Revised sampling regime for the Southern and Eastern Scalefish and Shark Fishery - Final Report


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## EXECUTIVE SUMMARY

Australia's Southern and Eastern Scalefish and Shark Fishery (SESSF) is a complex multispecies, multi-gear fishery which was formed in 2003 from the amalgamation of four different fisheries; the Commonwealth Trawl Sector (CTS); the Great Australian Bight Trawl Fishery (GABTF); the Gillnet, Hook and Trap Fishery (GHATF) and the East Coast Deepwater Trawl Fishery (ECDWTF). Management of the fishery is conducted by the Australian Fisheries Management Authority (AFMA).

Scientific monitoring of the SESSF is an important aspect of the fishery's management. It collects critical information on the age and size structure of the main target species to feed into stock assessments as well as data on species catch composition and vessel interactions with threatened / endangered / protected (TEP) species. There have been various formal scientific monitoring programs on the different sub-fisheries of the SESSF over the last 15 years. The most extensive and long running of these has focussed on the CTS and has been formally designed to collect specific biological information and data on discard rates in the trawl sector as part of the Integrated Scientific Monitoring Program (ISMP). More recently, monitoring programs have been introduced into the various other trawl and non-trawl fisheries under the umbrella of the ISMP, but these have never been formally designed or reviewed. Initially, these monitoring programs focussed mainly on collection of information on the main quota species but over the last decade, the focus and level of monitoring of the fishery has changed to take into account its impact on bycatch/byproduct species and TEP species.

Over the last five years there have been considerable changes to both the structure and size of the SESSF fishing fleet, largely although not entirely as a result of the Commonwealth's Structural Adjustment Package (SAP). The current design of the monitoring program was not established to cover such a level of change in the fishery and may not adequately cover all of the information now required. As a result, there has been a need for a redesign of the SESSF monitoring program that includes all sectors and reflects changed management objectives.

To this end, AFMA funded this study entitled "A revised sampling regime for the Southern and Eastern Scalefish and Shark Fishery". The aim of this study was to undertake a complete redesign of the current ISMP sampling regime for all sectors of the SESSF to ensure that it is representative, effective and statistically robust. To do this, the study was required to consider all sectors of the SESSF and update the monitoring design so that it reflected the current structure of the fishery post-SAP. It needed to review and update stratification of the monitoring program including consideration of various coefficients of variation (CVs) for discard rates and size/age composition of the total catch (retained and discarded) for the SESSF quota species, major non-quota species, other non-quota species and any other species/groups identified by the ISMP Review Committee. Further, it was required to design effective and statistically robust coverage for recording the total number and circumstances of interactions with species identified as high risk through the ecological risk assessment (ERA) process and all TEP species. An important output from the study was the provision of a range of sampling redesign options for the SESSF, including a table of costs outlining what CVs will be achieved with each option and comparison of cost versus precision levels so that AFMA can consider the most appropriate future monitoring of the fishery.

The data used for this report included: logbook catch and effort data; catch landing data; onboard ISMP data; port-based ISMP data; age and length data mainly for SESSF quota species; and data related to interactions with threatened species, including birds, syngnathids and seals.

Data audits were carried out, to assess the various fields involved, their quality with respect to missing values, blanks, empty spaces etc, and their relationship to each other was established for the purposes of the study. A statistical review of previous stratification and sample allocation decisions for the ISMP was carried out with respect to distributional assumptions, stratification considerations and CV estimation and calculation approaches.

A critical difference from the previous design of the ISMP, was that the current ISMP design review was to include, not just estimation of total catch (retained and discarded) for quota species within certain CVs, but also estimation of the total catch and CVs for major non-quota species, other non-quota species, species identified as "high risk" through the ERA process and the interactions with TEP species. This difference required a fundamental change in the analysis and design process, because many of the additional species and species groups are not usually retained in the commercial catch.

To achieve the above, it was necessary to identify which of the $500+$ species (or species groups) caught in the SESSF should be included in each of the above categories and reduce these to a workable number of groups termed 'Project Keys'. Obviously, identification of the quota species and species baskets was straight forward. The major non-quota species were identified as the species which, in addition to the quota species, cumulatively accounted for $80 \%$ of the total catch for each gear type. All remaining minor non-quota species (not included as high risk or TEP species) were grouped into: Chimaeras, Crustaceans, Dogfish, Echinoderms, Fish, Hagfish, Rays, Sawsharks, Molluscs, Sharks, Stingarees or Other. The above grouping resulted in a total of 68 Project Keys. The $\sim 30$ ERA "high risk" species were combined into 7 groups: Teleosts; Sharks; Upper Slope Dogfish; Other Dogfish; Rays/Skates; Molluscs; and Hagfish. TEP species were divided into five species groups: Seals (including sea lions), Sharks, Pipefish, Birds, Whales/Dolphins, and Reptiles (none recorded), of which by far the dominant groups are the seal group and the bird group.

Cluster analyses of the species composition of shots in each sub-fishery of the SESSF were conducted to elucidate appropriate spatial and temporal stratification for the ISMP design. Data on discard ratios were used to carry out additional analyses on these. This process resulted in 24 strata upon which determination of optimal survey designs was based.

Methods employed previously in estimating discard levels in the SESSF involved analysis of the discard proportion for each species at a stratum by stratum level using the landed catch, and then aggregation across strata to obtain an estimate, with associated CV, of a fishery-wide discarded proportion for each species under study. This was reasonable because it concentrated mainly on quota species which were generally retained and had good landings information. The present ISMP design, however, had to encompass a much larger range of species and species groups, many of which are discarded in high proportions. We therefore used a method to estimate the discarded catch mass in each stratum so that fishery-wide mean and variance on total discards could be obtained by simple addition of the stratum means and variances. This required more complexity in the calculation of the stratum-level statistics and these methods are described in detail in the report.

In previous designs, the target precision on discarding has been set in terms of a CV on the discarded proportion with reduced target CVs for increasing discard proportions. In the new design, the target was based on a flat $20 \% \mathrm{CV}$ on the total catch mass (not discard proportion). Under this target specification, for a single species, the objective underlying optimization of the sample size distribution to strata is relatively simple - to allocate strata so as to minimize the CV for that species. With multiple species groups, however, a number of different objectives are possible depending on the competing requirements for different species. The results are presented under five different optimization methods:

VWCV: minimize the value-weighted average CV over the 68 non high-risk project keys based on the approximate price per kg of the commercial species;
MinS: maximize the number of 68 non high-risk project keys which have CV below the target threshold of $20 \%$.
PROP: simply make an allocation of any given availability of sampling days in proportion to the number of sea-days expended by the fishery in each stratum;
CWCV-HR: minimize the catch-weighted average CV over the 7 high risk project keys; and,
MinS-HR: maximize the number of high-risk project keys which have CV below the target threshold of $20 \%$.
None of the above procedures pre-suppose any particular constraint on total sampling effort available, but each presents a continuum of benefit vs sampling cost from which a sampling level may be selected. Though we recognize that observers will select trips rather than shots, we note also that the cost of observation is dependent on trip length. We thus use the sea-day as the measure of sampling effort.

## Analysis of Quota, Non-quota and High Risk Species

The above optimizations have been modeled to show the CVs that are achieved for all 75 project keys based on 250 to 1500 sea-days at 250 sea-day increments. For presentation purposes, the CVs achieved are shown for each project key and have been grouped as:
Good CV 0-20\%, Medium CV 20-40\%, Poor CV > 40\%
Of all the commercial allocation methods, method VWCV performed the most poorly resulting in only 40,47 and 55 non high risk project keys with CVs below $20 \%$ for 250,500 and 1000 sea-days of sampling respectively. In addition, more than 1000 sea-days are required to increase the number of high risk project keys with good CVs from three to four. The method MinS performed better for non high risk species resulting in 52, 59 and 62 project keys with good CVs for 250,500 and 1000 sea-days of sampling, but performed just as poorly as method VWCV for high risk project keys. Method PROP performed poorly at low sample size (only 46 non high risk project keys had CVs below $20 \%$ with 250 sea-days), however performance increased greatly with an additional 250 sea-days where-after it performed nearly as well as method MinS. Method PROP performed well for high risk project keys, with all seven keys having good to medium CVs for 250 and 500 sea-days, and good CVs for 750+ sea-days.

Method CWCV-HR resulted in four of the seven high risk project keys having good CVs for 250-1000 sea-days, but only improves by one high risk project key at 1250-1500 sea-days. Just over half of the 68 non high risk project keys achieve good CVs with 250 sea-days of sampling, and this number increases only slightly with additional sea-days. Even at 1000 seadays, only 48 non high risk project keys achieve good CVs. This illustrates one of the tradeoffs to be made between the differing objectives of the program. Method MinS-HR achieved good CVs for five of the seven high risk project keys, but only 33 of the 68 non high risk project keys with 250 sea-days of sampling. The addition of 250 sea-days results in good CVs for all seven high risk project keys, and vastly improves the number of non high risk project keys with good CVs (to 54 keys), so much so that it is nearly on par with method PROP, and much better than method VMCV. A further increase in sampling to 750 sea-days makes little difference, however with 1000 sea-days, 61 non high risk project keys achieve good CVs.

| Method VWCV | Sea Days |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 250 | 500 | 750 | 1000 | 1250 | 1500 |
| 68 NON HIGH RISK PROJECT KEYS |  |  |  |  |  |  |
| Number with CV 20\% of below | 40 | 47 | 50 | 55 | 57 | 57 |
| Number with CV 20\% to 40\% | 17 | 13 | 13 | 10 | 8 | 9 |
| Number with CV above 40\% | 11 | 8 | 5 | 3 | 3 | 2 |
| 7 HIGH RISK PROJECT KEYS |  |  |  |  |  |  |
| High Risk Sharks | 6\% | 4\% | 4\% | 3\% | 3\% | 3\% |
| High Risk Teleosts | 7\% | 5\% | 4\% | 4\% | 3\% | 3\% |
| High Risk Molluscs | 8\% | 7\% | 7\% | 7\% | 5\% | 5\% |
| High Risk Dogfish Other | 40\% | 29\% | 23\% | 21\% | 18\% | 17\% |
| High Risk Hagfish | 33\% | 27\% | 26\% | 25\% | 24\% | 24\% |
| High Risk Skates / Rays | 42\% | 35\% | 33\% | 25\% | 24\% | 20\% |
| High Risk Upper Slope Dogfish | 87\% | 57\% | 48\% | 41\% | 37\% | 33\% |


| Method MinS | Sea Days |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 250 | 500 | 750 | 1000 | 1250 | 1500 |
| 68 NON HIGH RISK PROJECT KEYS |  |  |  |  |  |  |
| Number with CV 20\% of below | 52 | 59 | 61 | 62 | 65 | 65 |
| Number with CV 20\% to 40\% | 7 | 6 | 4 | 3 | 0 | 1 |
| Number with CV above 40\% | 9 | 3 | 3 | 3 | 3 | 2 |
| 7 HIGH RISK PROJECT KEYS |  |  |  |  |  |  |
| High Risk Dogfish Other | 4\% | 3\% | 3\% | 2\% | 2\% | 2\% |
| High Risk Hagfish | 5\% | 3\% | 3\% | 3\% | 3\% | 3\% |
| High Risk Molluscs | 8\% | 7\% | 7\% | 7\% | 7\% | 7\% |
| High Risk Sharks | 28\% | 27\% | 27\% | 27\% | 16\% | 16\% |
| High Risk Skates / Rays | 46\% | 25\% | 25\% | 25\% | 25\% | 25\% |
| High Risk Teleosts | 47\% | 26\% | 25\% | 25\% | 25\% | 25\% |
| High Risk Upper Slope Dogfish | 45\% | 43\% | 29\% | 28\% | 28\% | 28\% |


| Method PROP | Sea Days |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 250 | 500 | 750 | 1000 | 1250 | 1500 |
| 68 NON HIGH RISK PROJECT KEYS |  |  |  |  |  |  |
| Number with CV 20\% of below | 46 | 56 | 58 | 59 | 60 | 60 |
| Number with CV 20\% to 40\% | 13 | 6 | 5 | 5 | 5 | 6 |
| Number with CV above 40\% | 9 | 6 | 5 | 4 | 3 | 2 |
| 7 HIGH RISK PROJECT KEYS |  |  |  |  |  |  |
| High Risk Sharks | 3\% | 2\% | 2\% | 2\% | 1\% | 1\% |
| High Risk Teleosts | 4\% | 3\% | 2\% | 2\% | 2\% | 2\% |
| High Risk Molluscs | 7\% | 5\% | 5\% | 4\% | 4\% | 4\% |
| High Risk Hagfish | 29\% | 20\% | 17\% | 15\% | 13\% | 12\% |
| High Risk Upper Slope Dogfish | 31\% | 22\% | 17\% | 15\% | 14\% | 12\% |
| High Risk Dogfish Other | 33\% | 23\% | 19\% | 16\% | 15\% | 13\% |
| High Risk Skates / Rays | 33\% | 23\% | 19\% | 17\% | 15\% | 13\% |


| Method CWCV-HR | Sea Days |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 250 | 500 | 750 | 1000 | 1250 | 1500 |
| 68 NON HIGH RISK PROJECT KEYS |  |  |  |  |  |  |
| Number with CV 20\% of below | 36 | 42 | 44 | 48 | 49 | 50 |
| Number with CV 20\% to 40\% | 14 | 12 | 13 | 9 | 10 | 10 |
| Number with CV above 40\% | 18 | 14 | 11 | 11 | 9 | 8 |
| 7 HIGH RISK PROJECT KEYS |  |  |  |  |  |  |
| High Risk Teleosts | 4\% | 3\% | 3\% | 2\% | 2\% | 2\% |
| High Risk Sharks | 9\% | 5\% | 4\% | 4\% | 3\% | 3\% |
| High Risk Molluscs | 9\% | 6\% | 5\% | 5\% | 4\% | 4\% |
| High Risk Dogfish Other | 16\% | 11\% | 9\% | 8\% | 7\% | 6\% |
| High Risk Upper Slope Dogfish | 44\% | 31\% | 25\% | 22\% | 20\% | 18\% |
| High Risk Hagfish | 46\% | 34\% | 31\% | 29\% | 28\% | 27\% |
| High Risk Skates / Rays | 62\% | 41\% | 34\% | 29\% | 26\% | 24\% |


| Method MinS-HR | Sea Days |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 250 | 500 | 750 | 1000 | 1250 | 1500 |
| 68 NON HIGH RISK PROJECT KEYS |  |  |  |  |  |  |
| Number with CV 20\% of below | 33 | 54 | 55 | 61 | 64 | 65 |
| Number with CV 20\% to 40\% | 17 | 7 | 9 | 3 | 1 | 1 |
| Number with CV above 40\% | 18 | 7 | 4 | 4 | 3 | 2 |
| 7 HIGH RISK PROJECT KEYS |  |  |  |  |  |  |
| High Risk Sharks | 8\% | 4\% | 3\% | 2\% | 2\% | 2\% |
| High Risk Teleosts | 12\% | 6\% | 4\% | 4\% | 3\% | 3\% |
| High Risk Molluscs | 15\% | 9\% | 8\% | 7\% | 7\% | 7\% |
| High Risk Hagfish | 16\% | 15\% | 15\% | 15\% | 15\% | 15\% |
| High Risk Skates / Rays | 19\% | 17\% | 16\% | 15\% | 15\% | 15\% |
| High Risk Upper Slope Dogfish | 25\% | 19\% | 19\% | 19\% | 19\% | 19\% |
| High Risk Dogfish Other | 93\% | 18\% | 18\% | 18\% | 16\% | 16\% |

We were only provided with very broad figures from which to estimate the cost of the various sampling designs and without additional information it is impossible to accurately determine the costs of any of the sampling designs. To enable a relative comparison of the different designs, we have simply estimated relative costs based on the sea-day requirements alone, at $\$ 1000 /$ per sea-day. This will be the very minimum cost as it does not take into account travelling time and expenses and any land-day costs, not does it include any port-based data requirements. The actual cost for port-sampling will depend on where the port-based samplers are based and the amount of length frequency data that will be collected at sea, which is dependent on which sampling design is ultimately chosen.

Thus, all that can be provided in this report is an indication of the relative sampling costs, where a sampling strategy requiring 500 days can be estimated to cost a minimum of $\$ 0.5$ million, whereas one requiring 1500 days will cost a minimum of $\$ 1.5$ million.

## Analysis of fishery-TEP interactions

The collection of information on TEP interactions by previous observer programs was not based on a statistically robust sampling design, but rather during opportunistic sampling while recording estimates of the total weight and size composition of retained and discarded catches. While incidental capture of syngnathids and seals have been recorded in the catch composition since the inception of onboard observing in the former SEF, the ISMP was not required to record all TEP interactions observed (direct and indirect) until 2003.

Development of an optimised ISMP design for TEP species was undertaken at the stratum level for the SESSF, and was based on minimising the variance calculated from:

- the estimated total number of interacting TEP individuals for a particular species group;
- the variance of the mean number of interacting TEP individuals of that species observed for the stratum; and,
- the stratum sizes squared.

A lack of TEP interaction observations in most strata necessitated the introduction of additional underlying 'true mean interactions' (priors) for strata where there were no interactions for the period 2002-2008. Optimum design results are presented with and without allowance for the additional underlying interactions (referred to loosely as 'priors), and with the constraint that the total sampling intensity equals 400 sea-days.

Optimised for each of the five TEP species groups individually, CVs below $20 \%$ can be achieved for the bird and pipefish groups, and a CV of $24 \%$ achieved for seals. CVs were greater than $60 \%$ for the sharks and whales groups without introduced priors and about $120 \%$ with priors. Allocation of sampling effort to each stratum based on the average sampling required to optimise CVs for each species group results in higher CVs, but in most cases these increases were only small. CVs for the birds and pipefish groups increased to just above $21 \%$ and $19 \%$ respectively. A further (but mostly small) increase in CVs resulted from allocation of sampling pro-rata to the shot distribution evident in the fishery. CVs for the birds, pipefish and seals groups increased to $26 \%, 20 \%$ and $30 \%$ respectively, and remained at about $130 \%$ for both sharks and whales groups.

Results in this section should be treated with caution because of the paucity of data that was available for optimisation of TEP interactions.

| No prior introduced | Sea-days=400 |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Birds | Pipefish | Seals | Sharks | Whales |
| CV when optimised for each group | $15 \%$ | $14 \%$ | $24 \%$ | $67 \%$ | $84 \%$ |
| CV when optim ised is averaged for each group | $21 \%$ | $24 \%$ | $29 \%$ | $124 \%$ | $140 \%$ |
| CV when allocated pro rata to the shot distribution | $27 \%$ | $21 \%$ | $31 \%$ | $181 \%$ | $202 \%$ |


| Prior introduced | Sea-days=400 |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Birds | Pipefish | Seals | Sharks | Whales |
| CV when optim ised for each group | $17 \%$ | $16 \%$ | $24 \%$ | $120 \%$ | $120 \%$ |
| CV when optimised is averaged for each group | $21 \%$ | $19 \%$ | $27 \%$ | $136 \%$ | $128 \%$ |
| CV when allocated pro rata to the shot distribution | $26 \%$ | $20 \%$ | $30 \%$ | $137 \%$ | $131 \%$ |

## Length and age sampling

## Length

Methods were developed to make it possible to comment on the adequacy of the sampling for length and age. For this purpose it was assumed that the redesign of the SESSF ISMP would be primarily based on the retained and discarded catch information, and hence length- and age-related design considerations are an adjunct to the main drivers of the design. Given that context, the most important design issue considered for length was the trade-off between the number of fish sampled per shot or trip, and the number of shots or trips sampled. Shots are relevant to the at-sea sampling for length, while trip is applicable to the port-based length
sampling. While the sampling frame for the sea samples can be approximated as a standard random stratified sampling scheme with infinite population sizes (even though the population sizes are finite), with the strata as defined here for the ISMP, the sampling frame for the port sampling is less easily characterised on that basis. In the first instance the concept strata for port sampling is not the ISMP strata, but rather each trip that is sampled. In that case, the stratum size is related to the total number of fish landed. However, since not all trips are sampled, and since there are apparently no within trip replicates, the applicable sampling frame falls into the class of complex sampling frames. Thus the method developed here, assuming that each trip represents a simple random sample of a length frequency distribution from an infinite population size of trips, is described as a short-cut method. For the port sampling data only, the multinomial within trip variance and the trip-to-trip variance is decomposed so that the implications of different numbers of fish sampled per trip, and trips sampled in total, can be explored. These results suggest that if the total annual number of individuals sampled per year is held constant, then the precision of length frequency estimates is always improved by increasing the number of trips sampled (and of course reducing the number of fish sampled per trip). There are indications that the port-based ISMP data collected during 2007 and 2008 shifted to sampling more fish per trip on fewer trips. This strategy would reduce the precision of length frequency estimates and should be avoided.

The same calculations were not carried out for the at-sea length frequency data, but the same result is extremely likely, i.e. that sampling of length frequencies from more shots will significantly reduce MWCVs for the length frequency distributions, and not the sampling of more fish per shot. Cost considerations obviously impinge on the shot/trip versus fish measured trade-off and the additional expense of more shots/trips is likely at some point to kick-in, but this threshold requires more detailed economic information than was available for this study. It is possible that if increasing the number of shots is much more expensive than increasing the number of fish sampled, then increasing the number of fish sampled per trip is the only economically feasible option. If so, one needs to appreciate that the benefits will be considerable less than 'root N ' where N is the total number of fish sampled.

We do provide some information in the text on species specific MWCVs, and how to determine what sample size would be needed to achieve lower MWCVs. It is easy in this output to highlight species where the MWCVs are too large, for either the retained or discarded component of the catch.

The precision of the length frequency distributions from at-sea samples seems most inadequate overall for retained: gummy shark, gemfish, ribaldo. A large number of species from the discarded catch exhibit unacceptably high MWCVs for their length frequency distributions.

For the port sampling data the following species show MWCVs well in excess of $20 \%$ : redfish, school shark, blue-eye trevalla, alfonsino.

## Age

The adequacy of ageing samples was studied using either the at-sea or port-based length frequency distributions as fixed quantities, and then calculating the MWCV for the age frequency distributions from the cumulation of multinomial error arising from the age-length key proportions and number of fish aged per length class. Assuming that the length frequencies are error free in this way allows one to focus on the errors in the catch age frequencies arising from the age-length key. It is of course acknowledged that the total variance of the age frequency distributions estimates involves both the errors in the length frequency estimates and the age frequency estimates per length class.

This calculation assumes that the age sampling was length stratified, which may not be the case, but the method is regarded as a useful index of the adequacy of age sampling. It is then a simple matter to inspect the output and ascertain where the age sampling needs to be increased or decreased to achieve a particular MWCV - we suggested a target MWCV of $10 \%$, but the calculation is easily adapted to deal with other target MWCVs via the square root of N rule described in the text.

## ISMP Data Management

The data pre-processing and analysis phase of this project highlighted that numerous improvements can be made to data collection and storage. The quality of the data available from sampling during 2007 and 2008 was poor with respect to spatial and temporal coverage and consistency compared to previous years. Compounding this problem, these data were also in a different database structure to previous years which made consistency in data extraction difficult. The greatest issue with the data are the inconsistencies in recording of a unique key and species code, both of which imposed difficulties in matching CDR, GENLOG and ISMP data, resulting in inefficiencies, increasing the possibility of errors being introduced and reducing the amount of data available for analysis. It is recommended that a common unique key (such as 'vessel ID', 'trip end date' and 'trip end time') be recorded for every fishing and observer trip, that a shot ID code (such as 'shot start time' or 'shot number') be recorded in GENLOG and ISMP data, and that the Codes for Australian Aquatic Biota (CAAB) be used as the standard species code in all SESSF fishery datasets. All of these variables are recorded in various databases, but not consistently or adequately enough to allow specific linking between all of the CDR, GENLOG and ISMP databases. The TEP data and databases require significant improvements if they are to support any level of formal analysis in the future.

Additional improvements to data collection and storage were recommended include: creating a validation procedure for GENLOG data entry to reduce errors in start, end and shot dates, standardising collection of TEP interaction data, and increasing the detail of trawl gear specifications recorded.

Improvements recommended will increase efficiencies for future data processing (such as during annual stock assessments), and make these data sets more compatible with other data such as those collected during Fishery Independent Surveys.
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## 1. Introduction

### 1.1 The Fishery

Australia's Southern and Eastern Scalefish and Shark Fishery (SESSF) is a complex multispecies, multi-gear fishery which was formed in 2003 from the amalgamation of four different fisheries; the Commonwealth Trawl Fishery (CTF); the Great Australian Bight Trawl Fishery (GABTF); the Gillnet, Hook and Trap Fishery (GHATF) and the East Coast Deepwater Trawl Fishery (ECDWTF). The area of operation covers almost half of the Australian Fishing Zone (see Figure 1), ranging from $24^{\circ} 30^{\prime}$ S off Queensland to Cape Leeuwin in Western Australia and from shallow coastal waters to depths of over 1000 m . The SESSF caught approximately $22,000 t$ of fish during 2007 with a gross production valued at around $\$ 96$ million during the 2006-07 financial year (Morison, 2008). More than 100 species of finfish and invertebrates are routinely taken in the SESSF, supplying most of the fresh fish for markets in NSW, Victoria, Tasmania and South Australia, and some product for the export market.

Output controls on the trawl sector of the SESSF were first introduced in 1988 when a total allowable catch (TAC) was set for eastern gemfish. TACs were set for orange roughy in 1990 and a further 15 species or species groups in 1992 (blue-eye trevalla, blue grenadier, blue warehou, flathead, western gemfish, jackass morwong, John dory, ling, mirror dory, ocean perch, orange roughy, redfish, royal red prawn, school whiting, silver trevally and spotted warehou). Since then, more species or species groups have been allocated TACs, and with amalgamation with the three other sectors, there are now 34 species or species groups subject to TACs in the SESSF. These 34 species or species groups comprise about $80 \%$ of the fisheries total catch (Morrison, 2008). Other output controls include prohibited species and trip, bycatch and size limits for some species. Input controls used to manage the fishery include limited entry, gear restrictions and some area closures.

### 1.2 Bycatch and Discarding

Fisheries bycatch and discards have received considerable worldwide attention with concern expressed over potential ecological effects, interactions with other fisheries and perceived wastage (e.g. Saila 1983, Alverson et al. 1994, Murawski 1996, Mace 1997, Hall et al. 2000), and a recent review of discards in the world's marine fisheries estimated that approximately $8 \%$ of fish caught are discarded (Kelleher 2005). While Kelleher (2005) reported that trawl fisheries only accounted for about $22 \%$ of the global landings, they accounted for more than $50 \%$ of the total estimated discards. For stock assessment it is crucial that the total catch is estimated as this may be considerably higher than the landed catch (Howell and Langan 1987, Hilborn and Walters 1992, Alverson et al. 1994).

Methods to estimate the discarded portion of the catch usually involve at-sea observations (Howell and Langan 1987, Atkinson et al. 1994, Gerrior et al. 1994, Murawski et al. 1996, Smith et al. 1997), but alternate methods have also been developed (e.g. Casey 1996). Portbased sampling of commercial landings for both size and age are also important components for monitoring the landed catch.

### 1.3 Monitoring in the SESSF

Catch monitoring in the then South East Fishery (SEF) began towards the end of 1992 by the Scientific Monitoring Program (SMP), after the introduction of TACs prompted concerns that they would lead to an increase in discarding (as argued by Tilzey, 1994). The SMP was implemented by the Bureau of Resource Sciences (BRS) to collect data on catches in the South East Trawl Fishery with the primary objective to provide information (including discards and bycatch) for stock assessment of the Commonwealth managed fisheries.

Participants in this project were AFMA, BRS, CSIRO, Industry and State fisheries agencies from NSW, Victoria and Tasmania.

A review of the SMP in 1995 recommended that the various elements of the SMP should be combined into a single integrated program, an Integrated Scientific Monitoring Program (ISMP). The term "Integrated" for the purposes of this program was given to mean the monitoring of trawler catches at-sea, landed trawl and non-trawl catches and the inclusion of fishing activities / catches from NSW ports north of Eden, where as previously these ports were not covered.

In September 1995, a meeting was held between AFMA and the research agencies involved in the SMP, to discuss objectives and design for an interim ISMP for 1996 and 1997. AFMA agreed to undertake a co-ordination role for this program. In October 1995, BRS, NSW Fisheries Research Institute and the Marine and Freshwater Research Institute (MAFRI) Victoria were contracted to undertake the interim ISMP on behalf of AFMA. This program did not include at-sea monitoring of the South East Non Trawl Fishery (SEFNT).

In July 1996, AFMA sought tenders for the design of a new ISMP for the SEF. In August 1996 the National Institute of Water and Atmospheric Research Ltd New Zealand (NIWA) and MAFRI were jointly contracted to design a new ISMP for the fishery. Final design of the new ISMP was received by AFMA in May of 1997 and submitted to various independent reviewers for comment. Comments from reviewers were received in early June 1997 and a meeting between key design people and various reviewers was held in late July 1997. As an outcome of this meeting the statistical design of the program (Smith et al. 1997) was accepted, however there were concerns about the lack of detail relating to the practical implementation of the program. Reviewers acknowledged that the presentation of detailed operational procedures in the design report was not part of the agreed project brief. It was agreed that there was a need for this level of detail and that it should be addressed as part of the tender process.

MAFRI submitted the successful tender and conducted the ISMP from 1998 to mid 2000. In April 2000, AFMA again sought tenders to undertake the ISMP in the SEF for a further three years, including two additional strata - East Coast Deepwater Zone and the Victorian Inshore Trawl.

It was recognised that in a complex and dynamic fishery such as the SEF, any revision of the sampling design would need to be undertaken on a regular basis to ensure that the monitoring program adequately sampled the fishery, yet the ongoing process needed to be relatively automated to minimise the costs involved (Knuckey and Gason, 2001). This 'adaptive survey design strategy' used past information as well as the most recent logbook and ISMP data to revise the ISMP sampling strategy on an annual basis. This sampling design has since been used to annually monitor discards in the fishery (see Talman et al., 2003; Talman et al., 2004; Koopman et al., 2005; Koopman et al., 2006; Koopman et al., 2007).

In contrast to the trawl sector of the SEF, there was limited information available on the nontrawl sector (SENTF) prior to 2000. To obtain the information necessary to design a sampling strategy for the SENTF, a pilot study was undertaken by MAFRI from 1999 to 2000. Based on the results of this project (Knuckey et al. 2001), the SENTF was stratified by gear and region, and the discard rates and associated coefficients of variations (CVs) of important species were estimated. Simulation modelling was undertaken to determine the number of shots within each stratum that would be required to achieve these target CVs. Discard rates for most of the main target species in each stratum were found to be low or nonexistent. As a result, the ISMP only needed to undertake a low-level sampling regime to
achieve target CVs for discard rates and adequately represent the spatial and (to a lesser extent) temporal characteristics of the catch composition from the different non-trawl methods. Based on feedback from South East Non-Trawl Management Advisory Committee (SENTMAC), on-board observer days were allocated to monitor the different fishing methods in the five zones of the South East Non-trawl Fishery. MAFRI conducted the ISMP for the SENTF for the two years 2001-02 and 2002-03.

MAFRI conducted an FRDC-funded pilot study in the Great Australian Bight Trawl Fishery (GABTF) from 2000 to 2001. This study assessed and quantified bycatch in the GABTF and collected additional biological information for major species. Upon completion of the project in June 2001, industry, scientists, AFMA and other stakeholders felt it was necessary to continue monitoring the fishery through an onboard observer program. At Industry's request, MAFRI has conducted the ISMP for the GABTF for the two years 2001-02 and 2002-03.

During 2003, Primary Industries Research Victoria (PIRVIC - previously MAFRI), successfully applied for a further three year tender to conduct the ISMP for the trawl, Great Australian Bight and non-trawl sectors. PIRVIC carried out this work until December 2006. AFMA decided to bring the ISMP 'in-house" during 2006 and has done so since January 2007.

Over the last five years there have been considerable changes to both the structure of the SESSF fishing fleet (largely although not entirely as a result of the Structural Adjustment Package) and the expectation of the focus and level of monitoring of the fishery to take into account its impact on high-risk bycatch/byproduct species and threatened / endangered / protected (TEP) species. The current design of the ISMP was not established to cover such a level of change in the fishery and may not adequately cover all of the information now required. As a result, there has been a need to call for a redesign of the SESSF monitoring program that includes all sectors and reflects changed objectives of the management needs.

## 2. Objectives

The following sets out the objectives for the entire study.
Undertake a complete redesign of the current ISMP sampling regime for all sectors of the SESSF to ensure that it is representative, effective and statistically robust including consideration of the following requirements:
(a) Update the ISMP sampling regime to reflect changes to the fishing fleet and fishing practices in the SESSF since the Australian Government's "Securing our Fishing Future" structural readjustment package as well as the inclusion of the Gillnet and Shark hooks sector to the SESSF in 2003. In addition, incorporation of any new data sources should be considered including those identified by the ISMP Review Committee.
(b) Review and update the stratification and various coefficients of variation (CVs) within the current sampling design in light of changes to the fishing industry and to develop new CVs and sampling regime for monitoring of the shark fishery, interactions by fishing operators with threatened, endangered and protected (TEP) and high risk species.
(c) Review and update the stratification and various coefficients of variation (CVs) within the current sampling design in light of changes to the fishing industry and to develop new CVs and sampling regime for monitoring of the shark fishery, interactions by fishing operators with threatened, endangered and protected (TEP) and high risk species.
(d) Review and update the stratification and coefficients of variation (CVs) within the current sampling design for the size/age composition of the total catch (retained and discarded) for selected (quota) species.
(e) Update the ISMP sampling regime to provide effective and statistically robust coverage for estimating the total catch (retained and discarded) through at sea and in port sampling of species, caught by all sectors of the SESSF, specifically:
(i) quota species;
(ii) major non-quota species;
(iii) other non-quota species;
(iv) any other species/groups identified by the ISMP Review Committee.
(f) Update the ISMP sampling regime to provide effective and statistically robust coverage for recording the total number and circumstances of interactions (including the life status at the time of the interaction) by all sectors of the SESSF with species identified as:
(i) high risk through the ecological risk assessment process;
(ii) TEP species (not identified as high risk through the ecological risk assessment process).
(g) Incorporate into the ISMP sampling regime the collection of additional information relevant to fishing trips/operations including:
(i) fishing gear type used on fishing trip/operation including any gear modifications;
(ii) use of mitigation tools used during fishing trip/operation including any mitigation tool modifications;
(iii) the level of observer coverage during the fishing trip/operation;
(iv) offal management systems in place during the fishing operation/trip;
(v) other significant fish trip related information.
(h) Provision of a range of sampling redesign options including a table of costs outlining what CVs will be achieved with each option - cost vs. precision levels will determine the size of the sampling program.

