filter 1). The change in the indices only occurs in the years (1986-1988) which had the very high levels of gemfish catch/targeting (Figure 2).


Figure 2: Eastern-stock pink ling CPUE indices calculated using the three filters defined in the text: "Minor filtering" (filter 1), "Exclude gem>250" (filter 2), "gre>100 or trs>100" (filter 3).

When the grenadier and silver-warehou filter is used (filter 3) there is a bigger dip in the indices from 1991-1995 compared to the indices produced by the other two filters (Figure 2). This indicates that the signal from the "gre-trs fishery", for ling abundance, is different from the signal from a more loosely defined fishery. It doesn't mean, necessarily, that it is the "wrong" signal.

The choice of filter to use in the east is somewhat problematic because there is a lot of ling caught in trawls which retain the "oth" (other) species category. It is clear that some filtering should be done (i.e., records with "large" gemfish catches should be excluded).

My recommendation for the eastern base model is to use the following filter:
lig >0 and gem < 250 and (gre $>0$ or trs $>0$ or oth $>0$ ).

This is minor filtering in terms of the ling catch retained ( $92 \%$ before excluding gem, and $89 \%$ after excluding gem; including red only increases catches slightly so it has been dropped from the filter).

The results for the interaction models are not given because the inclusion of interactions barely altered the standardised indices.

## Time-blocking of vessel effects

The original method of determining vessels to use in the linking of time blocks was mainly based on each vessel's proportion of positive ling tows between the two blocks under consideration. Plots of the proportion of positive tows by year and boxplots of depth and duration were examined by eye to find vessels which appeared to have fairly constant behaviour across the two time blocks. This was a subjective method that could be hard for someone else to duplicate.

To test the sensitivity of CPUE indices to the choice of linking vessels, an alternative set of vessels was chosen using an objective rule based only on the proportion of positive ling tows (Tables 1 and 2). The rule was simply to take the three vessels which had the ratio of the mean annual proportion of positive ling tows closest to zero (which suggests that their ling targeting behaviour is little changed between the two blocks). An alternative statistic based on the difference in the (single) proportion of positive ling tows in each block was also calculated with a $95 \% \mathrm{CI}$ (Tables 1 and 2). This provides an alternative statistic to use in selecting the vessels with little change in their targeting (i.e., we are looking for vessels for which the confidence interval spans zero, going from a negative to a positive value).


Figure 3: Eastern-stock pink ling standardised CPUE indices when no time blocking is used and when three time blocks are used without any linking vessels (not recommended).

It is clear from Tables 1 and 2 that there are very few candidate vessels that appear to have consistent ling targeting across the time blocks. We already know that some vessel linking needs to be done across time blocks or the year and vessel effects are completely confounded.

There is a continuum between no time blocking, which corresponds to all vessels being used to link all time blocks, and the use of time blocks with no linking vessels. As expected, the use of no linking vessels allows the year effects (standardised CPUE indices) to go "wild" (Figure 3). The difference between the two lines in Figure 3 represents the full range of the continuum that CPUE indices can span given the choice of which vessels to link with and the number of linking vessels. Any linkage at all will quickly bring the indices down from the "wild" ones but how close they end up to the no-time-block indices depends on which vessels are used.


Figure 4: Eastern-stock pink ling standardised CPUE indices when no time blocking is used and when three time blocks are used with linking vessels chosen subjectively (original choice) or objectively by one particular method (see text).

There is a difference in the standardised CPUE indices depending on the method used to choose the linking vessels (Figure 4). The difference is all in the linkage choice for the middle time block, with the CPUE indices, for the objective method, being partway between the no-time-block indices and the indices from the subjective method (Figure 4). The change in linking vessels for the second time block appears not to matter much (as the same trend is apparent).

The need for time-blocking seems clear (as there were known events occurring which could/did alter ling targeting behaviour). However, the CPUE indices are sensitive to the choice of linking vessels for the first two time blocks. That makes it important to use an
objective method to choose these vessels. The sensitivity is in the level of increase in the indices from 2000-2012, but there is no sensitivity to the direction in which the indices move.

## My recommendation for the method to select linking vessels is to require that two criteria be meet:

- $0.9 \leq$ block ratio $\leq 1.1$
- block difference $\mathbf{9 5 \%}$ CI: lower limit $<0$ and upper limit $>0$

This method selects vessels 1, 2, and 4 (from Table 1) to link time blocks [1986-1999] and [2000-2006]. And, it selects vessels 1 and 2 (from Table 2) to link time blocks [2000-2006] and [2007-2012].

## Conclusion and recommendation

The choice of filter is important for the eastern stock CPUE standardisation. It is clear that some filtering (in addition to depth) is needed (e.g., to remove the confounding in 1986-1988 between the gemfish-targeting effect and the year effects). However, it is not yet clear exactly which filter should be used to provide the best chance of obtaining a legitimate biomass signal. To have no filtering is not a good option as it is at the extreme of a continuum. Therefore, I suggest for the base model that a small move is made in the right direction with a filter to remove the gemfish targeting but to include most of the ling catch in the eastern zones:

- $\operatorname{lig}>0$ and gem $<250$ and $($ gre $>0$ or trs $>0$ or oth $>0$ ).

On the question of time blocking, there is a clear need to do it to address the issues of the sale of ling quota from trawl fishers to line fishers that started in 1999-2000 and the structural adjustment from 2006-2007. Again, to have no time blocking is to be at one extreme of a continuum. Therefore, a move in the right direction is suggested. An objective method of choosing the linking vessels is needed and I recommend two criteria be used together:

- $0.9 \leq$ block ratio $\leq 1.1$
- block difference $\mathbf{9 5 \%}$ CI: lower limit $<0$ and upper limit $>0$

Table 1: Statistics for candidate vessels for linking blocks [1986-1999] and [2000-2006]. The means are for the proportion of positive ling tows in each block (Mean1, Mean 2); the ratio is Mean2/Mean1; Abslog is a measure of how close the ratio is to zero and the table is sorted on this value; the block difference and the associated $95 \%$ confidence interval is for the proportion of positive ling tows in the first block minus the proportion in the second block. Arbitrary vessel numbering for privacy. * vessel used in subjective linking. \# vessel used in objective linking.

| Vessel | Mean1 | Mean2 | Ratio | AbsLog | Block <br> difference | Lower <br> $\mathbf{9 5 \%} \mathbf{C l}$ | Upper <br> $\mathbf{9 5 \%} \mathbf{C l}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\# 1$ | 0.31 | 0.31 | 0.97 | 0.03 | 0.01 | -0.03 | 0.04 |
| $\#^{*} 2$ | 0.45 | 0.48 | 1.06 | 0.06 | -0.02 | -0.05 | 0.01 |
| $\# 3$ | 0.67 | 0.63 | 0.94 | 0.06 | 0.07 | 0.03 | 0.11 |
| 4 | 0.40 | 0.36 | 0.90 | 0.10 | 0.02 | -0.01 | 0.05 |
| ${ }^{2} 5$ | 0.29 | 0.26 | 0.90 | 0.10 | 0.04 | 0.00 | 0.07 |
| 6 | 0.54 | 0.64 | 1.19 | 0.17 | -0.08 | -0.10 | -0.06 |
| 7 | 0.52 | 0.64 | 1.22 | 0.20 | -0.12 | -0.15 | -0.08 |
| 8 | 0.29 | 0.35 | 1.23 | 0.21 | -0.05 | -0.08 | -0.02 |
| 9 | 0.36 | 0.53 | 1.47 | 0.38 | -0.19 | -0.23 | -0.15 |
| $* 10$ | 0.09 | 0.06 | 0.67 | 0.40 | 0.02 | 0.01 | 0.04 |
| 11 | 0.22 | 0.34 | 1.58 | 0.46 | -0.12 | -0.17 | -0.08 |
| 12 | 0.23 | 0.13 | 0.57 | 0.57 | 0.10 | 0.07 | 0.12 |
| 13 | 0.40 | 0.77 | 1.93 | 0.66 | -0.35 | -0.38 | -0.32 |
| 14 | 0.51 | 0.22 | 0.44 | 0.83 | 0.26 | 0.20 | 0.32 |

Table 1a: Additional statistics for Vessels 1-5: number of years in each time block, number of records in each time block, proportion of ling catch in each time block, and the proportion of ling records in zone 20 (including zone 60).

| Vessel | Number of years |  | Number of records |  | Proportion of catch |  | Proportion of records in zone |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Block 1 | Block 2 | Block 1 | Block 2 | Block 1 | Block 2 | Block 1 | Block 2 |
| 1 | 14 | 7 | 426 | 201 | 0.78 | 0.22 | 0.84 | 0.97 |
| 2 | 8 | 7 | 713 | 788 | 0.47 | 0.53 | 0.70 | 0.95 |
| 3 | 6 | 7 | 334 | 479 | 0.62 | 0.38 | 0.98 | 0.97 |
| 4 | 13 | 7 | 536 | 364 | 0.74 | 0.26 | 0.01 | 0.00 |
| 5 | 12 | 7 | 234 | 102 | 0.83 | 0.17 | 0.36 | 0.01 |

Table 2: Statistics for candidate vessels for linking blocks [2000-2006] and [2007-2012]. The means are for the proportion of positive ling tows in each block (Mean1, Mean 2); the ratio is Mean2/Mean1; Abslog is a measure of how close the ratio is to zero and the table is sorted on this value; the block difference and the associated $95 \%$ confidence interval is for the proportion of positive ling tows in the first block minus the proportion in the second block. Arbitrary vessel numbering for privacy. * vessel used in subjective linking. \# vessel used in objective linking.

| Vessel | Mean1 | Mean2 | Ratio | AbsLog | Block <br> difference | Lower <br> $\mathbf{9 5 \% ~ C I}$ | Upper <br> $\mathbf{9 5 \% ~ C I}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\# 1$ | 0.34 | 0.36 | 1.04 | 0.04 | 0.00 | -0.05 | 0.06 |
| $\# * 2$ | 0.63 | 0.6 | 0.95 | 0.05 | 0.02 | -0.02 | 0.06 |
| $\# 3$ | 0.48 | 0.45 | 0.93 | 0.07 | 0.05 | 0.00 | 0.09 |
| 4 | 0.48 | 0.52 | 1.09 | 0.09 | -0.03 | -0.06 | 0.00 |
| $* 5$ | 0.06 | 0.07 | 1.17 | 0.16 | -0.01 | -0.03 | 0.01 |
| 6 | 0.42 | 0.34 | 0.81 | 0.21 | 0.08 | 0.05 | 0.11 |
| 7 | 0.57 | 0.46 | 0.81 | 0.21 | 0.14 | 0.12 | 0.16 |
| 8 | 0.64 | 0.50 | 0.79 | 0.24 | 0.15 | 0.12 | 0.17 |
| 9 | 0.13 | 0.10 | 0.73 | 0.32 | 0.03 | 0.01 | 0.06 |
| 10 | 0.77 | 0.51 | 0.65 | 0.43 | 0.21 | 0.17 | 0.25 |
| 11 | 0.64 | 0.35 | 0.54 | 0.61 | 0.29 | 0.26 | 0.32 |
| 12 | 0.55 | 0.26 | 0.47 | 0.75 | 0.28 | 0.24 | 0.33 |

Table 2a: Additional statistics for Vessels 1-5: number of years in each time block, number of records in each time block, proportion of ling catch in each time block, and the proportion of ling records in zone 20 (including zone 60).

|  |  |  | Number of <br> records |  |  | Proportion of <br> Catch |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Vessel | Proportion of <br> records in zone <br> 20 |  |  |  |  |  |  |  |
|  | Block 1 | Block 2 | Block 1 | Block 2 | Block 1 | Block 2 | Block 1 | Block 2 |
| 1 | 7 | 6 | 124 | 87 | 0.71 | 0.29 | 1.00 | 1.00 |
| 2 | 7 | 6 | 479 | 431 | 0.43 | 0.57 | 0.97 | 1.00 |
| 3 | 5 | 6 | 249 | 392 | 0.39 | 0.61 | 0.69 | 0.81 |
| 4 | 7 | 6 | 788 | 841 | 0.49 | 0.51 | 0.95 | 0.83 |
| 5 | 7 | 6 | 63 | 51 | 0.65 | 0.35 | 0.00 | 0.02 |

# Analysis of trawl catch and effort data to produce standardised CPUE indices for pink ling 

P.L. Cordue<br>7 September 2013

## Introduction

Detailed trawl logbook data were analysed to determine suitable subsets of data to use in producing standardised ling CPUE indices for the eastern and western stocks. The complication of possible "regime shifts" in ling targeting caused by the movement of quota in 1999-2000 from trawl fishers to line fishers, and the structural adjustment in 2006, were addressed in the CPUE standardisation.

## Methods

Detailed logbook data were supplied by CSIRO for a list of 157 vessels that took about $99 \%$ of the ling catch over the period 1985-2013. All records were supplied for each vessel irrespective of whether they reported retained ling or not on a particular station. Each record contained estimated green-weight for about 30 species including an "OTH" category which contained all of the other species (see Klaer and Smith 2012 for details of the species and other fields).

The detailed data were imported into R and restricted to trawl stations only (some vessels had in some years used gear other than OT). For the descriptive analysis as many records as possible were used. For the production of CPUE indices, records were restricted to the years 1986-2012 and records with missing values in explanatory variables were excluded.

The east and west CPUE standardisations were done separately. However, in most of the descriptive analysis all zones were dealt with together. Summary plots were produced by zone to determine, for the "ling vessels", the main species caught in the zones and the seasonality of the catches (Appendix 1). Summary plots were also produced by species to look at trends and characteristics by depth, month, year, zone, and their association with ling (Appendix 2). Summary plots were also produced for individual vessels to examine whether their ling targeting behaviour may have changed over time.

The objective of the descriptive analysis was to determine suitable subsets of data to use in the CPUE standardisations. The form of the models were the same for both east and west:

$$
\log (\text { ling catch }) \sim \text { year }+ \text { month }+ \text { DorN }+ \text { hours }+ \text { depth }+ \text { latitude }+ \text { vessel }
$$

All explanatory variables were categorical:
year: 1986-2012
month: 1-12
DorN: four codes: day (D), night (N), mixed (M), and unknown (U)
hours: cut into a factor with 12 levels from $0.5-10 \mathrm{hrs}$ (west) and 10 levels from 0.5-8 hrs (east)
depth: cut into a factor with 11 levels from 200-750 m ( 50 m bins)
latitude: cut into a factor with 9 levels (east) and 8 levels (west)
vessel: individual effects for any vessel present in at least two years within a "block"

The potential changes in vessel effects were modelled by allowing most vessels to change their vessel effect between blocks of time:
east: 1986-1999, 2000-2006, 2007-2012
west: 1986-2006, 2007-2012.

The split from 2006-2007 is indicated by the timing of the structural adjustment and also by the logbook data where the number of vessels approximately halves from 2006 to 2007. The eastern split from 1999-2000 is indicated by advice from the industry that ling quota associated with eastern trawl vessels was sold to support the developing line fishery and the split is also supported by the logbook data (see Results below).

The link between blocks was maintained by requiring that some vessels retained a constant vessel effect over the break points. This is needed to ensure that vessel and year effects are not confounded within a block.

## Results

## Filters

The species summaries and zone summaries were used to define filtering criteria for the ling CPUE analysis. The objective was to define a consistent "fishery" in which ling was a "major" bycatch. The point being to try to get a consistent subset of data across years; although the depth filtering of 200-750 m has already done some of the job. Depths beyond 750 m were excluded because the catches are very intermittent across years (and are therefore unlikely to provide useful information on abundance). Short tows were also excluded as they are likely to be highly targeted.