## 3. General Methods

### 3.1 Data gathering

All data relevant to the revision of the sampling regime was gathered. Such data included information pertaining to the following fisheries:

- Commonwealth Trawl Sector (including the Victorian Inshore Trawl)
- Gillnet, Hook and Trap Sector
- East Coast Deep Water Sector
- Great Australian Bight Trawl Sector

For each of these fisheries the data types comprised, where available (also see Appendix A):

- Catch and effort data
- Catch landing data for quota species
- On-board ISMP data
- Port based ISMP data
- Central Ageing Facility age and length data sets for SESSF species (as mentioned in page 4 of "ISMP EOI List of Docs - Final.doc")
- Various shark databases (as mentioned in Page 4 of "ISMP EOI List of Docs Final.doc")
- Data related to interactions with threatened species, including birds, syngnathids and seals
- Data from Fishery Independent Surveys (FIS)


### 3.2 Data reconciliation

For all of the data, it was necessary to establish basic data connections and relationships (i.e. adequate digital reading of the data), as well as establishing a means to allocate data to strata, either strata used in ISMP studies historically, or candidate ISMP strata for consideration during the execution of the study. The identification of strata and variables for which CVs are to be estimated will involve certain critical data in the various datasets supplied.

For all data, data audits were carried out, to assess the various fields involved, their quality with respect to missing values, blanks, empty spaces etc, to determine how they are related to each other and to the objectives of the study.

Relation links had to be established between the different tables of relational databases as and when required.

Where required, disparate databases were merged on a sensible basis (e.g. port level data with at-sea level data), using the best methods available based either on a logical analysis of the situation, and/or guided by past practice for similar previous studies.

### 3.3 Statistical review

Parallel with the process of data reconciliation referred to above, a statistical review of previous stratification and sample allocation decisions for ISMP studies was carried out.

This review comprised

- Distributional assumptions
- Stratification considerations
- CV estimation and calculation approaches

The primary references were Knuckey and Gason (2001) and Smith et al. (1997), which document the reasoning behind previous sample designs and/or redesigns. Note that these studies were based on the 'main species' caught, while the EOI for this study specifically asked for CVs to be reported for 'other non-quota species' and 'any other species/groups

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identified by the ISMP Review Committee'. This requirement broadened the study beyond the scope of Knuckey and Gason (2001) and Smith et al. (1997).

Knuckey and Gason (2001) also base their design predominantly on discard rate (expressed as a proportion of the total catch coming up in the gear) and length frequency CVs, whereas we note that the EOI specifically states that CVs (and hence design requirements) need to be included for the retained catch as well.

The methods in Knuckey and Gason (2001) use a non-parametric re-sampling approach to the calculation of the distribution of the design variable of interest (e.g. the discard \%). Although not reported here, this study explored the merits of the non-parametric bootstrap, and although this method, as well as a method based on the binomial distributions and its derivatives (the beta distribution and the beta prime distribution), showed considerable promise, particularly for the treatment of shot to shot correlations within trips, we eventually relied more heavily on a simpler approach using the sample variance information in the historic ISMP data, and only in certain cases did we rely on the beta distribution and the beta prime distribution.

Documentation available as part of the EOI made reference to certain minimum CVs that were treated as targets in the development of optimal ISMP designs. There was a lack of clarity, however, in the EOI with regard to the unit to which a given CV requirement is applicable, whether to a zone, a fishery, or all fisheries for a given stock/species. Proposals for the treatment of these issues are made in this document.

Not all sub-fisheries in the SESSF which are part of this study have been included in the ISMP process to date. For those that have, it was necessary to review the arguments for the stratification schema adopted in the past, and consider whether revisions are appropriate. For those that have not, the merits of stratification needs to be considered. The latter includes East Coast Deepwater, shark gillnet and Vic Inshore Trawl.

### 3.4 Statistical analyses

Following on from the statistical review, the capability to calculate CVs for desired quantities at the desired level of resolution (e.g. zone, fishery, or all fisheries for a given stock/species) was developed, essentially for the following three data types:

1. Discarded catches
2. Age and length frequency distributions
3. Interactions with TEP species

We note that the EOI specifies that the following species groups should be included in the development of robust sampling designs:

- Group 1: Quota species
- Group 2: Major non-quota species
- Group 3: Other non-quota species
- Group 4: Any other species/groups identified by the ISMP Review Committee
- Group 5 (nature of interactions): Species identified as high risk through the ecological risk assessment process.
- Group 6: TEP species (not identified as high risk through the ecological risk assessment process).

This list significantly expands the scope of the design compared to the work reported in, for example, Knuckey and Gason (2001).

One of the first steps in the analysis of the relevant data therefore involves identifying species that should be included in the analyses at a species level, and others that should be dealt with at the level of a group of species.

The ultimate aim is to be able to show the impact of including each species or species group in consideration of the optimal design, in sequential fashion. It seems likely that expanding the study from species groups 1 and 2, to include groups 3 to 6 will significantly expand the sample size requirements. In such a situation we expect that management will need to understand how the progressive expansion of the scope of the study impacts on sample sizes.

### 3.5 Collection of additional relevant information

The EOI sets out some additional requirements that need to form part of this study, namely
"Incorporate into the ISMP sampling regime the collection of additional information relevant to fishing trips/operations including:

1. fishing gear type used on fishing trip/operation including any gear modifications;
2. use of mitigation tools used during fishing trip/operation including any mitigation tool modifications;
3. the level of observer coverage during the fishing trip/operation;
4. offal management systems in place during the fishing operation/trip;
5. other significant fish trip related information."

### 3.6 Discarded catch amount

Clearly the discarded catch can only be determined from the at-sea ISMP sampling data (perhaps supplemented by survey data). For quota species, a stratified mean discard $\%$ can be calculated, and then applied to the total (by fishery) catch for quota species. Smith et al (1997) provides a basis for the estimation of the CV of the stratified mean discard $\%$, and although there is no reason to question this, the statistical arguments underlying this were reviewed. The stratification schema also needs to be reviewed in light of the requirement that strata, in order to have any utility, need to reflect common discard patterns.

For non-quota species, the discarded catch can also only be determined from the at-sea ISMP sampling data (again possibly supplemented by survey data). A similar procedure to that described above for estimating discarded quota species catches is implemented to estimate the discarded non-quota species catches. This process was reviewed.

### 3.7 Retained and discarded length frequency distributions

The "Statement of Requirements" is not prescriptive about exactly which species need to be included in the sample redesign from the perspective of length frequency distributions. However, since the requirement does specify a need to review and update the existing ISMP, we take as our direction the current practices followed in ISMP for species sampled. This practice is to take length measurements for quota species.

### 3.8 Sampling for Age

The requirements for assessing the adequacy of sampling with respect to age distributions appear to involve, at a minimum, updating the work reported in Krusic-Golub and Gason
(2006, see Appendix 6 of Final report on ageing work on key SESSF species from the 2006 sampling year). Krusic-Golub and Gason's (2006) MWCV calculations cover the uncertainty in the final age distribution assuming a given true length frequency, and assuming correct ageing. The main source of error thus arises from the distribution of age classes found in each length class, in the case where age-length keys (ALKs) are used. Alternatively, if a straight age distribution is based on a random sample, then the variability is due to that distribution. Krusic-Golub and Gason (2006) also limit their simulations to scaling the overall sample up or down, and do not explore the possibility of alternative sample allocations across length classes in the case of ALKs. The work reported here is an update of the work in KrusicGolub and Gason (2006).

### 3.9 Interactions with TEP species

There is a requirement in the request for EOI to develop new CV s and a sampling regime for monitoring interactions by fishing operators with threatened, endangered, protected (TEP) and high risk species. For this we studied the precision of the estimates of number of shots interacting with TEP species and number of animals involved in those interactions in the framework of either a binomial distribution or a negative binomial distribution. The former would be suitable for data where an event either does or does not occur linked to a fishing activity like a shot (e.g. whales were present) while the latter would be appropriate where the counts of numbers of individuals associated with a shot was recorded (e.g. 23 whales were present at the time of the shot).

### 3.10 Port based sampling versus at-sea sampling

As an "integrated" monitoring program, the design of the ISMP includes both on-board and in-port monitoring. The at-sea component of the ISMP provides size distributions for retained and discarded catches by strata and other information which can only be collected at-sea (TEP interactions, fishing practices etc). This information can be supplemented by port-based sampling during which length frequency and age data on the retained catch and species composition can be collected in a more cost effective manner. The potential benefits of using port-based monitoring to supplement information collected at-sea was considered in the redesign of the ISMP.

## 4. Interim Report

An Interim Report was provided to AFMA on $9^{\text {th }}$ July 2009. AFMA provided feedback on the interim report in late August 2009, which has been used to modify some of the methodology presented in the interim report.

The timelines of this study were such that not all of the available data could be analysed for the Interim Report. However, this final report incorporates all known and available logbook, catch landing and ISMP data up to and including 2008.

The issues that AFMA addressed in their feedback on the interim report are as follows, with a summary of the responses that were made:

### 4.1 CV based on total catch

The question of (b) the appropriate method to be used to infer discard tonnages based on the ISMP records and (b) the precise definition of target levels of precision, was raised in the interim report. Past practice (refer previous ISMP annual reports or design studies) focus on the 'discard rate' $d$ as the quantity of interest, where $d$ is discarded divided by total catch. However, it is $d /(1-d)$ which must be multiplied by the retained catch in order to estimate the discarded tonnages from the catch disposal records, and the variance of the quantity $d /(1-d)$
becomes very large for large $d$. The interim report thus proposed that considering that the ultimate purpose was to quantify the 'total catch' including discards, it is more sensible to use statistical targets expressed as a C.V. on the total catch per se (previous ISMP studies refer to a C.V. on $d$ as a target precision level). An important need for this change has been that, compared to previous designs, the current redesign must consider many more species or species groups (other than quota species) that are largely or completely discarded and otherwise not well identified or recorded in logbook data. This proposal was accepted by AFMA and others consulted about the interim report.

### 4.2 Stratification proposals

The stratification proposals made here are to stratify the GAB by depth, but not into longitude groups. For the GHAT, the proposal is to use the zones of the SET as a geographic zonation, although no zonation west of 138 degrees longitude is proposed. We note that the stratification presently in use (AFMA Observer Program: ISMP Progress Report, Progress report for May 2009) is somewhat different, for example it includes a longitudinal split for the GAB (far West, West, Central and East), and it has a similar but slightly different zonation for the GHAT including some additional spatial stratification west of 138 degrees longitude. The 2007/2008 ISMP data are quite few and probably insufficient to argue between the zonation proposed in this document and the zonation reported in the AFMA report referred to above. Any feedback on this matter would be valuable, such as overriding issues underlying the stratification in the AFMA 2009 document referred to above. AFMA requested that for the GHAT, the strata be established in accordance with existing practice, and this change is thus incorporated into the final report. Some reservations were expressed about the shift from longitudinal to depth splits for strata in the GAB, however no clear position was expressed in relation to this. This final report thus uses the depth-wise strata for the GAB.

### 4.3 Grouping of species

Another area where feedback was sought in the interim report was the way that species were grouped for the purpose of the interim report. In that work, the $500+$ species that are at times recorded in the SESSF fishery, were reduced to a number of groups termed 'project keys'. The 'project key' is the same as a species designation for the major quota species and for important non-quota species, but thereafter species are grouped into broad taxonomic groups for all calculations and results. In all, species which are subject to capture are represented by 73 species and/or groups, the 'project key'. In addition, 5 species groups were defined for threatened, endangered or protected (TEP) species (Seals (Sea Lions included), Sharks, Pipefish, Birds, Whales/Dolphins, Reptiles (none recorded)), of which by far the dominant groups are the seal group and the bird group (from the point of view of species involved in interactions with fishing gear). The analyses looked at TEP species from two points of view (a) the number of shots involved in an interaction with TEP species and (b) the number of individuals for a TEP species involved in an interaction with the SESSF fishery. AFMA expressed agreement with this method of grouping, hence this approach remains for the final report.

### 4.4 Grouping of high risk species

The interim report contained a proposal to group the $\sim 30$ high risk species into 7 groups, i.e.

1. Teleosts
2. Sharks
3. Dogfish other
4. Rays/Skates
5. Endeavour Dogfish.
6. Molluscs
7. Hagfish

The following feedback was received from AFMA on this proposal (A-M Lynch, pers. Comm..)
"Secondly, looking at the proposed grouping for High Risk Species, a bit of clarification is required. The report lists 'Endeavour Dogfish' which is actually an old group term for Southern Dogfish, Endeavour Dogfish and Harrisson's Dogfish. It is suggested that another name such as High Risk upper slope dogfish ( 5 species) is used to define this group. It would be useful to group these dogfish species plus the Greeneye Dogfish (Squalus mitsukurii) and the Nilson's Deepsea Dogfish, also known as the Leafscale Dogfish (Centrophorus squamosus), all of which are in the current High Risk species list and are currently the focus of a management strategy that is being developed. Can you please let us know if this is possible?"

For this final report the grouping of the high risk species has been modified to accommodate the AFMA feedback received on the interim report.

1. Teleosts
2. Sharks
3. Dogfish other
4. Rays/Skates
5. Upper Slope Dogfish.
6. Molluses
7. Hagfish

### 4.5 Design optimisation considerations

The interim report was unable to resolve the importance rating of different species and/or species groups for the survey design optimisation. Different objective functions were considered and presented. The following feedback was received from AFMA in this regard:
"... the optimization methods to be provided should be as follows -

1) the catch weighted CV for non-commercial species is minimized; and the value weighted CV for commercial species is minimized
2) as many species as possible have CVs below a defined threshold."

These recommendations are adopted for this final report. However we include as well a number of other methods of allocation in sample sizes to strata, including:

- Allocation of shots pro-rata to the shot distribution in the commercial SESSF fishery.
- For the TEP species groups, we include a sample size distribution which is the average of the optimised allocations for each of 5 TEP species groups (initially run separately).

A further point of clarification requested of AFMA was:
With regard to the multi-gear / fishery nature of the SESSF, we assume a unified ISMP across all gear and fishery sectors and hence the targets dealt with are expressed at the SESSF level. If there are, however, sample size quotas per fishery (GAB, GHAT, SET) which predefine the split of sampling effort between these sectors, then the authors need to be notified so that it can be brought into the optimisation explicitly.

AFMA did not comment on this and hence the design is optimised at SESSF level.

### 4.6 Sampling unit

Finally AFMA expressed reservations about the use in the interim report of shot as the sampling unit for the revision of the ISMP design. AFMA suggested instead that the unit of sampling should be trip. The authors argued that provided the correct analysis was done, whether trip or shot was used as a sampling unit was not germane to the final quantitative results. However, they acknowledged that under the heading 'correct analysis', it was necessary, in addition to catering for the number of shots per trip, to address the extent of within trip correlation. A compromise proposed to AFMA was to use shot as the fundamental sampling unit, but to include in the analyses a fair quantification for the within trip covariance referred to. In addition, since costs are paramount in the determination of the optimal survey design, we have optimised the design using number of days as the optimisation variable, in the process days are internally converted to shots by the appropriate factor for each stratum. This compromise proposal was accepted by AFMA.

## 5. Methods and Results

Due to the multiplicity of calculations this report is not structured traditionally by methods and results, but rather methods and results are addressed for each topic in turn. An overall discussion section follows once the presentation is complete for all data types and topics.

### 5.1 Data gathering

The following basic information was supplied in April 2009:

1. CDR: Catch disposal records for the SESSF
2. GENLOG: Logbook data for the SESSF
3. ISMP: Observer data, including onboard Length/Frequency observation tables, catch composition and weight tables for retained and discarded catch as well as interaction observations for threatened and endangered species (TEP).
4. PORTLENGTH: Port length frequency information
5. CAF: Catch at age data

The scope of the data was as follows:

### 5.1.1 CDR data

These data covered, by fishery, the following years:

- SET (including ECDW and VIT): 1996-2009
- GAB: 2001-2008
- GHAT: 1997-2008


### 5.1.2 GENLOG data

These data covered, by fishery, the following years:

- GENLOG 1985-2009


### 5.1.3 ISMP data

These data covered, by fishery, the following years:

- SET ISMP 1992-2008
- GHAT ISMP 1999-2008
- GAB ISMP 2000-2008

The TEP (TEP Interactions table and the TEP abundance table) data in the ISMP dataset only covered the period 2004 to 2008.

The initial tranche of ISMP data (supplied in April 2009) covered, by fishery, the years 1992 to 2007. However, the amount of 2007 data included was very limited. A revised ISMP dataset was made available on 3 June 2009. This dataset includes ISMP data for 2007 and 2008, albeit in an entirely different format to the data provided previously in April 2009. Since the TEP Interaction and TEP Abundance tables of the ISMP dataset only appear to be substantially populated post 2004, the TEP component of the analyses reported upon here is effectively limited to the period 2004 to 2008 inclusive.

### 5.1.4 Port based length frequency information

- Port length frequency data: 1979-2009


### 5.1.5 CAF: Catch at age data

- These data cover the period 1979 to 2008, for about 15 different species.


### 5.2 Data reconciliation

Much of the work underlying this report was involved with pre-processing the data to a point suitable for higher level analyses (see Appendix A for full details), and is not really reflected in the pages that follow. A crucial step in this regard is the much referred to merge between the port catch disposal record (CDR) dataset and the at-sea shot by shot (GENLOG) dataset. The reliability of this merge is central to the reliability of the final results. As is also documented in previous studies, the merge in question must be achieved without the aid of a unique key linking the shots of a fishing trip with the landing and catch disposal information. While this posed some difficulties in previous studies, the potential for error is magnified here since this study treats the data at a finer species resolution than previous similar work (including non-quota and high risk species). Thus the errors that arise out of 'forcing the merge' are potentially more significant for amounts such as discard estimates for less frequently caught and/or recorded species. Such errors are not necessarily adequately quantified by means of CVs, but would be revealed by adopting varying approaches to 'forcing the merge'. Our concerns in this regard have only been addressed inasmuch as we limited the data to the post 2002 period, since there were higher level considerations that pointed to a risky CDR-GENLOG merge pre-2002. For example, a substantial proportion of the vessels in the CDR data do not appear in the GENLOG data pre-2002, and vice versa (see discussion Appendix A).

The CDR-GENLOG merge referred to above is achieved by a reconciliation of vessel IDs, trip start and end dates, and shot dates, as described in more detail in Appendix A. Whereas pre-2002 the potential for a reasonable proportion of shots in GENLOG to survive the merge is compromised somewhat by the mismatch between vessels in the CDR and GENLOG datasets (based on a macro view of the situation as outlined in Appendix A), post 2002 there are some challenges that arise due to the high level species resolution use made of the CDR
and GENLOG data. In short, for a given trip, a species may occur in the GENLOG shots assigned to that trip (CDR) but may be completely absent in the CDR data, or the other way around (i.e. absent in the GENLOG data but present in the CDR trip data). Although some of this mismatch could be due to misallocation of GENLOG shots to the wrong trip, or to the failure of allocation of a GENLOG shot because of incorrect or absent vessel information, an additional potential source of error is the variable degree of species identification that occurs for shots, landings, and indeed observer covered shots (and perhaps landings). In previous work on this topic the ramifications of this species level source of error was limited because of the focus of that work on $\sim 20$ quota species and an 'other' category. Many of the species mismatch problems and errors occur 'internally' in the 'other' category and are thus quite opaque. However, the study reported here is far more ambitious in the species resolution that is attempted, given the project requirements w.r.t. species, and thus encounters this kind of issue on a scale that would not have been encountered in previous work. In order to address this in a way that seemed to retain the spirit of the project requirements, but imposed some limit on the extent of errors from the speciation issue described, a particular species grouping approach has been adopted, as described below.

### 5.3 Coding of species

Among the challenges involved in the linking of the three datasets involved in this study, are those related to the different coding of species in these datasets. The issues that were addressed covered the following:

- The ISMP data contained CAAB code, common name and scientific name.
- The CDR data contain a 3-letter AFMA code, common name and scientific name.
- The GENLOG data contain AFMA code and CAAB code.
- Thus no common key was available on all 3 datasets.
- There are instances in which more than one AFMA code is linked to the same CAAB code in the GENLOG data.
- Certain species appear to be present in one or two data-sets but not all three.
- Certain species are grouped for quota purposes, yet individual species within these groups have differing risk status.
A first step in addressing these matters was to link AFMA codes to CAAB codes using an external reference, so that CAAB code is a common key between the three datasets.
A total of 729 unique CAAB codes appeared in one or more of the 3 data sets. Of these, 297 appeared in the GENLOG, 580 in the ISMP and 317 in the CDR. Some means of grouping species was required in order to achieve a reasonable correspondence between the 3 datasets.
For the purposes of the discard rate and total catch analyses, we created the following system of classification, resulting in 75 distinct "Project Keys"
- The full list of quota species is given in Table 44. Where quota for a single species is allocated by AFMA, the species is allocated a Project Key. Where quota is allocated to a basket of species, a Project Key is allocated to the basket. This applies to the Deepwater Shark basket ( 18 species), Saw-shark basket ( 3 species) and the Oreo basket (3 species).
- In order to identify major non-quota species for each gear type, species were ranked by total catch recorded in the GENLOG for years 2002 to 2008. From these sorted lists, species which cumulatively accounted for $80 \%$ of the total catch for that gear
type were labelled as "major non-quota" species. Each of these was allocated a Project Key. The list of major non-quota species is shown in Table 45.
- Twelve additional project keys were created to group all remaining species, which are considered to be minor non-quota. These grouped keys are: Chimaeras, Crustaceans, Dogfish, Echinoderms, Fish, Hagfish, Rays, Saw-sharks, Molluscs, Sharks, Stingarees or Other.
- Birds, mammals and reptiles are not included in this classification (but are treated separately in the sections dealing with TEP species).
- In total, 75 distinct "Project Keys" were developed using the above methods (Table 46)

The components of this study which deal with High Risk Species and with TEP species are treated separately. In these, the species involved are treated individually. These species appear in Table 47 and Table 48.

### 5.4 Analysis of sub-fisheries and potential strata using clustering techniques

This section discusses whether the present stratification is appropriate for forthcoming ISMPs. It uses the plots of frequencies of clusters by year, depth and longitude to determine whether the cluster structures are stable over time, and whether there are grounds from the data summarised at this level to create additional strata for the ISMP, or whether some strata could perhaps be coalesced with others.

Notwithstanding the foregoing discussion about species grouping, the cluster analyses reported on here predated the species grouping step of the study, and hence these are based on individual species data throughout. Appendix B provided the graphical details of the cluster analyses performed in this section.

### 5.4.1 Commonwealth Trawl Sector (CTS)

The strata for the CTS are well established and, following analysis which revealed that the spatial distribution and catch composition of the fishery post-SAP had not changed appreciably, we saw no reason to propose re-stratification for this fishery.

### 5.4.2 Great Australian Bight Trawl

For the GAB, the data employed in the cluster analysis was the available GENLOG data from 2002. Five clusters Appendix B-Figure 1 were identified in the GAB fishery, as follows:

1. Deepwater Flathead Cluster - B-Figure 2 shows the species composition of this cluster which is dominated by deepwater flathead ( $\sim 55 \%$ by weight on average).
2. Redfish Cluster - see B-Figure 3, this cluster is dominated by Bight redfish ( $\sim 50 \%$ by weight on average).
3. Grenadier Cluster - see B-Figure 4, dominated by blue grenadier ( $\sim 35 \%$ by weight on average).
4. Gemfish Cluster - see B-Figure 5 dominated by gemfish ( $\sim 74 \%$ by weight on average).
5. Orange Roughy Cluster - see B-Figure 6, dominated by orange roughy ( $>95 \%$ by weight on average).

Frequency plots of these clusters in the GENLOG dataset by year suggests a changing pattern with a decline in Deepwater Flathead Cluster shots and an increase in the frequency of Redfish Cluster shots (see B-Figure 7 and B-Figure 8). The other three clusters have declined in shot frequency over the period 2002 to 2008.

Frequency plots of these clusters, by longitude, show some evidence of heterogeneity (see BFigure 9 - B-Figure 12); however a separate plot versus depth reveals that this heterogeneity is actually due to depth (see B-Figure 13 - B-Figure 16).

As a result of this analysis, it is recommended that three depth strata be defined for the GAB component of the ISMP, at a depth split of:

- Inshore stratum: 0-250 metres
- Midshore stratum: 250-850 metres
- Offshore stratum: $>850$ metres

See B-Figure 17 - B-Figure 20.

### 5.4.3 East Coast Deepwater Trawl

No cluster analyses were carried out for this sector. The intention is to model this by means of a single cluster.

### 5.4.4 Victorian Inshore Trawl

No cluster analyses were carried out for this sector. The intention is to model this by means of a single cluster.

### 5.4.5 GHaT-Longline ( $\operatorname{AL} \& B L$ )

The following clusters were identified B-Figure 21:

1. Pink Ling Cluster - see B-Figure 22, the species composition of this cluster is dominated by pink ling ( $\sim 70 \%$ by weight on average).
2. Gummy Shark Cluster - see B-Figure 23, this cluster is dominated by gummy shark (~ $90 \%$ by weight on average).
3. Blue-Eye Trevalla Cluster - see B-Figure 24, dominated by blue-eye trevalla ( $\sim 70 \%$ by weight on average).
4. Mixed Cluster - see B-Figure 25, dominated by five species comprising more than $90 \%$ by weight: gummy shark, snapper, southern eagle ray, broadnose shark, school shark.
5. Hapuku Cluster - see B-Figure 26, dominated by hapuku ( $\sim 57 \%$ by weight on average).

A plot of cluster frequencies versus depth indicates a depth wise split of the clusters into two groups (B-Figure 33 - B-Figure 36), i.e.

- Shallow ( $<200$ metres) - Gummy Shark and Mixed Clusters.
- Deep (> 200 metres) - Pink Ling, Blue-Eye Trevalla and Hapuku Clusters.

Although this suggests a possible stratum based on a depth split, it turns out that the clusters group logically around gear type (see B-Figure 37 to B-Figure 44), i.e.

- Bottom longline (BL) - Gummy Shark and Mixed Clusters.
- Automatic longline and longline (AL) - Pink Ling, Blue-Eye Trevalla and Hapuku Clusters,
and so this issue is entirely addressed by stratification by gear rather than depth.


### 5.4.6 GHaT - Dropline

The following 2 clusters were identified (B-Figure 45):

1. Blue-Eye Trevalla Cluster - see B-Figure 46, the species composition of this cluster is dominated by blue-eye trevalla ( $\sim 90 \%$ by weight on average).
2. Hapuku Cluster - see B-Figure 47, this cluster is dominated by hapuku ( $\sim 35 \%$ by weight on average), with a $22 \%$ contribution from blue-eye trevalla.

Plots by longitude (B-Figure 50 - B-Figure 53) and depth (B-Figure 54 - B-Figure 57) do not suggest the need for additional stratification of the DL gear type within the GHAT fishery at this stage of the analysis.

### 5.4.7 GHaT - Demersal gillnet

The following 3 clusters were identified (B-Figure 58):

1. Gummy Shark Cluster - see B-Figure 59, the species composition of this cluster is dominated by gummy shark ( $\sim 90 \%$ by weight on average).
2. Gummy Mixed Cluster - see B-Figure 60 , this cluster is dominated by gummy shark ( $\sim 42 \%$ by weight on average), with a fair contribution from a range of other species as shown.
3. School Shark Cluster - see B-Figure 61, the species composition of this cluster is dominated by school shark ( $\sim 65 \%$ by weight on average).

Plots by longitude (B-Figure 64 - B-Figure 67) and depth (B-Figure 68 - B-Figure 71) do not suggest the need for additional stratification of the GN gear type within the GHAT fishery, although there is an increasing frequency of school shark shots with depth.

Despite the above, it was recommended by SharkRAG that the gillnet should be stratified by the shark zones previous used for School Shark. For this reason this stratification was applied to all the GHaT sectors.

### 5.5 Additional stratification considerations based on the discard ratio by cluster.

The ISMP data were used to carry out additional analyses on the clusters identified above. The ratio of discards to total catch (discarded + retained) was calculated regardless of species for each cluster. The F-test and the Kruskal-Walllis test were then carried out to test for statistically significant differences in this cluster-specific, but species-within-clusterindifferent, discard ratio, across the following variables:

- GAB: by season, i.e. winter, spring, summer, autumn
- GAB: by longitude group, i.e. < 127 longitude, 127 - 130 longitude, $>130$ longitude.
- GHAT: by season, i.e. winter, spring, summer, autumn
- GHAT: by zone, we used the SET zones $20,30,40,50$ and 60 , GAB (west of 138 longitude) and a far NE zone, which we simply relabelled as ' 70 ' for subsequent analyses.

The use of the Kruskal-Wallis test is recommended given the non-normal distribution of the discard ratios described above.

The results are presented in Table 50 to Table 53. Although we inspected the pairwise post hoc results, only the overall statistics are shown here. Because many comparisons are examined one needs to be aware of the potential for false positive results. For this reason we discount the single statistically significant results at the $5 \%$ level in Table 51 , for longitude grouping in the GAB. Similarly, we discount the single statistically significant result in the GAB for seasonal groups and analyses. As a result the stratification suggested for the GAB based on the cluster analyses remains at three strata, i.e.

Inshore, Midshore, Offshore.
Results for the GHAT by season are presented in Table 53. Only three gear codes are considered since the AL \& BL gear types are combined in the GHAT ISMP data (but separated in the GHAT GENLOG data). One cluster-gear type combination shows up as statistically significant for the analysis by season. Using this as an argument to stratify by season is discounted for the same reason as before (the chance of obtaining a false positive result).

On the other hand the GHAT analyses carried out on discard ratio per zone as defined show generally statistically significant results (see Table 52). It is therefore recommended that each gear type be further stratified by a geographic variable, which in the analyses defined above would be the following gear/geographic breakdown:

- AL: zones GAB, $10,20,30,40,50,60$, NORTH EAST
- BL: zones GAB, $10,20,30,40,50,60$, NORTH EAST
- DL: zones GAB, $10,20,30,40,50,60$, NORTH EAST
- GN: zones GAB, 10, 20, 30, 40, 50, 60 , NORTH EAST

As a result of feedback received on the interim report of this study, it is noted that the GHAT has a different historical spatial zonation pattern to the SET fishery. Thus the result obtained above for GHAT zoned via the SET zones is taken as a general finding that "gear - spatial" zonation is important and necessary for this fishery. For the GHAT we have thus modified our stratification recommendations to be consistent with the historic pattern of zonation, i.e. the following zones:

1. CSA
2. EBS
3. ESA
4. ET
5. NSW
6. SAV
7. WA
8. WBS
9. WSA
10. WT

In addition to that, the following considerations led to a further simplification of the strata for the purpose of optimal survey design calculations:

1. DL appears to be a very small fishery and has thus been subsumed into AL.
2. BL is subsumed under the heading AL, since it cannot be distinguished from AL in the ISMP data for the GHAT.

The following AL strata are concatenated for the purpose of optimal sample size calculations:

1. WA and WSA - reasoning - there is very little fishing in the GHAT taking place in WA, and there is no ISMP data for the GHAT in WA, so there is no basis to retain it as a separate stratum.
2. WT and ET - the reasoning is identical as for WA and WSA: there are no ISMP data for WT, and there is effectively no fishing in the GHAT taking place in WT.
3. SAV and WBS - very little fishing taking place in GHAT SAV, no ISMP data, hence merge SAV with an adjacent zone where there are ISMP data and there is fishing.
4. NSW and EBS - reasoning - the data suggest more shots sampled in the ISMP data than there are shots in GENLOG, a logical impossibility. Out of concern for basing strata conclusions on incorrect data we decided to merge NSW and EBS for the GHAT.

Note that there are no ISMP data in the 2002-2006 ISMP data used for this study to consider any particular stratification of the Victoria Inshore Trawl fishery, and given the relatively small size of this fishery, a single stratum is recommended.

There are also no ISMP data to use as the basis for stratification of the ECDW fishery, and therefore for this fishery a single stratum is recommended, as is also supported by the use of a single gear type in this fishery.

Furthermore, the strata for the CTF are well established and we saw no reason to propose restratification for this fishery.

Finally, there were no ISMP data available or 'sourceable' for the gillnet shark fishery within the limitations and timeframe and scope of this study. We understand that these were recorded as trawl fishing in the ISMP database (during 2007-2008). For the purpose of sample size calculations for optimal designs we therefore omit gillnet, but then post allocate sample size to that stratum on the basis of the number of shots in the GENLOG database.

The final recommendations for stratification of the SESSF ISMP for the determination of optimal survey designs are therefore the following 24 strata (omitting VIT since there are no ISMP data for VIT):

1. GAB: Inshore, Midshore, Offshore: 3 Strata.
2. GHAT: AL: WA+WSA, WT+ET, SAV+WBS, NSW+EBS, CSA, ESA: 6 Strata.
3. SET: ECDW_TR: 1 Stratum.
4. SET: BS_IN_TR, EDL_DS, EDL_IN_TR, EDL_OFF_TR, NSW_IN_TR, NSW_OFF_TR, NSW_RRP_TR, SW_BGS_TR, SW_ORO_TR, SW_TR, TAS_BGS_TR, TAS_E_TR, TAS_ORO_TR, TAS_W_TR: 14 Strata.

This does not suggest ignoring gillnet, but rather gillnet strata (and VIT, and the sub-gears of AL) will be reintroduced later, with sample sizes allocated pro rata to the number of GENLOG shots for the recent history of the fishery (2007 and 2008 are used).

### 5.6 Methods of estimation and optimal design for discarded catches

This section contains a complete treatment of the optimal design calculations based on the species or species group level for the retained and discarded catch amounts.

### 5.6.1 Principles underlying the methods for discard estimation.

Methods employed previously in estimating discard levels in the SESSF involved analysis of the discard proportion for each species at a stratum by stratum level, and then aggregation across strata to obtain an estimate, with associated CV, of a fishery-wide discarded proportion for each species under study. The use of proportions in such analyses brings with it a good deal of statistical complexity. In the aggregation across strata for example, the appropriate weighting on the stratum level proportions is the total catch (i.e. retained and discarded combined). However this total catch is itself subject to estimation variance (by virtue of its inclusion of the discards) and this then gives rise to SESSF-wide variance considerations which are far from straight-forward. Previous ISMP designs have avoided these complexities by using the landed catch as the weighting on strata instead of the total catch. This has been viewed as a reasonable approximation in those studies, since most of the species under study had low discard rates, and thus the retained catch ratios between strata was closely related to the corresponding ratios of total catch.
For this study, the scope of the ISMP has been enlarged to encompass a larger number of species and species groups. Many of these are discarded in high proportions. We have therefore been forced to depart from the previous approach based on the total catch, and thus we have used a method which involves estimating the discarded catch mass in each stratum.
The fishery-wide mean and variance on total discards can then be obtained by simple addition of the stratum means and variances. The price to be paid for the simplicity gained in the aggregation across strata, is the requirement of more complexity in the calculation of the stratum-level statistics (i.e. in the estimation of the discarded catch per species and stratum).
We have employed two methods to calculate these estimates from the ISMP data, each of which has merits and demerits in different circumstances:

- Method A is to estimate the discarded catch per shot in the ISMP and multiply this up by the number of shots recorded in the GENLOG.
- Method B is to estimate the discarded to retained catch ratio from the ISMP, and apply this to the total retained catch per stratum as recorded in the GENLOG and corrected by the CDR.
Method A is the simpler of the two methods and may be used more generally.
Method B on the other hand provides improvement in precision under conditions in which catch variability is high compared with the variability of the discard proportion. This requires a certain amount of complexity in analysis for 2 reasons:
i) If the discard rate $d$ is $100 \%$ for any particular shot, the discarded to retained catch ratio is not defined.
ii) Even if the discard is $<100 \%$ for all shots, the mean of the shot level
discard:retained catch ratio $E\left(\frac{d}{1-d}\right)$ is a biased estimator of the stratum level discarded:retained catch ratio. This is evident from a simple example of a stratum comprising 2 shots. One shot retains 1 kg discarding 99 kg , the $2^{\text {nd }}$ shot retains 99 kg and discards 1 kg . The stratum level discard: retained ratio is $100: 100=1: 1$ whereas the mean of the shot level discard: retained ratios is $(99: 1+1: 99) / 2=49.5: 1$.