In the west (Zones 40,50 ) there were obvious seasonal fisheries for GRE and TRS and during the ling spawning season (September-November) a reasonable proportion of ling (LIG) was also caught in Zone 40 (See Appendices). The filter used for the base model in the western ling CPUE analysis was:
where each species code denotes the species catch (kg). This filter retained $62 \%$ of the logbook ling catch and 66,000 records. Very similar results were obtained from a threshold of 0 ( $92 \%$ of the catch) and 200 ( $48 \%$ of the catch). The key point is that GRE and TRS are associated with ling in the west.

In the east (Zones 10, 20, 30 with Zone 60 included in Zone 20) the situation is more complex with a mixture of species, in particular the "mixed bag" OTH, making up the bulk of the catches in ling depths (see Appendices). There used to be an ORO fishery in Zone 30 and a GEM fishery in Zone 10. Also, there is an inshore FLT fishery in all of the eastern zones. A mix of species, associated with ling and its depth range, were used in the filter:

$$
\text { East: LIG }>0 \text { \& }(\text { RED }>100 \mid \text { OTH }>100 \mid \text { TRS }>100 \mid \text { GRE }>100)
$$

Various other filters were tried (e.g., higher and lower values than the threshold of 100; excluding moderate-large GEM catches) but they made little difference to the results. The original filter was at a threshold of 0 which accounted for $93 \%$ of the ling logbook catch in the eastern zones ( 406,000 records). At a threshold of 100 only $40 \%$ of the catch was accounted for but it was still 171,000 records - more than enough. At a threshold of 200 which was just $23 \%$ of the catch the analysis still gave very similar results.

## Eastern split 1999-2000

There has been much debate about a potential change in ling catchability in the eastern stock after the 1999 sell-off of ling quota from the trawl fleet. The 2010 assessment split the trawl CPUE time series into two segments with the break at 1999. However, the split was removed in 2011 because of an apparent lack of evidence for a real change.

The detailed logbook data does point to a change in ling targeting in the eastern zones, especially in Zones 10 and 20. The first indication is from catch per tow (including zeroes) for all of the trawl logbook records (Figure TB1). Ling catch per tow was trending slightly upwards in Zones 20 and 30 until 1999 when it started heading down. This is unlikely to be a biomass signal - it relates more to ling targeting and/or reporting of small ling catches. In the eastern and western zones, the proportion of tows that reported no retained ling were all trending down until 1999 (Figure TB2). In Zones 30, 40, and 50, the trends (of varying levels of steepness) continued beyond 1999. However, in Zone 10 there was a steep increase from 1999 to 2000; and in Zone 20, the proportion of zeroes stopped going down and moved up slightly.

For many species there seemed to be a general trending down in the proportion of zero tows over time (see Appendices) which suggests that there was better reporting of small catches. However, ling is not alone in reversing that trend in Zone 10 in 1999. The same reversal is seen to some degree for DOM, GRE, and REG.


Figure TB1: Arithmetic ling catch per tow by zone and year (" 1 " = Zone 10 , etc).


Figure TB2: The proportion of trawls for which zero retained ling was reported, by zone and year.

An analysis of individual vessels by year also indicated that 1999 to 2000 was a time of change in Zone 10, with the vast majority of vessels who fished regularly in Zone 10 showing an increase in the proportion of zero ling tows from 1999 to 2000.

There is little doubt that there was a move away from ling targeting in some of the eastern trawl fishery from 1999 to 2000 . However, that does not necessarily mean that a standardisation cannot deal with the change. If the movement away from targeting ling just involved a change in depth or zone then that could be accounted for. What cannot be accounted for are more subtle changes such as moving away from an area where too much ling is being caught while staying in the same zone and depth.

## Vessels that did not appear to change ling-targeting behaviour across time blocks

In the east and the west I searched for vessels which maintained a fairly constant depth range, tow duration, and proportion of zero ling tows across the time blocks. For the west, three vessels were found that spanned the 2006-2007 breakpoint. In the east, there were far more vessels available but it was difficult to find vessels that hadn't changed their behaviour. However, there was one vessel that spanned all three time blocks and two more which spanned the 1999-2000 breakpoint and an additional one which spanned 2006-2007. See Appendix 3 for the plots on which my assessment was made. The callsigns of the vessels are available but are not given here.

## CPUE standardisation results

In the east, the use of time blocks has a substantial effect on the CPUE indices (Figure 1). Essentially there is shift upwards of all of the points from 2000 onwards, with little change in the trend from 2000 to 2012. This suggests that there was a shift in ling targeting after 1999 and that it cannot be accounted for just by depth and area. There was little change in the trend after 2006 so it appears that the vessels remaining after the structural adjustment have not had a major change in their ling-targeting behaviour (or at least their reported retained ling catches).

The ISL and CSIRO indices have similar trends but they differ particularly in recent years with the ISL indices being higher from 2005-2008 (Figure 2). How much difference the use of one or the other will make in the stock assessment is hard to judge, but I suspect that the ISL indices will lead to higher estimated stock status. The CSIRO indices were calculated without time blocking which would seem to be an essential element to include.

For the east, the other effects show some interesting patterns. There is a very strong depth effect with the peak catch rates at about 500 m (Figure 3). The day-or-night effects are minor (Figure 4). The tow duration is important with catches increasing with longer duration although not in a linear fashion (as is assumed in the CSIRO model)(Figure 5). There is a very strong latitude effect with catches peaking in Zone 20 (Figure 6). There are moderate month effects with highest catches in June (Figure 7). The peak season could vary by zone/latitude but a latitude-season interaction did not change the year effects in an earlier model.

The vessel effects show only moderate variation and therefore changes in a single vessel's effect across time blocks are not unreasonable (Figure 8).

In the west, the use of the two time blocks had little effect on the CPUE indices (Figure 9). Also, the ISL indices are very similar to those produced by CSIRO (Figure 10). As in the east, there is a depth effect (but it is only moderate here)(Figure 11) and there is very little day-night effect (Figure 12). The tow duration effect is linear as assumed by CSIRO (Figure 13). The latitude effect is not strong but catches do peak in Zone 40 (Figure 14). The seasonal effect is moderate with catches peaking in September/October during the spawning season (Figure 15). Vessel effects show moderate variation with little change within vessels across time blocks (Figure 16).

## Eastern stock diagnostics

The use of the time blocks allows the model extra freedom to deal with the residual patterns, across time, within individual vessels. Because there was a preponderance of apparent decrease in vessel effectiveness after 1999, the year effects and vessel effects have shifted accordingly. The use of the time blocks shifted the 2000-2012 year effects up by 30-40\% compared to the effects without time blocks (Figure 17). The median vessel effects were reduced not nearly so much, shifting down by just $13 \%$ (Figure 18). Other changes in effects were examined and they were minor.

There were no obvious problems with the residuals in the CPUE model (Figures 19 and 20).

## Zone-based CPUE indices

At CSIRO's request the latitude factor was replaced with a zone factor in the ISL models and a combined-zone model was produced for each stock together with individual trends by zone (i.e., putting in a zone-year interaction). For both east and west the use of the latitude factor or the zone factor made almost no difference to the estimated year effects (Figures 21 and 22).

The individual-zone trends appear similar in the east although the indices for Zone 30 are very noisy (Figure 21). In the west the trends in Zones 40 and 50 obviously differ in recent years with Zone 40 trending upwards and Zone 50 being flat (Figure 22). In both the east and the west, the individual trends, when averaged in any sensible way will produce something very similar to the combined-zone models (so the different individual-zone trends is not a problem for a spatially-aggregated stock assessment model).

## Conclusion

The automated, generic, CSIRO approach is adequate for the western CPUE standardisation. However, in the east a more detailed analysis is needed, including filtering of records and, most importantly, an allowance for possible changes in vessel effects across the 1999-2000 boundary. The structural adjustment and the reduction in ling quotas is also a cause for concern in terms of the ability of CPUE to track abundance. However, it appears that there is nothing in the data that confirms this is a problem - though it still remains an issue.


Figure 1: Standardised eastern trawl CPUE when vessel effects are blocked and when they are not.


Figure 2: A comparison of the base ISL eastern trawl CPUE indices and the CSIRO indices.


Figure 3: Depth effects for the ISL base eastern tawl CPUE model .


Figure 4: Day or night effect for the ISL base eastern trawl CPUE model.


Figure 5: Trawl duration effect for the ISL base eastern trawl CPUE model.


Figure 6: Latitude effect for the ISL base eastern trawl CPUE model. The grey lines mark the boundaries of Zone 20.


Figure 7: Month effect for the ISL base eastern trawl CPUE model.


Figure 8: Vessel effects for the ISL base eastern trawl CPUE model. Consecutive black, red, and green dots are for the same vessel in three different time blocks. Consecutive black and red dots are for the same vessel in two time blocks. Open circles are for vessels with a single effect across blocks or occuring only in a single block.


Figure 9: Standardised western trawl CPUE when vessel effects are blocked and when they are not.


Figure 10: A comparison of the base ISL western trawl CPUE indices and the CSIRO indices.


Figure 11: Depth effects for the ISL base western tawl CPUE model .


Figure 12: Day or night effects for the ISL base western tawl CPUE model .


Figure 13: Tow duration effects for the ISL base western tawl CPUE model .


Figure 14: Latitude effects for the ISL base western tawl CPUE model. The grey line marks the boundary between Zones 40 and 50.


Figure 15: Month effects for the ISL base western tawl CPUE model .


Figure 16: Vessel effects for the ISL base western trawl CPUE model. Consecutive black and red dots are for the same vessel in the two different time blocks. Open circles are for vessels with a single effect across blocks or occuring only in a single block.


Figure 17: The percentage increase in each year effect when vessel-effect blocking was introduced for the eastern trawl CPUE standardiasion. The grey line is at $\mathbf{3 5 \%}$.


Figure 18: Boxplot of the annual vessel effects for the ISL base eastern trawl CPUE model. The red line is at the median across all records (1.32). The median vessel effects for each block are: 1.35 [1986-1999], 1.18 [2000-2006], 1.18 [2007-2012].


Figure 19: Standardised residuals vs predicted values for the ISL base eastern trawl CPUE model.


Figure 20: Standardised residuals compared to a $\mathbf{N}(0,1)$ density for the ISL base eastern trawl CPUE model.


Figure 21: A comparison of the base ISL eastern trawl CPUE indices and ISL zonebased indices for all zones and for individual zones.


Figure 22: A comparison of the base ISL western trawl CPUE indices and ISL zonebased indices for all zones and for individual zones.

## Appendix 1: Summary plots by zone

For each of Zones 10-80 two summary plots for catch by species, aggregated over all years, are given. The plots are for the 157 vessels which accounted for about $99 \%$ of the ling catch. The plots, from left to right, are mean catch by month and mean proportion-of-total-catch by month (given for the species which were above some cutoff on the maximum monthly value).













## Appendix 2: Species summary plots

These plots are in two sections. They are for catches associated with the 157 vessels which caught about $99 \%$ of the ling trawl catch. They do not cover catches from all trawl vessels just those associated with ling in this study.

The first section has four plots for each species: proportion of tows on which a zero catch was recorded for the species, by zone (10-50) and year; mean catch per tow by zone(10-50) and year; mean catch per tow by month, aggregated over years, for zones $10-80$; mean proportion-of-total-catch by month, aggregated over years, for zones 10-80.

The second section also has four plots for each species: mean catch per tow by depth class and zone (10-80) aggregated over years; mean catch per tow by depth class and month, aggregated over years; mean catch per tow by depth class and year; annual correlation coefficient for the species catch and ling catch by zone (10-50).



























Year









Year







Species: trs




## Appendix 2 (continued): four more plots for each species






























## Appendix 3: Plots for vessels selected to have constant vessel effects

For the west, three vessels were selected to have constant vessel effects over the whole time period - see the first three plots below.

For the east, there was one vessel that gave a weak link over the whole time period - it didn't catch much ling but at least it seemed fairly consistent. There were two additional vessels spanning the break at 1999-2000 and a single additional vessel covering 2006-2007.


West vessel 1.


## West vessel 2.




Year


## West vessel 3.



East vessel 1 (link across all three blocks).


East vessel 2 (1986-1999, 2000-2006 link).


East vessel 3 (1986-1999, 2000-2006 link).


East vessel 4 (2000-2006, 2007-2012 link)

# Preliminary stock assessment for pink ling: ISL models 

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15 September 2013

## Executive Summary

This document provides a summary and many details with regard to preliminary stock assessments of Australian pink ling (eastern and western stocks) performed by ISL for consideration at the September 2013 Slope RAG meeting. The document describes the development of an ISL "base" eastern model and a western "reference" model. These are for illustrative purposes as the RAG will make the choices as to which data and model structures are used in the 2013 base models used to provide management advice.

The eastern and western assessments have very different challenges to deal with in terms of providing a reliable stock assessment. For the east, the issues are mainly to do with data preparation, while in the west, the issues primarily involve model and data weighting decisions.

The analysis of length and age-at-length data (performed by ISL and described in other draft RAG documents) showed that the eastern data had to be stratified and scaled to some extent otherwise the model would be given misleading signals. There is a trade-off between how much stratification and scaling is done and how much data can be used in the model. The acceptance rules used in the ISL-base model appear to provide a good balance between the quality and quantity of data used. The analysis also showed that the assumption of lengthbased selection in the fisheries was not fully supported; and there was little support for the use of cohort-specific growth.

ISL's CPUE analysis, to some extent, was able to deal with the issues of the sale of trawl quota in the east in 1999 and the structural adjustment in 2006. The approach used to allow a change in vessel behaviour to alter the trawl CPUE indices had a strong effect on the eastern indices over the 1999-2000 breakpoint but little effect over the 2006-2007 breakpoint (in the east and the west). Consequently the choice of the ISL or CSIRO CPUE indices in the east has an impact on the assessment results, but it is not important in the west.

In the western assessment, there is a strong conflict between the trawl age-length data and the trawl CPUE indices. This indicates that a down-weighting of the trawl age-length data is required. However, when this is done the estimate of $M$ appears to be a bit too high. As there are little data in the western assessment to provide information on $M$, it is best to either fix $M$ at a "sensible" value, or to use a strongly informed prior. In both cases, it appears that the eastern assessment can provide the needed information on $M$.

## Introduction

This document provides a summary and many details with regard to preliminary stock assessments of Australian pink ling (eastern and western stocks) performed by ISL for consideration at the September 2013 Slope RAG meeting. The document describes the development of an ISL "base" eastern model and a western "reference" model. These are for illustrative purposes as the RAG will make the choices as to which data and model structures are used in the 2013 base models. The work could not have been done without the help and cooperation of CSIRO scientists. CSIRO supplied ISL with raw and processed data and there were numerous email discussions on modelling and data issues. My particular thanks to Andre Punt and Neil Klaer.