We adopt instead $\frac{E(d)}{1-E(d)}$ as the unbiased estimator, the variance implications of which are discussed below.

### 5.6.2 Distributional assumptions for the discard proportions.

Many species strata have discard rates which are dichotomous at the shot level (i.e. either $0 \%$ or $100 \%$ of the catch of that species for each shot is retained) and thus have variances which are determined precisely by the mean discard rate d. i.e. $\mathrm{V}(\mathrm{d})=\mathrm{d}(1-\mathrm{d})$.
If all species were dichotomous at the shot level and if discard rates for shots are uncorrelated within trip, then the true mean discard rate for the stratum may be assumed to be betadistributed, $\mathrm{d} /(1-\mathrm{d})$ is beta prime distributed with the same parameters. The variance of the discarded catch for the species-stratum can be theoretically determined using only the mean discard rate, number of shots sampled and total retained catch.

However not all species-strata satisfy these two conditions and so provision needs to be made for:
(a) Non-dichotomy at the shot level, which will reduce variance
(b) Within-trip correlation of shot-level discard rates, which will increase the variance of estimated mean discard rate (this is because it is impractical for observers to sample shots individually, but rather they will select trips and sample all shots from these trips).
Both of these features can in principle be addressed in the same way, by adjustments on the sample size. i.e. we view each stratum species as comprising packets of fish which are uncorrelated in their discard rates and each of which is dichotomous. In case (a) above, such packets will be smaller than the shot level, and in case (b) larger than the shot level.
This leads to the concept of an effective sample size, which may be more or less than the actual number of shots sampled depending on the relative extent of (a) or (b) and may be calculated empirically from the data using the standard deviation of the proportion. The discard proportion itself and the effective sample size then provide a theoretical basis for determining the variances of the posterior distributions for d (assumed to be beta-distributed) and $\mathrm{d} /(1-\mathrm{d})$ assumed to be beta-prime distributed.

An alternative to this theoretical approach which we have explored extensively is to bootstrap the ISMP data to obtain numerical estimates of the variances associated with both the discarded mass and the discarded proportions at a stratum level for a range of candidate sample sizes.

In favour of the bootstrap is its ability to capture the variance implications of within-trip correlations of both the discarded mass and discarded proportions by shot.

We have opted for the theoretical approach for two principle reasons:

- The number of quantities to be estimated (species*stratum combinations) is large compared with the quantity of data available and are thus subject to substantial noise which would in any case warrant some smoothing by means of theoretical priors.
- A simple bootstrap in effect assumes the sample distribution as the underlying reality. If the discard proportions are, as we assume, beta distributed, then the discard proportions realized from bootstrap re-samples are binomially distributed and thus do not fully capture the uncertainty in the underlying population mean itself.
The within-trip correlation is then handled by means of a decomposition of variance to between and within trip components, which enables the calculation of appropriate multipliers of the variances for each species and stratum (see Equations D8 to D12 below.)


### 5.6.3 Algebraic formulation of the estimation methods.

## Stratum level estimate, Method A:

From the ISMP data, we calculate the mean $\bar{D}_{H s t}^{s p}$ and variance $V_{H}\left(D_{s t}^{s p}\right)$ of the discarded catch in kg over shots that encounter that species.

$$
\begin{align*}
& \bar{D}_{H s t}^{s p}=\frac{1}{n_{s r}^{o} \cdot h_{s t}^{s p}} \sum_{i \in s t} D_{i}^{s p}  \tag{4.6.3-D1}\\
& V_{H}\left(D_{s t}^{s p}\right)=\frac{1}{n_{s r}^{O} \cdot h_{s t}^{s p}} \sum_{i \in s t}\left(D_{i}^{s p}-\bar{D}_{H s t}^{s p}\right)^{2} \tag{4.6.3-D2}
\end{align*}
$$

where $D_{i}^{s p}$ is the discarded catch in kg for species sp as recorded by the observer for shot i.
$h_{s t}^{s p}$ is the ISMP "hit rate" i.e. the proportion of shots observed in stratum $s t$ which encounter species $s p$.
$n_{s t}^{O}$ is the total number of shots observed in the stratum (as per the ISMP data.)
The mean and variance over all shots fired require an adjustment taking account of the hit rate
$\bar{D}_{A s t}^{s p}=h_{s t}^{s p} \cdot \bar{D}_{H s t}^{s p}$
$V_{A}\left(D_{s t}^{s p}\right)=h_{s t}^{s p} \cdot V_{H}\left(D_{s t}^{s p}\right)+\left\lfloor h_{s t}^{s p} .-\left(h_{. s t}^{s p}\right)^{2}\right]\left(D_{H s t}^{s p}\right)^{2}$
The variance of the mean adjusted discarded catch per shot fired is then
$V\left(\bar{D}_{A s t}^{s p}\right)=\frac{V\left(D_{A s t}^{s p}\right) \cdot \omega_{s t}^{s p}}{n^{o}{ }_{s t}}$
where $\omega_{s t}^{s p}$ is the variance multiplier due to within trip correlation.

The estimate of total discarded catch for species $s p$ and stratum $s t$ is then
$\hat{D}_{s t}^{s p}=n_{s t}^{F} \bar{D}_{A s t}^{s p}$
With variance:

$$
\begin{equation*}
V\left(\hat{D}_{s t}^{s p}\right)=\left(n_{s t}^{F}\right)^{2} V\left(\bar{D}_{A s t}^{s p}\right) \tag{4.6.3-D7}
\end{equation*}
$$

Where $n_{s t}^{F}$ is the total number of commercial shots fired in the stratum (as per the GENLOG data.)

## Accounting for non-independence of errors within trip.

Equation D5 above, if $\omega_{s t}^{s p}=1$, assumes independence of deviations in discarded catch between shots, whereas for some species in some strata there is in fact some measure of correlation of these errors within trip.
To this end, we model the discard catch for shot $i$ on trip $t$ in year $y$ as follows. For the sake of simplicity of notation we omit the subscripts $s p$ (species) and $s t$ (stratum) which are to be understood throughout.
$D_{t, i}=\mu_{y}+\xi_{t}+\varepsilon_{i}$
$\mu_{y_{i}} \quad$ is a (fixed) year-dependent intercept
$\xi_{t} \quad$ is a random batch effect for trip t with variance $\phi^{2}$
$\varepsilon_{i} \quad$ is a random error for shot i with variance $\sigma^{2}$
If N shots are sampled from T trips, then the expected variance of the mean discarded catch $\bar{D}$
assuming independence of the errors (i.e without accounting for trip) is:
$V_{1}(\bar{D})=\frac{V\left(D_{i}\right)}{n}=\frac{\phi^{2}+\sigma^{2}}{n}$
On the other hand, assuming the model (D8*), the variance of the mean is
$V_{2}(\bar{D})=\frac{\phi^{2}}{T}+\frac{\sigma^{2}}{n}$
The variance multiplier for use in equation D5 is then
$\omega=\frac{V_{2}(\bar{D})}{V_{1}(\bar{D})}=\frac{m r+1}{r+1}$
Where $m=\frac{N}{T}$
is the average number of shots per trip (which encounter the species in question)
And $r=\frac{\phi^{2}}{\sigma^{2}}$
is the ratio of between trip variance to within trip variance.
We have obtained the quantities $m$ for each species and stratum using ISMP data from years 2002 to 2006. (2007 and 2008 data could not be used for this exercise because trip could not always be identified.) The quantities $r$ are obtained using the same data, by a MINQUE decomposition of variance using SPSS17 software.

We have subsequently calculated the variance multipliers $\omega_{s t}^{s p}$ for each species s and stratum (which we assume to be year-invariant.) Note that for some species and strata $\mathrm{m}<1$ because not all trips encounter the species. In this case we apply the constraint $\omega_{s t}^{s p}>=1$.
The multipliers $\omega_{s t}^{s p}$ can be provided in a spreadsheet for reference and for application to future estimation of discards using this method.

## Stratum level estimate, Method B:

For each shot $i$ we define the discard rate as
$d_{i}^{s p}=\frac{D_{i}^{s p}}{D_{i}^{s p}+R_{i}^{s p}}$
For each species $s p$ and stratum $s t$, we calculate the mean and variance of the discard rate $d_{i}^{s p}$ for shots $i$ in the stratum which encounter the species.

$$
\begin{align*}
& \bar{d}_{s t}^{s p}=\frac{1}{n_{s t}^{O} \cdot h_{s t}^{s p}} \sum_{i \in s t} d_{i}^{s p}  \tag{4.6.3-D14}\\
& \mathrm{~V}\left(d_{i}^{s p}\right)=\frac{1}{n_{s t}^{O} \cdot h_{s t}^{s p}} \sum_{i \in s t}\left(d_{i}^{s p}-\bar{d}_{s t}^{s p}\right)^{2} \tag{4.6.3-D15}
\end{align*}
$$

The effective sample size is calculated as:
$n^{e f f}{ }_{s t}^{s p}=\frac{d_{i}^{s p}\left(1-d_{i}^{s p}\right) n_{s t}^{o} \cdot h_{s t}^{s p}}{V\left(d_{i}^{s p}\right)}$
We calculate the theoretical parameters of the beta distribution corresponding to the mean and standard error on $\bar{d}_{s t}^{s p}$
$\alpha=n^{\text {eff }}{ }_{s t} \cdot \bar{d}_{s t}^{s p}+1$
$\beta=n_{s t}^{\text {eff }}$ sp $.\left(1-\bar{d}_{s t}^{s p}\right)+1$
Use the beta parameters to calculate the theoretical variance of $\left(\frac{\bar{d}_{s t}^{s p}}{1-\bar{d}_{s t}^{s p}}\right)$ on the assumption that this is beta-prime distributed.
$. V\left(\frac{\bar{d}_{s t}^{s p}}{1-\bar{d}_{s t}^{s p}}\right)=\frac{\alpha(\alpha+\beta-1))}{(B-2)\left(\beta-1^{2}\right)}$
\{Note that If $\beta<2$, i.e. if $n^{\text {eff }} s t\left(1-\bar{d}_{s t}^{s p}\right)<=1$ the variance is infinite. In this case, Method A is forced.\}
The estimate of total discarded catch for species $s p$ and stratum $s t$ is then

$$
\begin{equation*}
\hat{D}_{s t}^{s p}=R_{s t}^{s p} \cdot \frac{\bar{d}_{s t}^{s p}}{1-\bar{d}_{s t}^{s p}} \tag{4.6.3-D20}
\end{equation*}
$$

With variance:

$$
\begin{equation*}
V\left(\hat{D}_{s t}^{s p}\right)=\left(R_{s t}^{s P}\right)^{2} V\left(\frac{\bar{d}_{s t}^{s p}}{1-\bar{d}_{s t}^{s p}}\right) \tag{4.6.3-D21}
\end{equation*}
$$

Where $\hat{R}_{s t}^{s p}$ is the total retained catch for species $s p$ and stratum $s t$ as per the GENLOG/CDR data.

## Aggregation across strata.

The SESSF-wide estimate of total discarded catch for species sp is
$\vec{D}^{s p}=\sum_{s t} \hat{D}_{s t}^{s p}$
With variance:
$V\left(\vec{D}^{s p}\right)=\sum_{s t} V\left(\hat{D}_{s t}^{s p}\right)$
The CV is given as

$$
\begin{equation*}
C . V .=\left(\sum_{s t} V\left(\hat{D}_{s t}^{s p}\right)\right)^{1 / 2} / \sum_{s t} \hat{D}_{s t}^{s p} \tag{4.6.3-D24}
\end{equation*}
$$

### 5.6.4 Conditions of selection of Method $\boldsymbol{A}$ vs Method B and implications for optimal sample size allocation to strata.

If we let
$\gamma=n_{s t}^{F} \frac{\sqrt{V_{A}\left(D_{A s t}^{s p}\right)}}{D_{A s t}^{s p}}$
Then from equations (D5) and (D7) above we obtain
$V\left(\hat{D}_{s t}^{s p}\right)=\frac{\left(\gamma_{s t}^{s p}\right) \cdot\left(D_{A s t}^{s p}\right)^{2}}{n_{s t}^{O}}$
The CV on the estimate for the stratum is then
$c v\left(\hat{D}_{s t}^{s p}\right)=\frac{\gamma_{s t}^{s p}}{\sqrt{n_{s t}^{o}}}$
which illustrates the inverse dependence of the variance of the stratum estimate on the sample size allocated to the stratum using Method A.

Note that this simple relationship does not apply with Method B, where the CV is
$\sqrt{\frac{\left(n_{e f f} d_{s t}^{s p}+1\right)\left(n_{e f f}+1\right)}{\left(n_{e f f}\left(1-d_{s t}^{s p}\right)-1\right) n_{e f f}{ }^{2} d_{s t}^{s p^{2}}}}$
Table 9 shows the theoretical CVs on the estimate using methods A (CV1) and B(CV2) for a species with discard proportion of $50 \%$ and gamma $=3$. Method B is more precise for sample sizes greater than 10 , but becomes considerably less precise very quickly as the sample size decreases towards zero. While this is a particular example, the features shown are typical for proportions between 0.1 and 0.9

For proportions outside this range, Method A is almost invariably more precise, even at large sample sizes.

In principle, for estimation purposes, the data should be allowed to make their own decision where possible, i.e. calculate the estimate of discards CV by both methods and accept the estimate with the lower CV.

If the discard proportion is very close to 1 , the variance by Method $B$ is undefined, so Method A has to be used.

In our application of the Methods A and B to the ISMP data from 2002 to 2006 (Interim Report), only $1 \%$ of species-stratum combinations preferred Method B, and in those cases the degree of preference was slight. Thus, for the purposes of sample size design we have adopted Method A throughout in the interests of simplicity (see Figure 2).
5.6.5 Results of calculations of mean and CV for discarded catch 2002-2008. Although ISMP data for years 2002-2006 were collected under the old design, we present here results of the calculations under the new stratification.

Table 1 shows, for each of the 32 strata, the number of shots fired (from the GENLOG data) and the number of shots observed (from the ISMP data.)
Note that no ISMP observations are available for the VIT or GN strata, which collectively make up $18 \%$ of all shots ( $25 \%$ of sea-days) recorded in the GENLOG.

Table 3 shows the estimated discarded catch and associated CV for the 27 strata for which data are available.

### 5.6.6 Interpolation of missing values.

Not all of the input quantities for equations D1 to D12 above are obtainable directly from the data at our disposal. Thus for purposes of optimization of sample size allocation to strata, it has been necessary to interpolate values in certain species*strata combinations.

In the Gillnet (GN) and Victoria Inshore Trawl (VIT) strata, we have no ISMP data for any of the project keys. No interpolation is attempted for these strata, but in the optimization procedures described below, these strata are allocated sea-days on a pro rata basis, in proportion to sea-days allocated to the rest of the fishery.
In addition, there are species-stratum combinations for which either GENLOG or ISMP data are not available. These are dealt with as follows:

- If a species-stratum does not appear in the ISMP or the GENLOG, we assume that the species is never caught in that stratum, thus the discards are zero with zero variance.
- If a species-stratum appears in the ISMP, but not the GENLOG, we assume that the discard rate is $100 \%$ regardless of the discard rate recorded by the ISMP observers.
- If the species-stratum appears in GENLOG but not ISMP we interpolate as follows:
- Assume that the discarded: retained catch ratio is the same as the average for the species over strata for which data are available.
- Multiply this assumed ratio by the retained catch as recorded in the GENLOG to infer a mean discarded catch.
- Assume that the standard deviation: mean ratio for discarded catch is the mean of this ratio for the species over strata for which data are available.
- Multiply this assumed ratio by the inferred mean discarded catch to infer a standard deviation on the discarded catch.
- Assume that the ISMP hit rate is the same as the GENLOG hit-rate.


### 5.6.7 Optimisation of ISMP design.

In previous designs, the target precision on discarding has been set in terms of a CV on the discarded proportion. This target level differed depending on the discard proportion itself.

| Proportion | Target CV |
| :--- | :---: |
| $>20 \%$ | $40 \%$ |
| $5-20 \%$ | $80 \%$ |
| $<5 \%$ | $150 \%$ |

These target CVs were in principle CVs on the discarded proportion. However, due to the nature of the calculations of these quantities (retained catch rather than total catch weighting in the aggregation of strata) these targets were effectively CVs on the discarded mass. The rationale behind the gradation of these targets was that less precision is required when the discarded proportion is low, because in this case the uncertainty is not large compared with the total catch. This feature is illustrated in Figure 3, in which the target CVs on discards are shown together with the implied CV on the total catch.
The effect of the gradation is to confine the CV on total catch to between 0 and $40 \%$ at all proportions. See Figure 3.

In the new design, we proposed to restate the targets instead at a flat $20 \% \mathrm{CV}$ on the total catch. Following feedback from AFMA after the interim report this has been revised to the following.

The target level of precision for each project key is:

- $20 \% \mathrm{CV}$ on the total catch, if the discarded proportion $>13.33 \%$
- $150 \% \mathrm{CV}$ on the discarded catch if the discarded proportion $\leq 13.33 \%$.

This is shown in Figure 4.
Given the above target specification, for a single species, the objective underlying optimization of the sample size distribution to strata is relatively simple - to allocate strata so as to minimize the CV for that species. However, with 75 species groups (Project Keys) involved, a number of different objectives are possible, in view of the potential for competing requirements for different species.
Following feedback from AFMA on the interim report of this study, we have settled on 3 optimization methods meeting 3 different objectives.


#### Abstract

Allocation Method VWCV: For any given number of sampling days available, minimize the value-weighted average CV over the 68 commercial project keys. For this purpose, the value weighting we have used for each of commercial species is the estimated discarded catch in kg X the approximate price per kg. Average prices (from Sydney Fish Market) for 25 project keys are shown in Table 2. These prices range from A\$ 1.75 to $\mathrm{A} \$ 9.83$ per kg. All other species are effectively catch-weighted with respect to each other. For the purposes of marrying the two weighting regimes, we impose a "dummy" price of $\$ A 1.00 \mathrm{per} \mathrm{kg}$ for all the remaining project keys.


#### Abstract

Allocation Method MinS: For any given number of sampling days available, minimize the number of species which have CV above the target threshold of $20 \%$. Note that changing this threshold may alter the resulting order of allocation, but we have not investigated that here.

The mechanism of this method is, at each increment - determine from among those species with CVs greater than the target CV - which species has the CV is closest to target. Add a sea-day in the stratum which most reduces the CV of that species.

Allocation Method PROP: This is simply an allocation of any given availability of sampling days in proportion to the number of sea-days expended by the fishery in each stratum.


Allocation Method CWCV-HR: For any given number of sampling days available, minimize the catch-weighted average CV over the 7 High Risk species groupings.
Allocation Method MinS-HR: For any given number of sampling days available, minimize the number of high-risk species groupings which have CV above the target threshold of $20 \%$. If all high risk project keys are below target, then minimize the number of commercial project keys above target
The mechanism for this method is, at each increment - determined from among those species with CVs greater than the target CV, which species has the CV is closest to target. Add a seaday in the stratum which most reduces the CV of that species.
Note that none of the above procedures above pre-suppose any particular constraint on total sampling effort available, but each presents a continuum of benefit vs sampling cost from which a sampling level may be selected.

Though we recognize that observers will select trips rather than shots, we note also that the cost of observation is dependent on trip length. We thus use the sea-day as the basic sampling unit. All sample sizes referred to below are expressed as numbers of sea-days per annum, though this is also translated into number of shots where appropriate. Implicit is the assumption that the observer will record all shots for each sea-day spent on board a vessel.
Each of the optimization procedures start with a single (seed) sea-day allocated to each stratum and all add single sea-days to the survey incrementally in 1 of 3 ways, depending on which of the above objectives is given priority:

- Method VWCV. At each increment - add a sea-day in the stratum which most reduces the value-weighted average CV across all species.
- Method MinS. At each increment - determine from among those species with CVs greater than the target CV, which species has the CV closest to target. Add a shot in the stratum which most reduces the CV of that species.
- Method PROP. At each increment - add a sea-day in the stratum which has the least proportional coverage by the ISMP as at the previous increment.
- Method CWCV-HR. At each increment - add a sea-day in the stratum which most reduces the catch-weighted average CV across the HIGH RISK species groups.
- Method MinS-HR. At each increment - determine from among those HIGH RISK species groups with CVs greater than the target CV, which species group has the CV closest to target. Add a shot in the stratum which most reduces the CV of that species.

Method VWCV, Method MinS and Method PROP are all used to optimize sampling for commercial species (commercial allocation), while Method CWCV-HR and Method MinSHR are used to optimize sampling for high risk species (high risk allocation).

For each of the methods, we have continued the process up to a total sample size of 2000 observed sea-days per annum. In all cases the expected CV for a given stratum size is calculated using estimation Method A.

Figure 5 and Figure 6 illustrate the progress of the 5 allocation methods towards meeting the various objectives as the sample size is incrementally increased.

Figure 7 to Figure 16 show the sample sizes distributions in terms of number of sea-days at 250 sea-day increments from 500 to 1500 days by each of the 5 methods.

These distributions are also presented in Table 4 to Table 8, which show in addition the equivalent number of shots sampled.
Figure 17 to Figure 21 illustrate for each of the 5 methods, for project key, the progress towards target at 250 day increments in the total sample sizes.
Notable in Figure 7 for example, is the disproportionate sampling effort preferred in the EDL_IN_TR and GAB Inshore strata. Both of these strata have a lot of species (EDL has 60 out of 75 , Gab Inshore has 62 out of 75 ). Both also have high average standard deviation to mean discard ratios of more than 1 . Note that the EDL_INS _TR stratum is favoured by the VWCV method, implying that a lot of the high value species reside there. This stratum gets a high proportion of sampling effort right from the word go.

The GAB-Inshore stratum on the other hand is favoured by the MinS method. It also only starts blowing out when the sea-days gets to about 400 - i.e. when we get down to the last 10 to 15 species which are still above target. Note, that for 13 project keys, more than half of the total estimated discards for that project key occur in the GAB Inshore stratum, so that in order to get precision on the estimate of discards for these keys, considerable sampling is required in that area.

In Figure 6 we note that the MinS-HR method achieves its objective of bringing the 7 high risk project keys below target within 500 sea-days. However, at this level of sampling 14 of the commercial project keys are still above target. The MinS (commercial) method on the other hand has only 9 commercial species above target, but also 4 of the 7 high risk project keys. This illustrates one of the trade-offs to be made between the differing objectives of the program. The 4 plots of Figure 5 and Figure 6 represent 4 distinct objective functions ( 2 commercial and 2 high risk). Each plot is underscored by the method which is specifically designed to meet that objective.

### 5.7 Analysis of fishery-TEP interactions

The TEP species are grouped as follows:

1. Seals (includes Sea Lions)
2. Sharks
3. Pipefish
4. Birds
5. Whales/Dolphins
6. Reptiles (no records of Reptilian interactions in the data)

We note that for the 2004-2008 ISMP data, there are no data Reptiles and so these two groups do not appear in any of the analyses that follow.

### 5.7.1 Extraction of the TEP interaction data from multiple tables

We experienced difficulties in reconciling information on TEP species which interact with the fishing gear and/or are captured by the gear itself. As a result, the interaction may appear in the 'TEP Interactions' and/or the 'ISMP catch' table. The additional absence of a unique key
linking a record in the 'TEP interactions' table and the 'ISMP catch' table posed significant data mining and data pre-processing challenges which are addressed here in only one of a number of possible ways.

In order to produce the TEP Interactions analysis, broken down by year and strata, it was necessary to match the records from the TEP Interactions table to records in the Trawl Catch table. In this way, it is possible to identify the strata where the interaction occurred. It should also be noted that since the format of the TEP interaction data received in April 2009 did not match the format of the TEP data received June 2009 - and each dataset exhibited distinct problems - two distinct exercises were required in order to produce a single, stratified table of TEP interactions for the period 2004-2008 inclusive.

### 5.7.2 Linking 2004-2006 TEP interaction data

The most obvious and reliable approach to this problem requires a record matching based on the vessel call sign, the shot date and the shot number. When we first attempted to merge the two datasets using this approach, we discovered two problems which were causing the majority of TEP Interaction records to drop out of the analysis (Table 9):

1. A vessel call sign which exists in the TEP interaction records does not exist in the ISMP Trawl database
2. The vessel call signs and shot dates can be matched, but there is no corresponding shot number in the Trawl table, so a TEP Interaction record cannot be matched to a Trawl.

The first of these problems cannot be remedied - we assume that the "unknown" vessels are capturing errors and drop these records from the analysis as there is no possibility of stratifying these records.

The second problem can be dealt with by using the following approach:

- Select records from the ISMP Trawl table such that a list of distinct combinations of vessel call sign, shot date and strata are created, and match the resulting table to the TEP records which fall into category 2 above.

This produces a stratified version of the TEP Interaction data which can then be appended to the previously matched TEP Interaction records. It should be noted that a potential problem with this approach arises if a particular vessel operated in more than one stratum on the same day - this scenario does affect 37 TEP Interaction records. This complication is handled by selecting the strata of the Trawl catch record with the closest shot number to the ambiguous record. By calculating the absolute difference between the TEP Interaction shot number and the Trawl table shot number it is possible to identify and select the record with the minimum difference - in this way we can determine the most likely strata for the ambiguous interaction records.

A further complication to the analysis of TEP interactions arises from the possibility of the TEP species group appearing in both the Interactions table and the Trawl Catch table improper handling of these records would lead to duplication of interactions. The only species groups for which this could occur are "Seals" and "Pipefish". Identification of these records is accomplished by creating a field which contains a 1 if the record originated from the TEP Interactions table or a zero if the record originated from the Trawl Catch table along with a similar marker to indicate if the record originated from the Trawl Catch table. Shots which contain a duplicate can then be identified by combining the species separated catch and interaction records and aggregating down to a shot level, retaining the maximum value of the two marker fields. At this point, a record with duplicated shots will contain values for both
marker fields which are greater than zero - using this information we can de-duplicate the records.

### 5.7.3 Linking 2006-2008 TEP interaction data

Unlike the pre 2007 data tranche, the 2006-2008 observer data did not contain unique trip identifiers, or shot numbers. This made it impossible to match the TEP interaction data based on shot date and shot number, as only shot date is common to both datasets. The only approach available in this case was to determine the strata that the vessel operated on a specific day, and assign any TEP interactions to those strata. This approach is vulnerable to two potential errors:

1. A TEP interaction record can appear on a date for which no fishing was recorded
2. A vessel have fished in one or more strata on a given day

The first of these problems cannot be solved, therefore all TEP interaction records for a vessel with no catch records on the same day were excluded from the analysis as they could not be stratified.

The second problem was resolved by using the latitude and longitude of the TEP interaction and catch records in order to calculate the distance between a stratified shot thrown on an ambiguous day and the TEP interactions for the same vessel and date. This calculation permitted stratification of the TEP interaction records by assigning it to the strata of the shot which was physically closest to the location of the interaction.

The resulting stratified 2006-2008 TEP interactions can then be appended to the 2002-2004 TEP interactions providing a single table with TEP interaction records for the period 20022008. See Table 9.

A description of the methods used for the analysis of the TEP interactions data follows. Although reference is made to seals, the same methods are used for all TEP species groups

### 5.7.4 Fishery-TEP interaction analyses

In the interim report, an analysis of the number of shots involved in an interaction with the TEP species / species group under consideration was presented. The question of the mean number of individuals interacting with the fishery, for any shot actually involved in an interaction, and the associated variance, was not analysed in any detail, although estimates of the number of interacting individuals were presented. In retrospect it is clear that separating the analysis for the TEP species groups in this way is problematic, hence a different approach has been developed for this final report.

The Poisson distribution provides a simple statistical framework for the joint analysis of the number of shots involved in interactions with TEP species groups, and the number of individuals per interaction. The Poisson distribution is however limited because the mean is equal to the variance. Since different species groups exhibit different degrees of 'clumpiness' in their distributions, a mass distribution function in which the variance can exceed the mean would provide a better vehicle for the determination of optimal ISMP designs. The obvious choice is the negative binomial distribution, although there are others such as the Sichel distribution which might be equally useful, but not as widely known.

The CVs that emerge from such an analysis are likely to be considerably larger for the same sample sizes than those that were reported in the interim report. The reason is that the CVs in the interim report were limited to considerations of the number of shots involved in
interactions with TEP species groups, whereas the ensuing analysis caters for the additional variance of the number of individuals per interaction.

Of course one might ask why not base the optimal design on the empirical means and variances? This option has considerable merit. There is though the difficulty that the empirical means and variances are an inadequate basis for a survey design, based as they are on a limited sample. For example, in many strata there are no records of interactions with particular TEP species groups. Does this mean that no interactions will ever occur in those strata and hence that no sampling should be directed there (under circumstances where the TEP determine the optimal sample design)? Clearly not - some provision should be made for the possibility of TEP interactions even in strata where no interactions were observed.

So a parametric distribution might be useful in that it provides a basis for, for example, smoothing the empirical information about means and variances, and perhaps for the setting of plausible priors for strata where no interactions have been recorded.

Although the negative binomial distribution is indicated as the standard formulation, we suggest that explicit use of this mass distribution function is not necessary and instead we focus on the degree of over-dispersion in the data (i.e. the ratio of the variance to the mean).

Conceptually we suggest that there is a level of clumping (number of individuals) that can be applied to particular TEP species groups and that at that level, the frequency of occurrence of groups will be Poisson distributed. If we let $\mathrm{V}_{\mathrm{c}}$ be the variance of the number of clumps observed, and $\mu_{\mathrm{c}}$ the mean number of clumps, for the Poisson distribution:
$\mathrm{V}_{\mathrm{c}}=\mu_{\mathrm{c}}$
Roughly then, the mean number of individuals ( f ) is $\mu_{\mathrm{f}}=\mathrm{S} \mu_{\mathrm{c}}$, and the variance of the number of individuals is $V_{f}=S^{2} V_{c}$. From this it follows that
$\mathrm{V}_{\mathrm{f}}=\mathrm{S} \mu_{\mathrm{f}}$
Thus the slope of a linear relationship (passing through the origin) between the variance and the mean of the number of individuals interacting with fishing gear is the average clump size. It is suggested that this average clump size is an intrinsic property for the TEP species group, and can form the basis for smoothing the data and perhaps for the introduction of priors for strata which have not shown any interactions with TEP species groups in the past.

Figure 36 - Figure 42 show plots of the stratum/year level means and variances at the individual level, for different species groups.

Two plots are shown for 'Seals' (Figure 36 and Figure 37) and 'Birds’ (Figure 39 and Figure 40 ), because the fit seems to have been distorted by a very high point in each case, and the second plot for each of these is carried out omitting this 'outlier'. The average clump sizes which are thus obtained are:

- Birds: 37.50
- Pipefish: 1.74
- Seals: 8.40
- Sharks: 1.50
- Whales/Dolphins: 292.90

These results point to Birds, Seals and Whales/Dolphins operating in medium to large groups, while Pipefish and Sharks would seem to be operating more as independent individuals and/or perhaps in pairs, at least in respect to how they interact with the fishing gear.

In order to carry out the ISMP design optimisation exercise it is necessary to use a measure of stratum size. We know that the gillnet gear type is not distinguished from trawl in the ISMP data. However, the two gear types are distinguishable in the GENLOG data. Thus, for the purpose of calculating the stratum sizes as number of shots per stratum, the SET shots in the GENLOG data have been increased by about $38 \%$, where $38 \%$ is the average number of gillnet shots in the SESSF as a whole, divided by the gillnet + trawl shots. Any sample size recommendations that emerge from the ISMP design optimisation therefore need to be disaggregated into gillnet and trawl at the end of the exercise.

A further consideration is that not all interactions with TEP species take place during a specific fishing event. To be more specific, it was not possible to link a record in the TEP interactions table with the catch table of the ISMP data. This has implications for the concept of what constitutes the population and the sample from a statistical standpoint, a topic which is outside the scope of this study. For this report we have simply treated a TEP interaction record as a 'shot' for the purpose of the analysis.

For the purpose of the determination of optimal designs driven solely by the TEP species groups and their interactions with fishing gear, the following steps were followed:

## Optimal ISMP Design Analysis A (no priors)

1. The analysis takes place at the stratum level for the SESSF.
2. Calculate the mean number of interacting TEP individuals per stratum over years 2002-2008.
3. Use the clump sizes to convert the mean number of interacting TEP individuals observed for the stratum to variances.
4. The variance of the mean is the variance calculated above, divided by the ISMP sample size for that stratum.
5. Calculate the stratum size as the number of GENLOG shots for that stratum, as an average between 2007 and 2008.
6. The estimated total number of interacting TEP individuals for a particular species group equals the sum product of the means and the stratum sizes.
7. The variance of the 'estimated total number of interacting TEP individuals for a particular species group' equals the sum product of the 'variance of means' and the 'stratum sizes squared'.
8. Choose sample sizes for each stratum subject to a constraint that they sum to $\mathrm{N}_{\text {tot }}$ and minimise the variance calculated above. By sample size is meant the number of seadays. The number of shots is based on the average number of shots per vessel per day per stratum from the GENLOG data (see Table 21).

This approach leads to the distribution of sample sizes shown in Table 22 and Table 23. The results are illustrative for $\mathrm{N}_{\text {tot }}=400$ sea-days, and fractional sea-days and shots have been permitted. The CVs for each TEP species group are shown in the output tables and figures as well. Note that the CVs are inversely proportional to the square root of sample size, provided that the distribution of shots per stratum does not change, just the overall scale of shots. Thus the sample size (say $\mathrm{N}_{20 \%}$ ) required to achieve a CV of, say $20 \%$, is:
$\mathrm{N}_{20 \%}=1000 \mathrm{x}\left[\mathrm{CV}\left(\mathrm{N}_{\text {tot }} 1000\right)\right]^{2} / 20^{2}$

Table 23 is a modification where additional underlying 'true mean interactions' were introduced for strata where there were no interactions for the period 2002-2008. These values were:

- Birds, Seals: $10 \%$ of the mean number of interactions per shot in strata where interactions did occur, calculated separately for the AL, SET or GAB sectors of the SESSF,
- Pipefish, Sharks and Whales / Dolphins: 10\% of the mean number of interactions per shot in strata where interactions did occur, regardless of SESSF stratum.

Table 22 and Table 23 also provide the average optimal shot allocation across TEP species groups subject to the constraint that the total sea-days $=400$.

Note as well that included in Table 22 and Table 23 are the results that are obtained if instead of using the optimization scheme to determine the shot and sea-day allocations, one were to instead base the shot allocation on a formulation in which shots are simply allocated pro-rata to the shot distribution evident in the fishery. The CV consequences of such an allocation scheme are shown.