## Methods

Numerous issues were considered during the preparation of this stock assessment (Cordue \& Punt draft). Issues not brought into question and which followed the approach used by CSIRO in 2012 were:

- Assumption of two stocks, east (Zones 10, 20, 30) and west (Zones 40, 50) (except Zone 60 is now included in Zone 20)
- Construction of catch history (except Zone 60 is now included in Zone 20)
- Two fisheries: trawl and non-trawl
- Discards not modelled
- Use of a two-sex population model
- Plus-group at age 30 in the population
- Construction and use of the ageing error matrix

In terms of other issues considered in Cordue \& Punt (draft) the following options were chosen in ISL's preferred models:

| Issue | Option chosen for base models |
| :--- | :--- |
| Construction of length frequencies for <br> spatially-aggregated models | Length frequency data were analysed and it <br> was concluded that stratification by depth <br> and/or zone was appropriate (see details of <br> exactly how LFs were prepared in the base <br> models below). (See Cordue, draft-1.) |
| Construction of CPUE indices from trawl <br> catch and effort data | The standard NZ approach of a detailed data <br> analysis followed by an appropriate <br> standardisation was followed. The ISL and <br> CSIRO indices are very similar in the west <br> and quite different in the east. (See Cordue, <br> draft-3.) |
| Preparation of age data sets | Age-length data were analysed and it was <br> concluded that some stratification and scaling <br> was required in the east (Cordue, draft-2). <br> The use of age-length keys was also deemed <br> appropriate for the unsexed age-length data. |
| Use of Kapala data | Excluded as it is very unlikely to be <br> representative of the eastern stock. |
| Use of non-trawl CPUE indices | Excluded as it is very unlikely to be tracking <br> biomass at the stock level. |
| Definition of Spawning Stock Biomass (1) | Female only |
| Definition of Spawning Stock Biomass (2) | Mid-year (i.e., after half of the mortality) |
| Estimation of growth (1) | Estimated inside the model |
| Estimation of growth (2) | No cohort specific growth as an analysis of <br> the age-length data showed that there were <br> few if any important cohort effects (Cordue, <br> draft-2). |
| Estimation of natural mortality | Estimated inside the model as a single <br> constant with a moderately informed prior |
| Weighting of trawl CPUE indices between <br> years | Constant CV of 15\% derived following <br> Francis (2011) |
| Weighting of data sets | Standard NZ approach following the spirit of <br> Francis (2011) |
| Spatial modelling changes in ling catchability (trawl | Dealt with in the CPUE analysis (see Cordue, <br> draft-3) |
| No spatial modelling was done as the single- <br> area models look appropriate. |  |
|  |  |

## New issues

Two additional issues arose since the distribution of Cordue and Punt (draft).

The first relates to the use of length-based selectivities for the trawl and non-trawl fisheries (as was done in the 2012 assessment). Specifying length-based selectivities for the trawl and
non-trawl fisheries is counter-indicated by the data which show that the two fisheries have very different age structure at given length (Cordue, draft-2). The trawl fishery catches younger fish at given length than the non-trawl fishery. If length-based selectivities are used then the age-length data are essentially assumed to be population age structure at length for both fisheries. For the ISL models, it was decided to use explicitly age-based selectivities for both fisheries. (Note, "length based" selectivities in age-structured models - which do not use "growth morphs" - are actually age-based selectivities as only numbers at age are maintained in the model over time and length is just produced as needed. In a sexed-model, using length selection is a way of saving parameters as the male-female differences in the corresponding age-selectivities come off the growth curves - i.e., length selection is a restricted subset of age selection.)

Also, an issue with regard to growth estimation was noticed after looking at CSIRO's 2013 model results. The CSIRO estimated growth curves, for the spatially-aggregated eastern model, were very different from the ISL estimates. I eventually deduced that the CSIRO model estimates were being driven by unsexed age-length data which had been fitted in their models. Only the sexed age-length data were used as individual observations of fish at age and length in the ISL models. As a good fit to the sexed age-length data is essential to get good growth estimates ,the approach used by CSIRO needs to modified to some extent. I suggested a down-weighting of the non-sexed age length data as a possibility.

## Stratification and scaling of LFs

Length frequency data, scaled from the sample numbers to the numbers in the sampled trawl/set/landing were supplied by CSIRO for each combination of east/west, trawl/nontrawl, zone, and depth stratum (for onboard samples).

Cordue (draft-1) showed that stratification and scaling of length data was needed for the eastern zones because of variation in length across zones and by depth within zone. The onboard length data for the eastern-trawl were stratified by depth ( $0-300 \mathrm{~m}, 300-500 \mathrm{~m}$, $500 \mathrm{~m}+$ ) and zone (SF10, SF20, SF30). For each year and zone, a depth stratum was required to have at least 2 operations and 30 fish to be used for scaling. Further, for each year at least $80 \%$ of the catch had to represented by scaled LFs otherwise that year was dropped. Scaling was by numbers of fish in the catch as estimated using the length-weight relationship and the LF in each depth stratum within each zone. There was adequate sampling for appropriate stratification for two years in the first time block (1998-99), six years in the middle time block, and three years in the last time block (Table 1) - which looks like an adequate number of LFs to estimate trawl-fishery selectivities (with the help of age-length data which play a role because of the age-based selectivities). The port samples for the eastern-trawl were not used as they could not be assigned to depth. This should not matter as there are plenty of measured fish included in the model - and we don't want length frequencies to do much more than inform on selectivities and, perhaps, year class strengths (YCS).

Table 1: Number of trawls sampled (onboard) and number of fish measured for eastern trawl length data. Highlighted fields indicate data that were used. See text for acceptance rules.

|  | Number of trawls sampled |  |  | Number of fish measured |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | SF10 | SF20 | SF30 | SF10 | SF20 | SF30 |
| 1993 | 0 | 11 | 0 | 0 | 370 | 0 |
| 1994 | 0 | 16 | 0 | 0 | 1026 | 0 |
| 1995 | 0 | 6 | 0 | 0 | 704 | 0 |
| 1996 | 0 | 28 | 0 | 0 | 2110 | 0 |
| 1997 | 0 | 20 | 1 | 0 | 1100 | 114 |
| 1998 | 36 | 26 | 7 | 3541 | 2649 | 728 |
| 1999 | 63 | 15 | 4 | 4814 | 2260 | 488 |
| 2000 | 34 | 6 | 6 | 2104 | 548 | 516 |
| 2001 | 46 | 22 | 5 | 2723 | 2276 | 514 |
| 2002 | 20 | 17 | 4 | 1126 | 1402 | 454 |
| 2003 | 19 | 20 | 2 | 1119 | 1528 | 237 |
| 2004 | 21 | 1 | 0 | 749 | 63 | 0 |
| 2005 | 26 | 23 | 3 | 1039 | 1509 | 217 |
| 2006 | 15 | 22 | 1 | 716 | 2408 | 9 |
| 2007 | 6 | 7 | 0 | 110 | 66 | 0 |
| 2008 | 5 | 7 | 1 | 200 | 161 | 5 |
| 2009 | 4 | 10 | 0 | 98 | 371 | 0 |
| 2010 | 9 | 15 | 1 | 254 | 385 | 7 |
| 2011 | 9 | 28 | 2 | 153 | 642 | 32 |
| 2012 | 9 | 11 | 8 | 220 | 212 | 200 |

The eastern-non-trawl LFs were stratified by zone only as the catch is just in Zones 20 and 30 which do not have the very strong depth effect seen in Zone 10 . Port and onboard samples were combined as it is assumed that they have the same selectivity pattern (following the 2012 stock assessment). The number of operations was calculated as onboard operations +3 times the port operations. Although the factor of 3 is fairly arbitrary it recognizes that port samples are of landings which contain a multiple number of shots. Years were accepted if they had at least 5 operations and 100 fish (Table 2).

Table 2: Number of calculated operations (see text) and number of fish measured for eastern non-trawl length data (port and onboard combined, Zones 20 and 30 combined). Highlighted fields indicate data that were used. See text for acceptance rules.

|  | Total <br> number of <br> operations | Total <br> number <br> of fish |
| ---: | ---: | ---: |
| 2000 | 3 | 10 |
| 2001 | 0 | 0 |
| 2002 | 64 | 4586 |
| 2003 | 81 | 4362 |
| 2004 | 78 | 5831 |
| 2005 | 27 | 1367 |
| 2006 | 14 | 788 |
| 2007 | 8 | 275 |
| 2008 | 9 | 256 |
| 2009 | 23 | 892 |
| 2010 | 36 | 1235 |
| 2011 | 31 | 1049 |
| 2012 | 48 | 1114 |

For the west, no stratification appeared to be needed and port and onboard samples were combined within gear-type (using operations $=$ onboard +3 times port). Almost all data were used with the acceptance rule of at least 5 operations and 100 fish (Table 3).

Table 3: Number of calculated operations (see text) and number of fish measured for western length data (port and onboard combined, Zones 40 and 50 combined). Highlighted fields indicate data that were used. See text for acceptance rules.

|  | Total number of <br> operations |  | Total number of <br> fish measured |  |
| ---: | ---: | ---: | ---: | ---: |
| 1992 | 15 | 0 | 539 | 0 |
| 1993 | 16 | 0 | 904 | 0 |
| 1994 | 17 | 0 | 1236 | 0 |
| 1995 | 68 | 0 | 4810 | 0 |
| 1996 | 51 | 3 | 2538 | 234 |
| 1997 | 81 | 0 | 4920 | 0 |
| 1998 | 44 | 0 | 3042 | 0 |
| 1999 | 31 | 1 | 1979 | 351 |
| 2000 | 62 | 9 | 2188 | 58 |
| 2001 | 105 | 13 | 3576 | 1566 |


| 2002 | 92 | 10 | 2507 | 2459 |
| ---: | ---: | ---: | ---: | ---: |
| 2003 | 97 | 31 | 3423 | 2382 |
| 2004 | 99 | 40 | 2769 | 4696 |
| 2005 | 68 | 8 | 1933 | 684 |
| 2006 | 27 | 5 | 937 | 397 |
| 2007 | 12 | 21 | 828 | 727 |
| 2008 | 5 | 0 | 133 | 0 |
| 2009 | 24 | 0 | 258 | 0 |
| 2010 | 49 | 10 | 494 | 212 |
| 2011 | 35 | 0 | 685 | 0 |
| 2012 | 43 | 22 | 827 | 805 |

## Stratification and scaling of age-length data

Cordue (draft-2) showed that stratification and scaling of age-length data was needed for the eastern zones because of variation in age at length across the zones. It was recommended that only Zones 10 and 20 be used for trawl (as Zone 30 has little trawl catch) and Zones 20 and 30 be used for non-trawl (as Zone 10 has almost no non-trawl catch). To increase the amount of data (and number of years) that were used, fairly lenient criteria were adopted. For easterntrawl the rule was at least 2 operations and 29 fish in each stratum. For western-trawl (which was not stratified) the rule was at least 4 operations and 90 fish in Zones 40 and 50 combined. This resulted in little loss of data with a good number of years represented and plenty of individual sexed age-length data for the model (to inform with regard to growth and YCS)(Table 4).

Table 4: Number of apparent sampling operations and number of sexed fish for trawl age-length data. Highlighted fields indicate data that were used (yellow = east, green = west). See text for acceptance rules.

|  | Number of operations |  |  |  |  |  | Number of fish (aged, sexed, measured) |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | SF10 | SF20 | SF30 | SF40 | SF50 | SF10 | SF20 | SF30 | SF40 | SF50 |
| 1979 | 4 | 2 | 0 | 0 | 0 | 139 | 241 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 23 | 0 | 0 |
| 1983 | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 146 | 0 | 0 |
| 1984 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 4 | 0 | 0 |
| 1985 | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 141 | 0 | 0 |
| 1986 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 31 | 0 | 0 |
| 1987 | 0 | 0 | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 487 |
| 1988 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 309 |
| 1989 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 185 |
| 1993 | 0 | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 46 | 0 |
| 1994 | 2 | 2 | 0 | 2 | 1 | 93 | 47 | 0 | 14 | 247 |
| 1995 | 3 | 2 | 0 | 1 | 3 | 92 | 195 | 0 | 1 | 314 |


| 1996 | 28 | 2 | 1 | 2 | 2 | 673 | 34 | 21 | 11 | 66 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1997 | 19 | 0 | 0 | 0 | 6 | 570 | 0 | 0 | 0 | 528 |
| 1998 | 16 | 3 | 0 | 5 | 4 | 418 | 125 | 0 | 109 | 95 |
| 1999 | 13 | 4 | 0 | 15 | 4 | 324 | 134 | 0 | 330 | 154 |
| 2000 | 4 | 1 | 0 | 3 | 1 | 109 | 21 | 0 | 44 | 27 |
| 2001 | 7 | 2 | 1 | 1 | 0 | 290 | 103 | 48 | 93 | 0 |
| 2002 | 6 | 0 | 2 | 2 | 2 | 220 | 0 | 100 | 99 | 99 |
| 2003 | 2 | 0 | 0 | 3 | 2 | 78 | 0 | 0 | 95 | 81 |
| 2004 | 3 | 0 | 1 | 3 | 13 | 61 | 0 | 6 | 144 | 332 |
| 2005 | 3 | 4 | 1 | 1 | 6 | 43 | 29 | 57 | 78 | 202 |
| 2006 | 0 | 0 | 3 | 1 | 12 | 0 | 0 | 27 | 50 | 424 |
| 2007 | 0 | 12 | 1 | 2 | 0 | 0 | 0 | 198 | 0 | 0 |
| 2008 | 0 | 6 | 0 | 2 | 0 | 0 | 0 | 0 | 45 | 0 |
| 2009 | 0 | 30 | 1 | 3 | 1 | 0 | 0 | 0 | 16 | 23 |
| 2010 | 11 | 36 | 0 | 6 | 2 | 89 | 113 | 0 | 95 | 0 |
| 2011 | 4 | 31 | 4 | 3 | 3 | 31 | 49 | 38 | 45 | 47 |
| 2012 | 2 | 18 | 2 | 2 | 4 | 21 | 125 | 42 | 24 | 72 |

The eastern-trawl age-length data were scaled by the relative average proportions of catch in Zones 10 and 20 in each of the time blocks used in the eastern trawl fishery. That is, a constant scalar was used within each time block rather than scaling to the annual catch proportions. The idea is to present a consistent set of data to the model within each timeblock (within which is assumed a constant selectivity). The catch proportions for each time block were:

|  | $1986-1999$ | $2000-2006$ | $2007-2012$ |
| :--- | ---: | ---: | ---: |
| SF10 | 0.35 | 0.21 | 0.19 |
| SF20 | 0.65 | 0.79 | 0.81 |

For the first time block the average was calculated from 1986 to 1999 to avoid the early period of relatively low catches that were dominated by Zone 10 (again, an attempt to present a consistent signal to the model).

For the non-trawl sexed age-length data no stratification was attempted as there were almost no years with adequate samples in both potential strata (east: Zones 20 and 30; west: zones 40 and 50). In both the east and west, 2 operations and 40 fish were required. This meant that most of the individual age-length measurements were used (Table 5).

Table 5: Number of apparent sampling operations and number of sexed fish for nontrawl age-length data. Highlighted fields indicate data that were used (yellow = east, green $=$ west). See text for acceptance rules.

|  | SF10 | SF20 | SF30 | SF40 | SF50 | SF10 | SF20 | SF30 | SF40 | SF50 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1995 | 1 | 1 | 0 | 0 | 0 | 12 | 2 | 0 | 0 | 0 |
| 1999 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 154 |
| 2000 | 1 | 0 | 1 | 1 | 0 | 94 | 0 | 248 | 3 | 0 |
| 2001 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 275 | 0 |
| 2002 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 50 | 0 |
| 2003 | 0 | 8 | 3 | 12 | 0 | 0 | 279 | 30 | 305 | 0 |
| 2004 | 0 | 4 | 0 | 2 | 0 | 0 | 93 | 0 | 48 | 0 |
| 2005 | 0 | 1 | 0 | 0 | 3 | 0 | 48 | 0 | 0 | 137 |
| 2006 | 0 | 0 | 0 | 1 | 2 | 0 | 0 | 0 | 124 | 246 |
| 2007 | 0 | 3 | 0 | 3 | 0 | 0 | 14 | 0 | 40 | 0 |
| 2008 | 0 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2009 | 0 | 20 | 4 | 1 | 1 | 0 | 147 | 0 | 8 | 0 |
| 2010 | 0 | 24 | 0 | 4 | 0 | 0 | 0 | 0 | 48 | 0 |
| 2011 | 0 | 24 | 9 | 0 | 0 | 0 | 23 | 34 | 0 | 0 |
| 2012 | 0 | 11 | 0 | 2 | 1 | 0 | 181 | 0 | 181 | 7 |

The number of operations were used as starting values for the effective sample sizes of the LFs. For the age-length data, the observations go in as individual fish, so that fish numbers determine the effective sample size. However, the data can still be down-weighted by maintaining the same proportions of age-at-length (and sex) but putting in fewer individual fish.