The results in Table 22 and Table 23 are for a subset of the final number of strata that is envisaged in the final implementation of the ISMP for the SESSF. Whereas the following strata are readily identifiable in the CDR-GENLOG data:

| AL CSA | EDL_OFF_TR | NSW_OFF_TR |
| :--- | :--- | :--- |
| AL EBS / NSW | GN CSA | NSW_RRP_TR |
| AL ESA | GN EBS / NSW | Offshore |
| AL ET / WT | GN ESA | SW_BGS_TR |
| AL WBS / SAV | GN ET / WT | SW_ORO_TR |
| AL WSA /WA | GN SAV | SW_TR |
| BS_IN_TR | GN WA / WSA GN WBS | TAS_BGS_TR |
| ECDW_TR | Inshore | TAS_E_TR |
| EDL_DS | Midshore | TAS_ORO_TR |
| EDL_IN_TR | NSW_IN_TR | TAS_W_TR |
|  |  | VIT |

The ISMP data are limited in the richness of the AL gear types that are identifiable. In addition gillnetting was not recorded as a separate gear type but was subsumed under trawling during 2007 and 2008 (SET). The strata that are identifiable in the historic ISMP data are as follows:

| AL CSA | EDL_DS | Offshore |
| :--- | :--- | :--- |
| AL EBS / NSW | EDL_IN_TR | SW_BGS_TR |
| AL ESA | EDL_OFF_TR | SW_ORO_TR |
| AL ET / WT | Inshore | SW_TR |
| AL WBS / SAV | Midshore | TAS_BGS_TR |
| AL WSA /WA | NSW_IN_TR | TAS_E_TR |
| BS_IN_TR | NSW_OFF_TR | TAS_ORO_TR |
| ECDW_TR | NSW_RRP_TR | TAS_W_TR |

In order to deal with the absence of gillnetting and VIT in the historic ISMP data, the stratum sizes for the SET strata were increased by $38 \%$ (based on information in the CDR-GENLOG data) prior to optimisation of designs using the more restricted set of strata listed immediately above. It is now necessary to disaggregate the 'design' allocation recommendations to the probable set of strata that are feasible in the actual implementation of the ISMP. The following initial considerations are required for this step:

1. Reduce the ISMP allocations to SET strata by $1 / 1.38$, and allocate these shots to gillnetting strata based pro-rata on the gillnet shots recorded for the SESSF in the CDR-GENLOG data.
2. Make an allocation to VIT based pro-rata on the VIT shots recorded in the SESSF in the CDR-GENLOG data.

During 2007 and 2008, the proportion of gillnet shots in each of the 7 gillnet strata considered here was as follows:

| GN CSA | 0.270 |
| :--- | ---: |
| GN EBS |  |
| NSW | 0.326 |
| GN ESA | 0.032 |
| GN ET / WT | 0.075 |
| GN SAV | 0.029 |
| GN WA / WSA | 0.189 |
| GN WBS | 0.080 |

In addition, during 2007 and 2008, VIT shots comprised $1.061 \%$ of all GENLOG shots in all other strata.

Table 24 and Table 25 report the ISMP stratum allocation information for TEP species groups adjusted for gillnet and VIT as described above. Figure 43 and Figure 44 provides plots of sample size in terms of shots, comparing the average optimal allocation across TEP species groups, with shots allocated pro rata to the actual CDR-GENLOG shots per strata for 2007 and 2008 combined.

We note at this point that in addition to the recommendation that gillnetting and VIT be represented as separate strata in the ISMP planning and implementation, this study will recommend, consistent with the findings from the initial cluster analyses presented much earlier in this document, that the shots allocated to the 'AL' fishery category reflected in Table 22, Table 23, Table 24 and Table 25 be disaggregated on a pro rata basis into the gear types AL+LL (automatic longlining and long lining), BL and DL, based on the expected breakdown of shots by these gear types per 'AL' stratum.

### 5.7.5 Description of analytical techniques used for annual estimates

The methods used for estimating the number of individuals interacting with SESSF fishing gear, 2002-2008, follow directly from the methods described above. One of the main difficulties which are encountered with this approach is that in certain years, for certain strata, no shots were sampled in the SESSF ISMP. There are thus no estimates of the number of species interacting with the SESSF in certain years and strata and thus for those years the overall estimate will exclude the unsampled strata. Furthermore, in some years, there were no observations of interactions with certain TEP species groups. In these cases, for the years in question, no estimate is given for the CV. The estimates are given in Table 26 and Table 27,
where Table 26 includes all individuals observed, whether dead or alive, while Table 27 if for dead individuals only.

### 5.8 Mapping shots to ports

This section is presented as an aid to planning the implementation of the ISMP. The main results presented here are clusters of ports which have a similar distribution of strata visited, presented separately for the AL, GN, SET, GAB and VIT fisheries.

The purpose of this section is to provide assistance with the selection of ports in order to achieve a predefined number of shots per stratum. In order to do so, the percentage of shots at each port falling into each of the 32 strata described above was calculated. Strata were then divided into the 'gear types' AL, GN, GAB, SET and VIT as indicated. Cluster analyses were carried out to cluster ports, for each of the groups AL, GN, GAB, SET and VIT using the percentages referred to (a K-means clustering algorithm was used). The number of clusters was set equal to the number of strata per group, i.e:

AL-6
GN-7
GAB-3
SET-15
VIT - 1
The following tables are presented:

- Table 28. A total of 108 ports appear in the data emerging from the CDR/GENLOG merge. A listing of the port names appearing in the CDR/GENLOG merge.
- Table 29. A listing of the 32 strata that arise if one uses the 25 strata from the survey design optimisation exercise, and the 7 gillnet strata shown. (Note that there are a total of 32 strata under the following conditions: (a) the 25 strata used in the survey design optimisation calculations, and (b) the 7 gillnet strata).
- Table 30. AL (6 clusters) - Cluster Membership
- Table 31. AL (6 clusters) - Mean Cluster Percentages
- Table 32. GN (7 clusters) - Cluster Membership
- Table 33. GN (7 clusters) - Mean Cluster Percentages
- Table 34. GAB (3 clusters) - Cluster Membership
- Table 35. GAB (3 clusters) - Mean Cluster Percentages
- Table 36 and Table 37. SET (15 clusters) - Cluster Membership
- Table 38. SET (15 clusters) - Mean Cluster Percentages
- Table 39. VIT (1 cluster) - Cluster Membership

The above results provide a basis for the selection of ports to achieve a particular number of shots.

### 5.9 Calculation of MWCVs for the ISMP at-sea length frequency data

The collection of length frequency data by observers at-sea forms a component of the ISMP. This section outlines the methods that are used to calculate the mean weighted CVs for relevant species using these data.

Let $n_{y, s, l, i}$ represent the number of individual fish sampled in year $y$, stratum $s$, length class $l$, shot $i$. We use the following dot notation to denote summation across particular subscripts, i.e.

$$
\begin{equation*}
n_{y, s,, i}=\sum_{l} n_{y, s, l, i} \tag{4.8-T21}
\end{equation*}
$$

so that $n_{y, s,, i}$ is the total number of individuals (of a given species of course) sampled on the $i$-th shot in stratum $s$ in year $y$. To calculate the estimated length frequency distribution, the length frequency distribution is first normalized for each shot in the ISMP:

$$
\begin{equation*}
p_{y, s, l, i}=n_{y, s, l, i} / n_{y, s,, ~} \tag{4.8-T22}
\end{equation*}
$$

A mean proportion $\bar{p}_{y, s, l}$ is then calculated for all shots in a particular year $y$ and stratum $s$ :

$$
\begin{equation*}
\bar{p}_{y, s, l}=p_{y, s, l, \bullet} / S_{y, s} \tag{4.8-T23}
\end{equation*}
$$

Where $S_{y, s}$ is the number of shots per year $y$ and stratum $s$ combination. At this stage $\bar{p}_{y, s, l}$ would be normalised, i.e.

$$
\begin{equation*}
\sum_{l} \bar{p}_{y, s, l}=1 \tag{4.8-T24}
\end{equation*}
$$

Associated with this mean is a standard deviation of the proportion $\bar{p}_{y, s, l}, S D_{y, s, l}$, which is calculated ignoring the small sample size correction of $S_{y, s} /\left(S_{y, s}-1\right)$ on the square of the sample standard deviation. For cases where $S_{y, s}=1$, we replace the 'null' value of $S D_{y, s, l}$ by the mean standard deviation for all values (years and strata) for the same length class where $S_{y, s}>1$ :
$S D_{y, s, l}\left(S_{y, s}=1\right)=\sum_{\forall y, s\left(S_{y, s}>1\right)} S D_{y, s, l} / \sum_{\forall y, s\left(S_{y, s}>1\right)} 1$
The stratum weight, $S W_{y, s}$ is year and stratum dependent, and is expressed on a mass basis. It is converted to a numbers basis $S N_{y, s}$ via the formula
$S N_{y, s}=S W_{y, s} / \bar{w}_{y, s}$
where

$$
\begin{equation*}
\bar{w}_{y, s, i}=\sum_{l} l^{3} \bar{p}_{y, s, l} \tag{4.8-T27}
\end{equation*}
$$

The estimate of the 'number of individuals' in year $y$ stratum $s$ in length class $l$ is $\hat{N}_{y, s, i}$, where:
$\hat{N}_{y, s, i}=S N_{y, s} \bar{p}_{y, s, l}$
And the variance of this estimate $V\left\{\hat{N}_{y, s, i}\right\}$ is given by

$$
\begin{equation*}
V\left\{\hat{N}_{y, s, i}\right\}=\left\{S N_{y, s}\right\}^{2}\left\{S D_{y, s, l}\right\}^{2} / S_{y, s} \tag{4.8-T29}
\end{equation*}
$$

The estimate of the 'number of individuals' in year $y, \hat{N}_{y, \cdot, i}$, is

$$
\begin{equation*}
\hat{N}_{y, \cdot i}=\sum_{s} \hat{N}_{y, s, i} \tag{4.8-T30}
\end{equation*}
$$

and the variance of this quantity $V\left\{\hat{N}_{y, i}\right\}$ is:

$$
\begin{equation*}
V\left\{\hat{N}_{y, \cdot, i}\right\}=\sum_{s}\left\{S N_{y, s}\right\}^{2}\left\{S D_{y, s, l}\right\}^{2} / S_{y, s} \tag{4.8-T31}
\end{equation*}
$$

The CV for a particular year $y$ and length class $l, C V_{y, l}$ is

$$
\begin{equation*}
C V_{y, l}=100 \times \frac{\left[\sum_{s}\left\{S N_{y, s}\right\}^{2}\left\{S D_{y, s, l}\right\}^{2} / S_{y, s}\right]^{1 / 2}}{\hat{N}_{y, \bullet, i}} \tag{4.8-T32}
\end{equation*}
$$

and the MWCV associated with year $y$ for the species under consideration is:
$M W C V_{y}=\sum_{s}\left\{C V_{y, l}\right\} \frac{\hat{N}_{y, \bullet}}{\sum_{l} \hat{N}_{y, \bullet}, i}$
A number of alternative calculation procedures have not been considered in this study. The main one involves a different prescription for averaging the shot level data to produce a year x stratum mean length frequency distribution. For example, a possibility is to weight each shot's length frequency distribution by the number of fish of that species sampled on that shot. Such an approach seems appropriate at low fish sample sizes per shot, where weighting by sample size is intended to down-weight shots which are likely to suffer from excessive variance due to low sample size. However, that can lead to overweighting of shots when a large number of fish are sampled for a given shot. Therefore, were such an approach is to be pursued, it should probably involve sample size weighting below a threshold of, say, 30 fish sampled and measured, and weighting by the threshold value (i.e. 30) when the actual sample size is above this threshold. For this interim report equal weighting of normalised shot level length frequency distributions is the approach followed, subject to the caveat that this might unduly overweight shots with very low sample sizes and underweight shots with larger sample sizes.

Table 40 and Table 41 provide the MWCV estimates described above in tabular format.

### 5.10 Calculation of MWCVs for the ageing data provided for this study, using the atsea ISMP sampling data.

Ageing data were provided for this study, and the terms of reference for this study require that the CVs for catch-at-age should be presented. The MWCVs that will be developed here are limited to the MWCVs arising from the sampling errors in the ALKs. We define $p_{y, a, l}$ as the proportion of fish in length class $l$ (in year $y$ ) which are from age class $a$. Thus for all $l$,

$$
\begin{equation*}
\sum_{a} p_{y, a, l}=1 \tag{4.10-T34}
\end{equation*}
$$

Let $p_{y, l}$ be the estimated proportion of fish in length class $l$ in year $y$ (the so-called length frequency distribution), i.e.

$$
\begin{equation*}
\sum_{l} p_{y, l}=1 \tag{4.10-T35}
\end{equation*}
$$

The estimated age frequency distribution (of the catch), $p_{y, a}$, is
$\sum_{l} p_{y, a, l} p_{y, l}=p_{y, a}$
We suggest an approximate approach to determine the MWCV of $p_{y, a}$ It involves treating the Maximum Likelihood Estimate (MLE) based proportions from the available sample as equal to the true underlying proportions $p_{y, a, l}$. In addition, we ignore the variance of $p_{y, l}$, which reflects on the adequacy of sampling for length. An advantage of this approach is that the resultant MWCV for the age frequency will be solely due to the adequacy of sampling for the age-length key, and will not be involved in the length frequency estimation error. Since these two quantities are sampled via different approaches, separation of the length frequency MWCV and the age-frequency MWCV assuming a fixed length frequency distribution provides separate insights into the adequacy of sampling for length and sampling for age. This is useful since the so-called 'at-sea' samples are not the only data available for the estimation of age frequencies for retained catches, port sample are also available for retained catches. However, the 'at-sea' data will be the only available information to estimate the age frequency distributions for the discarded portion of landings.

Under these circumstances the resultant variance of $p_{y, a}$ is $V\left(p_{y, a}\right)$, where:
$V\left(p_{y, a}\right)=V\left(\sum_{l} p_{y, a, l} p_{y, l}\right)=\sum_{l} p_{y, l}^{2} V\left(p_{y, a, l}\right)$
And, if $N_{y, l}$ fish are aged in length class $l$ in year $y$, then:
$V\left(p_{y, a, l}\right)=\frac{p_{y . a}\left(1-p_{y, a}\right)}{N_{y, l}}$
Therefore,

$$
\begin{equation*}
V\left(p_{y, a}\right)=V\left(\sum_{l} p_{y, a, l} p_{y, l}\right)=\sum_{l} p_{y, l}^{2} V\left(p_{y, a, l}\right) \tag{4.10-T39}
\end{equation*}
$$

Note that the above approach to calculating the MWCVs for the age frequency estimates makes the implicit assumption that the fish sampled for ageing are done on the basis of length class stratification, where the length strata are the 1 cm length classes used in the age data. Some caution needs to be used in interpreting the results obtained via this method, since it makes no allowance for the variance due to length classes for which there were no agesamples.

Values for $p_{y, l}$ can be sourced from either 'at-sea' samples, or from port samples. The first set of results that are presented here are for the 'at-sea' length frequency distributions, as derived from the previous section. In order to carry out the calculations referred to, overlap by year
and species is required between the age and length samples. The relevant species that are considered here are those which appear in the age data between 2002 and 2008 as well as in the length data, for at least one corresponding year, i.e.:

1. BIGHT REDFISH
2. BLUE GRENADIER
3. BLUE WAREHOU
4. BLUE-EYE TREVALLA
5. DEEPWATER FLATHEAD
6. EASTERN SCHOOL WHITING
7. GEMFISH
8. JACKASS MORWONG
9. ORANGE ROUGHY
10. PINK LING
11. REDFISH
12. SILVER WAREHOU
13. TIGER FLATHEAD

For the period 2002-2008, there were no at-sea length frequency data provided for HAPUKU in the data supplied for the project.

Age frequency MWCVs calculated as described above are presented in Table 15 (retained) and Table 16 (discarded).

The actual sample sizes achieved for ageing (retained and discarded combined) are given in Table 17.

The MWCVs are approximately proportional to the inverse of the square root of overall sample size (for the same shot distribution across strata), and therefore, if a target MWCV for the age error as defined above is set at $10 \%$, the required sample sizes can be calculated. These sample sizes are given in Table 18 (discarded) and Table 19 (retained). The ratios of the required sample sizes (for MWCV as defined to equal $10 \%$ ) divided by the actual sample sizes are given.

These results indicate that for the 13 species cited above, and either the discarded or retained catch-at-age frequency estimates, the sample size is more than is required to achieve an MWCV of $10 \%$ in 8 cases, while it is inadequate in 18 cases. The 8 cases where the sample size is sufficient are:

1. Discarded blue warehou
2. Discarded eastern school whiting
3. Discarded gemfish
4. Discarded pink ling
5. Discarded silver warehou
6. Retained blue grenadier
7. Retained blue warehou
8. Retained gemfish

### 5.11 Calculation of MWCVs for the ageing data provided for this study, using the port length frequency data

A similar exercise to that carried out above for the at-sea ISMP length frequency samples can be carried out for the port length frequency data. There are however, some differences in approach which have to do with the sampling procedure for port based data. The essential difference is that it is not possible to fully disaggregate a port length frequency sample back to its constituent strata. Stratified weighting of the port length samples to obtain a full catch weighted length frequency distribution cannot therefore be readily achieved using the ISMP strata, whether historic or as proposed by this study. Achieving a catch weighted length frequency distribution is therefore probably best done according to strata which are available at the port level, and which reflect the proportion of the catch weight in each strata. The data
contain variables such as 'Zone' and 'Grade' which can be used. For example, annual zonelevel retained catches are available from the CDR-GENLOG merge carried out here. However, catches split by 'Grade' are not readily available at the CDR-GENLOG merge level.

Furthermore, the definition of a sampling unit for a port length frequency sample is not straightforward. There is, for example, a sampling process taking place regarding fish sampled for length measurement, however the data does not contain replicates at this level so other than proposing a multinomial sampling process, there is no empirical basis for quantifying the variance or uncertainty associated with the port sampling process for a given landing.

Another sampling process takes place w.r.t. to the selection of landing to be sampled. The sampling process (e.g. proportional random) could be used to generate variances for length frequency samples at this level.

Some of these complications seem to lie outside the scope of this study. The approach taken here has been to calculate a number of fish per length class at the following level:

- CSIROcode
- Year
- Month
- Day
- Gear code
- Zone code
- Grade
- Port
- Catch weight
- Length class

The reason that catch weight has been included as a key is that there are instances in the data where the value of catch weight is not unique with respect to a key obtained from CSIROcode, Year, Month, Day, Gear code, Zone code, Grade, Port, and Length class. The authors have concluded that there were probably other keys involved for which catch weight was unique. These keys are apparently not present in the data (various exploratory analyses were carried out to try to ascertain which variables might be involved). Consequently catch weight itself was used as an additional key, as a proxy for these other keys.

The number of fish in the catch at the key level was calculated assuming a cubic relationship between fish length and weight - the weight conversion of these numbers would sum to equal the catch weight. Annual and species specific estimates of fish catch numbers were produced at the following level of data aggregation:

- CSIROcode
- Year
- Length class

These fish numbers were then normalized across all length classes to produce a normalised length frequency estimate.

The length frequency distributions so described were then used to calculate the MWCVs inherent in the ageing data for the retained catch, as described in the previous section. Only a
few species and years emerge after the merge between the port length frequency and ageing data available to this study. These are shown in Table 20, where the quantity shown is the MWCV associated with the 'multinomial ageing error'.

The results in Table 20 should be compared to those presented in Table 15. The only differences are that a different retained length frequency distribution is used as the basis for the calculations.

We note that the pink ling MWCV in 2008 was 10.11 , despite the fact that only 10 pink ling were aged in 2008. This rather misleading MWCV estimate comes about because the MWCV estimation procedure for ageing ignores the variance associated with lengths that are not sampled. A minimum requirement for age sampling would of course be that all length classes are sampled. This is certainly not the case for the 10 pink ling sampled for ageing in 2008. The MWCVs reported here should therefore be regarded as negatively biased and hence somewhat optimistic, particularly for very low numbers aged, say less than 100, and or where it is clear that the ageing data only covers a limited range of the entire retained catch length frequency distribution. Despite this the results in Table 20 do reliably reflect the trends in ageing sample size for jackass morwong, where the sample size dropped from 245 in 2006, to 120 in 2007, with a corresponding increase in the MWCV from 23.98 (see Table 15) to 31.68 (see Table 20 above). This problem is related to the fact that the MWCVs have been calculated under the assumption that the underlying sampling process is length stratified, which is only the case for certain species and years. A more comprehensive analysis incorporating the full impact of completely random sampling for age is outside the scope of this study.

### 5.12 Calculation of the optimal combination of numbers of fish sampled per landing, and number of landings sampled, for the port length frequency sampling process

An exercise was carried out to provide input into the determination of the optimal combination of numbers of fish sampled per landing, and number of landings sampled, for the port length frequency sampling. The initial aim of this calculation is to determine the respective contributions to the total MWCV due to the number of fish sampled per trip, and the number of trips sampled.

To do this exercise it is necessary to identify trip in the port length frequency data. This was done by using the following key (which includes a split by species):

- CSIROcode
- Year
- Month
- Day
- Gear code
- Zone code
- Grade
- Port
- Catch weight

There are potential problems with this key, such as 'Grade', but the use of grading as a stratification variable is limited, hence we felt that this provided a workable definition of trip.

For each trip x species, a mean length frequency was calculated - these length frequencies were all padded out with zeroes so that they have the same dimensions. A mean and variance can then be calculated for each length class (length classes are 1 cm wide for most species
although 0.5 cm is used in some cases such as redfish). This variance is the total variance of the proportion of fish sampled across trips, say Vtot, where it is implicit that this value differs for each length class. Calculation of the variance in this way without allowance for differences in total landings across trips is open to question but is a reasonable short cut approximation. The true sampling process involves a finite population, and possibly random selection of clusters, and a complete treatment would ultimately have to involve complex sampling theory, which is outside the scope of this study.

If Vtot was solely due to the multinomial error process involved in sampling fish, with the same true underlying length frequency distribution in catches from trip to trip, then it should closely approximate the theoretical multinomial variance from the known number of fish sampled per trip. This multinomial variance is simply $p(1-p) / f$ where $f$ is the number of fish sampled per trip. The variance due to trip is therefore approximated by the difference between Vtot and $p(1-p) / f$, which we will refer to as Vtrip, Vtrip $=$ Vtot $-\mathrm{p}(1-\mathrm{p}) / \mathrm{f}$. Vtrip can thus be calculated empirically for each species, year and length class, for the particular value of f that applied for 2008 (we use an average f across all trips for that year). We have limited our analysis to the 2008 data, or for species poorly sampled in 2008, the 2007 data, since these years reflect the most recent characteristics of the fishery. The variance of the mean proportion, Vmean, for any value of $f$ and for any number of trips $T$ is thus given by

Vmean $=($ Vtrip $+p(1-p) / f) / T$
And the MWCV is just the sum of
$\mathrm{px} 100[\text { Vmean }]^{1 / 2} / \mathrm{p}$
over all length classes.
It seems important to point out that in the port sampling data from the SESSF, Vtot exceeded $\mathrm{p}(1-\mathrm{p}) / \mathrm{f}$ (for the 2008 value of f ) by a factor of between 5 and 25, indicating that trip to trip variance unrelated to multinomial noise plays a very important role in the determination of the MWCV.

The main conclusion that follows from this is that if there is no difference in cost for all combinations of fand $T$ which have the same product fT (i.e. the same total number of fish sampled per year), then it is always better to increase the number of trips sampled, and reduce the number of fish sampled per trip, since the smallest MWCV will always be achieved for the largest possible value of $T$.

Naturally one expects there to be greater costs associated with increasing T subject to the constraint that fT is invariant. Although this matter is probably outside the scope of this study, if more detailed cost information becomes available, then it may be possible to carry out a more realistic optimisation calculation. In that case there may be a non-trivial optimal combination of f and T (i.e. one that does not simply favour the largest value of T possible).

We note that although the same analysis was not attempted for the at-sea length frequency data, we strongly suspect that it will conclude that the minimum MWCV is achieved by increasing the number of shots when the product of shots and fish sampled per shot is held constant (i.e. the total number of fish sampled at-sea remains invariant).

In relation to the above, it appears from Table 10 to Table 12 that the number of trips sampled per year is lower in 2007 and 2008 compared to 2002-2006, whereas the average number of fish sampled per landing is larger (by more than double). We are concerned that this may
simply reflect a problem with our definition of trip. However if this is a real change in sampling strategy then it represents a sampling trend in the wrong direction when minimum MWCV is the objective.

A limited amount of output has been prepared for a selection of species to illustrate the application of the methods described above. This output is summarised in Figure 22-Figure 35 , where the relationship between the MWCV and T is shown subject to the 2008 value of fT and a value of fT which is respectively twice or $50 \%$ as large.

### 5.13 Collection of additional information relevant to fishing trips/operations including

Thorough analysis of available data provided insight into the shortfalls in the information recorded as part of the routing fisheries and observer data collection programs. These shortfalls significantly increased the difficulty of analyses, and reduced the amount of data available for these analyses. Major issues are discussed in more detail in Appendix A.

The absence of a unique key to link CDR, GENLOG and observer data (at-sea and port) causes inefficiencies in data pre-processing that could be easily rectified. While merging these data sets has become more reliable since 2000, there is still considerable drop-out requiring "decision rules" that can introduce unnecessary error before analysis has even begun. We strongly advise - if this has not already been done - that such a key be established in the data recording process. An example of a unique key that would enable merging is including the fields: vessel ID, trip end date and trip end time in all datasets. Each fields, of course, should contain the same formatting in both datasets. Trip end time is required to take into account multiple landings on a single day. With the addition of shot number or shot time in the at-sea and GENLOG data, individual shots can also be matched between the data sets. Improvements in this area will increase the quality and quantity of data available for use in a wide range of analyses beyond this project.

The use of different species codes in CDR, GENLOG and ISMP data sets also required considerable effort to resolve. The ISMP data contained CAAB codes, the CDR data contained a 3-letter AFMA code and the GENLOG data contained both 3-letter AFMA code and CAAB codes. This was further complicated for instances in which more than one AFMA code is linked to the same CAAB code in the GENLOG data, and by species that appeared in one or two data-sets, but not all three. Australian Fish Names Standard AS SSA 5300 was approved in 2007 and incorporates the Codes for Australian Aquatic Biota (CAAB). We recommend that CAAB codes be used in all data sets that will make the merging of the data with each other easier, and also improve potential linking with other datasets (e.g. Fisheries Independent Surveys).

CDR data contained numerous instances of nonsensical issues that fall into six different categories:

- Trips for which end dates occur after the start date of the next trip for the same vessel overlapping trips
- Trips in which the end date is equal to the start date of the next trip for the same vessel special kind of overlapping trip
- Trips in which the end dates occur before start date - inverted trips
- Situations where shot dates lie in-between the end-date and start date of the next closest trips for the same vessel - crack situation
- Situations in which a trip is nested within another trip - nested trips
- Situations where two trips have the same start and/or end-dates - really a special combination of all of the above
These issues should be picked up and corrected during validation of data entry. If it is not already done, we recommend that a validation procedure is implemented to cross check start, end and shot dates and times for each vessel.

The observer data was delivered in a number of different structures including flat files, databases and Excel spreadsheets. Each tranche of data contained different fields and formats resulting in inefficiencies during the pre-processing stage, and increasing the risk of introducing errors. Consistent format and data storage should be employed for each of the datasets. This will also be beneficial for routine analysis (e.g. for stock assessment) and adhoc queries. A particular problem arose in trying to identify observer trips on gillnet vessels from trip on trawl vessels. It is unclear why this differentiation could not be made as there appears to be a number of fields available to accommodate gear code including "gear" and "TRAWLTYPE". Gear code recording procedures need to be improved for these data to be used in efficient, accurate analyses.

The TEP data was at times missing information on the life-state/fate of the interacting animal that might be required to more accurately assess impacts of fishing etc. This was particularly the case when interactions were recorded in the 'ISMP catch' table that contained no field for recording fate apart from 'Comments', which cannot easily be used in routine analyses.

There were two major difficulties in reconciling information on TEP species:

1. Interactions where the animal was caught were sometimes recorded in both the 'ISMP catch' table and the 'TEP interactions' table
2. A lack of unique key that could be used to match the 'TEP interactions' table and the 'ISMP catch' table

The first problem can be overcome by including a decision rule into the observer Operation and Procedure Manual (or equivalent) by specifying that either all interactions with TEP species get recorded on the 'TEP interactions' table only, or that extractive TEP interactions be recorded in the 'ISMP catch' table only. In light of the previous paragraph, it is preferable that the former option is adopted to enable information on the life-state/fate of the animal to be recorded. The second problem can only be overcome by the inclusion of a clear unique key that can be used to merge the 'TEP interactions' table 'ISMP catch' table, as mentioned near the top of this section.

While detailed gear information are routinely collected for the GHAT sector, the SET data lacks sufficient gear details to enable comparison of catch by gear type for example. Such information would be useful in examining impact of gear changes on discarding or of the use of BRDs and offal management on interactions with TEP species. Additional fields recommended to be collected as part of the at-sea observing are listed below:

Ground gear type
Disc size
Codend mesh size
Net type
Headrope length
Footrope length
Wing spread
Maximum wing mesh size

Sweep length
Bridle length
Length of rope per side
Mesh type
Maximum wing mesh size
Bycatch reduction device (e.g. square mesh panel/SED/Tori line)
Offal discharge management (e.g. bulk discharge/discharge while steaming)

## 6. Discussion

### 6.1 Data pre-processing

In preparing the data for analysis to meet this project's requirement a myriad of data preprocessing challenges were encountered. Some, such as the CDR-GENLOG merge, are a well known feature of the data and of this type of analysis; others were unique to this project, such as the difficulties of linking wildlife interaction records to fishing activities. In hindsight there are clear modifications to data recording procedures which would significantly enhance the value of the available data for analyses such as are carried out here. Examples of such modifications include the creation of a key that is able to link landing records (CDR data) with shot level fishing activity records (GENLOG data). Others are, for example, improvements in the gear code recording procedures - it is not possible to identify gillnet fishing in the ISMP data since there was, apparently no provision for this gear code to be recorded during the execution of the ISMP. Indeed recommendations for how to best make this modification forms part of the deliverables for this project. A compilation of all such issues does not appear here but can be gleaned from the Appendix dealing with data preprocessing (Appendix A).

It seems appropriate to offer a word of caution. The reader will have noted that the data preprocessing step of this project involved a substantial technical input. It also incorporated numerous assumptions centred predominantly around the CDR-GENLOG merge, and around other 'forced merges' which were required in the absence of keys linking operationally linked but data-wise unlinked tables. Much subsequent statistical analyses have been carried out using the data flowing from these processes. It is important to bear in mind that there are errors, perhaps better referred to as biases, which have been introduced by the data preprocessing steps. Their scale and direction have not been assessed for this document, and in some cases this is not possible.

### 6.2 ISMP Stratification

The re-stratification of the SESSF ISMP reported upon here has focussed on the GAB and GHAT components of the ISMP, given the longer history of the ISMP for the CTF, and the more established basis for stratification in that sector. The results suggest that the GAB needs to include some stratification by depth, while the GHAT needs stratification by gear and by a geographic zonation variable. The last mentioned is thus support for the current or similar geographic stratification scheme presently in use for the GHAT. The variable used to make recommendations on the stratification was (a) the species composition of catches, and (b) a gross discard measure for different catch species assemblages (clusters) identified in the SESSF fishery. Further stratification using other variables could be attempted, but was not carried out here.

The issue of stratification is a complex one in the context of this project. Broadly one envisages two sets of strata:

- For purposes of estimation - the historic strata used in the ISMP
- For purposes of a revised design for the SESSF ISMP - a new set of strata, as proposed broadly above

For this document we have however taken the liberty of presenting historic estimates using the new strata proposed for the revised ISMP (in effect mixing the two concepts above). In addition, the second category of strata must be separated further into:

- A set of strata used to determine the optimal sample size distribution across strata
- A set of strata which is the recommendation for strata to be used in the actual implementation of the revised SESSF ISMP

The reason for the difference in the two sets of strata referred to above is that certain strata which should be included in the revised SESSF ISMP cannot be identified in the ISMP data because:

1. As mentioned above gillnetting cannot be identified in the 2007/08 ISMP data and gillnet shots are recorded as trawl shots and hence subsumed in the SET strata.
2. VIT (Victoria Inshore Trawl) could not necessarily be identified as a separate stratum in the SESSF ISMP.
3. In many cases, the different types of line fishing (AL, LL, DL, BL) cannot be identified in the ISMP data.

It is thus necessary to concatenate certain strata which should ultimately be used in the revised SESSF ISMP, for the purpose of the design optimisation calculations. The way that this is done is as follows:

In order to deal with the absence of gillnetting and VIT in the historic ISMP data, the stratum sizes for the SET strata were increased by $38 \%$ (based on information in the CDR-GENLOG data) prior to optimisation of designs using the more restricted set of strata listed immediately above. It is now necessary to disaggregate the 'design' allocation recommendations to the probable set of strata that are feasible in the actual implementation of the ISMP. The following initial considerations are required for this step:

1. Reduce the ISMP allocations to SET strata by $1 / 1.38$, and allocate these shots to gillnetting strata based pro rata on the gillnet shots recorded for the SESSF in the CDR-GENLOG data.
2. Make an allocation to VIT based pro rata on the VIT shots recorded in the SESSF in the CDR-GENLOG data.

During 2007 and 2008, the proportion of gillnet shots in each of the 7 gillnet strata considered here was as follows:

| GN CSA | 0.270 |
| :--- | ---: |
| GN EBS / | 0.326 |
| NSW | 0.032 |
| GN ESA | 0.075 |
| GN ET / WT | 0.029 |
| GN SAV | 0.189 |
| GN WA / WSA | 0.080 |
| GN WBS |  |

In addition, during 2007 and 2008, VIT shots comprised $1.061 \%$ of all GENLOG shots in all other strata. Strata are, subsequent to the optimal design calculations, disaggregated into the final recommended strata, on a basis pro rata to shot frequency in the SESSF in 2007 and 2008 - to reflect the very recent structure of the fishery.

### 6.3 Optimal ISMP design considerations for total catch - commercial and high risk species

The analysis of discarding on a mass basis represents an important aspect of this study. Considerable conceptual complex was evident when we reviewed methods used in previous similar studies. The estimation of the tonnage discarded, based on ISMP records of retained and discarded amounts, has many facets to it. Does one base the estimate on the ratio of discarded to total catch, or on the mean discarded amount per shot? The pros and cons between these methods are numerous and by no means straightforward. They depend, for example, on whether estimates involve extrapolation into unsampled strata, and on the discarded amount expressed as a percentage. A composite method was eventually developed and formed the basis of the ISMP redesign analysis described here.

This report describes a progressive sample size allocation scheme which sequentially allocates samples, one at a time, based on predefined objectives.

Depending on the overall objective supplied to the ISMP optimisation algorithm, the progress of a sample allocation process can be tracked. This process can be stopped at an economically desired total number of shots, and so it becomes a useful tool for the development of the ISMP survey and sample size allocations per stratum. There are numerous areas though that could benefit from feedback from those responsible for the ISMP and its design and we have pointed out these areas at the beginning of this document. The main areas are the objectives for the design of which an important component is the relative importance of difference species.

Notable in Figure 7 for example, is the disproportionate sampling effort preferred in the EDL_IN_TR and GAB Inshore strata. Both of these strata contain many Project Keys (EDL_IN_TR has 60 out of 75, GAB Inshore has 62 out of 75 ). Both also have high average standard $\overline{d e v i a t i o n ~ t o ~ m e a n ~ d i s c a r d ~ r a t i o s ~ o f ~ m o r e ~ t h a n ~ 1 . ~ N o t e ~ t h a t ~ t h e ~ E D L \_I N S ~ \& T R ~ s t r a t u m ~}$ is favoured by the VWCV method, implying that a lot of the high value species reside there. This stratum gets a high proportion of sampling effort right from the word go.
The GAB Inshore stratum on the other hand is favoured by the MinS method. It also only starts blowing out when the sea-days gets to about 400 (i.e. when we get down to the last 10 to 15 species which are still above target). Note, that for 13 project keys, more than half of the total estimated discards for that project key occur in the GAB Inshore stratum, so that in
order to get precision on the estimate of discards for these keys, considerable sampling is required in that area.

In Figure 6 we note that the MinS-HR method achieves its objective of bringing the 7 high risk project keys below target within 500 sea-days. However, at this level of sampling 14 of the commercial project keys are still above target. The MinS (commercial) method on the other hand has only 9 commercial species above target, but also 4 of the 7 high risk project keys. This illustrates one of the trade-offs to be made between the differing objectives of the program. The 4 plots of Figure 5 and Figure 6 represent 4 distinct objective functions (2 commercial and 2 high risk). Each plot is underscored by the method which is specifically designed to meet that objective.

### 6.4 Costs of the sampling designs

We were only provided with very broad figures from which to estimate the cost of the various sampling designs. AFMA indicated that observer costs were currently $\$ 730.00$ per land-day based on 117 land days and $\$ 1005.00$ per sea-day based on 587 sea days. We understand that these costs include all AFMA overheads, but it was not clear if this included costs to analyse and report on the ISMP data. Without additional information on the locations of the observers relative to the ports and the method of determining the conditions under which land-days are paid, it is impossible to accurately determine the costs of any of the sampling designs. To enable a relative comparison of the different designs, we have simply estimated relative costs based on the sea-day requirements alone, based on $\$ 1000$ / per sea-day. Thus, a sampling strategy requiring 500 days can be estimated to cost $\$ 500,000$. It must be stated that this will be the very minimum cost as it does not take into account travelling time and expenses and any land-day costs. It should also be noted that these estimated costs are based on the current observer program and it could be reasonable to expect that greater at-sea observer coverage would give greater economies of scale.

The AFMA costs for port sampling were stated at $\$ 25$ per hour, during which it was expected that approximately 60 length frequency or 12 pair of otoliths could be measured per hour. It is not clear how these output figures were derived, but they appear to be a very conservative estimate of what can be achieved. Again, we are unable to estimate the actual cost requirements for port-sampling without knowledge of where the port-based samplers will be based and the amount of length frequency data that will be collected at sea. This latter factor is also dependent on which sampling design is ultimately chosen.

### 6.5 Length and age frequency MWCVs

## Lengths

Methods were developed to make it possible to comment on the adequacy of the sampling for length and age. For this purpose it was assumed that the redesign of the SESSF ISMP would be primarily based on the retained and discarded catch information, and hence length and age related design considerations are an adjunct to the main drivers of the design. Given that context, the most important design issue considered for length was the trade-off between the number of fish sampled per shot or trip, and the number of shots or trips sampled. Shots are relevant to the at-sea sampling for length, while trip is applicable to the port based length sampling. While the sampling frame for the at-sea samples can be approximated as a standard random stratified sampling scheme with infinite population sizes (even though the population sizes are finite), with the strata as defined here for the ISMP, the sampling frame for the port sampling is less easily characterised on that basis. In the first instance the concept strata for port sampling are not the ISMP strata, but rather each trip that is sampled. In that case the stratum size is related to the total number of fish landed. However, since not all trips are
sampled, and since there are apparently no within trip replicates, the applicable sampling frame falls into the class of complex sampling frames. Thus the methods developed here, assuming that each trip represents a simple random sample of a length frequency distribution from an infinite population size of trips, is described as a short cut method. For the port sampling data only, the multinomial within trip variance and the trip to trip variance is decomposed so that the implications of different numbers of fish sampled per trip and trips sampled in total can be explored. These results suggest that if the total annual number of individuals sampled per year is held constant, then the precision of length frequency estimates is always improved by increasing the number of trips sampled (and of course reducing the number of fish sampled per trip). There are some indications in the data that in 2007 and 2008 the sampling strategy for port samples shifted from sampling more fish per trip on fewer trips. This strategy, if correctly inferred from the data, would reduce the precision of length frequency estimates.

The same calculations were not carried out for the at-sea length frequency data, but the same result is extremely likely (i.e. that more shots is the way to go to significantly reduce MWCVs for the length frequency distributions, and not more fish measured per shot). Cost considerations obviously impinge on the shot/trip versus fish measured trade-off and the additional expense of more shots/trips is likely at some point to kick-in, but this threshold requires more detailed economic information than was available for this study. It is possible that increasing the number of shots is so much more expensive than increasing the number of fish sampled that the latter is the only economically feasible option. If so one needs to appreciate that the benefits will be considerable less than 'root N ' where N is the total number of fish sampled.

We do provide some information in the text on species specific MWCVs, and how to determine what sample size would be needed to achieve lower MWCVs. It is easy in this output to highlight species where the MWCVs are too large, for either the retained or discarded component of the catch.

The precision of the length frequency distributions from at-sea samples seems most inadequate overall for retained: gummy shark, gemfish, ribaldo; and for discarded a large number of species exhibit unacceptably high MWCVs .

For the port sampling data the following species show MWCVs well in excess of $20 \%$ : redfish, school shark, blue-eye trevalla and alfonsino.

## Age

The adequacy of age sampling was studied using either the at-sea or port based length frequency distributions as fixed quantities, and then calculating the MWCV for the age frequency distributions from the cumulation of multinomial error arising from the age length key proportions and number of fish aged per length class. Assuming that the length frequencies are error free in this way allows one to focus on the errors in the catch age frequencies arising from the ALK. It is of course acknowledged that the total variance of the age frequency distributions estimates involves both the errors in the length frequency estimates and the age frequency estimates per length class.

This calculation assumes that the age sampling was length stratified which may not always have been the case, but the method is regarded as a useful index of the adequacy of age sampling. It is then a simple matter to inspect the output and ascertain where the age sampling needs to be increased or decreased to achieve a particular MWCV - we suggested
a target MWCV of $10 \%$, but the calculation is easily adapted to deal with other target MWCVs via the square root of N rule described in the text.

The number of ageing samples is most inadequate for retained species: Bight redfish, jackass morwong, orange roughy (2004 and 2005 only available), and for discarded species: Bight redfish, jackass morwong, orange roughy (2004 and 2005 only available).

### 6.6 Optimal ISMP design considerations for fishery-TEP interactions

The TEP interaction data and analysis work submitted here focuses on an analysis of the number of TEP species group individuals interacting with the SESSF. We also present estimates of the number of deaths occurring as part of TEP-SESSF interactions. The statistical model that is used is one in which the ratio of variance to mean for a TEP species group is invariant over space and time, although the mean may vary over space and time. This ratio is closely related to the typical or average group size for TEP species groups encountered by the SESSF, where the numbers obtained were:

Birds: 37.5
Pipefish: $\quad 1.74$
Seals: $\quad 8.38$
Sharks: $\quad 1.50$
Whales/Dolphins: 292.91.
Consideration of the number of individuals involved in an 'interacting' shot is dealt with by multiplication of the mean number of individuals per interacting shot and the total number of interacting shots. This produces estimates of total individual numbers interacting with the fishery. In addition, using comments captured as free text in some of the data (ISMP TEP Interaction table), estimates of TEP mortalities recorded by ISMP observers were made and these were multiplied up to produce estimates of total annual mortalities in the TEP.

The results for the optimal design for TEP species groups are produced separately for each TEP species group. The results presented here are limited to illustrative results produced for 400 sea-days. Fractional days and shots are permitted in these calculations. The sample sizes required to achieve a $\mathrm{CV}_{\text {target }}$ different to the value shown here, CV , is obtained by multiplying 400 sea-days (or the corresponding number of shots) by the square of $\mathrm{CV} / \mathrm{CV}_{\text {target. }}$ Given this, it is clear that it is only for 'Birds', 'Pipefish' and 'Seals' that the optimal design achieves a target CV less to or close to $20 \%$. The sample size would have to increase by at least an order of magnitude to hope to achieve 20\% for the Sharks and Whales/Dolphins groups.

At this stage the important issue is the distribution of shots for the said 400 sea-days. Optimised results are shown in Table 22 to Table 25. A decision on the preferred design will have to involve a compromise between different TEP species groups (aside from the compromise with the optimised design results for the commercial and high risk species groups). We offer here one simple compromise, averaging the optimal number of days and shots per stratum for each species group across all species groups. Table 22 shows that these average optimal shot allocations are markedly different to the shot allocations which are simply proportional to the shots in the fishery (from the GENLOG data).

It is clear from these results that the optimal shot distribution differs markedly between the different species groups. In many cases these are being driven by the stratum where there happens to have been a chance sighting of a particular TEP species individual. Many strata show no observations of TEP species groups in the ISMP between 2002 and 2008. It seems unlikely that there were in fact no interactions in the fishery, hence the use of a background
level of SESSF-TEP interactions as an alternative basis for optimal design calculations (see Table 23 and Table 25). Here again we offer as a simple compromise, averaging the optimal number of days and shots per stratum for each species group across all species groups. Table 23 shows that this compromise is now closer to the proportional allocation to shots in GENLOG than for Table 22. This is because the inclusion of some underlying background level of TEP interactions tends to spread out the average optimal shot allocations.

The CV implications of these different results obviously need to be weighed up. Both Table 22 and Table 23 show the extent to which the optimal CVs increase when compromising on the average shot allocations, and again when moving to the proportional shot allocation approach. Since optimal cannot be achieved for each TEP species group, the important comparison is between something like average and proportional. Table 23 suggests that the increase in CV going from average to proportional is most marked for birds and seals, but immaterial for the other TEP species groups.

It seems appropriate to offer a word of caution. The reader will have noted that the data preprocessing step of this project involved a substantial technical input. It also incorporated numerous assumptions centred predominantly around the CDR - GENLOG merge, and around other 'forced merges' which were required in the absence of keys linking operationally linked but data-wise unlinked tables. Much subsequent statistical analyses have been carried out using the data flowing from these processes. It is important to bear in mind that there are errors, perhaps better referred to as biases, which have been introduced by the data pre-processing steps. Their scale and direction have not been assessed, and in some cases this is not possible.

### 6.7 Data collection

Many of the difficulties experienced during the pre-processing phase of this project could have been avoided if the CDR, GENLOG and ISMP data were more consistent. Merging these data sets was difficult because of the lack of consistent unique keys and species codes, and these issues are discussed more thoroughly in Section 5.12 and Appendix A. It is recommended that a common unique key (such as 'vessel ID', 'trip end date' and 'trip end time') be recorded for every fishing and observer trip. In addition, the recording of a shot ID code (such as 'shot start time' or 'shot number') in GENLOG and ISMP data would aid the merging of these data sets to the shot level. The Codes for Australian Aquatic Biota (CAAB) is an Australian Standard, and should be adopted as the standard species code in all SESSF fishery datasets. These two improvements will result in more efficient data processing in the future, reduce the risk of introducing errors, and increase the amount of quality data available for analysis. These benefits will be realised in applications such as routine stock assessments and data queries.

Other improvements recommended include creating a validation procedure for GENLOG data entry to reduce errors in start, end and shot dates, standardising collection of TEP interaction data, and increasing the detail of trawl gear specifications recorded.

## 7. Acknowledgments

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## 9. Figures and Tables



Figure 1 - Area of the Southern and Eastern Scalefish and Shark (© Commonwealth of Australia 2005).


Figure 2-Theoretical Comparison of CVs using Methods A and B for estimation of Discarded Catch. CV1 uses method A with an assumed gamma value of 3. CV2 uses method $B$ with an assumed discarded proportion of 0.5 .


Figure 3. Comparison of CVs for the discards and the equivalent CV for the total catch.


Figure 4. Comparison between a the new target CV n the total catch (no symbols, solid line), the equivalent CV on the discards (square symbols) and the target CVs on discards used in previous ISMP redesign studies (diamond symbols.)



Figure 5. Progress of each of the 5 allocation methods by number of observer sea-days allocated. The statistic shown is the value-weighted average $C V$ for the $\mathbf{6 8}$ commercial project keys (upper panel) and for the 7 HIGH RISK project keys (lower panel). The allocation methods used are: VWCV (which minimises the value-weighted average CV), MinS (which minimises the number of project keys with CVs above target); Prop (allocates sampling effort in proportion to the size of the fishery in each stratum); CWCV-HR (minimises the catch-weighted average CV for the HIGH RISK species), MinS-HR (minimises the number of HIGH RISK species groups with CVs above target.)


Figure 6. Progress of each of the 5 allocation methods by number of observer sea-days allocated. The statistic shown is the number of commercial project keys out of 68 (upper panel) and number of HIGH RISK project keys out of 7 (lower panel). The allocation methods used are: VWCV (which minimises the value-weighted average CV), MinS (which minimises the number of project keys with CVs above target); Prop (allocates sampling effort in proportion to the size of the fishery in each stratum); CWCV-HR (minimises the catch-weighted average CV for the HIGH RISK species), MinS-HR (minimises the number of HIGH RISK species groups with CVs above target.)


Figure 7 - Sample size distribution across strata for a total sample size of 500 observer sea-days by Methods: VWCV, MinS and Prop.


Figure 8 - Sample size distribution across strata for a total sample size of $\mathbf{5 0 0}$ observer sea-days by Methods: CWCV-HR, MinS-HR and Prop.


Figure 9 - Sample size distribution across strata for a total sample size of 750 observer sea-days by Methods VWCV, MinS and Prop.


Figure 10 - Sample size distribution across strata for a total sample size of $\mathbf{7 5 0}$ observer sea-days by Methods CWCV-HR, MinS-HR and Prop.


Figure 11 - Sample size distribution across strata for a total sample size of $\mathbf{1 0 0 0}$ observer sea-days by Methods VWCV, MinS and Prop.


Figure 12 - Sample size distribution across strata for a total sample size of $\mathbf{1 0 0 0}$ observer sea-days by Methods CWCV-HR, MinS-HR and Prop.


Figure 13 - Sample size distribution across strata for a total sample size of $\mathbf{1 2 5 0}$ observer sea-days by Methods VWCV, MinS and Prop.


Figure 14 - Sample size distribution across strata for a total sample size of $\mathbf{1 2 5 0}$ observer sea days by Methods CWCV-HR, MinS-HR and Prop.


Figure 15 - Sample size distribution across strata for a total sample size of $\mathbf{1 5 0 0}$ observer sea-days by Methods VWCV, MinS and Prop.


Figure 16 - Sample size distribution across strata for a total sample size of 1500 observer sea days by Method CWCV-HR, MinS-HR and Prop.

|  | Sea-days | CV (\%) on Total Catch: |  |  | Method VWCV |  | 1500 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 250 | 500 | 750 | 1000 | 1250 |  |
| OCEAN BLUE-EYE TREVALLA |  | 0 | 0 | O | O | 0 | O |
| Sawsharks |  | O | O | O | O | O | O |
| STRIPED TRUMPETER |  | 0 | 0 | 0 | 0 | O | 0 |
| WHISKERY SHARK |  | 0 | 0 | 0 | 0 | 0 | 0 |
| BIGHT REDFISH |  | 1 | 0 | O | 0 | 0 | 0 |
| PINK LING |  | 1 | 1 | O | O | O | O |
| YELLOWSPOTTED BOARFISH |  | 1 | 1 | 1 | 1 | 1 | 0 |
| GUMMY SHARK |  | 1 | 1 | 1 | 1 | 1 | 1 |
| ORANGE ROUGHY |  | 2 | 1 | 1 | 1 | 1 | 1 |
| HAPUKU |  | 2 | 2 | 1 | 1 | 1 | 1 |
| BLUE-EYE TREVALLA |  | 2 | 2 | 1 | 1 | 1 | 1 |
| BLUE GRENADIER |  | 3 | 2 | 2 | 1 | 1 | 1 |
| STARGAZER (UNSPECIFIED) |  | 3 | 2 | 2 | 2 | 1 | 1 |
| BROADNOSE SHARK |  | 2 | 2 | 2 | 2 | 2 | 2 |
| JOHN DORY |  | 3 | 2 | 2 | 2 | 1 | 1 |
| ROYAL RED PRAWN |  | 3 | 2 | 2 | 2 | 2 | 2 |
| ALFONSINO |  | 3 | 3 | 2 | 2 | 2 | 1 |
| DEEPWATER FLATHEAD |  | 4 | 3 | 2 | 2 | 2 | 2 |
| TIGER FLATHEAD |  | 4 | 3 | 2 | 2 | 2 | 2 |
| SILVER TREVALLY |  | 4 | 3 | 2 | 2 | 2 | 2 |
| SCHOOL SHARK |  | 3 | 3 | 3 | 2 | 2 | 2 |
| BOARFISH (UNSPECIFIED) |  | 4 | 3 | 3 | 3 | 2 | 2 |
| EASTERN SCHOOL WHITING |  | 6 | 4 | 4 | 3 | 3 | 2 |
| OCTOPUS |  | 7 | 5 | 4 | 3 | 3 | 3 |
| SQUID (GENERAL) |  | 8 | 5 | 4 | 4 | 3 | 3 |
| Molluscs |  | 5 | 4 | 4 | 4 | 3 | 3 |
| GOULD SQUID |  | 8 | 6 | 5 | 4 | 4 | 3 |
| BRONZE WHALER |  | 7 | 6 | 5 | 4 | 4 | 3 |
| BLUE MORWONG |  | 8 | 6 | 5 | 4 | 4 | 4 |
| JACKASS MORWONG |  | 12 | 8 | 7 | 6 | 5 | 5 |
| SNAPPER |  | 13 | 9 | 7 | 6 | 6 | 5 |
| MIRROR DORY |  | 12 | 9 | 7 | 6 | 6 | 5 |
| INSHORE OCEAN PERCH |  | 13 | 9 | 8 | 7 | 6 | 5 |
| SMOOTH HAMMERHEAD |  | 13 | 9 | 8 | 7 | 6 | 6 |
| BLUESTRIPED GOATFISH |  | 16 | 10 | 8 | 7 | 7 | 6 |
| SILVER WAREHOU |  | 15 | 11 | 9 | 7 | 7 | 6 |
| RED GURNARD |  | 15 | 11 | 9 | 7 | 7 | 6 |
| RIBALDO |  | 13 | 10 | 8 | 8 | 7 | 7 |
| TRIGGERFISH \& LEATHERJACKET |  | 17 | 12 | 10 | 8 | 7 | 7 |
| GEMFISH |  | 19 | 14 | 11 | 9 | 8 | 8 |
| SAWSHARK BASKET |  | 21 | 15 | 12 | 10 | 9 | 8 |
| SOUTHERN EAGLE RAY |  | 23 | 16 | 13 | 12 | 10 | 10 |
| Echinoderms |  | 26 | 18 | 15 | 13 | 12 | 11 |
| ELEPHANTFISH |  | 28 | 18 | 15 | 13 | 12 | 11 |
| SILVER DORY |  | 25 | 19 | 15 | 13 | 12 | 11 |
| Fish |  | 22 | 19 | 15 | 14 | 12 | 11 |
| SMOOTH OREODORY |  | 25 | 20 | 15 | 14 | 12 | 11 |
| KNIFEJAW |  | 30 | 21 | 17 | 15 | 13 | 12 |
| BLUE WAREHOU |  | 30 | 24 | 18 | 17 | 15 | 14 |
| Crustaceans |  | 27 | 21 | 18 | 17 | 17 | 16 |
| Chimaeras |  | 27 | 23 | 21 | 17 | 17 | 15 |
| DEEPWATER SHARK BASKET |  | 35 | 25 | 20 | 17 | 16 | 14 |
| Dogfish |  | 34 | 26 | 21 | 18 | 17 | 16 |
| OREO BASKET |  | 32 | 27 | 21 | 19 | 17 | 16 |
| REDFISH |  | 37 | 28 | 22 | 19 | 17 | 16 |
| KING DORY |  | 46 | 31 | 25 | 22 | 20 | 18 |
| ORNATE ANGELSHARK |  | 45 | 31 | 26 | 22 | 20 | 18 |
| DRAUGHTBOARD SHARK |  | 39 | 38 | 31 | 27 | 24 | 23 |
| AUSTRALIAN ANGELSHARK |  | 33 | 32 | 32 | 32 | 32 | 32 |
| OCEAN JACKET |  | 55 | 38 | 31 | 27 | 24 | 22 |
| Stingarees |  | 60 | 41 | 34 | 29 | 26 | 24 |
| BARRACOUTA |  | 53 | 43 | 33 | 30 | 26 | 25 |
| GUITARFISH (UNSPECIFIED) |  | 66 | 46 | 38 | 33 | 29 | 27 |
| Sharks |  | 46 | 45 | 41 | 34 | 33 | 30 |
| Rays |  | 76 | 53 | 43 | 37 | 33 | 31 |
| FROSTFISH |  | 85 | 65 | 52 | 45 | 40 | 37 |
| LATCHET |  | 165 | 115 | 93 | 80 | 72 | 66 |
| Hagfish |  | 112 | 110 | 109 | 109 | 108 | 108 |
| Number of project keys with CVs O to 20\% |  | 40 | 47 | 50 | 55 | 57 | 57 |
| Number of project keys with CVs 21 to 40\% |  | 17 | 13 | 13 | 10 | 8 | 9 |
| Number of Species above 40\% |  | 11 | 8 | 5 | 3 | 3 | 2 |
| HIGH RISK SPECIES |  |  |  |  |  |  |  |
| High Risk Sharks |  | 6 | 4 | 4 | 3 | 3 | 3 |
| High Risk Teleosts |  | 7 | 5 | 4 | 4 | 3 | 3 |
| High Risk Molluscs |  | 8 | 7 | 7 | 7 | 5 | 5 |
| High Risk Dogfish Other |  | 40 | 29 | 23 | 21 | 18 | 17 |
| High Risk Hagfish |  | 33 | 27 | 26 | 25 | 24 | 24 |
| High Risk Skates / Rays |  | 42 | 35 | 33 | 25 | 24 | 20 |
| High Risk Upper Slope Dogfish |  | 87 | 57 | 48 | 41 | 37 | 33 |

Figure 17 - Progress of CVs (expressed as percentages) for each of the 75 project keys for different sample sizes, at increments of 250 sea-days, allocated to strata using Method VWCV - which minimises the value-weighted average CV.

|  | Sea-days | CV (\%) on Total Catch: |  |  | Method Mins |  | 1500 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 250 | 500 | 750 | 1000 | 1250 |  |
| OCEAN BLUE-EYE TREVALLA |  | 0 | - | - | - | O | - |
| Sawsharks |  | 0 | O | 0 | O | O | O |
| STRIPED TRUMPETER |  | 0 | O | O | O | 0 | O |
| WHISKERY SHARK |  | 0 | 0 | 0 | 0 | 0 | 0 |
| BIGHT REDFISH |  | 1 | 0 | 0 | O | 0 | 0 |
| YELLOWSPOTTED BOARFISH |  | 2 | 1 | O | O | O | o |
| PINK LING |  | 1 | O | O | 0 | O | O |
| BLUE-EYE TREVALLA |  | 1 | 1 | 1 | 1 | 1 | 1 |
| GUMMY SHARK |  | 1 | 1 | 1 | 1 | 1 | 1 |
| ORANGE ROUGHY |  | 2 | 1 | 1 | 1 | 1 | 1 |
| HAPUKU |  | 1 | 1 | 1 | 1 | 1 | 1 |
| ROYAL RED PRAWN |  | 2 | 1 | 1 | 1 | 1 | 1 |
| BROADNOSE SHARK |  | 1 | 1 | 1 | 1 | 1 | 1 |
| ALFONSINO |  | 2 | 1 | 1 | 1 | 1 | 1 |
| JOHN DORY |  | 2 | 2 | 2 | 2 | 2 | 2 |
| BLUE GRENADIER |  | 3 | 2 | 2 | 2 | 2 | 2 |
| STARGAZER (UNSPECIFIED) |  | 2 | 2 | 2 | 2 | 2 | 2 |
| TIGER FLATHEAD |  | 3 | 2 | 2 | 2 | 2 | 2 |
| EASTERN SCHOOL WHITING |  | 3 | 3 | 2 | 2 | 2 | 2 |
| BLUE MORWONG |  | 10 | 4 | 3 | 2 | 2 | 2 |
| DEEPWATER FLATHEAD |  | 4 | 3 | 2 | 2 | 2 | 1 |
| SCHOOL SHARK |  | 3 | 3 | 3 | 3 | 3 | 3 |
| GOULD SQUID |  | 11 | 4 | 3 | 3 | 2 | 2 |
| SILVER TREVALLY |  | 5 | 3 | 3 | 3 | 3 | 3 |
| BOARFISH (UNSPECIFIED) |  | 3 | 3 | 3 | 3 | 3 | 3 |
| Molluscs |  | 4 | 4 | 4 | 4 | 4 | 4 |
| OCTOPUS |  | 7 | 4 | 4 | 4 | 4 | 4 |
| BRONZE WHALER |  | 5 | 4 | 4 | 4 | 4 | 4 |
| JACKASS MORWONG |  | 15 | 6 | 5 | 4 | 4 | 3 |
| SMOOTH HAMMERHEAD |  | 7 | 7 | 5 | 4 | 4 | 4 |
| SNAPPER |  | 7 | 7 | 5 | 5 | 5 | 5 |
| BLUESTRIPED GOATFISH |  | 8 | 8 | 5 | 5 | 5 | 5 |
| RIBALDO |  | 9 | 6 | 6 | 6 | 5 | 5 |
| SQUID (GENERAL) |  | 6 | 6 | 6 | 6 | 3 | 3 |
| INSHORE OCEAN PERCH |  | 14 | 7 | 7 | 7 | 5 | 5 |
| MIRROR DORY |  | 11 | 7 | 7 | 7 | 6 | 6 |
| SMOOTH OREODORY |  | 11 | 9 | 9 | 9 | 6 | 6 |
| RED GURNARD |  | 16 | 10 | 9 | 9 | 8 | 7 |
| SOUTHERN EAGLERAY |  | 14 | 12 | 10 | 9 | 9 | 9 |
| SILVER WAREHOU |  | 13 | 11 | 11 | 11 | 7 | 7 |
| OREO BASKET |  | 13 | 11 | 11 | 11 | 8 | 8 |
| GEMFISH |  | 14 | 11 | 11 | 11 | 10 | 10 |
| ELEPHANTFISH |  | 16 | 15 | 12 | 12 | 10 | 10 |
| Fish |  | 16 | 13 | 12 | 12 | 11 | 11 |
| Dogfish |  | 19 | 16 | 14 | 13 | 11 | 11 |
| Chimaeras |  | 14 | 14 | 14 | 14 | 13 | 13 |
| TRIGGERFISH \& LEATHERJACKET |  | 20 | 14 | 14 | 14 | 13 | 13 |
| AUSTRALIAN ANGELSHARK |  | 15 | 15 | 15 | 15 | 14 | 14 |
| BLUE WAREHOU |  | 18 | 16 | 16 | 16 | 11 | 11 |
| DRAUGHTBOARD SHARK |  | 17 | 17 | 17 | 17 | 17 | 17 |
| SILVER DORY |  | 18 | 17 | 17 | 17 | 11 | 11 |
| Sharks |  | 20 | 18 | 18 | 18 | 18 | 18 |
| SAWSHARK BASKET |  | 29 | 11 | 8 | 6 | 6 | 6 |
| BARRACOUTA |  | 24 | 17 | 17 | 17 | 13 | 13 |
| Echinoderms |  | 28 | 18 | 17 | 17 | 16 | 16 |
| DEEPWATER SHARK BASKET |  | 26 | 17 | 17 | 17 | 9 | 9 |
| Crustaceans |  | 35 | 20 | 20 | 20 | 19 | 19 |
| REDFISH |  | 27 | 20 | 20 | 20 | 16 | 16 |
| KNIFEJAW |  | 41 | 16 | 10 | 8 | 8 | 7 |
| ORNATE ANGELSHARK |  | 62 | 23 | 15 | 12 | 12 | 10 |
| OCEAN JACKET |  | 76 | 29 | 19 | 15 | 15 | 13 |
| Stingarees |  | 78 | 31 | 21 | 17 | 17 | 15 |
| KING DORY |  | 35 | 35 | 35 | 35 | 20 | 20 |
| Rays |  | 104 | 39 | 26 | 20 | 20 | 17 |
| FROSTFISH |  | 51 | 36 | 36 | 36 | 20 | 20 |
| LATCHET |  | 225 | 85 | 55 | 43 | 43 | 36 |
| Hagfish |  | 118 | 68 | 68 | 68 | 67 | 67 |
| GUITARFISH (UNSPECIFIED) |  | 101 | 72 | 72 | 72 | 72 | 72 |
| Number of project keys with CVs O to 20\% |  | 52 | 59 | 61 | 62 | 65 | 65 |
| Number of project keys with CVs 21 to 40\% |  | 7 | 6 | 4 | 3 | 0 | 1 |
| Number of Species above 40\% |  | 9 | 3 | 3 | 3 | 3 | 2 |
| HIGH RISK SPECIES |  |  |  |  |  |  |  |
| High Risk Dogfish Other |  | 4 | 3 | 3 | 2 | 2 | 2 |
| High Risk Hagfish |  | 5 | 3 | 3 | 3 | 3 | 3 |
| High Risk Molluscs |  | 8 | 7 | 7 | 7 | 7 | 7 |
| High Risk Sharks |  | 28 | 27 | 27 | 27 | 16 | 16 |
| High Risk Skates / Rays |  | 46 | 25 | 25 | 25 | 25 | 25 |
| High Risk Teleosts |  | 47 | 26 | 25 | 25 | 25 | 25 |
| High Risk Upper Slope Dogfish |  | 45 | 43 | 29 | 28 | 28 | 28 |

Figure 18 - Progress of CVs (expressed as percentages) for each of the 75 project keys for different sample sizes, at increments of 250 sea-days, allocated to strata using Method MinS - which minimises the number of commercial project keys (out of 68) that are above the target CV.

|  | Sea-days | CV (\%) on Total Catch: |  |  | Method PROP |  | 1500 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 250 | 500 | 750 | 1000 | 1250 |  |
| OCEAN BLUE-EYE TREVALLA |  | O | 0 | O | O | 0 | O |
| Sawsharks |  | O | O | O | O | O | O |
| STRIPED TRUMPETER |  | 0 | 0 | 0 | 0 | 0 | 0 |
| WHISKERY SHARK |  | 0 | 0 | 0 | 0 | 0 | 0 |
| PINK LING |  | 1 | 0 | O | 0 | 0 | 0 |
| BIGHT REDFISH |  | 1 | O | O | O | O | O |
| GUMMY SHARK |  | 1 | 1 | 1 | 0 | 0 | 0 |
| BLUE-EYE TREVALLA |  | 1 | 1 | 1 | 0 | 0 | 0 |
| HAPUKU |  | 1 | 1 | 1 | 1 | 1 | O |
| YELLOWSPOTTED BOARFISH |  | 1 | 1 | 1 | 1 | 1 | 1 |
| ROYAL RED PRAWN |  | 1 | 1 | 1 | 1 | 1 | 1 |
| ORANGEROUGHY |  | 2 | 1 | 1 | 1 | 1 | 1 |
| BROADNOSE SHARK |  | 2 | 1 | 1 | 1 | 1 | 1 |
| JOHN DORY |  | 2 | 1 | 1 | 1 | 1 | 1 |
| ALFONSINO |  | 2 | 1 | 1 | 1 | 1 | 1 |
| BLUE GRENADIER |  | 2 | 1 | 1 | 1 | 1 | 1 |
| EASTERN SCHOOL WHITING |  | 2 | 2 | 1 | 1 | 1 | 1 |
| TIGER FLATHEAD |  | 3 | 2 | 1 | 1 | 1 | 1 |
| SCHOOL SHARK |  | 3 | 2 | 2 | 1 | 1 | 1 |
| STARGAZER (UNSPECIFIED) |  | 3 | 2 | 2 | 2 | 1 | 1 |
| BOARFISH (UNSPECIFIED) |  | 3 | 2 | 2 | 2 | 2 | 2 |
| DEEPWATER FLATHEAD |  | 4 | 3 | 2 | 2 | 2 | 2 |
| OCTOPUS |  | 5 | 3 | 3 | 2 | 2 | 2 |
| SNAPPER |  | 5 | 3 | 3 | 2 | 2 | 2 |
| Molluscs |  | 4 | 3 | 3 | 2 | 2 | 2 |
| SILVER TREVALLY |  | 5 | 4 | 3 | 3 | 2 | 2 |
| SMOOTH HAMMERHEAD |  | 5 | 4 | 3 | 3 | 2 | 2 |
| BLUESTRIPED GOATFISH |  | 5 | 4 | 3 | 3 | 2 | 2 |
| BRONZE WHALER |  | 7 | 5 | 4 | 3 | 3 | 3 |
| SQUID (GENERAL) |  | 7 | 5 | 4 | 4 | 3 | 3 |
| RIBALDO |  | 8 | 6 | 5 | 4 | 4 | 3 |
| MIRROR DORY |  | 8 | 6 | 5 | 4 | 4 | 3 |
| BLUE MORWONG |  | 9 | 6 | 5 | 4 | 4 | 4 |
| GOULD SQUID |  | 10 | 7 | 6 | 5 | 4 | 4 |
| INSHORE OCEAN PERCH |  | 12 | 8 | 7 | 6 | 5 | 5 |
| SILVER WAREHOU |  | 13 | 9 | 7 | 6 | 6 | 5 |
| SMOOTH OREODORY |  | 14 | 9 | 8 | 7 | 6 | 5 |
| SOUTHERN EAGLE RAY |  | 14 | 10 | 8 | 7 | 6 | 6 |
| ELEPHANTFISH |  | 14 | 10 | 8 | 7 | 6 | 6 |
| JACKASS MORWONG |  | 14 | 10 | 8 | 7 | 6 | 6 |
| RED GURNARD |  | 17 | 12 | 10 | 8 | 7 | 7 |
| OREO BASKET |  | 17 | 12 | 10 | 9 | 8 | 7 |
| Fish |  | 17 | 12 | 10 | 9 | 8 | 7 |
| GEMFISH |  | 17 | 12 | 10 | 9 | 8 | 7 |
| Dogfish |  | 18 | 13 | 10 | 9 | 8 | 7 |
| DRAUGHTBOARD SHARK |  | 20 | 14 | 12 | 10 | 9 | 8 |
| BLUE WAREHOU |  | 21 | 14 | 12 | 10 | 9 | 8 |
| SILVER DORY |  | 22 | 15 | 13 | 11 | 10 | 9 |
| TRIGGERFISH \& LEATHERJACKET |  | 23 | 16 | 13 | 11 | 10 | 9 |
| Crustaceans |  | 23 | 16 | 13 | 12 | 10 | 9 |
| DEEPWATER SHARK BASKET |  | 25 | 17 | 14 | 12 | 11 | 10 |
| REDFISH |  | 25 | 17 | 14 | 12 | 11 | 10 |
| Chimaeras |  | 23 | 16 | 14 | 12 | 11 | 10 |
| SAWSHARK BASKET |  | 26 | 18 | 15 | 13 | 11 | 10 |
| Echinoderms |  | 28 | 20 | 16 | 14 | 12 | 11 |
| AUSTRALIAN ANGELSHARK |  | 23 | 19 | 16 | 15 | 13 | 12 |
| BARRACOUTA |  | 29 | 20 | 17 | 14 | 13 | 12 |
| Sharks |  | 28 | 20 | 17 | 15 | 13 | 12 |
| KNIFEJAW |  | 37 | 26 | 21 | 18 | 16 | 15 |
| KING DORY |  | 42 | 29 | 24 | 21 | 18 | 17 |
| FROSTFISH |  | 52 | 36 | 30 | 26 | 23 | 21 |
| ORNATE ANGELSHARK |  | 56 | 39 | 32 | 27 | 24 | 22 |
| OCEAN JACKET |  | 69 | 47 | 39 | 33 | 30 | 27 |
| Stingarees |  | 71 | 49 | 40 | 34 | 31 | 28 |
| Hagfish |  | 82 | 58 | 47 | 41 | 37 | 33 |
| Rays |  | 95 | 65 | 53 | 46 | 41 | 37 |
| GUITARFISH (UNSPECIFIED) |  | 118 | 82 | 68 | 58 | 52 | 47 |
| LATCHET |  | 206 | 141 | 116 | 99 | 89 | 81 |
| Number of project keys with CVs O to 20\% |  | 46 | 56 | 58 | 59 | 60 | 60 |
| Number of project keys with CVs 21 to 40\% |  | 13 | 6 | 5 | 5 | 5 | 6 |
| Number of Species above 40\% |  | 9 | 6 | 5 | 4 | 3 | 2 |
| HIGH RISK SPECIES |  |  |  |  |  |  |  |
| High Risk Sharks |  | 3 | 2 | 2 | 2 | 1 | 1 |
| High Risk Teleosts |  | 4 | 3 | 2 | 2 | 2 | 2 |
| High Risk Molluscs |  | 7 | 5 | 5 | 4 | 4 | 4 |
| High Risk Hagfish |  | 29 | 20 | 17 | 15 | 13 | 12 |
| High Risk Upper Slope Dogfish |  | 31 | 22 | 17 | 15 | 14 | 12 |
| High Risk Dogfish Other |  | 33 | 23 | 19 | 16 | 15 | 13 |
| High Risk Skates / Rays |  | 33 | 23 | 19 | 17 | 15 | 13 |