For the eastern assessment, only the middle-time-block LFs were tuned, but sample sizes in the other blocks were scaled by the same factor (it is desirable to have at least 5 years in a block for tuning). The non-trawl LFs were not tuned as they would have had so little weight on them that the predicted mean lengths would have been badly biased low (there is a conflict with some of the age-length data through the age-based selectivities). For them the effective sample sizes were fixed at two times the number of operations.

For the western assessment, both time blocks were tuned for the trawl data. As for the east, the non-trawl data required two times the number of operations to keep the estimated mean length at selection fairly unbiased. Also, a serious conflict was seen between the trawl agelength data and the trawl CPUE indices which required the weight on the trawl age-length data to be reduced (fish numbers were halved).

## Construction of age frequencies

In the eastern assessment, some of the stratified length frequencies in recent years were converted to age frequencies using the unsexed age-length data from Zone 20 (and, for 2012,
some unused sexed age-length data). This was considered reasonable despite the variation of age at length between the eastern zones because Zone 20 has older fish at length than Zone 10 and younger fish at length than Zone 30 (Cordue, draft-2) and also it has most of the catch in recent times. At least 200 fish were required in each year. Length bins of 5 cm were used and proportions-at-age across all years and eastern zones (within gear type) were assumed for bins which had no year-specific data. Good numbers of fish were available for 2010-2012 for trawl and for 2008-2011 for non-trawl and a high proportion of each length frequency was covered with year-specific data:

| Eastern <br> fishery | Year | Proportion <br> of LF <br> covered (\%) | Number <br> of fish |
| :--- | ---: | ---: | ---: |
|  | 2010 | 97 | 501 |
| Trawl | 2011 | 95 | 532 |
|  | 2012 | 95 | 447 |
|  |  |  |  |
|  | 2008 | 95 | 585 |
| Non-trawl | 2009 | 100 | 353 |
|  | 2010 | 94 | 333 |
|  | 2011 | 100 | 511 |

The effective sample sizes in the model were fixed at a tenth of the number of fish (not really enough years in each time series to apply the methods of Francis, 2011).

## Model structure

A single-area model with a single time-step was used for both stocks. Ages (1-30+), sex, and maturity were in the partition (although the latter is irrelevant as there were no fisheries that preferentially selected mature fish). The two fisheries (trawl and non-trawl) were assumed to be year-round and mortality was modelled using the Pope approximation to Baranov (CASAL's standard option). Further details of the models are:

| Model years | $1970-2013$ | Stock status assessed mid- <br> year 2013 |
| :--- | :--- | :--- |
| Biomass parameterisation | $B_{0}$ | Estimated parameter. $R_{0}$ is <br> derived. |
| Recruitment parameterisation | Haist, lognormal prior, <br> sigmaR =0.7 | Also, a moderate penalty on <br> year class strengths (YCS) <br> averaging to 1. |
| YCS estimated (i.e., <br> recruitment deviations) | East: $1969-1977$, 1983-2009 <br> West: 1975-2009 initially, <br> but 2009 dropped as its very <br> high estimate was based on <br> little data. | Required to have at least 8 <br> observations in the sexed <br> age-length data used |
| Steepness | 0.75 | As used in 2012. A <br> conservative value - it could |


|  |  | be higher. Fixed. |
| :--- | :--- | :--- |
| Maturity | Logistic at age: <br> $\mathrm{a}_{50}=5 \mathrm{yr}$ ato95 $=2 \mathrm{yr}$ | Approximates the length- <br> based curve used in the 2012 <br> assessment. Fixed. |
| Trawl selectivities | Three blocks in the east: <br> $1970-99$, 2000-2006, 2007- <br> 2013. Two in the west: 1970- <br> 2006, 2007-2013. Separate <br> male and female selectivities, <br> double normal at age. | Estimated in the model. <br> Timing of blocks indicated <br> by events and confirmed by <br> data analysis. Different from <br> the block timing used by <br> CSIRO. |
| Non-trawl selectivities | Logistic at age, separate for <br> males and females. | Estimated in the model |
| Growth | Separate male and female <br> von Bertalanffy | Estimated in the model |
| Length-weight relationship | a 2.93e-9 <br> b 3.139 | Fixed at 2012 assessment <br> values. (cm to tonnes) |

## Results

## Eastern stock

## Data preparation

It is hard to know how much difference good data preparation will make to stock assessment results. However, the only way to find out is to do it. A number of small changes, each favouring good practice, can have a strong cumulative effect.

The stratification of the eastern-trawl LFs may perhaps not make much difference (as a single component of change) but it does present a more consistent signal to the model than it would have got otherwise (Figure TB1). The timing of a 1999 to 2000 change in selectivity is confirmed as is a shift to higher mean lengths in recent years (a probable shift in selectivity as a consequence of the structural adjustment). In comparison, the unstratified LFs present a noisy picture although the first two points would perhaps be omitted as they have very low sample sizes.


Figure TB1: Mean length for unstratified east-trawl LFs (open symbols) and depth and zone stratified LFs (solid symbols).

## MPD fits

The fit to the ISL trawl CPUE series is good (Figure 1). The 1986 point is missed, but after that the fit is very good. The fit to the non-trawl age frequencies is also good given that they jump around and the effective sample sizes are less than 60 (Figure 2). The same comments apply to the trawl age frequencies (Figure 3). For the length frequencies the effective sample sizes are larger. The fits are reasonable except that the predictions fail to get as "peaky" as the observations and there is a tendency for the predicted means (and modes) to be a bit higher than the observations (Figures 4 and 5).

For the age-length data, that are input as individual measurements, it is difficult to produce plots that show how well the data are fitted. Numerous conditional age-at-length plots can be produced but the sheer quantity of them reduces their usefulness. However, at a minimum the age-length data should be seen to be consistent with the estimated growth curves. This is the case for the ISL models where only sexed age-length data were used as individual measurements. The fit to the male age-length data looks excellent (Figure 6) as does the fit to the female age-length data although the $L_{i n f}$ looks like it is getting a bit high (Figure 7). This has no consequence as there is little biomass beyond about age 20 even in the virgin population.

## Likelihood profiles

It is sometimes useful to check likelihood profiles as models are developed to make sure that composition data are not driving the biomass signals entirely and also to see if there appears to be a sound basis for estimating parameters such as natural mortality $(M)$.

The likelihood profile for $B_{0}$ shows a well-defined minimum at about 6000 t (Figure 8). In terms of individual components most of them find a minimum in the 6000-8000 t range; the exceptions are the age-length data from trawl and the trawl CPUE (Figure 8). The trawl CPUE has its minimum negative-log-likelihood at a low level probably because it is quite up-and-down which means it is easier to fit when there is low biomass (as that can be pushed around more easily by variation in recruitment). Just about everything is saying "don't go too low", and the three components that say "don't go too high" include the CPUE (Figure 8). Overall, the balance doesn't look too bad.

For $M$ there is a serious conflict between the trawl age-length data and the trawl length frequencies (Figure 9). The trawl length frequencies oppose all other data sets as it favours a value of $M$ at about 0.15 (Figure 9). Most other components favour a value in the 0.24-0.30 range. An informed prior was put on $M$ to discourage values above 0.3 (the prior is moderately informed being normal with mean $=0.2$ and $\mathrm{cv}=0.2$ (it is based on NZ ling and the standard default for $M$ of 0.2).

## MPD estimates

The estimated fishing selectivities show marked differences between gears and sexes and, for trawl, between time blocks (Figure 10). The non-trawl was forced to be asymptotic and the trawl selectivities were allowed to be domed - and the estimates are very domed (Figure 10). Males are selected preferentially compared to females for the non-trawl fishery and the trawl fishery in the last time block, which is inconsistent with length-based selection (as females grow faster - Figure 11). The shift in the mode of the trawl selectivities over time shows a move to progressively larger fish (Figure 10).

The pattern of recruitment estimated suggests moderate recruitment variability with no very large YCS and only one very weak YCS in 1977 (Figure 12). There are no extended periods of low or strong recruitment estimated (Figure 12).

The estimated female spawning biomass trajectory shows a steady decline from a local peak in 1980 to a level of about $25-30 \% B_{0}$ since the mid 1990s (Figure 13).

## Audit trail from the 2012 base model

In the 2012 eastern base model the catch histories for trawl and non-trawl were accidently swapped (I noticed this when looking at the SS3 data and control files which were supplied by CSIRO). The base model had an estimate for 2013 beginning-of-year stock depletion of $26 \% B_{0}$. When the catches are correctly entered the estimate is $19 \% B_{0}$ (I fixed the catches and reran the SS3 model).

It is important to understand what changes in data and/or model structure can move us from the $19 \% B_{0}$ estimate up to the $30 \% B_{0}$ estimate in the ISL base model. Therefore, a series of runs were done with intermediate steps between the 2012 SS3/CSIRO model (with correct catches) and the 2013 CASAL/ISL base model (Table 6) The results are revealing (Table 6, Figure 14).

The first step was a switch from SS3 to a CASAL run with the ISL base-model structure, the most recent catch history estimates, no trawl-caught port length samples, and a constant CV of $15 \%$ on the trawl CPUE. The comparison is $19 \%$ (SS3) to $12 \%$ (CASAL: step 1). The difference is not large and is due mainly, I think, to minor differences between SS3 and CASAL, the differences in the selectivity parameterisations, and the use of mid-year stock depletion instead of beginning-of-year (SS3). Dropping the port data had little effect as I have done other runs including the port data and got similar results.

The second step was to drop the Kapala data and switch to the properly stratified composition data. This made a big difference with a jump from $12 \%$ to $20 \%$. Sensitivity tests suggest that the exclusion of the Kapala data has little impact and hence this change was due to the different composition data and the change in relative weightings associated with it. I believe that the change in composition data was the most important component because the two runs have very similar estimates of biomass and natural mortality (Table 6). When comparing the estimates from the two runs I found that the big changes were for selectivity and YCS.

Dropping the non-trawl CPUE boosted the estimate a little bit from $20 \%$ to $23 \%$. The biggest change was when the ISL trawl CPUE series was introduced and this increased the 2012 stock-status estimate from $23 \%$ to $32 \%$ (Table 6). The inclusion of the newly available data in 2013 reduced the estimate of 2012 stock status from $32 \%$ to $31 \%$.

Table 6: Eastern stock: audit trail from SS3 2012 base model to 2013 ISL base model. Estimated female virgin biomass ( $B_{0}$ ), stock status in the 2012 assessment year, and natural mortality are given for a series of runs linking the two models. * SS3 virgin biomass estimates are not comparable with CASAL virgin biomass estimates because of a different maturity definition.

| Model run | $\boldsymbol{B}_{\boldsymbol{0}}(\mathbf{t})$ | Stock status <br> $\left(\boldsymbol{\%} \boldsymbol{B}_{\boldsymbol{0}}\right)$ | $\boldsymbol{M}\left(\mathbf{y r}^{\mathbf{- 1}}\right)$ |
| :--- | ---: | ---: | ---: |
|  |  |  |  |
| SS3 2012 base with data to the end of 2011 | $7296^{*}$ | 26 | 0.24 |
| SS3: trawl and non-trawl catches the right way round | $6926^{*}$ | 19 | 0.24 |
| As above, CASAL with ISL base-model structure | 5734 | 12 | 0.26 |
| As above, drop Kapala data and use ISL age \& length | 5675 | 20 | 0.24 |
| As above, drop non-trawl CPUE | 5752 | 23 | 0.24 |
| As above, switch to ISL trawl CPUE | 6039 | 32 | 0.25 |
| As above, add in new data = ISL base model | 5932 | 31 | 0.25 |

## MPD sensitivities

Various sensitivities to the base model were run (Table 7, Figure 15). The sensitivity of stock status to the trawl CPUE time series was confirmed with the lowest estimate of stock status, among all the runs tried, occurring when the CSIRO CPUE series was used (Table 7, Figure 15). Alternative maturity schedules had almost no effect on current stock status, but did change the trajectory in earlier years (Figure 15). The inclusion of the Kapala data had little effect, but including the non-trawl CPUE reduced stock status from $30 \%$ to $26 \%$. Lower and higher values of $M$ (which cannot be ruled out) produced lower and higher values of stock status ( $24 \%$ and $38 \% B_{0}$ ).

Table 7: Eastern stock: ISL-base-model sensitivities. Estimated female virgin biomass $\left(B_{0}\right)$, stock status in the 2013 assessment year (mid-year $\left.B_{13} / B_{0}\right)$, and natural mortality are given for a series of runs which differ from the base model in only one respect.

| Model run | $\boldsymbol{B}_{\boldsymbol{0}}(\mathbf{t})$ | Stock status <br> $\left(\boldsymbol{\%} \boldsymbol{B}_{\mathbf{0}}\right)$ | $\boldsymbol{M}\left(\mathbf{y r}^{\mathbf{- 1}}\right)$ |
| :--- | ---: | ---: | ---: |
|  |  |  |  |
| ISL base model | 5932 | 30 | 0.25 |
| Using CSIRO trawl CPUE | 5644 | 22 | 0.25 |
| Including Kapala data | 5710 | 32 | 0.25 |
| Including non-trawl CPUE | 5517 | 26 | 0.26 |
| With fixed and lower $M$ | 6604 | 24 | 0.22 |
| With fixed and higher $M$ | 5602 | 38 | 0.28 |
| With maturity ogive shifted 1 year younger | 6429 | 30 | 0.25 |
| With maturity ogive shifted 1 year older | 5271 | 31 | 0.26 |

## Preliminary MCMC results

For the eastern assessment there was time to run a preliminary MCMC to obtain estimates of the posterior distribution. The median of the marginal posterior distribution, for each parameter/derived-parameter of interest, is usually used as the point estimate for that parameter. It has superior statistical properties as an estimator compared to the MPD. This is intuitively reasonable, as it provides an estimate of centrality for the whole distribution rather than just the location of the peak (i.e., the median is in the middle of the most likely place to find the true value).

Three independent chains of 1.7 million were each run. One in every one thousand samples were retained. Plots of the three chains showed good agreement between the main variables of interest (virgin and current biomass, stock status and natural mortality). A burn-in of 400,000 was indicated by plots of the objective function and this was used. The remaining parts of each of the three chains were combined.

The YCS are fairly well determined and recent recruitment is a bit below average, with poor recruitment estimated for 2009 (Figure 27). Fishing selectivities are fairly well determined with definite doming for the trawl fishery and an increase in the age of full selection over the time blocks (Figure 28).

Virgin biomass was estimated in the 5000-6500 t range (Table 8, Figure 29) with current female mid-year spawning biomass from 1400-2800 t (Table 8). The current stock status/depletion has a median of $35 \% B_{0}$ with a $95 \%$ CI of $25-46 \% B_{0}$. The MCMC median is a bit higher than the MPD estimate (Figure 30). Natural mortality is fairly tightly estimated with a median at 0.27 (Table 8).

The spawning-stock biomass trajectory shows a peak above virgin levels in the 1980s followed by a steady decline to about $30 \% B_{0}$ in the late 1990 s and a slow increase from then to about $35 \% B_{0}$ (Figure 31).

Table 8: MCMC estimates for the ISL eastern base model. The point estimate (median of marginal posterior distribution) and the $\mathbf{9 5 \%}$ credibility interval are given for virgin biomass, current biomass stock status, and natural mortality.

|  | $\boldsymbol{B}_{\boldsymbol{0}}(\mathbf{t})$ | $\boldsymbol{B}_{2013}(\mathbf{t})$ | Stock status <br> $\left(\boldsymbol{\%} \boldsymbol{B}_{\boldsymbol{0}}\right)$ | $\boldsymbol{M}\left(\mathbf{y}^{\mathbf{- 1}}\right)$ |
| ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |
| Median | 5738 | 1990 | 35 | 0.27 |
| $95 \% \mathrm{CI}$ | $5140-6422$ | $1366-2825$ | $25-46$ | $0.24-0.30$ |

## Western stock

## Data preparation

For the western stock, data preparation was not an issue as there seemed to be little variation in mean length or age-at-length by zone or depth. Also, the detailed CPUE analysis yielded almost identical indices to those produced by CSIRO's generic approach.