Figure 19 - Progress of CVs (expressed as percentages) for each of the 75 project keys for different sample sizes, at increments of 250 sea-days, allocated to strata using Method Prop - which allocates sea-days to strata in proportion to the size of the fishery.

|  | Sea-days | CV (\%) on Total Catch: |  |  | Method CWCV-HR |  | 1500 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 250 | 500 | 750 | 1000 | 1250 |  |
| OCEAN BLUE-EYE TREVALLA |  | 0 | 0 | 0 | 0 | 0 | 0 |
| Sawsharks |  | o | O | - | - | - | o |
| STRIPED TRUMPETER |  | O | O | 0 | 0 | 0 | 0 |
| WHISKERY SHARK |  | 1 | 1 | 0 | 0 | 0 | 0 |
| PINK LING |  | 1 | 1 | 1 | 1 | O | 0 |
| BIGHT REDFISH |  | 2 | 1 | 1 | 1 | 1 | 1 |
| HAPUKU |  | 1 | 1 | 1 | 1 | 1 | 1 |
| BOARFISH (UNSPECIFIED) |  | 2 | 2 | 1 | 1 | 1 | 1 |
| BLUE-EYE TREVALLA |  | 2 | 2 | 1 | 1 | 1 | 1 |
| ROYAL RED PRAWN |  | 2 | 1 | 1 | 1 | 1 | 1 |
| BLUE GRENADIER |  | 3 | 2 | 2 | 1 | 1 | 1 |
| GUMMY SHARK |  | 2 | 2 | 1 | 1 | 1 | 1 |
| ALFONSINO |  | 3 | 2 | 2 | 1 | 1 | 1 |
| EASTERN SCHOOL WHITING |  | 3 | 2 | 2 | 2 | 1 | 1 |
| ORANGEROUGHY |  | 3 | 3 | 2 | 2 | 2 | 2 |
| YELLOWSPOTTED BOARFISH |  | 4 | 3 | 2 | 2 | 2 | 1 |
| SQUID (GENERAL) |  | 4 | 3 | 2 | 2 | 2 | 2 |
| JOHN DORY |  | 4 | 3 | 2 | 2 | 2 | 2 |
| BROADNOSE SHARK |  | 2 | 2 | 2 | 2 | 2 | 2 |
| STARGAZER (UNSPECIFIED) |  | 5 | 3 | 3 | 2 | 2 | 2 |
| Molluscs |  | 7 | 4 | 4 | 3 | 3 | 3 |
| OCTOPUS |  | 8 | 5 | 4 | 4 | 3 | 3 |
| BLUESTRIPED GOATFISH |  | 8 | 5 | 4 | 4 | 3 | 3 |
| TIGER FLATHEAD |  | 9 | 5 | 4 | 4 | 3 | 3 |
| SNAPPER |  | 8 | 5 | 4 | 4 | 3 | 3 |
| DEEPWATER FLATHEAD |  | 8 | 6 | 5 | 4 | 4 | 3 |
| SMOOTH OREODORY |  | 8 | 6 | 5 | 4 | 4 | 4 |
| SMOOTH HAMMERHEAD |  | 7 | 5 | 5 | 4 | 4 | 4 |
| SCHOOL SHARK |  | 5 | 5 | 5 | 5 | 5 | 5 |
| RIBALDO |  | 10 | 7 | 6 | 5 | 5 | 4 |
| ELEPHANTFISH |  | 13 | 9 | 8 | 7 | 6 | 5 |
| OREO BASKET |  | 12 | 9 | 7 | 7 | 6 | 5 |
| MIRROR DORY |  | 16 | 11 | 9 | 8 | 7 | 6 |
| SILVER TREVALLY |  | 19 | 11 | 10 | 9 | 7 | 7 |
| KING DORY |  | 18 | 12 | 10 | 9 | 8 | 7 |
| SILVER WAREHOU |  | 18 | 13 | 11 | 9 | 8 | 7 |
| GEMFISH |  | 24 | 17 | 13 | 11 | 10 | 9 |
| BLUE MORWONG |  | 22 | 15 | 13 | 11 | 10 | 9 |
| BRONZE WHALER |  | 25 | 14 | 12 | 11 | 9 | 9 |
| GOULD SQUID |  | 24 | 17 | 14 | 12 | 11 | 10 |
| INSHORE OCEAN PERCH |  | 28 | 19 | 15 | 13 | 12 | 11 |
| BLUE WAREHOU |  | 32 | 19 | 16 | 15 | 12 | 12 |
| Fish |  | 26 | 21 | 17 | 16 | 14 | 14 |
| SILVER DORY |  | 37 | 22 | 19 | 17 | 14 | 13 |
| JACKASS MORWONG |  | 37 | 25 | 21 | 18 | 16 | 15 |
| Dogfish |  | 36 | 25 | 20 | 18 | 16 | 14 |
| Chimaeras |  | 35 | 26 | 21 | 19 | 18 | 17 |
| SOUTHERN EAGLERAY |  | 43 | 25 | 22 | 19 | 16 | 15 |
| Crustaceans |  | 36 | 26 | 23 | 21 | 19 | 18 |
| RED GURNARD |  | 52 | 32 | 27 | 24 | 21 | 19 |
| DRAUGHTBOARD SHARK |  | 39 | 38 | 29 | 29 | 26 | 24 |
| AUSTRALIAN ANGELSHARK |  | 33 | 32 | 32 | 32 | 32 | 32 |
| BARRACOUTA |  | 55 | 34 | 29 | 25 | 22 | 20 |
| REDFISH |  | 58 | 35 | 30 | 26 | 23 | 21 |
| Sharks |  | 52 | 41 | 32 | 30 | 26 | 25 |
| SAWSHARK BASKET |  | 63 | 45 | 36 | 32 | 28 | 26 |
| DEEPWATER SHARK BASKET |  | 69 | 49 | 40 | 35 | 31 | 28 |
| TRIGGERFISH \& LEATHERJACKET |  | 91 | 54 | 46 | 41 | 35 | 33 |
| Echinoderms |  | 102 | 61 | 52 | 47 | 40 | 37 |
| KNIFEJAW |  | 92 | 65 | 53 | 46 | 41 | 37 |
| FROSTFISH |  | 103 | 73 | 59 | 51 | 46 | 42 |
| ORNATE ANGELSHARK |  | 138 | 98 | 80 | 69 | 62 | 56 |
| OCEAN JACKET |  | 170 | 120 | 98 | 85 | 76 | 69 |
| Stingarees |  | 177 | 124 | 102 | 88 | 79 | 72 |
| Hagfish |  | 118 | 112 | 111 | 110 | 109 | 109 |
| Rays |  | 233 | 164 | 134 | 116 | 104 | 95 |
| GUITARFISH (UNSPECIFIED) |  | 486 | 283 | 245 | 220 | 187 | 175 |
| LATCHET |  | 503 | 356 | 291 | 252 | 225 | 206 |
| Number of project keys with CVs O to 20\% |  | 36 | 42 | 44 | 48 | 49 | 50 |
| Number of project keys with CVs 21 to 40\% |  | 14 | 12 | 13 | 9 | 10 | 10 |
| Number of Species above 40\% |  | 18 | 14 | 11 | 11 | 9 | 8 |
| HIGH RISK SPECIES |  |  |  |  |  |  |  |
| High Risk Teleosts |  | 4 | 3 | 3 | 2 | 2 | 2 |
| High Risk Sharks |  | 9 | 5 | 4 | 4 | 3 | 3 |
| High Risk Molluscs |  | 9 | 6 | 5 | 5 | 4 | 4 |
| High Risk Dogfish Other |  | 16 | 11 | 9 | 8 | 7 | 6 |
| High Risk Upper Slope Dogfish |  | 44 | 31 | 25 | 22 | 20 | 18 |
| High Risk Hagfish |  | 46 | 34 | 31 | 29 | 28 | 27 |
| High Risk Skates/Rays |  | 62 | 41 | 34 | 29 | 26 | 24 |

Figure 20 - Progress of CVs (expressed as percentages) for each of the 75 project keys for different sample sizes, at increments of 250 sea-days, allocated to strata using Method CWCV-HR - which minimises the catch-weighted average CV of the 7 HIGH RISK project keys.

|  | Sea-days | CV (\%) on Total Catch: |  |  | Method Mins-HR |  | 1500 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 250 | 500 | 750 | 1000 | 1250 |  |
| OCEAN BLUE-EYE TREVALLA |  | 0 | 0 | O | O | 0 | O |
| Sawsharks |  | O | O | O | O | O | O |
| STRIPED TRUMPETER |  | 0 | 0 | 0 | 0 | O | 0 |
| WHISKERY SHARK |  | 0 | 0 | 0 | 0 | 0 | 0 |
| BIGHT REDFISH |  | 1 | 0 | 0 | 0 | 0 | 0 |
| YELLOWSPOTTED BOARFISH |  | 1 | 1 | O | O | O | O |
| PINK LING |  | 1 | 0 | 0 | 0 | 0 | 0 |
| HAPUKU |  | 3 | 1 | 1 | 1 | 1 | 1 |
| BLUE-EYE TREVALLA |  | 1 | 1 | 1 | 1 | 1 | 1 |
| GUMMY SHARK |  | 1 | 1 | 1 | 1 | 1 | 1 |
| ORANGEROUGHY |  | 2 | 1 | 1 | 1 | 1 | 1 |
| ALFONSINO |  | 4 | 1 | 1 | 1 | 1 | 1 |
| BLUE GRENADIER |  | 3 | 1 | 1 | 1 | 1 | 1 |
| ROYAL RED PRAWN |  | 2 | 1 | 1 | 1 | 1 | 1 |
| JOHN DORY |  | 5 | 2 | 2 | 1 | 1 | 1 |
| BROADNOSE SHARK |  | 2 | 2 | 2 | 2 | 2 | 2 |
| DEEPWATER FLATHEAD |  | 8 | 2 | 2 | 2 | 2 | 1 |
| EASTERN SCHOOL WHITING |  | 3 | 2 | 2 | 2 | 2 | 2 |
| TIGER FLATHEAD |  | 9 | 4 | 2 | 2 | 2 | 2 |
| STARGAZER (UNSPECIFIED) |  | 4 | 2 | 2 | 2 | 2 | 2 |
| OCTOPUS |  | 3 | 3 | 3 | 2 | 2 | 2 |
| GOULD SQUID |  | 9 | 5 | 3 | 3 | 2 | 2 |
| BOARFISH (UNSPECIFIED) |  | 7 | 3 | 3 | 3 | 3 | 3 |
| SCHOOL SHARK |  | 5 | 3 | 3 | 3 | 3 | 3 |
| SILVER TREVALLY |  | 18 | 7 | 5 | 3 | 3 | 3 |
| BLUE MORWONG |  | 7 | 5 | 4 | 3 | 2 | 2 |
| BLUESTRIPED GOATFISH |  | 4 | 3 | 3 | 3 | 3 | 3 |
| SNAPPER |  | 8 | 4 | 4 | 3 | 3 | 3 |
| Molluscs |  | 7 | 4 | 4 | 4 | 4 | 4 |
| SMOOTH HAMMERHEAD |  | 5 | 4 | 4 | 4 | 3 | 3 |
| RIBALDO |  | 13 | 6 | 6 | 5 | 4 | 4 |
| MIRROR DORY |  | 15 | 11 | 10 | 7 | 5 | 5 |
| GEMFISH |  | 18 | 12 | 12 | 11 | 10 | 10 |
| SQUID (GENERAL) |  | 21 | 4 | 4 | 4 | 4 | 4 |
| JACKASS MORWONG |  | 20 | 10 | 6 | 5 | 4 | 4 |
| BRONZE WHALER |  | 25 | 10 | 7 | 5 | 4 | 4 |
| SAWSHARK BASKET |  | 20 | 14 | 8 | 6 | 6 | 5 |
| INSHORE OCEAN PERCH |  | 21 | 9 | 8 | 7 | 5 | 4 |
| SILVER WAREHOU |  | 34 | 11 | 11 | 7 | 6 | 6 |
| ELEPHANTFISH |  | 31 | 8 | 8 | 8 | 8 | 8 |
| SMOOTH OREODORY |  | 28 | 10 | 10 | 9 | 7 | 6 |
| KNIFEJAW |  | 29 | 20 | 11 | 9 | 8 | 7 |
| OREO BASKET |  | 30 | 14 | 14 | 12 | 10 | 9 |
| Crustaceans |  | 29 | 14 | 13 | 12 | 12 | 12 |
| Fish |  | 36 | 18 | 16 | 14 | 13 | 13 |
| Dogfish |  | 38 | 19 | 17 | 16 | 11 | 10 |
| Chimaeras |  | 24 | 18 | 18 | 17 | 15 | 15 |
| DRAUGHTBOARD SHARK |  | 30 | 18 | 18 | 18 | 18 | 18 |
| Sharks |  | 29 | 19 | 19 | 19 | 19 | 19 |
| RED GURNARD |  | 48 | 18 | 13 | 9 | 8 | 8 |
| SOUTHERN EAGLERAY |  | 44 | 17 | 12 | 9 | 9 | 9 |
| SILVER DORY |  | 66 | 18 | 14 | 12 | 12 | 11 |
| BLUE WAREHOU |  | 58 | 19 | 16 | 13 | 12 | 12 |
| KING DORY |  | 121 | 20 | 20 | 20 | 20 | 20 |
| ORNATE ANGELSHARK |  | 44 | 30 | 17 | 13 | 12 | 10 |
| TRIGGERFISH \& LEATHERJACKET |  | 92 | 37 | 26 | 17 | 16 | 16 |
| OCEAN JACKET |  | 60 | 38 | 21 | 17 | 15 | 14 |
| DEEPWATER SHARK BASKET |  | 40 | 40 | 40 | 19 | 11 | 9 |
| BARRACOUTA |  | 63 | 30 | 27 | 20 | 16 | 14 |
| Stingarees |  | 69 | 40 | 24 | 18 | 17 | 15 |
| Echinoderms |  | 97 | 41 | 30 | 19 | 17 | 17 |
| AUSTRALIAN ANGELSHARK |  | 36 | 32 | 32 | 32 | 18 | 18 |
| REDFISH |  | 79 | 37 | 28 | 26 | 17 | 17 |
| Rays |  | 74 | 50 | 28 | 22 | 20 | 18 |
| FROSTFISH |  | 86 | 68 | 68 | 46 | 23 | 20 |
| LATCHET |  | 160 | 108 | 60 | 48 | 43 | 38 |
| Hagfish |  | 48 | 47 | 47 | 47 | 46 | 46 |
| GUITARFISH (UNSPECIFIED) |  | 487 | 186 | 125 | 84 | 84 | 84 |
| Number of project keys with CVs O to 20\% |  | 33 | 54 | 55 | 61 | 64 | 65 |
| Number of project keys with CVs 21 to 40\% |  | 17 | 7 | 9 | 3 | 1 | 1 |
| Number of Species above 40\% |  | 18 | 7 | 4 | 4 | 3 | 2 |
| HIGH RISK SPECIES |  |  |  |  |  |  |  |
| High Risk Sharks |  | 8 | 4 | 3 | 2 | 2 | 2 |
| High Risk Teleosts |  | 12 | 6 | 4 | 4 | 3 | 3 |
| High Risk Molluscs |  | 15 | 9 | 8 | 7 | 7 | 7 |
| High Risk Hagfish |  | 16 | 15 | 15 | 15 | 15 | 15 |
| High Risk Skates/Rays |  | 19 | 17 | 16 | 15 | 15 | 15 |
| High Risk Upper Slope Dogfish |  | 25 | 19 | 19 | 19 | 19 | 19 |
| High Risk Dogfish Other |  | 93 | 18 | 18 | 18 | 16 | 16 |

Figure 21 - Progress of CVs (expressed as percentages) for each of the 75 project keys for different sample sizes, at increments of 250 sea-days, allocated to strata using Method MinS-HR - which minimises the number of HIGH RISK project keys (out of 7) that are above the target CV.


Figure 22. Three different tradeoffs between number of redfish sampled per landing (trip) and the total number of landings sampled. Each trade-off represents a different total number of fish sampled per year. The 2008 data are used and the triangle represents the 2008 estimate, while the central curve is for the total number of fish sampled in 2008. The upper curve involves twice the number of fish and the lower curve half the number of fish sampled per year.


Figure 23. Three different tradeoffs between number of elephant fish sampled per landing (trip) and the total number of landings sampled.


Figure 24. Three different tradeoffs between number of gemfish sampled per landing (trip) and the total number of landings sampled.


Figure 25. Three different tradeoffs between number of blue grenadier sampled per landing (trip) and the total number of landings sampled.


Figure 26. Three different tradeoffs between number of school shark sampled per landing (trip) and the total number of landings sampled.


Figure 27. Three different tradeoffs between number of gummy shark sampled per landing (trip) and the total number of landings sampled.


Figure 28. Three different tradeoffs between number of eastern school whiting sampled per landing (trip) and the total number of landings sampled. 2007 data were used as the species was more intensively sampled.


Figure 29. Three different tradeoffs between number of tiger flathead sampled per landing (trip) and the total number of landings sampled. 2007 data were used as the species was more intensively sampled. 2007 data where the species was more intensively sampled.


Figure 30. Three different tradeoffs between number of pink ling sampled per landing (trip) and the total number of landings sampled.


Figure 31. Three different tradeoffs between number of mirror dory sampled per landing (trip) and the total number of landings sampled.


Figure 32. Three different tradeoffs between number of John dory sampled per landing (trip) and the total number of landings sampled.


Figure 33. Three different tradeoffs between number of jackass morwong sampled per landing (trip) and the total number of landings sampled.


Figure 34. Three different tradeoffs between number of blue-eye trevalla sampled per landing (trip) and the total number of landings sampled.


Figure 35. Three different tradeoffs between number of alfonsino sampled per landing (trip) and the total number of landings sampled.


Figure 36. Plot of the variance versus the mean number for individuals interacting with the fishing gear, where the level of aggregation for the calculation of means and variances was year and stratum. This plot is for the TEP group 'Seals'.


Figure 37. Plot of the variance versus the mean number for individuals interacting with the fishing gear, where the level of aggregation for the calculation of means and variances was year and stratum. This plot is for the TEP group 'Seals'. This plot is a subset of the earlier plot for 'Seals' (Fig. M5a) where the high point is omitted.


Figure 38. Plot of the variance versus the mean number for individuals interacting with the fishing gear, where the level of aggregation for the calculation of means and variances was year and stratum. This plot is for the TEP group 'Pipefish'.


Figure 39. Plot of the variance versus the mean number for individuals interacting with the fishing gear, where the level of aggregation for the calculation of means and variances was year and stratum. This plot is for the TEP group 'Birds'.


Figure 40. Plot of the variance versus the mean number for individuals interacting with the fishing gear, where the level of aggregation for the calculation of means and variances was year and stratum. This plot is for the TEP group 'Birds'. This plot is a subset of the earlier plot (Fig. M7a) for 'Birds' where the high point is omitted.


Figure 41. Plot of the variance versus the mean number for individuals interacting with the fishing gear, where the level of aggregation for the calculation of means and variances was year and stratum. This plot is for the TEP group 'Sharks'.


Figure 42. Plot of the variance versus the mean number for individuals interacting with the fishing gear, where the level of aggregation for the calculation of means and variances was year and stratum. This plot is for the TEP group 'Whales / Dolphins'.


Figure 43. Average optimal shot allocation across TEP species groups, compared to the shot allocation proportional to the number of shots per stratum for 2007 and 2008.


Figure 44. Average optimal shot allocation across TEP species groups, compared to the shot allocation proportional to the number of shots per stratum for 2007 and 2008. These values correspond to Table M6d, i.e. they make some provision in the design optimisation for an underlying true interaction in strata where no interaction was recorded over the period 2002 to 2008 . In this output fractional shots are permitted.

Table 1 - Average annual number of shots fired in each of the 32 strata, as obtained from GENLOG data and average annual number of shots observed by the ISMP for years 2002 to 2008. Both of these quantities are also expressed as proportions. Also shown is the average number of shots recorded in the GENLOG per sea-day for each stratum.

| Code | Stratum | GENLOG <br> shots per <br> year | ISMP <br> shots <br> per <br> year | Proportion <br> of shots <br> (GENLOG) | Proportio <br> nof shots <br> (ISMP) | Average <br> number of <br> shots per <br> sea-day |
| :--- | :--- | :---: | :---: | :---: | :---: | :---: |
| 1 | VIT | 306 |  | $0.9 \%$ |  | 3.6 |
| 2 | BS IN TR | 107 | 16 | $0.3 \%$ | $1.2 \%$ | 3.5 |
| 3 | ECDW TR | 45 | 49 | $0.1 \%$ | $3.8 \%$ | 2.4 |
| 4 | EDL DS | 6400 | 69 | $18.5 \%$ | $5.3 \%$ | 4.6 |
| 5 | EDL IN TR | 3189 | 93 | $9.2 \%$ | $7.2 \%$ | 3.1 |
| 6 | EDL OFF TR | 1567 | 64 | $4.5 \%$ | $4.9 \%$ | 2.7 |
| 7 | NSW IN TR | 2862 | 214 | $8.3 \%$ | $16.5 \%$ | 2.5 |
| 8 | NSW OFF TR | 1081 | 75 | $3.1 \%$ | $5.8 \%$ | 2.1 |
| 9 | NSW RRP TR | 178 | 10 | $0.5 \%$ | $0.8 \%$ | 1.8 |
| 10 | SW BGS TR | 258 | 21 | $0.7 \%$ | $1.6 \%$ | 2.0 |
| 11 | SW ORO TR | 8 | 3 | $0.0 \%$ | $0.2 \%$ | 2.3 |
| 12 | SW TR | 2501 | 120 | $7.2 \%$ | $9.3 \%$ | 2.6 |
| 13 | TAS BGS TR | 325 | 16 | $0.9 \%$ | $1.2 \%$ | 2.5 |
| 14 | TAS E TR | 1292 | 39 | $3.7 \%$ | $3.0 \%$ | 3.0 |
| 15 | TAS ORO TR | 95 | 18 | $0.3 \%$ | $1.4 \%$ | 2.1 |
| 16 | TAS W TR | 1156 | 25 | $3.3 \%$ | $1.9 \%$ | 2.5 |
| 17 | AL CSA | 436 | 45 | $1.3 \%$ | $3.4 \%$ | 1.7 |
| 18 | AL EBS / | 233 | 82 | $0.7 \%$ | $6.3 \%$ | 1.2 |
| 19 | AL ESA | 67 | 4 | $0.2 \%$ | $0.3 \%$ | 1.5 |
| 20 | AL ET / WT | 319 | 52 | $0.9 \%$ | $4.0 \%$ | 1.1 |
| 21 | AL WBS / | 119 | 37 | $0.3 \%$ | $2.8 \%$ | 1.3 |
| 22 | AL WSA / WA | 117 | 72 | $0.3 \%$ | $5.5 \%$ | 1.5 |
| 23 | GN CSA | 2183 | 4 | $6.3 \%$ | $0.3 \%$ | 2.2 |
| 24 | GN EBS / | 2650 |  | $7.7 \%$ |  | 1.8 |
| 25 | GN ESA | 259 |  | $0.7 \%$ |  | 1.9 |
| 26 | GN ET / WT | 621 |  | $1.8 \%$ |  | 1.4 |
| 27 | GN SAV | 232 |  | $0.7 \%$ |  | 1.6 |
| 28 | GN WA/ WSA | 1559 |  | $4.5 \%$ |  | 2.4 |
| 29 | GN WBS | 672 |  | $1.9 \%$ |  | 1.6 |
| 30 | Inshore | 3446 | 131 | $10.0 \%$ | $10.1 \%$ | 3.3 |
| 31 | Midshore | 193 | 19 | $0.6 \%$ | $1.5 \%$ | 2.0 |
| 32 | Offshore | 102 | 20 | $0.3 \%$ | $1.5 \%$ | 1.7 |
|  |  |  |  |  |  |  |

Table 2 - Assumed price in \$A per kg for purposes of value weighting of species in the value-weighted average $\mathbf{C V}$ - which is used as the objective function in one of the optimization procedures.

|  | $\$$ A per kg |
| :--- | :---: |
| John dory | 9.83 |
| King dory | 9.83 |
| Blue-eye trevalla | 9.42 |
| Ocean blue-eye trevalla | 9.42 |
| Gummy shark | 7.01 |
| Orange roughy | 6.13 |
| Silver trevally | 5.23 |
| Inshore ocean perch | 5.07 |
| Pink ling | 5.03 |
| Alfonsino | 4.66 |
| Gemfish | 4.34 |
| Tiger flathead | 4.24 |
| Oreo basket | 3.95 |
| Blue grenadier | 3.44 |
| Eastern school whiting | 3.26 |
| Smooth oreodory | 3.11 |
| Jackass morwong | 2.86 |
| Bight redfish | 2.72 |
| Redfish | 2.72 |
| Mirror dory | 2.57 |
| Silver dory | 2.57 |
| Sawshark basket | 2.25 |
| Blue warehou | 2.17 |
| Silver warehou | 1.91 |
| Squid | 1.75 |
| All other species | 1 |
|  |  |

Table 3 - Estimate of discarded catch and associated CVs for each of the 75 project keys for years 2002 to 2006. Only 27 of the 32 strata are included in this analysis. Excluded strata are the VIT and GN strata.

|  | RETAINED | ESTIMATE OF DISCARDE D CATCH | S.E. OF DISCARDS | EStIMATE OF TOTAL CATCH | DISCARDED PROPORTION | $\begin{gathered} \text { CV } \\ \text { (DISCARDS) } \end{gathered}$ | $\begin{aligned} & \text { CV (TOTAL } \\ & \text { CATCH) } \\ & \hline \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Alfonsino | 214,911 | 9,085 | 2,274 | 223,997 | 4\% | 25.0\% | 1.0\% |
| Australian angelshark | 179,549 | 71,731 | 25,993 | 251,281 | 29\% | 36.2\% | 10.3\% |
| Barracouta | 171,783 | 591,890 | 191,826 | 763,673 | 78\% | 32.4\% | 25.1\% |
| Bight redfish | 1,006,150 | 14,482 | 2,148 | 1,020,632 | 1\% | 14.8\% | 0.2\% |
| Blue grenadier | 6,444,957 | 392,465 | 102,615 | 6,837,422 | 6\% | 26.1\% | 1.5\% |
| Blue morwong | 83,696 | 61,107 | 6,584 | 144,802 | 42\% | 10.8\% | 4.5\% |
| Blue warehou | 336,987 | 361,750 | 104,732 | 698,737 | 52\% | 29.0\% | 15.0\% |
| Blue-eye trevalla | 640,453 | 9,996 | 648 | 650,449 | 2\% | 6.5\% | 0.1\% |
| Bluestriped goatfish | 29,133 | 6,058 | 1,314 | 35,192 | 17\% | 21.7\% | 3.7\% |
| Boarfish (unspecified) | 56,833 | 9,622 | 1,240 | 66,455 | 14\% | 12.9\% | 1.9\% |
| Broadnose shark | 48,998 | 11,038 | 377 | 60,037 | 18\% | 3.4\% | 0.6\% |
| Bronze whaler | 26,131 | 26,411 | 3,919 | 52,542 | 50\% | 14.8\% | 7.5\% |
| Chimaeras | 25,004 | 13,836 | 2,273 | 38,840 | 36\% | 16.4\% | 5.9\% |
| Crustaceans | 47,800 | 57,565 | 21,600 | 105,365 | 55\% | 37.5\% | 20.5\% |
| Deepwater flathead | 1,897,768 | 101,426 | 15,212 | 1,999,195 | 5\% | 15.0\% | 0.8\% |
| Deepwater shark basket | 313,080 | 128,494 | 84,615 | 441,574 | 29\% | 65.9\% | 19.2\% |
| Dogfish | 93,887 | 243,444 | 76,478 | 337,331 | 72\% | 31.4\% | 22.7\% |
| Draughtboard shark | 36,298 | 91,493 | 9,865 | 127,791 | 72\% | 10.8\% | 7.7\% |
| Eastern school whiting | 534,215 | 49,374 | 11,510 | 583,589 | 8\% | 23.3\% | 2.0\% |
| Echinoderms | - | 33,670 | 6,776 | 33,670 | 100\% | 20.1\% | 20.1\% |
| Elephantfish | 112,448 | 128,266 | 156,872 | 240,715 | 53\% | 122.3\% | 65.2\% |
| Fish | 592,139 | 2,372,557 | 126,352 | 2,964,697 | 80\% | 5.3\% | 4.3\% |
| Frostfish | 151,761 | 511,716 | 226,991 | 663,476 | 77\% | 44.4\% | 34.2\% |
| Gemfish | 740,981 | 310,627 | 60,536 | 1,051,608 | 30\% | 19.5\% | 5.8\% |
| Gould squid | 1,031,701 | 119,739 | 15,821 | 1,151,440 | 10\% | 13.2\% | 1.4\% |
| Guitarfish (unspecified) | 68,723 | 676,987 | 617,975 | 745,710 | 91\% | 91.3\% | 82.9\% |
| Gummy shark | 2,428,197 | 865,705 | 10,784 | 3,293,901 | 26\% | 1.2\% | 0.3\% |
| Hagfish | - | 572 | 280 | 572 | 100\% | 49.0\% | 49.0\% |
| Hapuku | 165,150 | 8,602 | 940 | 173,752 | 5\% | 10.9\% | 0.5\% |
| Inshore ocean perch | 377,668 | 205,827 | 49,677 | 583,496 | 35\% | 24.1\% | 8.5\% |
| Jackass morwong | 1,194,220 | 363,997 | 57,758 | 1,558,217 | 23\% | 15.9\% | 3.7\% |
| John dory | 154,713 | 15,622 | 1,414 | 170,336 | 9\% | 9.1\% | 0.8\% |
| King dory | 295,878 | 86,391 | 87,289 | 382,268 | 23\% | 101.0\% | 22.8\% |
| Knifejaw | 116,888 | 41,095 | 4,107 | 157,983 | 26\% | 10.0\% | 2.6\% |
| Latchet | 324,043 | 1,095,845 | 152,822 | 1,419,888 | 77\% | 13.9\% | 10.8\% |
| Mirror dory | 544,068 | 200,756 | 39,896 | 744,824 | 27\% | 19.9\% | 5.4\% |
| Molluscs | 171,510 | 29,017 | 3,974 | 200,527 | 14\% | 13.7\% | 2.0\% |
| Ocean blue-eye trevalla | 1,601 | - | - | 1,601 | 0\% | 0.0\% | 0.0\% |
| Ocean jacket | 714,251 | 542,984 | 122,117 | 1,257,235 | 43\% | 22.50\% | 9.70\% |
| Octopus | 81,272 | 22,592 | 3,814 | 103,864 | 22\% | 16.90\% | 3.70\% |
| Orange roughy | 1,903,671 | 38,016 | 24,328 | 1,941,687 | 2\% | 64.00\% | 1.30\% |
| Oreo basket | 217,103 | 114,259 | 114,625 | 331,362 | 34\% | 100.30\% | 34.60\% |
| Ornate angelshark | 163,587 | 83,421 | 9,263 | 247,007 | 34\% | 11.10\% | 3.80\% |

Table continues over page...