## MPD fits

The fit to the trawl CPUE indices in the reference model is very good except for a strong residual pattern associated with the steep rise and decline of the indices during the 1990s and early 2000s (Figure 16). The predicted biomass does not rise or decline as steeply as the indices but it does peak in the same year (1998).

The fits to the length frequencies are very good, especially for trawl, given the low effective sample sizes used (Figures 17 and 18). The fits to the sexed age-length data look good except that the male growth curve has slightly more curvature through the middle than the data show - perhaps a slight failing caused by requiring the von Bertalanffy parameterisation (Figures 19 and 20).

## MPD estimates

The estimated fishing selectivities, as for the eastern stock, show a departure from just length based selection, with males preferentially selected at young ages in the non-trawl fishery (Figure 21). There is also a significant shift in estimated selection in the two trawl time blocks (Figure 21). As in the east, and as expected, females grow to a larger size than males (Figure 22) although the estimated mean-lengths at infinity are smaller than in the east (see Figure 11). The estimated YCS, as for the east, showed only moderate variability, with just a couple of YCS being above 2 and no very low ones (Figure 23).

The estimated female spawning-biomass trajectory shows two periods of increase and decline, and the biomass stays between $80-120 \%$ of its virgin levels up until the mid 2000s and is still estimated at $70 \% B_{0}$ in 2013 (Figure 24). The high estimates of stock status are driven by natural mortality which was estimated at 0.3 (Table 9).

## MPD sensitivities

A serious of sensitivity runs for the western reference model were revealing. The results are not sensitive to the use of the CSIRO or ISL trawl CPUE indices or the use or not of the nontrawl CPUE indices (Table 9, Figure 25). They are also insensitive to the use of younger or older maturity.

There is a dramatic sensitivity to the weight placed on the trawl age-length data, with greater weight decreasing estimated stock status (Table 9, Figure 25). In the reference model effective numbers for the trawl age-length data were halved, and in the sensitivity run the full numbers were used. In the sensitivity run, not only is there a huge decrease in estimated stock status ( $71 \% B_{0}$ down to $47 \% B_{0}$ ) but there is an associated large decrease in the estimate of $M$
( 0.3 down to 0.22 ). It is clear that the weight on the trawl age-length data had to be reduced as it was preventing a good fit the trawl CPUE indices (Table 9, Figure 26). However, the reduction in weight on the age-length data allowed the estimate of $M$ to "blow out" a bit and led to the somewhat implausible estimate of stock status.

Fixing, rather than estimating $M$, shows that the model can still fit the trawl CPUE indices well at lower values of $M$ - although the fit is not quite as good (Table 9). Essentially the model is trying to fit the steep decline from 1998. If there is high weight on the trawl agelength data then it cannot fit the CPUE decline until a year after 1998 (Figure 26). However, when the weight is taken off that data, it can fit the decline a year earlier and it can estimate a higher value of $M$ (which also helps fit the decline because it can then kill fish more quickly). The solution to this problem is to either fix $M$ (at a "sensible" value) or estimate it with a strongly informed prior. Since the eastern assessment appears to provide a good estimate of $M$, the posterior from the eastern assessment is the obvious prior to use for the western assessment (this is currently ISL's preferred option for a western base model - but discussions at the RAG meeting may change this view).

Table 9: Western stock: ISL-reference-model sensitivities. Estimated female virgin biomass ( $B_{0}$ ), stock status in the 2013 assessment year (mid-year $B_{13} / B_{0}$ ), and natural mortality are given for a series of runs which differ from the base model in only one respect.

| Model run | $\boldsymbol{B}_{\boldsymbol{0}}(\mathbf{t})$ | Stock status <br> $\left(\% \boldsymbol{B}_{0}\right)$ | $\boldsymbol{M}\left(\mathbf{y r}^{\mathbf{- 1}}\right)$ | Trawl <br> CPUE neg- <br> log-like. |
| :--- | ---: | ---: | ---: | ---: |
|  |  |  |  |  |
| ISL reference model | 12222 | 71 | 0.30 | -40.6 |
| Using CSIRO trawl CPUE | 12148 | 71 | 0.30 | -40.7 |
| Including non-trawl CPUE | 12186 | 70 | 0.30 | -40.8 |
| With fixed and lower $M$ | 5965 | 58 | 0.26 | -39.7 |
| With fixed and even lower $M$ | 4890 | 48 | 0.22 | -38.3 |
| With maturity ogive shifted 1 year younger | 13277 | 74 | 0.30 | -40.7 |
| With maturity ogive shifted 1 year older | 10121 | 68 | 0.30 | -40.5 |
| Twice the effective sample size on trawl age-length | 3872 | 47 | 0.21 | -28.1 |

## Conclusions

The eastern and western assessments have very different challenges to deal with in terms of providing a reliable stock assessment. For the east, the issues are mainly to do with data preparation, while in the west, the issues primarily involve model and data weighting decisions.

The analysis of length and age-at-length data showed that the eastern data had to be stratified and scaled to some extent otherwise the model would be given misleading signals. There is a trade-off between how much stratification and scaling is done and how much data can be
used in the model. The acceptance rules used in the ISL-base model appear to provide a good balance between the quality and quantity of data used. The analysis also showed that the assumption of length-based selection in the fisheries was not supported; nor was the use of cohort-specific growth.

The CPUE analysis, to some extent, was able to deal with the issue of the sale of trawl quota in the east in 1999 and the structural adjustment in 2006. The approach used to allow a change in vessel behaviour to alter the trawl CPUE indices had a strong effect on the eastern indices over the 1999-2000 breakpoint but little effect over the 2006-2007 breakpoint. Consequently the choice of the ISL or CSIRO CPUE indices in the east has an impact on the assessment results, but it is not important in the west.

In the western assessment, there is a strong conflict between the trawl age-length data and the trawl CPUE indices. This indicates that a down-weighting of the trawl age-length data is required. However, when this is done the estimate of $M$ appears to be a bit too high. As there are little data in the western assessment to provide information on $M$, it is best to either fix $M$ at a "sensible" value, or to use a strongly informed prior. In both cases, it appears that the eastern assessment can provide the needed information on $M$.

## Acknowledgements

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Figure 1: Eastern base model fit to trawl CPUE indices. Dashed lines are 95\% CIs.


Figure 2: Eastern base model fit to non-trawl age frequencies.


Figure 3: Eastern base model fit to trawl age frequencies.


Figure 4: Eastern base model fit to non-trawl length frequencies (2002-2005).


Figure 4 (cont): Eastern base model fit to non-trawl length frequencies (2006, 2007, 2012).


Figure 5: Eastern base model fit to trawl length frequencies (1998-2001).


Figure 5 (cont): Eastern base model fit to trawl length frequencies (2002-2006).


Figure 6: Eastern base model: comparison of estimated male growth curve to all male age-length data, caught by line, in the eastern zones.


Figure 7: Eastern base model: comparison of estimated female growth curve to all female age-length data, caught by line, in the eastern zones.


Figure 8: Eastern base model: likelihood profile for $B_{0}$ showing each of the grouped components (ALK = age-length data). The dashed line shows the shape of the total negative log likelihood.


Figure 9: Eastern base model: likelihood profile for $M$ showing each of the grouped components (ALK = age-length data). The dashed line shows the shape of the total negative log likelihood.


Figure 10: Eastern base model: estimated fishery selectivities (age-based, sex specific, three time blocks for trawl).


Figure 11: Eastern base model: estimated growth curves (sex specific).


Figure 12: Eastern base model: estimated "true year class strengths" (annual recruitment divided by virgin recruitment).


Figure 13: Eastern base model: estimated stock status (annual SSB divided by virgin SSB; female biomass only; mid-year). Grey lines at 0.3 and 0.4 , red line at 0.2 .


Figure 14: Eastern base model: stock status audit trail for CASAL models, starting with compatible data used in the 2012 assessment and the ISL base model structure (see text for details of steps 1-4). Grey lines at $\mathbf{0 . 3}$ and $\mathbf{0 . 4}$, red lines at 0.1 and $\mathbf{0 . 2}$.


Figure 15: Eastern base model: estimated stock status for the base model and sensitivity runs. Grey lines at 0.3 and 0.4 , red lines at 0.1 and 0.2 .


Figure 16: Western reference model: fit to ISL trawl CPUE indices. The dashed lines show 95\% CIs.


Figure 17: Western reference model fit to non-trawl length frequencies (2001-2004).


Figure 17 (cont): Western reference model fit to non-trawl length frequencies (20052007, 2010).


Figure 17 (cont): Western reference model fit to non-trawl length frequencies (2012).


Figure 18: Western reference model fit to trawl length frequencies (1992-1995).


Figure 18 (cont): Western reference model fit to trawl length frequencies (1996-1999).


Figure 18 (cont): Western reference model fit to trawl length frequencies (2000-2003).


Figure 18 (cont): Western reference model fit to trawl length frequencies (2004-2007).


Figure 18 (cont): Western reference model fit to trawl length frequencies (2008-2011).


Figure 18 (cont): Western reference model fit to trawl length frequencies (2012).


Figure 19: Western reference model: comparison of estimated male growth curve to all male age-length data, caught by line, in the western zones.


Figure 20: Western reference model: comparison of estimated female growth curve to all female age-length data, caught by line, in the western zones.


Figure 21: Western reference model: estimated fishery selectivities (age-based, sex specific, two time blocks for trawl).


Figure 22: Western reference model: estimated growth curves (sex specific).


Figure 23: Western reference model: estimated "true year class strengths" (annual recruitment divided by virgin recruitment).


Figure 24: Western reference model: estimated stock status (annual SSB divided by virgin SSB; female biomass only; mid-year). Grey lines at 0.3 and 0.4 , red line at 0.2 .


Figure 25: Western reference model: estimated stock status for the reference model and sensitivity runs. Grey lines at 0.5 and 0.7 , red line at 0.2 .


Figure 26: Western reference model: fit to ISL trawl CPUE indices for the reference model (half weight on trawl age-length data) and the run with full weight on the trawl age-length data . The dashed lines show $95 \%$ CIs.


Figure 27: Eastern ISL base model: posterior distribution of "true" YCS ( $\boldsymbol{R}_{y} / \boldsymbol{R}_{\boldsymbol{f}}$; over the full range of MCMC samples). Each box contains $50 \%$ of the samples, the line within each box marks the median.


Figure 28: Eastern ISL base model: posterior distributions for fishing selectivities (over the full range of MCMC samples). Each box contains $50 \%$ of the samples, the line within each box marks the median.


Figure 29: Eastern ISL base model: posterior distribution of $\boldsymbol{B}_{0}$ with the median marked in black and the MPD estimate in red.


Figure 30: Eastern ISL base model: posterior distribution of stock status $\left(B_{13} / B_{0}\right)$ with the median marked in black and the MPD estimate in red.


Figure 31: Eastern ISL base model: spawning-stock biomass trajectory. Each box contains $50 \%$ of the samples (in that year), the line within each box marks the median, and the dashed lines extend over the full range of MCMC samples. Green line at $\mathbf{3 0 \%}$.

## Tabulated issues for stock assessment of pink ling 2013

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7 September 2013

## Introduction

This report documents the main issues discussed by André and Patrick during the 2013 pink ling assessment with regard to the spatiallyaggregated models. The report is not comprehensive in that some (hopefully) minor issues are not documented and not all aspects of the assessment are covered.

There is also a section on spatial modeling which André put together but the issues there have not been discussed.

In particular, the following components of the assessment were not brought into issue and the approach taken by CSIRO in 2012 was continued:

- Assumption of two stocks, east (Zones 10, 20, 30) and west (Zones 40, 50) (except Zone 60 is now included in Zone 20)
- Construction of catch history (except Zone 60 is now included in Zone 20)
- Two fisheries: trawl and non-trawl
- Discards not modelled
- Construction of non-trawl CPUE indices
- Use of a two-sex population model
- Plus-group at age 30 in the population
- Construction and use of the ageing error matrix

The major issues which arose during the assessment were: (TO BE FILLED IN FURTHER WHEN WE KNOW FOR SURE)

- The preparation and use of composition data (pages $3,4,6$ )
- The derivation of standardised trawl CPUE for the eastern stock (page 5)
- The use of the trawl CPUE indices - whether to split series and/or ignore some points (page 16)
- Relative and absolute weighting of different data sets in the stock assessment models (page 15)

The bulk of this report consists of a series of tabulated issues. For each issue the nature of the issue is stated and several options are given, each with its advantages (pros) and disadvantages (cons). A statement is made on the importance of the issue and this is judged in terms of how much difference the choice of option could make to the results. Andre and Patrick individually graded each of the options and chose a single option as their individual preferred choice. Generally, two gradings were made by each person, an initial grading which was chosen before reviewing diagnostics and a final grading chosen after diagnostics were prepared and examined. Important diagnostics are referenced in the tables and the reader is directed to the appendices or another draft document. Do not pay too much attention to the number of records listed in tables in the appendices. These may differ from the actual number of records used in later analysis as different filtering criteria were used.

## Where final grades are not filled in, it means that the person has not yet reviewed/seen the diagnostics.

A grade of "Not a good option" is used sparingly in the initial grading because it means that the person believes there are serious technical issues with the option (there are minor technical issues with almost every option). However, in the final grading, after diagnostics have been examined, "Not a good option" also includes cases where the diagnostics suggest that the option should not be used in this assessment. The assignment of a grade is independent of whether the choice of option is considered important or not.

## Key to tables

Names: $\quad$ An $=$ André Punt, $\mathrm{P}=$ Patrick Cordue
Fill colour: Black $=$ Preferred option, Green $=$ Acceptable option, Red $=$ Not a good option.
In the 2012 column, black indicates the assumption used in the 2012 east and west base models.