## Table 3 Continued...

|  | RETAINED | ESTIMATE OF DISCARDED CATCH | S.E. OF DISCARDS | ESTIMATE OF TOTAL CATCH | DISCARDED PROPORTION | CV <br> (DISCARDS) | CV (TOTAL CATCH) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Pink ling | 1,791,126 | 68,778 | 5,739 | 1,859,904 | 4\% | 8.30\% | 0.30\% |
| Rays | 91,588 | 452,075 | 54,217 | 543,663 | 83\% | 12.00\% | 10.00\% |
| Red gurnard | 348,939 | 277,205 | 60,723 | 626,144 | 44\% | 21.90\% | 9.70\% |
| Redfish | 402,305 | 355,210 | 283,632 | 757,515 | 47\% | 79.80\% | 37.40\% |
| Ribaldo | 225,755 | 58,303 | 18,091 | 284,058 | 21\% | 31.00\% | 6.40\% |
| Royal red prawn | 184,400 | 7,037 | 1,832 | 191,436 | 4\% | 26.00\% | 1.00\% |
| Sawshark basket | 462,336 | 159,585 | 21,793 | 621,921 | 26\% | 13.70\% | 3.50\% |
| Sawsharks | 8,499 | 95 | - | 8,593 | 1\% | 0.00\% | 0.00\% |
| School shark | 235,863 | 38,042 | 2,211 | 273,905 | 14\% | 5.80\% | 0.80\% |
| Sharks | 47,950 | 53,007 | 3,705 | 100,957 | 53\% | 7.00\% | 3.70\% |
| Silver dory | 103,929 | 133,248 | 28,273 | 237,176 | 56\% | 21.20\% | 11.90\% |
| Silver trevally | 167,200 | 36,551 | 6,932 | 203,752 | 18\% | 19.00\% | 3.40\% |
| Silver warehou | 3,279,278 | 748,531 | 215,474 | 4,027,809 | 19\% | 28.80\% | 5.30\% |
| Smooth hammerhead | 9,814 | 3,273 | 778 | 13,087 | 25\% | 23.80\% | 5.90\% |
| Smooth oreodory | 263,587 | 88,449 | 39,804 | 352,035 | 25\% | 45.00\% | 11.30\% |
| Snapper | 112,362 | 43,331 | 9,213 | 155,693 | 28\% | 21.30\% | 5.90\% |
| Southern eagle ray | 24,220 | 23,503 | 6,005 | 47,724 | 49\% | 25.50\% | 12.60\% |
| Squid (general) | 670,546 | 224,760 | 39,267 | 895,306 | 25\% | 17.50\% | 4.40\% |
| Stargazer (unspecified) | 288,201 | 24,983 | 4,548 | 313,184 | 8\% | 18.20\% | 1.50\% |
| Stingarees | - | 484,373 | 88,133 | 484,373 | 100\% | 18.20\% | 18.20\% |
| Striped trumpeter | 7,324 | - | - | 7,324 | 0\% | 0.00\% | 0.00\% |
| Tiger flathead | 4,008,944 | 760,303 | 98,773 | 4,769,247 | 16\% | 13.00\% | 2.10\% |
| Triggerfish \& leatherjacket | 130,090 | 75,917 | 35,872 | 206,007 | 37\% | 47.30\% | 17.40\% |
| Whiskery shark | 35,058 | 4,887 | 95 | 39,945 | 12\% | 2.00\% | 0.20\% |
| Yellowspotted boarfish | 159,561 | 10,242 | 1,686 | 169,803 | 6\% | 16.50\% | 1.00\% |
| High risk dogfish other | 191,289 | 498,834 | 144,982 | 690,124 | 72\% | 29.10\% | 21.00\% |
| High risk hagfish | - | 2,711 | 506 | 2,711 | 100\% | 18.70\% | 18.70\% |
| High risk molluscs | 180,758 | 51,162 | 8,383 | 231,920 | 22\% | 16.40\% | 3.60\% |
| High risk sharks | 56,197 | 54,980 | 3,639 | 111,177 | 49\% | 6.60\% | 3.30\% |
| High risk skates / rays | - | 48,629 | 10,063 | 48,629 | 100\% | 20.70\% | 20.70\% |
| High risk teleosts | 629,000 | 251,623 | 25,535 | 880,622 | 29\% | 10.10\% | 2.90\% |
| High risk upper slope | 7,788 | 70,400 | 29,978 | 78,188 | 90\% | 42.60\% | 38.30\% |
| AVERAGE CV ON TOTAL CATCH |  |  |  |  |  |  | 10.3\% |
| VALUE-WEIGHTED CV ON TOTAL CATCH |  |  |  |  |  |  | 9.6\% |

Table 4-Sample size distribution across strata for a total sample size of 500 observer sea-days by Methods: VWCV (which minimises the value-weighted average CV); MinS (which minimises the number of project keys with CVs above target); Prop (which allocates sampling effort in proportion to the size of the fishery in each stratum); CWCV-HR which minimises the catch-weighted average CV for the high risk species groupings and MinS-HR which minimizes the number of high risk species groups above target. Distributions are shown in terms of sea-days and in terms of the estimated equivalent number of shots sampled.

Total Sample Size = 500 observer seadays.

|  | Number of Sea-Days |  |  |  |  | Number of Shots |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Method VWCV | Method MinS | Method Prop | Method CWCV HR | Method MinS HR | Method VWCV | Method Mins | Metho <br> d Prop | Method CWCV HR | Method MinS HR |
| VIT | 3 | 3 | 3 | 3 | 3 | 11 | 11 | 11 | 11 | 11 |
| BS_IN_TR | 1 | 1 | 2 | 3 | 1 | 3 | 3 | 7 | 10 | 3 |
| ECDW_TR | 1 | 1 | 1 | 1 | 1 | 2 | 2 | 2 | 2 | 2 |
| EDL_DS | 7 | 12 | 50 | 25 | 64 | 32 | 55 | 229 | 115 | 293 |
| EDL_IN_TR | 156 | 55 | 37 | 3 | 7 | 487 | 172 | 115 | 9 | 22 |
| EDL_OFF_TR | 24 | 14 | 22 | 9 | 46 | 64 | 37 | 58 | 24 | 122 |
| NSW_IN_TR | 16 | 15 | 42 | 20 | 4 | 40 | 37 | 104 | 49 | 10 |
| NSW_OFF_TR | 5 | 17 | 19 | 23 | 8 | 11 | 36 | 40 | 49 | 17 |
| NSW_RRP_TR | 1 | 3 | 4 | 1 | 6 | 2 | 5 | 7 | 2 | 11 |
| SW_BGS_TR | 1 | 2 | 5 | 8 | 2 | 2 | 4 | 10 | 16 | 4 |
| SW_ORO_TR | 1 | 1 | 1 | 1 | 1 | 2 | 2 | 2 | 2 | 2 |
| SW_TR | 30 | 24 | 34 | 190 | 74 | 79 | 63 | 89 | 500 | 195 |
| TAS_BGS_TR | 1 | 3 | 5 | 1 | 1 | 2 | 7 | 12 | 2 | 2 |
| TAS_E_TR | 8 | 17 | 16 | 2 | 3 | 24 | 51 | 48 | 6 | 9 |
| TAS_ORO_TR | 1 | 3 | 2 | 1 | 1 | 2 | 6 | 4 | 2 | 2 |
| TAS_W_TR | 3 | 26 | 17 | 39 | 12 | 7 | 65 | 42 | 97 | 30 |
| AL CSA | 1 | 7 | 10 | 1 | 6 | 2 | 12 | 17 | 2 | 10 |
| AL EBS / NSW | 1 | 10 | 8 | 3 | 15 | 1 | 12 | 9 | 3 | 17 |
| AL ESA | 1 | 1 | 2 | 2 | 1 | 2 | 2 | 3 | 3 | 2 |
| AL ET / WT | 12 | 4 | 11 | 1 | 3 | 13 | 4 | 12 | 1 | 3 |
| AL WBS / SAV | 1 | 3 | 4 | 1 | 2 | 1 | 4 | 5 | 1 | 3 |
| AL WSA / WA | 1 | 6 | 3 | 1 | 5 | 1 | 9 | 4 | 1 | 7 |
| GN CSA | 38 | 36 | 37 | 35 | 38 | 83 | 78 | 81 | 76 | 83 |
| GN EBS / NSW | 56 | 55 | 55 | 53 | 57 | 99 | 97 | 97 | 93 | 100 |
| GN ESA | 5 | 5 | 5 | 5 | 5 | 9 | 9 | 9 | 9 | 9 |
| GN ET / WT | 17 | 16 | 16 | 16 | 17 | 24 | 23 | 23 | 23 | 24 |
| GN SAV | 5 | 5 | 5 | 5 | 5 | 8 | 8 | 8 | 8 | 8 |
| GN WA / WSA | 25 | 24 | 24 | 23 | 25 | 59 | 57 | 57 | 54 | 59 |
| GN WBS | 16 | 15 | 16 | 15 | 16 | 25 | 24 | 25 | 24 | 25 |
| GAB Inshore | 58 | 104 | 38 | 6 | 65 | 190 | 340 | 124 | 20 | 213 |
| GAB Midshore | 4 | 6 | 4 | 2 | 5 | 8 | 12 | 8 | 4 | 10 |
| GAB Offshore | 1 | 5 | 3 | 1 | 1 | 2 | 9 | 5 | 2 | 2 |
| total | 501 | 499 | 501 | 500 | 500 | 1297 | 1255 | 1269 | 1222 | 1310 |

[^0]Table 5-Sample size distribution across strata for a total sample size of $\mathbf{7 5 0}$ observer sea-days by Methods: VWCV MinS, Prop, CWCV-HR, MinS-HR. Distributions are shown in terms of Sea-days and in terms of the estimated equivalent number of shots sampled.

Total Sample Size = $\mathbf{7 5 0}$ observer seadays.

|  | Number of Sea-Days |  |  |  |  | Number of Shots |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Method VWCV | Method Mins | Metho d Prop | Method CWCV HR | Method MinS HR | Method vwCV | Method Mins | Metho d Prop | Method CWCV HR | Method MinS HR |
| VIT | 5 | 5 | 5 | 5 | 5 | 18 | 18 | 18 | 18 | 18 |
| BS_IN_TR | 1 | 1 | 2 | 5 | 1 | 3 | 3 | 7 | 17 | 3 |
| ECDW_TR | 1 | 1 | 2 | 1 | 1 | 2 | 2 | 5 | 2 | 2 |
| EDL_DS | 10 | 27 | 76 | 38 | 64 | 46 | 124 | 348 | 174 | 293 |
| EDL_IN_TR | 233 | 55 | 55 | 4 | 16 | 727 | 172 | 172 | 12 | 50 |
| EDL_OFF_TR | 36 | 14 | 32 | 13 | 46 | 95 | 37 | 85 | 34 | 122 |
| NSW_IN_TR | 24 | 15 | 63 | 29 | 12 | 59 | 37 | 156 | 72 | 30 |
| NSW_OFF_TR | 8 | 17 | 28 | 35 | 8 | 17 | 36 | 59 | 74 | 17 |
| NSW_RRP_TR | 1 | 3 | 6 | 1 | 6 | 2 | 5 | 11 | 2 | 11 |
| SW_BGS_TR | 2 | 2 | 8 | 12 | 2 | 4 | 4 | 16 | 23 | 4 |
| SW_ORO_TR | 1 | 1 | 1 | 2 | 1 | 2 | 2 | 2 | 5 | 2 |
| SW_TR | 46 | 24 | 52 | 286 | 74 | 121 | 63 | 137 | 752 | 195 |
| TAS_BGS_TR | 2 | 3 | 8 | 2 | 1 | 5 | 7 | 20 | 5 | 2 |
| TAS_E_TR | 12 | 17 | 24 | 3 | 3 | 36 | 51 | 72 | 9 | 9 |
| TAS_ORO_TR | 2 | 3 | 3 | 1 | 1 | 4 | 6 | 6 | 2 | 2 |
| TAS_W_TR | 5 | 26 | 25 | 59 | 12 | 12 | 65 | 62 | 147 | 30 |
| AL CSA | 2 | 7 | 15 | 2 | 6 | 3 | 12 | 25 | 3 | 10 |
| AL EBS / NSW | 1 | 10 | 11 | 4 | 15 | 1 | 12 | 13 | 5 | 17 |
| AL ESA | 1 | 1 | 3 | 2 | 1 | 2 | 2 | 5 | 3 | 2 |
| AL ET / WT | 17 | 4 | 17 | 1 | 3 | 18 | 4 | 18 | 1 | 3 |
| AL WBS / SAV | 1 | 3 | 5 | 1 | 2 | 1 | 4 | 7 | 1 | 3 |
| AL WSA / WA | 1 | 6 | 5 | 2 | 5 | 1 | 9 | 7 | 3 | 7 |
| GN CSA | 56 | 57 | 55 | 53 | 58 | 122 | 124 | 120 | 116 | 126 |
| GN EBS / NSW | 85 | 86 | 83 | 80 | 87 | 150 | 151 | 146 | 141 | 153 |
| GN ESA | 8 | 8 | 8 | 7 | 8 | 15 | 15 | 15 | 13 | 15 |
| GN ET / WT | 25 | 25 | 24 | 23 | 26 | 35 | 35 | 34 | 32 | 37 |
| GN SAV | 8 | 8 | 8 | 8 | 8 | 13 | 13 | 13 | 13 | 13 |
| GN WA / WSA | 37 | 38 | 36 | 35 | 38 | 87 | 90 | 85 | 83 | 90 |
| GN WBS | 24 | 24 | 23 | 22 | 24 | 38 | 38 | 37 | 35 | 38 |
| GAB Inshore | 88 | 248 | 57 | 9 | 210 | 288 | 811 | 186 | 29 | 687 |
| GAB Midshore | 6 | 6 | 6 | 4 | 5 | 12 | 12 | 12 | 8 | 10 |
| GAB Offshore | 1 | 5 | 4 | 1 | 1 | 2 | 9 | 7 | 2 | 2 |
| total | 750 | 750 | 750 | 750 | 750 | 1943 | 1973 | 1903 | 1837 | 2003 |

[^1]Table 6 - Sample size distribution across strata for a total sample size of 1000 observer sea-days by Methods: VWCV, MinS, Prop, CWCV-HR, MinS-HR. Distributions are shown in terms of Sea-days and in terms of the estimated equivalent number of shots sampled.

Total Sample Size $=1000$ observer seadays.

|  | Number of Sea-Days |  |  |  |  | Number of Shots |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Method VWCV | Method MinS | Metho d Prop | Method CWCV HR | Method MinS HR | Method VWCV | Method MinS | Metho d Prop | Method CWCV HR | Method MinS HR |
| VIT | 6 | 7 | 6 | 6 | 7 | 21 | 25 | 21 | 21 | 25 |
| BS_IN_TR | 1 | 1 | 3 | 6 | 1 | 3 | 3 | 10 | 21 | 3 |
| ECDW_TR | 1 | 1 | 2 | 1 | 1 | 2 | 2 | 5 | 2 | 2 |
| EDL_DS | 14 | 29 | 102 | 50 | 64 | 64 | 133 | 467 | 229 | 293 |
| EDL_IN_TR | 313 | 55 | 75 | 5 | 37 | 977 | 172 | 234 | 16 | 115 |
| EDL_OFF_TR | 48 | 14 | 43 | 18 | 46 | 127 | 37 | 114 | 48 | 122 |
| NSW_IN_TR | 32 | 15 | 84 | 39 | 13 | 79 | 37 | 207 | 96 | 32 |
| NSW_OFF_TR | 11 | 17 | 37 | 47 | 8 | 23 | 36 | 78 | 100 | 17 |
| NSW_RRP_TR | 1 | 3 | 8 | 1 | 6 | 2 | 5 | 14 | 2 | 11 |
| SW_BGS_TR | 2 | 2 | 10 | 16 | 2 | 4 | 4 | 20 | 31 | 4 |
| SW_ORO_TR | 1 | 1 | 1 | 3 | 1 | 2 | 2 | 2 | 7 | 2 |
| SW_TR | 61 | 24 | 69 | 384 | 74 | 160 | 63 | 181 | 1010 | 195 |
| TAS_BGS_TR | 2 | 3 | 10 | 2 | 2 | 5 | 7 | 25 | 5 | 5 |
| TAS_E_TR | 16 | 17 | 32 | 4 | 13 | 48 | 51 | 95 | 12 | 39 |
| TAS_ORO_TR | 2 | 3 | 4 | 1 | 1 | 4 | 6 | 9 | 2 | 2 |
| TAS_W_TR | 6 | 26 | 34 | 79 | 18 | 15 | 65 | 85 | 197 | 45 |
| AL CSA | 2 | 7 | 20 | 2 | 6 | 3 | 12 | 33 | 3 | 10 |
| AL EBS / NSW | 2 | 10 | 15 | 6 | 15 | 2 | 12 | 17 | 7 | 17 |
| AL ESA | 1 | 1 | 4 | 3 | 1 | 2 | 2 | 6 | 5 | 2 |
| AL ET / WT | 23 | 4 | 22 | 1 | 3 | 24 | 4 | 23 | 1 | 3 |
| AL WBS / SAV | 2 | 3 | 7 | 1 | 2 | 3 | 4 | 9 | 1 | 3 |
| AL WSA / WA | 1 | 6 | 6 | 2 | 5 | 1 | 9 | 9 | 3 | 7 |
| GN CSA | 75 | 77 | 74 | 71 | 78 | 164 | 168 | 161 | 155 | 170 |
| GN EBS / NSW | 113 | 116 | 111 | 107 | 117 | 199 | 204 | 195 | 188 | 206 |
| GN ESA | 10 | 11 | 10 | 10 | 11 | 19 | 20 | 19 | 19 | 20 |
| GN ET / WT | 33 | 34 | 32 | 31 | 34 | 47 | 48 | 45 | 44 | 48 |
| GN SAV | 11 | 11 | 10 | 10 | 11 | 18 | 18 | 16 | 16 | 18 |
| GN WA / WSA | 50 | 51 | 49 | 47 | 51 | 118 | 120 | 116 | 111 | 120 |
| GN WBS | 32 | 33 | 31 | 30 | 33 | 51 | 52 | 49 | 48 | 52 |
| GAB Inshore | 118 | 406 | 77 | 12 | 333 | 386 | 1328 | 252 | 39 | 1089 |
| GAB Midshore | 9 | 6 | 7 | 5 | 5 | 18 | 12 | 14 | 10 | 10 |
| GAB Offshore | 1 | 5 | 5 | 1 | 1 | 2 | 9 | 9 | 2 | 2 |
| total | 1000 | 999 | 1000 | 1001 | 1000 | 2594 | 2671 | 2542 | 2450 | 2690 |

[^2]Table 7 - Sample size distribution across strata for a total sample size of $\mathbf{1 2 5 0}$ observer sea-days by Methods: VWCV, MinS, Prop, CWCV-HR, MinS-HR. Distributions are shown in terms of Sea-days and in terms of the estimated equivalent number of shots sampled.

Total Sample Size = 1250 observer seadays.

|  | Number of Sea-Days |  |  |  |  | Number of Shots |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Method VWCV | Method Mins | Metho <br> d Prop | Method CWCV HR | Method MinS HR | Method VWCV | Method MinS | Metho <br> d Prop | Method CWCV HR | Method MinS HR |
| VIT | 8 | 8 | 8 | 8 | 8 | 28 | 28 | 28 | 28 | 28 |
| BS_IN_TR | 2 | 1 | 3 | 8 | 1 | 7 | 3 | 10 | 28 | 3 |
| ECDW_TR | 1 | 1 | 2 | 2 | 1 | 2 | 2 | 5 | 5 | 2 |
| EDL_DS | 17 | 29 | 127 | 63 | 64 | 78 | 133 | 582 | 289 | 293 |
| EDL_IN_TR | 392 | 55 | 93 | 7 | 37 | 1223 | 172 | 290 | 22 | 115 |
| EDL_OFF_TR | 61 | 14 | 54 | 22 | 46 | 162 | 37 | 143 | 58 | 122 |
| NSW_IN_TR | 40 | 15 | 106 | 49 | 16 | 99 | 37 | 262 | 121 | 40 |
| NSW_OFF_TR | 14 | 67 | 47 | 59 | 46 | 30 | 142 | 100 | 125 | 98 |
| NSW_RRP_TR | 1 | 3 | 10 | 1 | 6 | 2 | 5 | 18 | 2 | 11 |
| SW_BGS_TR | 2 | 2 | 13 | 20 | 2 | 4 | 4 | 25 | 39 | 4 |
| SW_ORO_TR | 1 | 1 | 1 | 3 | 1 | 2 | 2 | 2 | 7 | 2 |
| SW_TR | 76 | 74 | 87 | 478 | 74 | 200 | 195 | 229 | 1257 | 195 |
| TAS_BGS_TR | 3 | 3 | 12 | 3 | 2 | 7 | 7 | 30 | 7 | 5 |
| TAS_E_TR | 20 | 67 | 40 | 5 | 47 | 60 | 200 | 119 | 15 | 140 |
| TAS_ORO_TR | 3 | 3 | 5 | 1 | 1 | 6 | 6 | 11 | 2 | 2 |
| TAS_W_TR | 8 | 42 | 43 | 98 | 29 | 20 | 105 | 107 | 244 | 72 |
| AL CSA | 3 | 7 | 24 | 3 | 6 | 5 | 12 | 40 | 5 | 10 |
| AL EBS / NSW | 2 | 10 | 19 | 7 | 15 | 2 | 12 | 22 | 8 | 17 |
| AL ESA | 1 | 1 | 4 | 4 | 1 | 2 | 2 | 6 | 6 | 2 |
| AL ET / WT | 29 | 4 | 28 | 1 | 3 | 31 | 4 | 30 | 1 | 3 |
| AL WBS / SAV | 2 | 3 | 9 | 1 | 2 | 3 | 4 | 12 | 1 | 3 |
| AL WSA / WA | 1 | 6 | 8 | 3 | 5 | 1 | 9 | 12 | 4 | 7 |
| GN CSA | 94 | 95 | 92 | 89 | 96 | 205 | 207 | 201 | 194 | 209 |
| GN EBS / NSW | 141 | 142 | 138 | 133 | 145 | 248 | 250 | 243 | 234 | 255 |
| GN ESA | 13 | 13 | 13 | 12 | 13 | 24 | 24 | 24 | 22 | 24 |
| GN ET / WT | 41 | 42 | 40 | 39 | 42 | 58 | 59 | 56 | 55 | 59 |
| GN SAV | 13 | 13 | 13 | 12 | 14 | 21 | 21 | 21 | 20 | 23 |
| GN WA / WSA | 62 | 62 | 61 | 58 | 64 | 146 | 146 | 144 | 137 | 151 |
| GN WBS | 40 | 40 | 39 | 37 | 41 | 64 | 64 | 62 | 59 | 65 |
| GAB Inshore | 147 | 414 | 96 | 15 | 414 | 481 | 1354 | 314 | 49 | 1354 |
| GAB Midshore | 11 | 6 | 9 | 6 | 5 | 22 | 12 | 18 | 12 | 10 |
| GAB Offshore | 1 | 5 | 6 | 1 | 3 | 2 | 9 | 10 | 2 | 5 |
| total | 1250 | 1248 | 1250 | 1248 | 1250 | 3245 | 3267 | 3175 | 3059 | 3331 |

[^3]Table 8 - Sample size distribution across strata for a total sample size of 1500 observer sea-days by Methods: VWCV, MinS, Prop, CWCV-HR, MinS-HR. Distributions are shown in terms of Sea-days and in terms of the estimated equivalent number of shots sampled.

Total Sample Size = 1500 observer seadays.

|  | Number of Sea-Days |  |  |  |  | Number of Shots |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Method VWCV | Method Mins | Metho <br> d Prop | Method CWCV HR | Method MinS HR | Method VWCV | Method MinS | Metho <br> d Prop | Method CWCV HR | Method MinS HR |
| VIT | 10 | 10 | 9 | 9 | 10 | 36 | 36 | 32 | 32 | 36 |
| BS_IN_TR | 2 | 1 | 4 | 10 | 1 | 7 | 3 | 14 | 35 | 3 |
| ECDW_TR | 1 | 1 | 3 | 2 | 1 | 2 | 2 | 7 | 5 | 2 |
| EDL_DS | 21 | 29 | 153 | 76 | 64 | 96 | 133 | 701 | 348 | 293 |
| EDL_IN_TR | 470 | 55 | 112 | 8 | 37 | 1466 | 172 | 349 | 25 | 115 |
| EDL_OFF_TR | 73 | 14 | 65 | 27 | 46 | 193 | 37 | 172 | 72 | 122 |
| NSW_IN_TR | 48 | 15 | 127 | 59 | 16 | 119 | 37 | 314 | 146 | 40 |
| NSW_OFF_TR | 16 | 67 | 56 | 70 | 63 | 34 | 142 | 119 | 148 | 134 |
| NSW_RRP_TR | 1 | 3 | 12 | 1 | 6 | 2 | 5 | 21 | 2 | 11 |
| SW_BGS_TR | 3 | 2 | 15 | 24 | 2 | 6 | 4 | 29 | 47 | 4 |
| SW_ORO_TR | 1 | 1 | 1 | 4 | 1 | 2 | 2 | 2 | 9 | 2 |
| SW_TR | 92 | 74 | 105 | 575 | 74 | 242 | 195 | 276 | 1512 | 195 |
| TAS_BGS_TR | 4 | 3 | 15 | 3 | 2 | 10 | 7 | 37 | 7 | 5 |
| TAS_E_TR | 24 | 67 | 48 | 6 | 64 | 72 | 200 | 143 | 18 | 191 |
| TAS_ORO_TR | 4 | 3 | 5 | 1 | 1 | 9 | 6 | 11 | 2 | 2 |
| TAS_W_TR | 9 | 42 | 51 | 118 | 40 | 22 | 105 | 127 | 294 | 100 |
| AL CSA | 3 | 7 | 29 | 4 | 6 | 5 | 12 | 48 | 7 | 10 |
| AL EBS / NSW | 3 | 10 | 23 | 9 | 15 | 3 | 12 | 26 | 10 | 17 |
| AL ESA | 1 | 1 | 5 | 5 | 1 | 2 | 2 | 8 | 8 | 2 |
| AL ET / WT | 35 | 4 | 33 | 1 | 3 | 37 | 4 | 35 | 1 | 3 |
| AL WBS / SAV | 2 | 3 | 10 | 1 | 2 | 3 | 4 | 13 | 1 | 3 |
| AL WSA / WA | 1 | 6 | 9 | 3 | 5 | 1 | 9 | 13 | 4 | 7 |
| GN CSA | 113 | 115 | 110 | 106 | 116 | 246 | 251 | 240 | 231 | 253 |
| GN EBS / NSW | 169 | 173 | 166 | 160 | 174 | 297 | 304 | 292 | 282 | 306 |
| GN ESA | 16 | 16 | 15 | 15 | 16 | 30 | 30 | 28 | 28 | 30 |
| GN ET / WT | 50 | 51 | 49 | 47 | 51 | 71 | 72 | 69 | 66 | 72 |
| GN SAV | 16 | 16 | 16 | 15 | 16 | 26 | 26 | 26 | 25 | 26 |
| GN WA / WSA | 74 | 76 | 73 | 70 | 76 | 175 | 179 | 172 | 165 | 179 |
| GN WBS | 48 | 48 | 47 | 45 | 49 | 76 | 76 | 75 | 72 | 78 |
| GAB Inshore | 176 | 575 | 116 | 18 | 533 | 576 | 1880 | 379 | 59 | 1743 |
| GAB Midshore | 13 | 6 | 11 | 7 | 5 | 27 | 12 | 22 | 14 | 10 |
| GAB Offshore | 1 | 5 | 7 | 1 | 3 | 2 | 9 | 12 | 2 | 5 |
| total | 1500 | 1499 | 1500 | 1500 | 1499 | 3894 | 3968 | 3814 | 3676 | 3998 |

[^4]Table 9. Extent to which records in the TEP Interaction table can be matched to records in the ISMP catch table using vessel as the linking key.

|  | Un <br> Matched <br> Vessel | No <br> Matching <br> Shot | Matched <br> Shot | Total <br> Interactions |
| :--- | :---: | :---: | :---: | :---: |
| Birds | 19 | 182 | 0 | $\mathbf{2 0 1}$ |
| Pipefish | 1 | 1 | 1 | $\mathbf{3}$ |
| Seals | 9 | 100 | 0 | $\mathbf{1 0 9}$ |
| Sharks | 0 | 0 | 1 | $\mathbf{1}$ |
| Whales/Dolphins | 0 | 6 | 1 | $\mathbf{7}$ |

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Table 10. The number of trips sampled per species and year for length frequency distribution information, at the point of landing - the difference between species is assumed to be due partly to hit rate but may also reflect sampling decisions.

|  | CSIRO | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2007/2008 | 2002-2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Gummy shark | 37017001 | 25 | 13 | 22 | 36 | 2 | 22 | 40 | 31 | 20 |
| School shark | 37017008 | 7 | 4 | 6 | 7 | 1 | 18 | 18 | 18 | 5 |
| Elephantfish | 37043001 | 0 | 0 | 0 | 0 | 0 | 11 | 27 | 19 | 0 |
| Ribaldo | 37224002 | 0 | 0 | 0 | 0 | 0 | 1 | 9 | 5 | 0 |
| Blue grenadier | 37227001 | 75 | 39 | 43 | 50 | 31 | 10 | 18 | 14 | 48 |
| Pink ling | 37228002 | 59 | 35 | 41 | 46 | 28 | 24 | 13 | 19 | 42 |
| Orange roughy | 37255009 | 16 | 26 | 19 | 17 | 4 | 0 | 0 |  | 16 |
| Alfonsino | 37258002 | 2 | 6 | 1 | 0 | 0 | 6 | 11 | 9 | 2 |
| Redfish | 37258003 | 48 | 42 | 60 | 56 | 41 | 7 | 8 | 8 | 49 |
| Bight redfish | 37258004 | 0 | 16 | 13 | 14 | 2 | 0 | 0 |  | 9 |
| Mirror dory | 37264003 | 47 | 46 | 38 | 44 | 31 | 16 | 16 | 16 | 41 |
| John dory | 37264004 | 60 | 45 | 61 | 46 | 33 | 13 | 22 | 18 | 49 |
| Oreo basket | 37266001 | 2 | 8 | 0 | 1 | 0 | 0 | 3 | 2 | 2 |
| Smooth oreodory | 37266003 | 0 | 0 | 3 | 0 | 0 | 0 | 0 |  | 1 |
| Inshore ocean perch | 37287001 | 13 | 15 | 19 | 10 | 16 | 0 | 0 |  | 15 |
| Tiger flathead | 37296001 | 136 | 82 | 156 | 133 | 79 | 36 | 9 | 23 | 117 |
| Deepwater flathead | 37296002 | 0 | 25 | 38 | 40 | 9 | 0 | 0 |  | 22 |
| Eastern school whiting | 37330014 | 28 | 22 | 55 | 40 | 40 | 28 | 8 | 18 | 37 |
| Silver trevally | 37337062 | 21 | 5 | 32 | 25 | 45 | 13 | 4 | 9 | 26 |
| Boarfish (unspecified) | 37367003 | 0 | 1 | 0 | 0 | 1 | 0 | 0 |  | 0 |
| Jackass morwong | 37377003 | 80 | 64 | 60 | 67 | 89 | 26 | 20 | 23 | 72 |
| Gemfish | 37439002 | 70 | 55 | 50 | 41 | 14 | 6 | 22 | 14 | 46 |
| Blue-eye trevalla | 37445001 | 80 | 40 | 31 | 12 | 11 | 4 | 10 | 7 | 35 |
| Blue warehou | 37445005 | 54 | 32 | 31 | 25 | 32 | 5 | 8 | 7 | 35 |
| Silver warehou | 37445006 | 164 | 88 | 79 | 83 | 60 | 10 | 12 | 11 | 95 |

ble 11. The average number of fish sampled per trip by species and year for length frequency distribution information, at the point of
ding.

|  | CSIRO | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ | $\mathbf{2 0 0 7 / 2 0 0 8}$ | $\mathbf{2 0 0 2 - 2 0 0 6}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Gummy shark | 37017001 | 98 | 89 | 155 | 280 | 413 | 508 | 587 | $\mathbf{5 4 8}$ | $\mathbf{2 0 7}$ |
| School shark | 37017008 | 21 | 18 | 27 | 63 | 18 | 158 | 179 | $\mathbf{1 6 9}$ | 29 |
| Elephantfish | 37043001 |  |  |  |  |  | 348 | 436 | $\mathbf{3 9 2}$ |  |
| Ribaldo | 37224002 |  |  |  |  |  | 404 | 560 | $\mathbf{4 8 2}$ |  |
| Blue grenadier | 37227001 | 126 | 129 | 103 | 126 | 93 | 375 | 428 | $\mathbf{4 0 2}$ | 115 |
| Pink ling | 37228002 | 92 | 51 | 57 | 62 | 76 | 498 | 586 | $\mathbf{5 4 2}$ | 68 |
| Orange roughy | 37255009 | 92 | 191 | 127 | 119 | 178 |  |  |  | 141 |
| Alfonsino | 37258002 | 29 | 47 | 128 |  |  | 430 | 364 | $\mathbf{3 9 7}$ | 68 |
| Redfish | 37258003 | 103 | 99 | 169 | 131 | 262 | 916 | 167 | $\mathbf{5 4 1}$ | 153 |
| Bight redfish | 37258004 |  | 101 | 100 | 86 | 88 |  |  |  | 93 |
| Mirror dory | 37264003 | 140 | 64 | 58 | 68 | 71 | 439 | 423 | $\mathbf{4 3 1}$ | 80 |
| John dory | 37264004 | 117 | 98 | 108 | 76 | 96 | 408 | 471 | $\mathbf{4 4 0}$ | 99 |
| Oreo basket | 37266001 | 63 | 62 |  | 115 |  |  | 478 | $\mathbf{2 3 9}$ | 80 |
| Smooth oreodory | 37266003 |  |  | 39 |  |  |  |  |  | 8 |
| In. Ocean perch | 37287001 | 125 | 97 | 77 | 54 | 122 |  |  |  | 39 |
| Tiger flathead | 37296001 | 110 | 88 | 103 | 130 | 199 | 757 | 596 | $\mathbf{6 7 6}$ | 95 |
| D.w. flathead | 37296002 |  | 101 | 147 | 224 | 186 |  |  |  | 126 |
| E. School whiting | 37330014 | 88 | 88 | 221 | 215 | 422 | 770 | 514 | $\mathbf{6 4 2}$ | 164 |
| Silver trevally | 37337062 | 69 | 114 | 80 | 105 | 337 | 630 | 465 | $\mathbf{5 4 8}$ | 207 |
| Boarfish | 37367003 |  | 13 |  |  | 45 |  |  |  | 141 |
| Jackass morwong | 37377003 | 136 | 97 | 96 | 134 | 186 | 556 | 419 | $\mathbf{4 8 7}$ | 29 |
| Gemfish | 37439002 | 100 | 87 | 73 | 50 | 43 | 209 | 337 | $\mathbf{2 7 3}$ | 130 |
| Blue-eye trevalla | 37445001 | 80 | 67 | 66 | 95 | 226 | 378 | 424 | $\mathbf{4 0 1}$ | 70 |
| Blue warehou | 37445005 | 119 | 101 | 110 | 94 | 105 | 570 | 215 | $\mathbf{3 9 2}$ | 107 |
| Silver warehou | 37445006 | 134 | 102 | 99 | 149 | 147 | 613 | 451 | $\mathbf{5 3 2}$ | 106 |
|  |  |  |  |  |  |  |  | 126 |  |  |

Table 12. The total number of fish sampled per year for length frequency distribution information, at the point of landing.