## Data Issues

| Issue | Options | Pros | Cons |
| :---: | :---: | :---: | :---: |
| Construction of length frequencies for spatiallyaggregated models: possible scaling and stratification by zone, sampling method (port including Sydney Market, onboard), fishing method (trawl, non-trawl), and depth. | Scale sample to catch/landing and combine across zones. Separate LFs by sampling-method and fishing method. (Simple) | Simple approach which uses all of the data. | Unlikely to be representative of the catch because of unbalanced sampling across zones. |
|  | As above but stratify by zone. (Simple zone stratification) | Still relatively simple and uses all of the data. | Still doesn't deal with unbalanced sampling due to zero samples in some zones in some years. |
| Importance depends on average magnitude of effective sample sizes (N). Can impact on point estimates and reference points through changes in estimated selectivity. | Conduct a quick analysis to determine if there are substantial size differences between zones within fishing method. Stratify accordingly. Exclude years when inadequate sampling across strata. Separate LFs by sampling and fishing methods. (Analyze and stratify) | Standard approach which deals with unbalanced sampling across zones. | Extra analysis time needed. Produces four LF time series whereas perhaps only two are needed. Some data may be excluded. |
| Moderate-Major (high weight); Minor-Moderate (low weight). <br> Low weight: $\mathrm{N} \sim 10$ <br> High weight: $\mathrm{N} \sim 100$ | As above but check to see if separation of sampling methods is justified. If not, then separate LFs for fishing method only. Also check on need for stratification by depth. (Perhaps combine port and onboard; check depth) | The most comprehensive approach which deals with the unbalanced sampling and may produce just two LF time series. | Extra analysis time needed. Potential difficulty in finding a suitable starting vector of sample sizes if onboard and port combined. Some data may be excluded. |


| Issue | Options (short description) | $\underset{\sim}{\underset{\sim}{N}}$ | Initial preferences |  |  |  |  | Diagnostics checked | Final preferences |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Construction of length frequencies for spatiallyaggregated models |  |  | An |  | M | N | P |  | An |  | M |  | P |
|  | Simple |  |  |  |  |  |  | Descriptive analysis and linear |  |  |  |  |  |
|  | Simple zone stratification |  |  |  |  |  |  | modelling of length (see Appendix 1 and Appendix 2) |  |  |  |  |  |
|  | Analyze and stratify |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Perhaps combine port and onboard; check depth |  |  |  |  |  |  |  |  |  |  |  |  |


| Issue | Options | Pros | Cons |
| :--- | :--- | :--- | :--- |
| Construction of CPUE indices <br> from trawl catch and effort data <br> for use as biomass indices. | Standard CSIRO approach using <br> zone and month level positive- <br> catch data with depth filtering | A quick approach which may be <br> adequate and can be applied <br> generically to most SESSF stocks. | It doesn't provide much <br> understanding of the "fisheries" <br> and may not deliver the best <br> biomass indices. |
| Importance depends on how <br> different the CPUE indices are <br> from the two approaches. | Standard NZ approach: a <br> detailed descriptive analysis, <br> selective filtering, and use of fine- <br> scale data in the standardization. | Potentially provides some <br> understanding of what "fisheries" <br> are operating and hence gives a <br> good chance of filtering the data <br> appropriately. | Requires much more analysis time <br> and may not provide biomass <br> indices that are any different from <br> the quicker CSIRO approach. |
| Potentially: Minor, Moderate, or <br> Major |  |  |  |


| Issue | Options (short description) | $\stackrel{\text { N }}{\text { N }}$ |  | ial | ref | ren |  | Diagnostics checked | Final preferences |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Construction of CPUE indices from trawl catch and effort data |  |  | An | ${ }^{\text {At }}$ | M | N | ${ }^{\text {P }}$ |  | ${ }^{\text {An }}$ | ${ }^{\text {At }}$ | M | N | P |
|  | Standard CSIRO |  |  |  |  |  |  | Analysis of catch and effort data |  |  |  |  |  |
|  | Standard NZ approach |  |  |  |  |  |  | (see cpue.docx) - the CSIRO approach is fine for the west, but |  |  |  |  |  |
|  |  |  |  |  |  |  |  | perhaps not for the east. |  |  |  |  |  |


| Issue | Options | Pros | Cons |
| :--- | :--- | :--- | :--- |
| Preparation of age data sets. | Use all age data as age-at-length <br> without any stratification or <br> scaling | A common SS3 approach. | No stratification or scaling, so <br> unbalanced sampling may mean <br> that data are unrepresentative. |
| Importance is hard to judge <br> before results are seen and it will <br> depend on data weightings. | Use all age data as age-at-length <br> but attempt some stratification <br> and scaling | Data may be more representative <br> than if no stratification and <br> scaling. | Requires extra work and given the <br> small sample sizes and <br> unbalanced nature of the data may <br> be fruitless. |
| Potentially: Minor, Moderate, or <br> Major | Use age data as age-length keys <br> which are applied to LFs to <br> produce catch-at-age, by method, <br> in years when sampling is <br> adequate. Possibly age-length <br> keys by zone. | A standard approach which allows <br> data to be stratified and scaled. <br> Not all age data will be used but <br> remainder could be fitted as age- <br> at-length. | None associated with <br> construction. Complicates <br> modelling a bit as double use of <br> LF data must be avoided. There <br> are years with age but no length <br> data and length-classes with no <br> age data. |
|  |  |  |  |



| Issue | Options | Pros | Cons |
| :--- | :--- | :--- | :--- |
| Use of Kapala data | Use biomass times series and LFs | Fishery independent data covering <br> a broad range of years. | Data collected from only a small <br> part of Zone 10 and so very <br> unlikely to be representative of <br> the eastern stock. Length data are <br> Importance: Minor-Moderate <br> (because there are little data and <br> large CVs are likely to be applied <br> to the biomass indices) |


| Issue | Options (short description) | Nิ | Initial preferences |  |  |  |  | Diagnostics checked | Final preferences |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Use of Kapala data |  |  | An | At | M | N | P |  | An | At | M | N | P |
|  | Use biomass times series and LFs |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Exclude the data |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |


| Issue | Options | Pros | Cons |
| :--- | :--- | :--- | :--- |
| Use of non-trawl CPUE indices <br> as biomass indices | Calculate and use as in the past <br> as biomass indices | Provides another potential <br> biomass signal. The non-trawl <br> fishery is a major component of <br> the fishery as ling is a key target <br> species. | It may not be a genuine biomass <br> index as it is derived from a small <br> number of vessels and sets <br> targeting a relatively small area |
| Importance: Minor-Moderate <br> (depending on how much weight <br> goes on the indices) | Exclude the data | Follows the philosophy of quality <br> over quantity if the data are <br> considered of poor quality. | May exclude a valid signal. |
|  |  |  |  |


| Issue | Options (short description) | $\stackrel{\text { N }}{\underset{\sim}{c}}$ | Initial preferences |  |  |  |  | Diagnostics checked | Final preferences |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Use of non-trawl |  |  | An | At | M | N | P |  | An |  | M | N | P |
| CPUE indices | Calculate and use as in the past |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Exclude the data |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |

## Modelling issues

| Issue | Options | Pros | Cons |
| :--- | :--- | :--- | :--- |
| Definition of Spawning Stock <br> Biomass (1): sexes | Male + female spawning biomass | Standard NZ approach. | An unlikely proxy for expected <br> fertilized egg production. Females <br> are larger than males and if gear is <br> length-selective females could <br> suffer higher exploitation. |
|  |  | Female spawning biomass | Simple and a reasonable proxy for <br> expected fertilized egg <br> production. |
|  |  |  |  |
|  |  | The best option if good fecundity <br> data are available. | Ignores effect of males. Little or <br> no data available to define egg <br> production. |
|  | Expected egg production |  |  |



| Issue | Options | Pros | Cons |
| :--- | :--- | :--- | :--- |
| Definition of Spawning Stock <br> Biomass (2): timing. | Beginning of year | SS3 default for a single-season <br> model | Assumes that fishing mortality <br> during the year has no effect on <br> the spawning potential of the <br> Importance depends on level of <br> stock. Gives stocks some extra <br> resilience to high $F$. |
| for stock depletion estimates <br> and on various factors with regard <br> to reference points. | Middle of spawning season <br> (after half the mortality; mid-year <br> if no explicit spawning season). | A logical assumption given that <br> spawning potential in each year is <br> likely to be related to the <br> "average" spawning biomass <br> engaged in spawning. | May lead to age-0 being 6 months <br> long in SS3 models? |
| Minor for stock depletion <br> estimates. <br> Minor-Moderate for reference <br> points. |  |  |  |


| Issue | Options (short description) | N | Initial preferences |  |  |  |  | Diagnostics checked | Final preferences |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Definition of |  |  | An | At | M | N | P |  | An | At | M | N | P |
| Spawning Stock | Beginning of year |  |  |  |  |  |  |  |  |  |  |  |  |
| Biomass (2) | Middle of spawning season |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |


| Issue | Options | Pros | Cons |
| :---: | :---: | :---: | :---: |
| Estimation of growth (1)Importance: Minor-Moderate | Estimate growth inside the base model | A natural approach which is likely to give the best estimate of growth as the model can account for selectivities and recruitment variability. | It can mean that data used in the model to allow good estimation of growth can dominate biomass indices (i.e., in the objective function). |
|  | Estimate growth outside the model | A simple approach which avoids faulty biomass signals from "growth data" distorting model results. | Leads to poor growth estimates if selectivity means that a Lee effect is present, which when fixed in the model may distort model results. |
|  | Estimate growth in a special model run and fix growth in the base model. (Two runs) | Could be the best of both worlds as the special run can put extra weight on the "growth data" to provide the best growth estimate. | Requires two runs and does not allow growth uncertainty to be included in overall uncertainty. |


| Issue | Options (short description) | $\underset{\tilde{N}}{\underset{\sim}{c}}$ | Initial preferences |  |  |  |  | Diagnostics checked | Final preferences |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Estimation of growth (1) |  |  | An | ${ }^{\text {At }}$ | M | N | P |  | An |  | M | N | P |
|  | Inside |  |  |  |  |  |  | SS3 model fits |  |  |  |  |  |
|  | Outside |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Two runs |  |  |  |  |  |  |  |  |  |  |  |  |


| Issue | Options | Pros | Cons |
| :--- | :--- | :--- | :--- |
| Estimation of growth (2): cohort <br> specific $k$ | Estimate cohort specific $\boldsymbol{k}$ for <br> some cohorts. | Attempts to account for variation <br> and trends in growth trajectories <br> across cohorts. | Requires well-balanced and <br> properly scaled age-length data to <br> have any chance of capturing real <br> trends. |
|  | Estimate a single $\boldsymbol{k}$ across all <br> cohorts | Normally adequate. | Not adequate if there are major <br> trends in growth. |
|  |  |  |  |


| Issue | Options (short description) | N | Initial preferences |  |  |  |  | Diagnostics checked | Final preferences |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Estimation of growth (2) |  |  | An | ${ }^{\text {At }}$ | M | N | P |  | An | At | M | N | P |
|  | Cohort specific $k$ |  |  |  |  |  |  | Analysis of age-length data (see |  |  |  |  |  |
|  | Single $k$ |  |  |  |  |  |  | Appendix 3). <br> SS3 model fits. |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |


| Issue | Options | Pros | Cons |
| :--- | :--- | :--- | :--- |
| Estimation of natural mortality | Estimate outside the model as a <br> single constant and fix; do <br> sensitivities to low and high <br> values | Simple and transparent. | Might do better inside the model. <br> Depending on data availability the <br> estimate from outside the model <br> might just be an educated guess. |
| Importance: Minor-Moderate | Estimate inside the model as a <br> single constant | Allows all of the data to <br> contribute to the estimation. <br> Allowance for selectivities and <br> YCS are made. | The data may not have much <br> information on $M$ - but the model <br> will give an estimate anyway. |
|  |  | ws for a single constant. No <br> reason not to if a sexed model | As for a single constant but could <br> lead to unrealistic or imprecise <br> estimates. |
|  | Estimate inside the model as sex <br> specific | Needs excellent age data <br> otherwise too complicated. |  |
|  | Estimate inside the model as age <br> and sex specific (e.g., double <br> exponential by age) | Realistic scenario |  |


| Issue | Options (short description) | $\stackrel{\text { N1 }}{ }$ | Initial preferences |  |  |  |  | Diagnostics checked | Final preferences |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Estimation of natural mortality |  |  | An | ${ }^{\text {At }}$ | M | N | P |  | ${ }^{\text {An }}$ | ${ }^{\text {At }}$ | M | N | P |
|  | Estimate outside the model |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Single constant |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Sex specific |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Age and sex specific |  |  |  |  |  |  |  |  |  |  |  |  |


| Issue | Options | Pros | Cons |
| :--- | :--- | :--- | :--- |
| Weighting of trawl CPUE <br> indices between years | Compare CPUE trends between <br> zones and assign CVs to <br> individual points on the basis of <br> similarities. (CSIRO 2012) | Was done this way in 2012. <br> Attempts to accounts to some <br> extent for between-zone variation <br> in trends in CPUE | Hard to understand why this <br> approach would provide sensible <br> results. The CVs assigned to some <br> years for the east CPUE indices <br> were very low (2-4\%). |
|  | Assign CPUE indices the same <br> CV across years. | A simple approach justified on the <br> basis that most of the CV should <br> be coming from process error (and <br> not observation error). | Some points in the time series are <br> more or less reliable than others. |
|  |  |  |  |
|  |  |  |  |


| Issue | Options (short description) | $\stackrel{\text { N1 }}{ }$ | Initial preferences |  |  |  |  | Diagnostics checked | Final preferences |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Weighting of trawl CPUE indices between years |  |  | An | ${ }^{\text {At }}$ | M | N | P |  | An | ${ }^{\text {at }}$ | M | N | P |
|  | CSIRO 2012 |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Same CV across years |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |


| Issue | Options | Pros | Cons |
| :--- | :--- | :--- | :--- |
| Weighting of data sets | Standard CSIRO approach: <br> emphasis on biomass indices and <br> iterative reweighting of effective <br> sample sizes and CVs | Accepted approach in Australia. <br> Aimed at fitting biomass indices <br> well. | Does not account for correlations <br> in composition data. |
| Importance depends on how <br> much the biomass signals from <br> different data sets agree or <br> disagree and the emphasis the <br> different methods put on the data <br> sets. | Standard NZ approach: emphasis <br> on biomass indices and iterative <br> reweighting of effective samples <br> sizes following Francis method <br> (gives low weights to LFs and <br> AFs). | Accepted approach in NZ. Aimed <br> at fitting biomass indices well. <br> Accounts for correlations in <br> composition data. | It is just one approach - there is <br> no perfect solution. As with other <br> iterative procedures, results <br> depend on how many parameters <br> are estimated and model structure. |
| Both methods emphasize the fit to <br> the biomass indices so importance <br> should be Minor-Moderate |  |  |  |
|  |  |  |  |


| Issue | Options (short description) | $\underset{\sim}{\underset{\sim}{c}}$ | Initial preferences |  |  |  |  | Diagnostics checked | Final preferences |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Weighting of data |  |  | An | ${ }^{\text {At }}$ | M | N | P |  | An | At | M | N | P |
|  | Standard CSIRO |  |  |  |  |  |  | SS3 model fits to CPUE indices |  |  |  |  |  |
|  | Standard NZ |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |


| Issue | Options | Pros | Cons |
| :---: | :---: | :---: | :---: |
| The potential changes in ling catchability caused by the loss of quota in 1999-2000 in the eastern trawl fishery and the structural adjustment in 20062007 for both east and west. | Assume no change but allow time-block fishery selectivities | Simple. Allows full time series of indices to be used. | There may have been changes in catchability. |
|  | Deal with the issue in the CPUE analysis | Allows the CPUE data to decide the issue. Allows full time series of indices to be used. | There is some potential for the analysis to get it wrong. |
| Importance is hard to judge before results are seen and it will depend on data weightings. | Split the eastern trawl CPUE series (different qs 1986-1999, 2000-2012) | Simple and addresses the issue which industry has repeatedly raised. | Reduces the power of the time series and the model may estimate an unrealistic change in $q$. |
| Potentially: Minor, Moderate, or Major | Ignore CPUE indices from 2007 onwards (Ignore from 2007) | Simple and follows the "quality data only" philosophy. | No biomass indices in recent years. Could be an over-reaction to problems with CPUE. |
|  | Ignore CPUE indices from 2000 onwards (east) and 2007 onwards (west) (Ignore many points) | Simple and follows the "quality data only" philosophy. | Could be a big over-reaction to problems with CPUE. |


| Issue <br> Potential changes in ling catchability | Options (short description) | $\underset{\sim}{\underset{\sim}{N}}$ | Initial preferences |  |  |  |  | Diagnostics checked | Final preferences |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | An | ${ }^{\text {At }}$ | M | N | P |  | An | At | M | N | P |
|  | Assume no change |  |  |  |  |  |  | Detailed CPUE analysis (see |  |  |  |  |  |
|  | CPUE analysis |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Split the eastern trawl CPUE |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Ignore from 2007 |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Ignore many points |  |  |  |  |  |  |  |  |  |  |  |  |

## Spatial modelling

| Issue | Options | Pros | Cons |
| :--- | :--- | :--- | :--- |
| Number of zones | Standard: <br> Zones 10,20, 30 (East) <br> Zones 40, 50 (west) | This is the way the data are <br> assembled and have been <br> analyzed in the past | There are no a priori reasons why <br> these spatial strata are appropriate |
|  | Alternative: <br> Something else | One would need to reanalyze the <br> data spatially and find spatial <br> strata which appear more <br> homogenous | Insufficient time to do a spatial <br> analysis of catch, effort, length- <br> frequency, and age data and re- <br> assemble all the data inputs |
|  |  |  |  |
|  |  |  |  |


| Issue | Options (short description) | तิ | Initial preferences |  |  |  |  | Diagnostics checked | Final preferences |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Number of zones |  |  | An | At | M | N | P |  | An | At | M | N | P |
|  | Standard |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Alternative |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |


| Issue | Options | Pros | Cons |
| :--- | :--- | :--- | :--- |
| Importance: Major | Larvae only | $\begin{array}{l}\text { This is consistent with what is } \\ \text { understood about ling biology; } \\ \text { fewer parameters }\end{array}$ | Some adult movement may occur |
|  | Adults and larvae | Some adult movement may occur | $\begin{array}{l}\text { More complicated model; not } \\ \text { consistent with general } \\ \text { understanding of ling biology; } \\ \text { more parameters }\end{array}$ |
|  |  | Full mixing | $\begin{array}{l}\text { No additional parameters to } \\ \text { model movement (can split fleets } \\ \text { across areas using different } \\ \text { selectivities). Can be seen as } \\ \text { pragmatic solution given noisy } \\ \text { data. }\end{array}$ | \(\left.\begin{array}{l}The assumption of full mixing <br>

among areas is unlikely for ling.\end{array}\right]\)

| Issue | Options (short description) | Nิ | Initial preferences |  |  |  |  | Diagnostics checked | Final preferences |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Movement assumptions |  |  | ${ }^{\text {An }}$ | At | M | N | P |  | An | ${ }^{\text {At }}$ | M | N | P |
|  | Larvae only |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Adults and larvae |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Full mixing |  |  |  |  |  |  |  |  |  |  |  |  |


| Issue | Options | Pros | Cons |
| :---: | :---: | :---: | :---: |
| Area-dependent growth <br> Importance: Moderate-Major | Not area-dependent | Simple, fewer parameters | Growth may differ spatially |
|  | Growth parameters area-specific | More realistic | Additional parameters; may be fitting to noise |


| Issue | Options (short description) | $\stackrel{\text { Ñ }}{\text { స̃ }}$ | Initial preferences |  |  |  |  | Diagnostics checked | Final preferences |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Area-dependent |  |  | An | ${ }^{\text {At }}$ | M | N | P |  | An |  |  |  | P |
| growth | Not area-dependent |  |  |  |  |  |  | SS3 estimates which varied a lot |  |  |  |  |  |
|  | Growth parameters area-specific |  |  |  |  |  |  | between areas |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |


| Issue | Options | Pros | Cons |
| :---: | :---: | :---: | :---: |
| Time-dependence in larval distribution <br> Importance: Major | Time independent | Simple, few parameters | Ignoring time-dependence in larval distribution could lead to poor fits and implicitly that local spawning is not very important |
|  | Time dependence, no priors | More realistic | More parameters, may be unstable if there is insufficient data in some spatial area |
|  | Time dependence, with priors | More realistic, less likely to be unstable | More parameters, there may be problems developing priors |
|  |  |  |  |


| Issue | Options (short description) | ヘิ | Initial preferences |  |  |  |  | Diagnostics checked | Final preferences |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Time-dependence in larval distribution |  |  | ${ }^{\text {An }}$ | ${ }^{\text {At }}$ |  |  | P |  | $\mathrm{An}^{\text {n }}$ |  |  |  | P |
|  | Time independent |  |  |  |  |  |  | Time-independent leads to |  |  |  |  |  |
|  | Time dependence, no priors |  |  |  |  |  |  | poorer fits than perfect mixing. |  |  |  |  |  |
|  | Time dependence, with priors |  |  |  |  |  |  |  |  |  |  |  |  |

## Appendix 1: Analysis of pink ling length data by depth and zone

P.L. Cordue

20 August 2013

## Introduction

The raw length data from port and onboard sampling were analyzed with regard to various factors including depth (onboard data) and zone (onboard and port data). The primary questions being addressed were whether the length structure of ling could be expected to vary across zones (within stock) and/or with depth. This question is relevant to which length data should be used in the stock assessment and how it should be scaled and stratified.

## Methods

Data from the almost flat files "LIGonboardlf.txt" and "LIGportlf.txt" were loaded into R for analysis as dataframes. Each row corresponded to fish of a certain length sampled from a particular tow/set/landing. The dataframes were expanded into fully-flat files by duplicating each record according to the total number of fish measured at the given length. Maximum depths recorded as zero were converted to missing values. The zone was calculated for each recorded position in the onboard sampling data (using CSIRO's standard SEF zone definition and algorithm).

Various histograms and boxplots were produced to examine the data, mainly just for the trawl catches. For the onboard data two linear models were used to explore which factors were the main determinants of length.

The first model had no interactions:
length $\sim$ year + month + zone + maxdepth + length-code + gear
where maxdepth was fitted as a cubic and other variables were categorical.

The second model was as above except that zone-maxdepth interactions were fitted. The "length-code" indicates what type of length measurement was taken (e.g., total, standard).

For the port data a single model was used:
length $\sim$ year + month + zone + grade + length - code + gear
and all variables were categorical.

## Results

In the eastern zones, the port-sampling data showed a clear difference in the raw lengths across zones for the trawl fleet (gear code = "OT") with smaller lengths in Zone 10 catches, and similar larger lengths in Zones 20 and 30 (Figure 1). However, in the onboard-sampling data it was Zones 10 and 20 that had similar sized smaller fish with Zone 30 having larger fish (Figure 2).

For the trawl catch in the western zones, there was quite a lot of variability across years within zone and there were no clear differences between Zones 40 and 50 or between port and onboard sampling (Figures 3 and 4).

The depth distribution of the trawl sampling varied considerably across years for the eastern zones (Figure 5) but was much less variable for the western zones (Figure 6). This indicates that it could be very misleading, for the eastern stock, to simply examine raw trawl length data by zone as any apparent differences could actually be driven by changes in depth distribution (assuming there is a depth effect).

The only way to get at the likely effects is through some type of model standardization - such as linear modelling.

The models had adjusted $\mathrm{R}^{2}$ of $46 \%$ (onboard non-interaction), $47 \%$ (onboard interaction), and $37 \%$ (port). Almost everything was highly significant because of the large number of records and the (false) assumption of independent errors. However, the models are just used to check for "important" factors with regard to length structure in the catch and significance levels are not relevant. Confidence intervals were calculated but they were all very tight and are not presented because they over-estimate the level of precision.

The results are focused on the onboard models as they can account for depth effects whereas the port model cannot.

A major effect that needed to be accounted for was gear and both onboard models estimated similar effects (Figure 7). Year effects were fitted to allow for recruitment variability. Both onboard models estimated very similar year effects (Figures 8).

The non-interaction model estimated a strong depth effect with a steady increase in the expected length of ling from $50-1000 \mathrm{~m}$ (Figure 9). The interaction model strongly suggested that there were important depth effects within zone with a major effect in Zone 10 (Figure 10). The relative zone effects differ by depth in the interaction model but a comparison of the zone effects between the two models at a maximum depth of 400 m shows similar effects (Figure 11). The port model estimated a similar pattern of zone effects as well (Figure 12).

## Conclusions

There are important zone effects and important depth effects within some of the zones.

Zone 10 shows a strong effect, typically having smaller ling than zones 20 and 30. There is a very strong depth effect in Zone 10 and a strong depth effect in Zone 20. For the eastern stock assessment, trawl length data should be stratified by depth within zone. Because of the strong depth effects in zones 10 and 20, port data should not be used unless it can be assigned to a depth range. For non-trawl data there is less need for stratification as there is no catch in Zone 10.

In the western zones, the depth effects are not very strong and the zones have similar sized ling. For the stock assessment it is probably unnecessary to stratify by zone or depth and the port data can probably be combined with the onboard data.


Figure 1: Boxplots of raw lengths by eastern zone and year for the port sampling of ling (OT only).


Figure 2: Boxplots of raw lengths by eastern zone and year for the onboard sampling of ling (OT only).


Figure 3: Boxplots of raw lengths by western zone and year for the port sampling of ling (OT only).


Figure 4: Boxplots of raw lengths by western zone and year for the onboard sampling of ling (OT only).


Year
Figure 5: Boxplots of max. depth by eastern zone and year for the onboard sampling of ling (OT only).


Figure 6: Boxplots of max. depth by western zone and year for the onboard sampling of ling (OT only).


Figure 7: Expected length of ling by gear according to the onboard linear models (standardized to June 2000, zone 10, max. depth 400 m , total length).


Figure 8: Expected length of ling by year according to the onboard linear models (standardized to June, zone 20, max. depth 400 m , gear OT, total length).


Figure 9: Expected length of ling by maximum depth according to the onboard noninteraction model (standardized to June 2000, zone 20, gear OT, total length).


Figure 10: Expected length of ling by maximum depth within zone according to the onboard interaction model (standardized to June 2000, gear OT, total length). Effects plotted over $\mathbf{9 9 \%}$ of the depth range.


Figure 11: Expected length of ling by zone according to the onboard linear models (standardized to June 2000, max. depth 400 m, gear OT, total length).


Figure 12: Expected length of ling by zone according to the onboard non-interaction model and the port model (standardized to June 2000, max. depth 400 m , gear OT, grade ALL, total length).

# Appendix 2: Stratification of trawl length data in the eastern zones 

P.L. Cordue

20 August 2013

## Methods

For the eastern zones I looked at how much trawl length data would be available each year under three different stratifications: none, zone only, and depth and zone (by depth within Zones 10 and 20, no depth strata in Zone 30). The depth strata were $0-300 \mathrm{~m}, 300-500 \mathrm{~m}$, and $500^{+}$m.

It was required that at least 30 fish were within each cell for scaling purposes and that a year would be excluded from the stock assessment if more than $10 \%$ of the annual catch could not be adequately scaled (i.e., if cells with less than 30 fish corresponded to more than $10 \%$ of the catch).

## Results

No stratification allows all of the data to be used from on-board and port sampling, but it is not a viable stock assessment option as there are zone and depth effects and the sampling is very unbalanced. However, it serves as a baseline from which to measure how many years of data are excluded if some form of stratification is used. With no-stratification, the years 19932012 are covered by on-board sampling (Table 1) and the years 1998-2012 are covered by port sampling (assuming that at least 4 sampled landings are required in each year)(Table 2). That is, on-board 20 years, port 15 years.

For the on-board sampling, stratification by zone covers 13 out of the 20 possible years (Table 1). Adding in the depth stratification within Zones 10 and 20 eliminates another 5 years leaving 8 of the years from 1998 to 2012 covered (Table 1).

The port sampling did not adequately cover Zone 10 in recent years and stratification by zone reduces the number of years covered from 15 down to 8 (spanning 1998 to 2008)(Table 2). As no depth information is (directly) recorded for the port sampling it would not be easy to stratify by depth and zone and it is assumed that no years could be used in the assessment under that stratification.

Table 1: Total number of fish measurements available each year from on-board sampling of trawl to use in the eastern stock assessment under different levels of stratification. In the depth and zone stratification Zone 30 is not stratified by depth. In the stratifications a year is excluded from use if more than $10 \%$ of the annual catch cannot be scaled with at least 30 fish in each cell.

| Year | Number of <br> tows sampled | No <br> stratification | Stratification <br> by zone | Stratification <br> by depth and <br> zon |
| :--- | ---: | ---: | ---: | ---: |
|  |  |  |  |  |
| 1992 | 1 | 1 | - | - |
| 1993 | 20 | 497 | - | - |
| 1994 | 30 | 1103 | - | - |
| 1995 | 18 | 1013 | - | - |
| 1996 | 68 | 2438 | - | - |
| 1997 | 70 | 2180 | - | - |
| 1998 | 79 | 7097 | 7097 | 7097 |
| 1999 | 82 | 7562 | 7562 | - |
| 2000 | 48 | 3179 | 3179 | 3179 |
| 2001 | 76 | 5546 | 5546 | 5546 |
| 2002 | 41 | 2982 | 2982 | 2982 |
| 2003 | 41 | 2884 | 2884 | 2884 |
| 2004 | 22 | 812 | 812 | - |
| 2005 | 53 | 2765 | 2765 | 2765 |
| 2006 | 38 | 3133 | 3133 | 3133 |
| 2007 | 35 | 531 | - | - |
| 2008 | 23 | 829 | - | - |
| 2009 | 16 | 507 | 507 | - |
| 2010 | 38 | 750 | 750 | - |
| 2011 | 46 | 846 | 846 | - |
| 2012 | 34 | 660 | 660 | 660 |

Table 2: Total number of fish measurements available each year from port sampling of trawl to use in the eastern stock assessment under different levels of stratification. In the depth and zone stratification Zone 30 is not stratified by depth. In the stratifications a year is excluded from use if more than $10 \%$ of the annual catch cannot be scaled with at least 30 fish in each cell.

| Year | Number of <br> landings <br> sampled | No <br> stratification | Stratification <br> by zone | Stratification <br> by depth and <br> zone |
| :--- | ---: | ---: | ---: | ---: |
|  |  |  |  |  |
| 1991 | 2 | 100 | - | - |
| 1992 | 1 | 54 | - | - |
| 1993 | - | - | - | - |
| 1994 | - | - | - | - |
| 1995 | 3 | 248 | - | - |
| 1996 | - | - | - | - |
| 1997 | 1 | 102 | - | - |
| 1998 | 17 | 1417 | 1417 | - |
| 1999 | 25 | 1482 | 1482 | - |
| 2000 | 9 | 489 | - | - |
| 2001 | 14 | 345 | - | - |
| 2002 | 30 | 2813 | 2813 | - |
| 2003 | 6 | 214 | - | - |
| 2004 | 7 | 201 | 201 | - |
| 2005 | 26 | 1447 | 1447 | - |
| 2006 | 20 | 1123 | 1123 | - |
| 2007 | 12 | 1364 | 1364 | - |
| 2008 | 8 | 1771 | 1771 | - |
| 2009 | 48 | 2507 | - | - |
| 2010 | 62 | 2510 | - | - |
| 2011 | 59 | 2448 | - | - |
| 2012 | 41 | 1915 | - | - |

## Appendix 3: A look at the pink ling age-length data

P.L. Cordue

26 August 2013

## Introduction

The age-length data were explored with regard to sample sizes and differences in proportion-at-age for given length across zones and gear type. Descriptive plots and categorical linear models were used. Also, the question of whether there are important cohort effects with regard to growth was explored with categorical linear models.

## Methods

Data with missing lengths or ages were excluded from the analysis. The number of tows/sets/landings that were sampled for age-length data was determined by distinguishing the records according to: year, month, batch, zone, gear, area, and vessel. Cohort was defined as year - age.

Six broad length classes were defined with breaks at: $0,50,70,80,90,100,150 \mathrm{~cm}$. The proportion-at-age was calculated for each zone (10-50) within each length class, for the trawl and non-trawl data separately. Lowess lines were fitted to the length-at-age data within zone for the trawl and non-trawl data separately.

A linear model to explain age was fitted to check if the zone effects were significant:

$$
\text { age } \sim \text { year }+ \text { month }+ \text { zone }+ \text { gear }+ \text { sex }+ \text { length-class }
$$

where all explanatory variables were categorical.

Various linear models to explain length were fitted to check for the importance of year and cohort effects with regard to growth:

$$
\begin{aligned}
& \text { length } \sim \text { month }+ \text { zone }+ \text { gear }+ \text { sex }+ \text { age } \\
& \text { length } \sim \text { month }+ \text { zone }+ \text { gear }+ \text { sex }+ \text { age }+ \text { year } \\
& \text { length } \sim \text { month }+ \text { zone }+ \text { gear }+ \text { sex }+ \text { age }+ \text { cohort } \\
& \text { length } \sim \text { month }+ \text { zone }+ \text { gear }+ \text { sex }+ \text { age }+ \text { year }+ \text { cohort }
\end{aligned}
$$

where all explanatory variables were categorical.

## Results

## Sample sizes

The sampling effort across zones is very patchy for trawl (Table 1) and non-trawl (Table 2). Within the non-trawl data, the sampling is also very patchy across the different gear types (Table 3). The sampling is so sparse, in terms of trawls/landings sampled, that there are no years for the east or west when at least 5 age-length trawl samples were obtained from every zone (Table 1). The same statement is almost true for the non-trawl sampling, except that in 2011 there were at least 5 samples in every eastern zone (ignoring Zone 10 where there is almost no non-trawl catch)(Table 2). The only reasonable level of sampling occurs for trawl: zone 10 (1996-1999), zone 20 (2007-2012), zone 50 (2004-2006); and non-trawl: zone 20 (2008-2012). Of course, the lack of over-lapping years means that the sampling is very unbalanced by zone.

In terms of the number of fish aged, the unbalanced nature of the sampling is the same but there are reasonable numbers of fish in a number of zones in a number of years (Tables 4 and 5). In the eastern trawl data, the number of fish ages is very unbalanced by zone with an early emphasis on Zone 30 (1983-1985), a shift to Zone 10 (1996-2002), and a final shift to Zone 20 in the later years (2007-2011). Relatively speaking the western trawl data is quite balanced (Table 4).

The non-trawl data in the east is almost all from Zone 20, except in 2000 when there are good numbers of fish from Zones 10 and 30 as well (Table 5) - a very unusual year as there is almost no non-trawl catch from Zone 10. For the non-trawl in the west, there is only one year (2006) with reasonable numbers of fish from both zones (Table 5).

## Length-at-age

The lowess fits to the trawl data by zone show similar fits for zones 30,40 , and 50 , with quite different fits in Zones 10 and 20 (Figure 1). The fits are a product of the trawl selectivities in each zone, the type of sampling done in each zone, and possibly even relative exploitation rates (if selectivities are by length). While suggestive of growth patterns, there are possible confounding effects. The non-trawl data had much more variable fits but were again suggestive of some possible differences across zones (Figure 2).

## Conditional proportion-at-age

When the data were split into the broad length classes for analysis pooled over years and sex, there were adequate numbers of aged fish in each length class and zone for the trawl data (Table 6). However, for the non-trawl there were only adequate numbers for the middle four length classes and zones 20-50 (Table 7).

The proportions-at-age within length class for the trawl data show marked differences between zones (Figure 3). Zones 10 and 20 stand out as having younger fish than the other zones (for the four middle length classes), with Zone 10 in particular having the youngest
fish. For the non-trawl data Zones 20 and 30 show marked differences from each other while Zones 40 and 50 appear fairly similar to each other (Figure 4).

As a check to see if the zone differences, in the pooled analysis, were not just driven by year and sex differences, a linear model was used to explain age in terms of the full set of explanatory variables. Everything was highly significant, as expected, including the zone effects (and gear effects). As with the descriptive plots, Zone 10 had the youngest mean-age at length, Zone 20 the next youngest, and then Zones 30-50 had almost identical mean-age at length (Figure 5). The difference in mean age between Zone 10 and Zone 30 was 0.75 years.

## Looking for important cohort effects

Stepwise forward linear regression was used to see if there were any important cohort effects for length-at-age. The model results showed that year effects were more important than cohort effects (Tables 8 and 9). This was true for models fitted to all zones and when separate models were fitted to only the eastern or the western zones. In all but one of the models, all of the individual cohort effects failed to achieve a high level of significance, whereas year effects often did (Table 9).

## Conclusions and recommendations

The analysis is strongly suggestive of important differences between zones and between fishing methods in proportion-at-age for given length in the catch. The western zones (40 and 50) appear to have very similar mean-age at length, but the eastern zones show increasing mean-age at length for Zones 10, 20, and 30.

The sampling is very unbalanced especially for the eastern trawl data. Therefore, it is probably important for stock assessment to stratify the age-length data to some extent.

I suggest the following approach:

- eastern stock:
- trawl: stratify and scale by zone but ignore Zone 30 (little catch)
- non-trawl: stratify and scale by zone but ignore Zone 10 (almost no catch)
- western stock:
- trawl: no stratification
- non-trawl: no stratification

In terms of estimating growth, the data are too unbalanced to estimate cohort effects as these are confounded with other effects that are difficult to eliminate (e.g., month, non-trawl gear types, depth, and year).

Table 1: Number of tows/landings sampled for age-length data in the trawl fishery by zone and year.

|  | SF10 | SF20 | SF30 | SF40 | SF50 |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1979 | 4 | 2 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 1 | 0 | 0 |
| 1983 | 0 | 0 | 5 | 0 | 0 |
| 1984 | 0 | 0 | 2 | 0 | 0 |
| 1985 | 0 | 0 | 5 | 0 | 0 |
| 1986 | 0 | 0 | 1 | 0 | 0 |
| 1987 | 0 | 0 | 0 | 0 | 5 |
| 1988 | 0 | 0 | 0 | 0 | 3 |
| 1989 | 0 | 0 | 0 | 0 | 2 |
| 1993 | 0 | 0 | 0 | 5 | 0 |
| 1994 | 2 | 2 | 0 | 2 | 1 |
| 1995 | 3 | 2 | 0 | 1 | 3 |
| 1996 | 29 | 2 | 1 | 2 | 2 |
| 1997 | 19 | 0 | 0 | 0 | 6 |
| 1998 | 16 | 3 | 0 | 5 | 4 |
| 1999 | 14 | 4 | 0 | 15 | 4 |
| 2000 | 4 | 1 | 0 | 3 | 1 |
| 2001 | 7 | 2 | 1 | 1 | 0 |
| 2002 | 6 | 0 | 2 | 2 | 2 |
| 2003 | 2 | 0 | 0 | 3 | 2 |
| 2004 | 3 | 0 | 1 | 3 | 13 |
| 2005 | 3 | 3 | 1 | 1 | 6 |
| 2006 | 0 | 0 | 3 | 1 | 12 |
| 2007 | 0 | 14 | 1 | 2 | 0 |
| 2008 | 0 | 6 | 0 | 2 | 0 |
| 2009 | 0 | 30 | 1 | 3 | 1 |
| 2010 | 11 | 36 | 0 | 6 | 2 |
| 2011 | 4 | 31 | 4 | 3 | 3 |
| 2012 | 2 | 18 | 2 | 2 | 4 |
|  |  |  |  |  |  |

Table 2: Number of sets/landings sampled for age-length data in the non-trawl fishery by zone and year.

|  | SF10 | SF20 | SF30 | SF40 | SF50 |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1994 | 0 | 1 | 0 | 0 | 0 |
| 1995 | 1 | 3 | 0 | 0 | 1 |
| 1996 | 0 | 0 | 0 | 0 | 1 |
| 1997 | 0 | 0 | 0 | 0 | 0 |
| 1998 | 0 | 0 | 0 | 0 | 0 |
| 1999 | 0 | 3 | 0 | 1 | 1 |
| 2000 | 1 | 2 | 1 | 1 | 0 |
| 2001 | 0 | 1 | 0 | 4 | 0 |
| 2002 | 0 | 4 | 0 | 2 | 0 |
| 2003 | 0 | 8 | 4 | 12 | 0 |
| 2004 | 0 | 6 | 1 | 2 | 0 |
| 2005 | 0 | 1 | 0 | 0 | 3 |
| 2006 | 0 | 0 | 0 | 1 | 3 |
| 2007 | 1 | 5 | 0 | 4 | 0 |
| 2008 | 0 | 13 | 0 | 0 | 0 |
| 2009 | 0 | 24 | 4 | 1 | 1 |
| 2010 | 0 | 24 | 0 | 5 | 0 |
| 2011 | 0 | 29 | 9 | 0 | 0 |
| 2012 | 0 | 13 | 0 | 2 | 1 |

Table 3: Number of sets/landings sampled for age-length data in the non-trawl fishery by gear and year (for all zones including GAB).

|  | ALL | DL | DS | Hook | MN | MWT | Tr | UNKNO |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1994 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 1995 | 0 | 0 | 2 | 2 | 0 | 0 | 0 | 1 |
| 1996 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 1999 | 0 | 0 | 0 | 2 | 0 | 0 | 3 | 0 |
| 2000 | 0 | 0 | 1 | 3 | 1 | 0 | 0 | 0 |
| 2001 | 0 | 0 | 0 | 3 | 1 | 0 | 1 | 0 |
| 2002 | 0 | 0 | 4 | 1 | 0 | 0 | 1 | 0 |
| 2003 | 0 | 0 | 0 | 24 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 1 | 8 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 5 | 0 | 0 | 0 | 0 |
| 2006 | 0 | 0 | 0 | 3 | 0 | 1 | 0 | 0 |
| 2007 | 0 | 0 | 2 | 8 | 0 | 0 | 0 | 0 |
| 2008 | 0 | 0 | 0 | 15 | 0 | 0 | 0 | 0 |
| 2009 | 27 | 3 | 4 | 4 | 0 | 0 | 0 | 0 |


| 2010 | 0 | 0 | 0 | 36 | 0 | 0 | 0 | 0 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2011 | 0 | 0 | 5 | 46 | 0 | 0 | 0 | 0 |
| 2012 | 17 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |

Table 4: Number of fish ages by zone and year for the trawl data

|  | SF10 | SF20 | SF30 | SF40 | SF50 |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1979 | 153 | 247 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 23 | 0 | 0 |
| 1983 | 0 | 0 | 148 | 0 | 0 |
| 1984 | 0 | 0 | 143 | 0 | 0 |
| 1985 | 0 | 0 | 143 | 0 | 0 |
| 1986 | 0 | 0 | 34 | 0 | 0 |
| 1987 | 0 | 0 | 0 | 0 | 564 |
| 1988 | 0 | 0 | 0 | 0 | 326 |
| 1989 | 0 | 0 | 0 | 0 | 190 |
| 1993 | 0 | 0 | 0 | 50 | 0 |
| 1994 | 129 | 89 | 0 | 14 | 247 |
| 1995 | 138 | 203 | 0 | 1 | 322 |
| 1996 | 792 | 34 | 31 | 11 | 68 |
| 1997 | 589 | 0 | 0 | 0 | 553 |
| 1998 | 542 | 127 | 0 | 109 | 100 |
| 1999 | 406 | 134 | 0 | 331 | 165 |
| 2000 | 110 | 21 | 0 | 44 | 27 |
| 2001 | 295 | 112 | 48 | 93 | 0 |
| 2002 | 244 | 0 | 100 | 99 | 99 |
| 2003 | 78 | 0 | 0 | 95 | 81 |
| 2004 | 61 | 0 | 6 | 144 | 334 |
| 2005 | 102 | 65 | 59 | 78 | 202 |
| 2006 | 0 | 0 | 27 | 50 | 424 |
| 2007 | 0 | 262 | 198 | 130 | 0 |
| 2008 | 0 | 325 | 0 | 45 | 0 |
| 2009 | 0 | 599 | 17 | 40 | 25 |
| 2010 | 158 | 545 | 0 | 95 | 34 |
| 2011 | 47 | 532 | 71 | 48 | 47 |
| 2012 | 21 | 25 | 62 | 0 | 0 |
|  |  |  |  |  |  |

Table 5: Number of fish ages by zone and year for the non-trawl data.

|  | SF10 | SF20 | SF30 | SF40 | SF50 |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1994 | 0 | 17 | 0 | 0 | 0 |
| 1995 | 12 | 31 | 0 | 0 | 155 |
| 1996 | 0 | 0 | 0 | 0 | 70 |
| 1997 | 0 | 0 | 0 | 0 | 0 |
| 1998 | 0 | 0 | 0 | 0 | 0 |
| 1999 | 0 | 134 | 0 | 0 | 155 |
| 2000 | 94 | 124 | 248 | 3 | 0 |
| 2001 | 0 | 2 | 0 | 323 | 0 |
| 2002 | 0 | 90 | 0 | 99 | 0 |
| 2003 | 0 | 279 | 31 | 322 | 0 |
| 2004 | 0 | 95 | 0 | 48 | 0 |
| 2005 | 0 | 51 | 0 | 0 | 137 |
| 2006 | 0 | 0 | 0 | 125 | 276 |
| 2007 | 0 | 98 | 0 | 95 | 0 |
| 2008 | 0 | 585 | 0 | 0 | 0 |
| 2009 | 0 | 486 | 20 | 8 | 15 |
| 2010 | 0 | 333 | 0 | 48 | 0 |
| 2011 | 0 | 537 | 109 | 0 | 0 |
| 2012 | 0 | 0 | 0 | 0 | 0 |

Table 6: Number of fish ages per length class and zone for trawl gear (OT).

|  | $(0,50]$ | $(50,70]$ | $(70,80]$ | $(80,90]$ | $(90,100]$ | $(100,150]$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| SF10 | 1003 | 2322 | 312 | 129 | 59 | 40 |
| SF20 | 225 | 2179 | 690 | 160 | 44 | 22 |
| SF30 | 27 | 349 | 316 | 271 | 108 | 39 |
| SF40 | 43 | 440 | 493 | 312 | 142 | 47 |
| SF50 | 161 | 1992 | 743 | 528 | 264 | 120 |

Table 7: Number of fish ages per length class and zone for non-trawl gear.

|  | $(0,50]$ | $(50,70]$ | $(70,80]$ | $(80,90]$ | $(90,100]$ | $(100,150]$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| SF10 | 0 | 3 | 19 | 58 | 25 | 1 |
| SF20 | 63 | 967 | 772 | 515 | 307 | 238 |
| SF30 | 1 | 57 | 151 | 116 | 49 | 34 |
| SF40 | 0 | 86 | 316 | 292 | 218 | 159 |
| SF50 | 114 | 199 | 182 | 195 | 100 | 18 |

Table 8: Adjusted $\mathbf{R}^{\mathbf{2}}$ for linear models explaining length in terms of month, zone, gear, sex, and age (Reference model) with the addition of year or cohort or both. Results are shown for all zones and just eastern zones or just western zones.

| Model | All zones | East only | West only |
| :--- | :--- | :--- | :--- |
|  |  |  |  |
| Reference | 0.781 | 0.778 | 0.771 |
| + year | 0.803 | 0.813 | 0.783 |
| +cohort | 0.785 | 0.788 | 0.776 |
| + year + cohort | 0.807 | 0.819 | 0.787 |

Table 9: Number of individual year or cohort effects significant at the 0.001 level for linear models explaining length in terms of month, zone, gear, sex, and age (Reference model) with the addition of year or cohort or both. Results are shown for all zones and just eastern zones or just western zones.

| (Model) effect | All zones | East only | West only |
| :--- | ---: | ---: | ---: |
|  |  |  |  |
| (Reference) | - | - | - |
| (+ year) year | 19 | 14 | 11 |
| (+cohort) cohort | 0 | 0 | 0 |
| (+ year + cohort $)$ <br> year | 14 | 10 | 3 |
| (+ year + cohort) <br> cohort | 2 | 0 | 0 |



Figure 1: Lowess fits to length at age by zone for trawl data pooled across sex and year.


Figure 2: Lowess fits to length at age by zone for non-trawl data pooled across sex and year.


Figure 3: Proportion-at-age by zone and length class (first 3 classes) for the trawl data.




Figure 3 (ctd): Proportion-at-age by zone and length class (last $\mathbf{3}$ classes) for the trawl data.


Figure 4: Proportion-at-age by zone and length class (middle 4 classes) for the nontrawl data.


Figure 5: Predicted mean age by length class for a linear model explaining age by year, month, sex, gear, zone, length class. Standardized to July 2000, Zone 10, trawl, unknown sex.