| Quota species | CSIRO | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Gummy shark | 37017001 | 2442 | 1155 | 3405 | 10086 | 826 | 11178 | 23485 |
| School shark | 37017008 | 150 | 72 | 161 | 440 | 18 | 2848 | 3221 |
| Elephantfish | $\mathbf{3 7 0 4 3 0 0 1}$ |  |  |  |  |  | $\mathbf{3 8 2 4}$ | $\mathbf{1 1 7 8 2}$ |
| Ribaldo | 37224002 |  |  |  |  |  | 404 | 5043 |
| Blue grenadier | 37227001 | 9416 | 5023 | 4413 | 6310 | 2874 | 3748 | 7709 |
| Pink ling | 37228002 | 5411 | 1788 | 2325 | 2861 | 2140 | 11952 | 7612 |
| Orange roughy | 37255009 | 1466 | 4973 | 2416 | 2021 | 711 |  |  |
| Alfonsino | 37258002 | 58 | 282 | 128 |  |  | 2580 | 4008 |
| Redfish | 37258003 | 4925 | 4178 | 10132 | 7316 | 10741 | 6412 | 1332 |
| Bight redfish | 37258004 |  | 1608 | 1298 | 1200 | 175 |  |  |
| Mirror dory | 37264003 | 6600 | 2922 | 2209 | 2993 | 2209 | 7028 | 6772 |
| John dory | 37264004 | 7004 | 4397 | 6618 | 3494 | 3180 | 5308 | 10371 |
| Oreo basket | 37266001 | 126 | 494 |  | 115 |  |  | 1433 |
| Smooth oreodory | 37266003 |  |  | 117 |  |  |  |  |
| Inshore ocean perch | 37287001 | 1627 | 1458 | 1468 | 539 | 1947 |  |  |
| Tiger flathead | $\mathbf{3 7 2 9 6 0 0 1}$ | $\mathbf{1 4 9 7 8}$ | $\mathbf{7 2 3 1}$ | $\mathbf{1 6 0 4 7}$ | $\mathbf{1 7 3 4 6}$ | $\mathbf{1 5 7 2 7}$ | $\mathbf{2 7 2 6 0}$ | $\mathbf{5 3 6 0}$ |
| Deepwater flathead | 37296002 |  | 2514 | 5577 | 8954 | 1678 |  |  |
| Eastern school whiting | $\mathbf{3 7 3 3 0 0 1 4}$ | $\mathbf{2 4 5 3}$ | $\mathbf{1 9 3 0}$ | $\mathbf{1 2 1 6 6}$ | $\mathbf{8 6 1 2}$ | $\mathbf{1 6 8 6 6}$ | $\mathbf{2 1 5 7 2}$ | $\mathbf{4 1 1 2}$ |
| Silver trevally | 37337062 | 1439 | 571 | 2557 | 2636 | 15170 | 8192 | 1860 |
| Boarfish (unspecified) | 37367003 |  | 13 |  |  | 45 |  |  |
| Jackass morwong | $\mathbf{3 7 3 7 7 0 0 3}$ | $\mathbf{1 0 8 8 5}$ | $\mathbf{6 2 3 1}$ | $\mathbf{5 7 7 3}$ | $\mathbf{9 0 0 4}$ | $\mathbf{1 6 5 7 0}$ | $\mathbf{1 4 4 6 0}$ | $\mathbf{8 3 7 4}$ |
| Gemfish | 37439002 | 6987 | 4781 | 3637 | 2037 | 605 | 1256 | 7404 |
| Blue-eye trevalla | 37445001 | 6436 | 2679 | 2032 | 1142 | 2491 | 1512 | 4235 |
| Blue warehou | 37445005 | 6400 | 3230 | 3415 | 2341 | 3360 | 2848 | 1719 |
| Silver warehou | $\mathbf{3 7 4 4 5 0 0 6}$ | $\mathbf{2 2 0 4 3}$ | $\mathbf{8 9 9 9}$ | $\mathbf{7 7 8 3}$ | $\mathbf{1 2 3 2 6}$ | $\mathbf{8 8 4 9}$ | $\mathbf{6 1 3 2}$ | $\mathbf{5 4 1 2}$ |

Table 13. A summary of the ISMP shots recorded by year and by strata.

| Strata | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AL CSA | 0 | 0 | 51 | 10 | 60 | 43 | 0 |
| AL EBS / NSW | 175 | 168 | 58 | 10 | 4 | 23 | 87 |
| AL ESA | 0 | 0 | 0 | 0 | 0 | 4 | 0 |
| AL ET / WT | 71 | 75 | 64 | 23 | 36 | 4 | 45 |
| AL WBS / SAV | 14 | 35 | 55 | 5 | 31 | 78 | 3 |
| AL WSA / WA | 0 | 0 | 0 | 91 | 0 | 31 | 90 |
| BS_IN_TR | 12 | 30 | 6 | 1 | 0 | 0 | 17 |
| ECDW_TR | 69 | 0 | 25 | 0 | 0 | 0 | 0 |
| EDL_DS | 34 | 115 | 48 | 70 | 129 | 47 | 13 |
| EDL_IN_TR | 81 | 100 | 47 | 133 | 117 | 68 | 103 |
| EDL_OFF_TR | 62 | 73 | 76 | 74 | 66 | 31 | 56 |
| GN | 4 | 0 | 0 | 0 | 0 | 0 | 0 |
| Inshore | 113 | 89 | 123 | 203 | 133 | 132 | 120 |
| Midshore | 6 | 23 | 29 | 1 | 25 | 12 | 12 |
| NSW_IN_TR | 208 | 296 | 297 | 272 | 213 | 47 | 121 |
| NSW_OFF_TR | 75 | 94 | 127 | 81 | 88 | 9 | 16 |
| NSW_RRP_TR | 26 | 11 | 4 | 5 | 4 | 6 | 0 |
| Offshore | 23 | 19 | 21 | 11 | 15 | 26 | 17 |
| SW_BGS_TR | 12 | 14 | 31 | 30 | 15 | 0 | 0 |
| SW_ORO_TR | 2 | 6 | 1 | 2 | 0 | 0 | 0 |
| SW_TR | 144 | 97 | 97 | 187 | 120 | 77 | 101 |
| TAS_BGS_TR | 9 | 33 | 8 | 19 | 5 | 0 | 0 |
| TAS_E_TR | 64 | 22 | 35 | 21 | 66 | 0 | 10 |
| TAS_ORO_TR | 9 | 9 | 25 | 25 | 22 | 0 | 0 |
| TAS_W_TR | 40 | 31 | 1 | 28 | 9 | 26 | 3 |

Table 14. A summary of the GENLOG shots recorded by year and by strata, after the CDR/GENLOG merge.

| Strata | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AL CSA | 144 | 317 | 661 | 567 | 747 | 500 | 381 |
| AL EBS / NSW | 184 | 497 | 677 | 395 | 292 | 274 | 181 |
| AL ESA | 46 | 154 | 205 | 456 | 226 | 84 | 51 |
| AL ET / WT | 29 | 492 | 445 | 370 | 348 | 345 | 288 |
| AL WBS / SAV | 131 | 339 | 685 | 382 | 231 | 128 | 108 |
| AL WSA / WA | 24 | 46 | 223 | 338 | 453 | 117 | 118 |
| BS_IN_TR | 157 | 165 | 411 | 749 | 602 | 45 | 177 |
| ECDW_TR | 88 | 167 | 230 | 60 | 1 | 75 | 1 |
| EDL_DS | 8939 | 10111 | 8323 | 7681 | 6489 | 6394 | 6405 |
| EDL_IN_TR | 5353 | 4889 | 4853 | 5588 | 4254 | 3186 | 3193 |
| EDL_OFF_TR | 3114 | 3623 | 2816 | 2606 | 2075 | 1462 | 1667 |
| GN CSA | 2840 | 2860 | 3060 | 2838 | 2994 | 2125 | 2244 |
| GN EBS / NSW | 4346 | 4758 | 4363 | 3743 | 3326 | 2772 | 2509 |
| GN ESA | 429 | 342 | 393 | 263 | 171 | 266 | 252 |
| GN ET / WT | 687 | 685 | 801 | 537 | 591 | 772 | 440 |
| GN SAV | 687 | 564 | 501 | 597 | 531 | 290 | 173 |
| GN WA / WSA | 1584 | 1770 | 1695 | 1423 | 2249 | 2010 | 1051 |
| GN WBS | 1668 | 1671 | 1764 | 1456 | 910 | 832 | 461 |
| Inshore | 2284 | 3810 | 4696 | 5325 | 5165 | 4044 | 2659 |
| Midshore | 0 | 169 | 474 | 649 | 488 | 313 | 48 |
| NSW_IN_TR | 6103 | 7491 | 7849 | 6151 | 4843 | 2834 | 2889 |
| NSW OFF TR | 2225 | 2395 | 2132 | 2024 | 1515 | 1045 | 1114 |
| NSW_RRP_TR | 1442 | 644 | 441 | 437 | 323 | 211 | 134 |
| Offshore | 2 | 205 | 246 | 177 | 192 | 123 | 78 |
| SW BGS TR | 353 | 317 | 501 | 414 | 435 | 315 | 191 |
| SW ORO TR | 141 | 90 | 93 | 79 | 55 | 8 | 0 |
| SW TR | 5238 | 4255 | 4966 | 3902 | 3660 | 2871 | 2153 |
| TAS_BGS_TR | 1095 | 1082 | 831 | 445 | 482 | 332 | 318 |
| TAS_E_TR | 3494 | 2840 | 2978 | 2391 | 1951 | 1254 | 1328 |
| TAS_ORO TR | 654 | 300 | 245 | 270 | 245 | 68 | 116 |
| TAS_W TR | 3463 | 2646 | 1944 | 1451 | 1057 | 1343 | 955 |
| VIT | 415 | 174 | 314 | 739 | 421 | 342 | 263 |
|  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |

Table 15. A table of the MWCVs for the age frequency estimates obtained for the retained catch using the ageing data provided, and the ISMP at-sea length frequency data provided.

| RETAINED CATCH - AGE FREQUENCY MWCVs |  |  |  |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :--- |
|  | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ |  |
| Bight redfish | 34.48 | 36.50 | 27.14 | 26.95 | 25.25 | 28.69 |  |  |
| Blue grenadier | 6.10 | 9.12 | 7.88 | 7.56 | 6.85 | 12.42 |  |  |
| Blue warehou | 6.88 | 10.12 | 7.54 | 7.76 | 7.62 | 20.66 |  |  |
| Blue-eye trevalla | 6.64 | 14.61 | 15.65 | 15.63 | 12.67 | 14.80 |  |  |
| Deepwater flathead | 12.99 | 24.36 | 19.09 | 11.79 | 13.01 | 11.96 |  |  |
| Eastern school whiting | 6.41 | 7.28 | 5.82 | 8.69 | 14.01 | 14.68 | 10.98 |  |
| Gemfish | 6.57 | 8.54 | 7.26 | 9.10 | 9.04 | 10.53 |  |  |
| Jackass morwong | 24.05 | 21.30 | 20.54 | 21.76 | 24.00 | 33.44 |  |  |
| Orange roughy |  |  | 21.92 | 25.38 |  |  |  |  |
| Pink ling | 10.61 | 11.05 | 12.74 | 9.65 | 11.40 | 10.65 |  |  |
| Redfish | 18.88 | 18.96 | 13.57 |  |  |  |  |  |
| Silver warehou | 10.68 | 13.53 | 13.25 | 10.73 | 16.41 | 16.52 | 22.51 |  |
| Tiger flathead | 10.01 | 13.02 | 13.17 | 13.72 |  |  |  |  |

Table 16. A table of the MWCVs for the age frequency estimates obtained for the retained catch using the ageing data provided, and the ISMP at-sea length frequency data provided.

| DISCARDED CATCH - AGE FREQUENCY MWCVs |  |  |  |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :--- |
|  | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ |  |
| Bight redfish | 34.48 | 36.50 | 27.14 | 26.95 | 25.25 | 28.69 |  |  |
| Blue grenadier | 6.10 | 9.12 | 7.88 | 7.56 | 6.85 | 12.42 |  |  |
| Blue warehou | 6.88 | 10.12 | 7.54 | 7.76 | 7.62 | 20.66 |  |  |
| Blue-eye trevalla | 6.64 | 14.61 | 15.65 | 15.63 | 12.67 | 14.80 |  |  |
| Deepwater flathead | 12.99 | 24.36 | 19.09 | 11.79 | 13.01 | 11.96 |  |  |
| Eastern school whiting | 6.41 | 7.28 | 5.82 | 8.69 | 14.01 | 14.68 | 10.98 |  |
| Gemfish | 6.57 | 8.54 | 7.26 | 9.10 | 9.04 | 10.53 |  |  |
| Jackass morwong | 24.05 | 21.30 | 20.54 | 21.76 | 24.00 | 33.44 |  |  |
| Orange roughy |  |  | 21.92 | 25.38 |  |  |  |  |
| Pink ling | 10.61 | 11.05 | 12.74 | 9.65 | 11.40 | 10.65 |  |  |
| Redfish | 18.88 | 18.96 | 13.57 |  |  |  |  |  |
| Silver warehou | 10.68 | 13.53 | 13.25 | 10.73 | 16.41 | 16.52 | 22.51 |  |
| Tiger flathead | 10.01 | 13.02 | 13.17 | 13.72 |  |  |  |  |

Table 17. Presentation of the number of fish sampled for ageing in the SESSF, regardless of whether retained or discarded.

|  | DISCARDED AND RETAINED SAMPLE SIZES FOR AGEING |  |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ |
| Bight redfish | 307 | 245 | 605 | 581 | 540 | 293 | 0 |
| Blue grenadier | 1687 | 1514 | 1484 | 1815 | 1607 | 1574 | 0 |
| Blue warehou | 750 | 297 | 1118 | 743 | 742 | 108 | 15 |
| Blue-eye trevalla | 637 | 1013 | 601 | 1066 | 894 | 410 | 0 |
| Deepwater flathead | 555 | 87 | 207 | 685 | 476 | 459 | 0 |
| Gemfish | 494 | 381 | 469 | 1257 | 800 | 666 | 0 |
| Hapuku | 0 | 0 | 0 | 143 | 210 | 0 | 0 |
| Jackass morwong | 379 | 250 | 557 | 471 | 245 | 120 | 0 |
| Pink ling | 707 | 886 | 688 | 793 | 902 | 1084 | 10 |
| Redfish | 672 | 658 | 684 | 0 | 0 | 0 | 0 |
| Orange roughy | 0 | 0 | 2033 | 800 | 0 | 0 | 0 |
| Eastern school whiting | 560 | 471 | 649 | 393 | 314 | 421 | 68 |
| Silver warehou | 646 | 444 | 639 | 625 | 395 | 316 | 132 |
| Tiger flathead | 1302 | 102 | 326 | 891 | 0 | 0 | 0 |

Table 18. Ageing sample sizes required in order to obtain an MWCV for each of the species listed of $\mathbf{1 0 \%}$, for the retained length frequency distribution.

| Retained | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Bight redfish | 3650 | 3264 | 4455 | 4219 | 3443 | 2412 |  |
| Blue grenadier | 627 | 1258 | 922 | 1038 | 755 | 2429 |  |
| Blue warehou | 355 | 304 | 635 | 448 | 430 | 461 |  |
| Blue-eye trevalla | 281 | 2163 | 1472 | 2604 | 1435 | 898 |  |
| Deepwater flathead | 936 | 516 | 754 | 952 | 805 | 657 |  |
| Eastern school whiting | 203 | 202 | 159 | 949 | 1571 | 1435 |  |
| Gemfish | 163 | 182 | 293 | 390 | 200 | 133 |  |
| Jackass morwong | 4089 | 4022 | 2903 | 3754 | 5196 | 12123 |  |
| Orange roughy |  |  | 3285 |  |  |  |  |
| Pink ling | 1996 | 1692 | 1195 |  |  |  |  |
| Redfish | 737 | 812 | 1122 | 719 | 1064 | 862 | 669 |
| Silver warehou | 1305 | 173 | 566 | 1677 |  |  |  |
| Tiger flathead |  |  |  |  |  |  |  |

Table 19. Ageing sample sizes required in order to obtain an MWCV for each of the species listed of $\mathbf{1 0 \%}$, for the discarded length frequency distribution.

| Discarded | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Bight redfish |  |  | 5228 |  |  |  |  |
| Blue grenadier | 67 | 22051 | 34 | 524 | 657 |  |  |
| Blue warehou | 464 | 368 | 230 | 184 |  |  |  |
| Blue-eye trevalla |  |  | 3921 |  |  |  |  |
| Deepwater flathead | 64 | 157 | 129 | 1482 | 0 | 1927 |  |
| Eastern school whiting | 213 | 160 | 299 | 97 | 181 | 594 |  |
| Gemfish | 3372 | 3182 | 2151 | 3950 | 6786 |  |  |
| Jackass morwong |  |  | 39584 |  |  |  |  |
| Orange roughy |  |  |  |  |  |  |  |
| Pink ling | 999 | 998 | 796 |  |  |  |  |
| Redfish | 678 | 338 | 468 | 530 | 127 |  |  |
| Silver warehou | 2657 | 154 | 429 | 759 |  |  |  |
| Tiger flathead |  |  |  |  |  |  |  |

Table 20. MWCVs associated with the multinomial ageing error from the age-length key, using the length frequency derived from port sampling as a quantity known without error (this quantity obviously has error associated with it, but as discussed elsewhere in this document, the assumption that the If distribution is error free is a device to determine the adequacy of the age sampling).

|  | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Bight redfish |  | 38.14 | 25.06 | 28.98 | 33.81 |  |  |
| Blue grenadier | 26.44 | 14.27 | 5.36 | 6.15 | 6.34 | 5.98 |  |
| Blue warehou | 6.89 | 12.37 | 6.86 | 11.11 | 9.14 | 5.70 |  |
| Blue-eye trevalla | 8.64 | 10.46 | 11.35 | 13.92 | 9.26 | 13.33 |  |
| Deepwater flathead |  | 25.74 | 18.75 | 11.49 | 13.24 |  |  |
| Eastern school whiting | 5.79 | 8.38 | 7.95 | 8.85 | 14.05 | 8.46 | 15.88 |
| Gemfish | 6.13 | 7.62 | 7.24 | 6.46 | 9.32 | 7.78 |  |
| Jackass morwong | 26.78 | 23.74 | 20.39 | 20.60 | 23.14 | 31.68 |  |
| Orange roughy |  |  | 34.55 | 44.37 |  |  |  |
| Pink ling | 12.34 | 12.88 | 13.18 | 11.02 | 11.05 | 8.42 | 10.11 |
| Redfish | 19.24 | 21.97 | 13.87 |  |  |  |  |
| Silver warehou | 10.58 | 12.71 | 11.48 | 10.52 | 21.52 | 17.08 | 20.68 |
| Tiger flathead | 8.55 | 11.91 | 13.38 | 15.83 |  |  |  |

Table 21. Average number of shots per vessel-day per stratum, as derived from the GENLOG data after exclusion of records where vessels straddle two or more strata in the same day.

| AL CSA | 1.648 |
| :---: | :---: |
| AL EBS / NSW | 1.149 |
| AL ESA | 1.534 |
| AL ET / WT | 1.063 |
| AL WBS / SAV | 1.306 |
| AL WSA / WA | 1.468 |
| BS IN TR | 3.485 |
| ECDW TR | 2.400 |
| EDL DS | 4.577 |
| EDL_IN_TR | 3.123 |
| EDL OFF TR | 2.650 |
| GN CSA | 2.185 |
| GN EBS / NSW | 1.756 |
| GN ESA | 1.859 |
| GN ET / WT | 1.414 |
| GN SAV | 1.641 |
| GN WA / WSA | 2.356 |
| GN WBS | 1.593 |
| Inshore | 3.273 |
| Midshore | 2.036 |
| Offshore | 1.707 |
| NSW_IN_TR | 2.471 |
| NSW OFF TR | 2.124 |
| NSW RRP TR | 1.766 |
| SW BGS TR | 1.945 |
| SW ORO TR | 2.298 |
| SW TR | 2.629 |
| TAS BGS_TR | 2.460 |
| TAS_E_TR | 2.979 |
| TAS ORO TR | 2.139 |
| TAS_W TR | 2.492 |
| VIT | 3.558 |

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|  | Birds |  |
| :--- | :---: | :---: |
|  | Days | Shots |
| AL CSA | 2.08 | 3.43 |
| AL EBS / NSW | 0.00 | 0.00 |
| AL ESA | 0.00 | 0.00 |
| AL ET / WT | 0.00 | 0.00 |
| AL WBS / SAV | 0.54 | 0.71 |
| AL WSA / WA | 1.14 | 1.68 |
| BS_IN_TR | 0.00 | 0.00 |
| ECDW_TR | 0.00 | 0.00 |
| EDL_DS | 0.00 | 0.00 |
| EDL_IN_TR | 7.54 | 23.56 |
| EDL_OFF_TR | 6.98 | 18.49 |
| Inshore | 3.38 | 11.07 |
| Midshore | 5.76 | 11.72 |
| NSW_IN_TR | 35.18 | 86.94 |
| NSW_OFF_TR | 3.60 | 7.65 |
| NSW_RRP_TR | 0.00 | 0.00 |
| Offshore | 4.87 | 8.31 |
| SW_BGS_TR | 22.81 | 44.37 |
| SW_ORO_TR | 0.00 | 0.00 |
| SW_TR | 177.70 | 467.23 |
| TAS_BGS_TR | 0.00 | 0.00 |
| TAS_E_TR | 26.57 | 79.16 |
| TAS_ORO_TR | 0.00 | 0.00 |
| TAS_W_TR | 101.84 | 253.78 |
| N | 400 | 1018 |
| CV opt | 14.68 |  |
| CV ave | 21.37 |  |
| CV prop | 27.04 |  |


| Shots <br> prop |
| :---: |
| 14.86 |
| 7.67 |
| 2.28 |
| 10.68 |
| 3.98 |
| 3.96 |
| 5.10 |
| 1.82 |
| 298.80 |
| 148.92 |
| 72.92 |
| 113.06 |
| 6.09 |
| 133.57 |
| 50.36 |
| 8.10 |
| 3.39 |
| 11.89 |
| 0.19 |
| 117.76 |
| 15.18 |
| 60.23 |
| 4.26 |
| 53.90 |
| $\mathbf{1 1 4 9 . 0 0}$ |

Table 23. Optimal design results for TEP species groups. For each species group, the optimal number of days is shown per stratum. The associated optimal number of shots is calculated from the optimal days, using the average number of shots per vessel per day reported in Table \%. The CVs associated with the optimal design are shown, as are the CVs associated with a sample day allocation equal to the average over all species groups. For this version of results additional priors were introduced for strata where there were no interactions for the period 2002 - 2008. These priors were: birds, seals: $10 \%$ of the mean number of interactions per shot in strata where interactions did occur, calculated separately for the AL, SET or GAB sectors of the SESSF, pipefish, sharks and whales / dolphins: $10 \%$ of the mean number of interactions per shot in strata where interactions did occur, regardless of SESSF stratum.


Table 24. Shot allocations modified from Table 22 to include the VIT and gillnet strata on a pro rata basis as described in the text (w.r.t. to the CDR-GENLOG shots in 2007 and 2008). The conversion between shots and days would need to be implemented to see the below in terms of sea-days.

|  | Birds | Pipefish | Seals | Sharks | Whales | Average | Proportion |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Shots | Shots | Shots | Shots | Shots | Shots | Shots |
| AL CSA | 3.43 | 0 | 0 | 0 | 0 | 0.69 | 14.86 |
| AL EBS / NSW | 0 | 0 | 0 | 0 | 0 | 0 | 7.67 |
| AL ESA | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.28 |
| AL ET / WT | 0.00 | 0.00 | 6.37 | 0.00 | 0.00 | 1.27 | 10.68 |
| AL WBS / SAV | 0.71 | 0.00 | 0.00 | 0.00 | 0.00 | 0.14 | 3.98 |
| AL WSA / WA | 1.68 | 0.00 | 0.87 | 0.00 | 0.00 | 0.51 | 3.96 |
| BS_IN_TR | 0.00 | 0.00 | 5.61 | 0.00 | 0.00 | 1.12 | 3.70 |
| ECDW_TR | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 1.32 |
| EDL_DS | 0.00 | 575.75 | 100.33 | 0.00 | 0.00 | 135.22 | 216.52 |
| EDL_IN_TR | 17.07 | 284.54 | 138.40 | 0.00 | 0.00 | 88.00 | 107.91 |
| EDL_OFF_TR | 13.40 | 0.00 | 26.87 | 0.00 | 0.00 | 8.05 | 52.84 |
| Inshore | 11.07 | 56.04 | 36.26 | 0.00 | 271.84 | 75.04 | 113.06 |
| Midshore | 11.72 | 0.00 | 0.00 | 0.00 | 0.00 | 2.34 | 6.09 |
| NSW_IN_TR | 63.00 | 149.59 | 98.14 | 716.25 | 0.00 | 205.40 | 96.79 |
| NSW_OFF_TR | 5.54 | 0.00 | 14.36 | 0.00 | 0.00 | 3.98 | 36.49 |
| NSW_RRP_TR | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 5.87 |
| Offshore | 8.31 | 0.00 | 0.00 | 0.00 | 50.46 | 11.75 | 3.39 |
| SW BGS_TR | 32.15 | 0.00 | 26.99 | 0.00 | 0.00 | 11.83 | 8.62 |
| SW_ORO_TR | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.14 |
| SW TR | 338.57 | 0.00 | 218.87 | 0.00 | 512.09 | 213.91 | 85.33 |
| TAS_BGS_TR | 0.00 | 0.00 | 14.18 | 0.00 | 0.00 | 2.83 | 11.00 |
| TAS_E_TR | 57.36 | 0.00 | 67.80 | 0.00 | 40.23 | 33.08 | 43.64 |
| TAS_ORO_TR | 0.00 | 0.00 | 3.54 | 0.00 | 0.00 | 0.71 | 3.09 |
| TAS_W TR | 183.90 | 0.00 | 63.04 | 0.00 | 0.00 | 49.38 | 39.06 |
| VIT | 10.80 | 15.38 | 11.85 | 10.48 | 11.51 | 12.01 | 12.19 |
| GN CSA | 72.88 | 103.51 | 79.76 | 73.42 | 56.61 | 77.24 | 73.01 |
| GN EBS / NSW | 88.09 | 125.12 | 96.41 | 88.74 | 68.43 | 93.36 | 88.26 |
| GN ESA | 8.64 | 12.27 | 9.46 | 8.70 | 6.71 | 9.16 | 8.66 |
| GN ET / WT | 20.22 | 28.72 | 22.13 | 20.37 | 15.71 | 21.43 | 20.25 |
| GN SAV | 7.72 | 10.97 | 8.45 | 7.78 | 6.00 | 8.18 | 7.74 |
| GN WA / WSA | 51.06 | 72.52 | 55.88 | 51.44 | 39.67 | 54.11 | 51.15 |
| GN WBS | 21.57 | 30.64 | 23.60 | 21.73 | 16.75 | 22.86 | 21.61 |

Table 25. Shot allocations modified from Table 23 to include the VIT and gillnet strata on a pro-rata basis as described in the text (w.r.t. to the CDR-GENLOG shots in 2007 and 2008).

|  | Birds | Pipefish | Seals | Sharks | Whales | Average | Proportion |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Shots | Shots | Shots | Shots | Shots | Shots | Shots |
| AL CSA | 2.81 | 5.28 | 2.55 | 8.98 | 10.85 | 6.09 | 14.86 |
| AL EBS / NSW | 0.52 | 2.28 | 1.1 | 3.87 | 4.68 | 2.49 | 7.67 |
| AL ESA | 0.18 | 0.78 | 0.38 | 1.33 | 1.60 | 0.85 | 2.28 |
| AL ET / WT | 0.69 | 3.05 | 6.27 | 5.18 | 6.26 | 4.29 | 10.68 |
| AL WBS / SAV | 0.58 | 1.26 | 0.61 | 2.14 | 2.59 | 1.43 | 3.98 |
| AL WSA / WA | 1.37 | 1.33 | 0.86 | 2.26 | 2.73 | 1.71 | 3.96 |
| BS_IN_TR | 3.08 | 1.93 | 5.52 | 3.29 | 3.98 | 3.56 | 3.70 |
| ECDW_TR | 0.88 | 0.55 | 0.66 | 0.93 | 1.13 | 0.83 | 1.32 |
| EDL_DS | 203.57 | 457.97 | 98.84 | 217.49 | 262.64 | 248.10 | 216.52 |
| EDL_IN_TR | 13.97 | 226.31 | 136.33 | 89.54 | 108.12 | 114.86 | 107.91 |
| EDL_OFF_TR | 10.96 | 23.78 | 26.47 | 40.46 | 48.86 | 30.11 | 52.84 |
| Inshore | 9.06 | 44.59 | 35.72 | 96.33 | 81.02 | 53.34 | 113.06 |
| Midshore | 9.59 | 2.40 | 0.48 | 4.09 | 4.94 | 4.30 | 6.09 |
| NSW_IN_TR | 51.57 | 119.01 | 96.67 | 225.96 | 86.29 | 115.90 | 96.79 |
| NSW_OFF_TR | 4.54 | 14.69 | 14.14 | 24.99 | 30.18 | 17.71 | 36.49 |
| NSW_RRP_TR | 3.41 | 2.14 | 2.58 | 3.64 | 4.40 | 3.23 | 5.87 |
| Offshore | 6.80 | 1.23 | 0.24 | 2.09 | 15.04 | 5.08 | 3.39 |
| SW BGS_TR | 26.32 | 3.30 | 26.57 | 5.61 | 6.77 | 13.71 | 8.62 |
| SW_ORO_TR | 0.09 | 0.06 | 0.07 | 0.09 | 0.12 | 0.09 | 0.14 |
| SW_TR | 277.19 | 38.02 | 215.53 | 64.70 | 152.18 | 149.52 | 85.33 |
| TAS_BGS_TR | 7.58 | 4.76 | 13.97 | 8.09 | 9.78 | 8.83 | 11.00 |
| TAS_E_TR | 46.94 | 20.80 | 66.78 | 35.40 | 11.99 | 36.38 | 43.64 |
| TAS_ORO_TR | 2.00 | 1.25 | 3.48 | 2.14 | 2.58 | 2.29 | 3.09 |
| TAS_W_TR | 150.49 | 16.93 | 62.10 | 28.81 | 34.80 | 58.62 | 39.06 |
| VIT | 10.80 | 15.38 | 11.85 | 10.48 | 11.51 | 12.01 | 12.19 |
| GN CSA | 82.27 | 95.48 | 78.90 | 76.99 | 78.29 | 82.39 | 73.01 |
| GN EBS / NSW | 99.44 | 115.41 | 95.37 | 93.07 | 94.63 | 99.58 | 88.26 |
| GN ESA | 9.75 | 11.32 | 9.35 | 9.13 | 9.28 | 9.77 | 8.66 |
| GN ET / WT | 22.82 | 26.49 | 21.89 | 21.36 | 21.72 | 22.85 | 20.25 |
| GN SAV | 8.72 | 10.12 | 8.36 | 8.16 | 8.30 | 8.73 | 7.74 |
| GN WA / WSA | 57.64 | 66.89 | 55.28 | 53.94 | 54.85 | 57.72 | 51.15 |
| GN WBS | 24.35 | 28.26 | 23.35 | 22.79 | 23.17 | 24.38 | 21.61 |
|  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |

Table 26. Estimates of the number of individuals interacting with the SESSF, by year and TEP species group, dead and alive combined.

|  | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Seals | 1518.2 | 326.8 | 3272.9 | 7039.1 | 7341.1 | 6390.0 | 2866.7 |
| CV | 34.2 | 37.0 | 24.3 | 22.6 | 28.1 | 34.5 | 29.0 |
| Birds | 0 | 0 | 23078.1 | 20261.8 | 91913.8 | 7891.0 | 704.9 |
| CV | - | - | 23.6 | 32.6 | 26.2 | 83.3 | 81.0 |
| Pipefish | 0 | 1610.5 | 4378.7 | 447.4 | 63.0 | 1809.4 | 2746.9 |
| CV |  | 44.4 | 38.8 | 46.1 | 100.0 | 78.1 | 33.8 |
| Sharks | 0 | 33.1 | 105.1 | 0 | 31.5 | 0 | 0 |
| CV |  | 100.0 | 74.4 |  | 100.0 |  |  |
| Whales | 0 | 0 | 4649.8 | 1311.6 | 13029.1 | 30.4 | 357.3 |
| CV |  |  | 83.0 | 100.0 | 112.7 | 100.4 | 100.0 |

Table 27. Estimates of the number of individuals interacting with the SESSF, by year and TEP species group, dead only.

|  | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Seals | 1035.5 | 229.8 | 551.6 | 1291.8 | 739.1 | 1061.7 | 1983.9 |
| CV | 47.1 | 42.5 | 53.2 | 28.1 | 28.3 | 33.8 | 39.4 |
| Birds | 0 | 0 | 311.6 | 538.0 | 1341.3 | 323.1 | 589.2 |
| CV |  |  | 62.8 | 41.1 | 31.8 | 54.4 | 95.6 |
| Pipefish | 0 | 1610.5 | 3338.7 | 447.4 | 63.0 | 1809.4 | 2746.9 |
| CV |  | 44.4 | 35.9 | 46.1 | 100.0 | 78.1 | 33.8 |
| Sharks | 0 | 33.1 | 70.1 | 0 | 31.5 | 0 | 0 |
| CV |  | 100.0 | 70.6 |  | 100.0 |  |  |
| Whales | 0 | 0 | 0 | 0 | 0 | 30.4 | 357.3 |
| CV |  |  |  |  |  | 100.4 | 100.0 |

Table 28. A list of the port names which are to be found in the data arising from the CDR/GENLOG merge.

| 1 | ADELAIDE | 29 | DUNALLEY | 57 | NELSON BAY | 85 | SORRENTO |
| :---: | :--- | :---: | :--- | :---: | :--- | :--- | :--- |
| 2 | ALBANY | 30 | EAGLEHAWK NECK | 58 | NELSON NZ | 86 | SOUTHEND |
| 3 | AMERICAN RIVER | 31 | EDEN | 59 | NEWCASTLE | 87 | SOUTHPORT |
| 4 | APOLLO BAY | 32 | EDINBURGH | 60 | NEWHAVEN | 88 | SOUTHPORT TAS |
| 5 | AUCKLAND | 33 | EMU POINT | 61 | NORTH ARMS | 89 | STANLEY |
| 6 | BANDY CREEK | 34 | ESPERANCE | 62 | NUBEENA | 90 | ST HELENS |
| 7 | BATEMANS BAY | 35 | EUCLA | 63 | PIRATES BAY | 91 | STRAHAN |
| 8 | BEACHPORT | 36 | FOWLERS BAY | 64 | POINT TURTON | 92 | STREAKY BAY |
| 9 | BEAUTY POINT | 37 | FREMANTLE | 65 | PONDALOWIE <br> BAY | 93 | SUNSET COVE |
| 10 | BELL BAY | 38 | GEORGETOWN | 66 | PORT ADELAIDE | 94 | SYDNEY |
| 11 | BERMAGUI | 39 | GRASSY | 67 | PORT ALBERT | 95 | TERRIGAL |
| 12 | BICHENO | 40 | GREENWELL POINT | 68 | PORT ARTHUR | 96 | THEVENARD |
| 13 | BLACKFELLOWS | 41 | HOBART | 69 | PORT FAIRY | 97 | TRIABUNNA |
| 14 | BRIDPORT | 42 | KANGAROO POINT | 70 | PORT FRANKLIN | 98 | ULLADULLA |
| 15 | BRISBANE | 43 | KETTERING | 71 | PORT HUON | 99 | VENUS BAY |
| 16 | BROKENDOWN | 44 | KILLECRANKIE | 72 | PORT JACKSON | 100 | VICTOR <br> HARBOUR <br> 17 BURNIE |

Table 29. A listing of the $\mathbf{3 2}$ strata that arise if one uses the $\mathbf{2 5}$ strata from the survey design optimisation exercise, VIT, and the 7 gillnet strata shown.

| 1 | AL CSA | AL | 17 | BS_IN_TR | SET |
| :---: | :--- | :---: | :---: | :--- | :---: |
| 2 | AL EBS / NSW | AL | 18 | ECDW_TR | SET |
| 3 | AL ESA | AL | 19 | EDL_DS | SET |
| 4 | AL ET / WT | AL | 20 | EDL_IN_TR | SET |
| 5 | AL WBS / SAV | AL | 21 | EDL_OFF_TR | SET |
| 6 | AL WSA / WA | AL | 22 | NSW_IN_TR | SET |
| 7 | GN CSA | GN | 23 | NSW_OFF_TR | SET |
| 8 | GN EBS / NSW | GN | 24 | NSW_RRP_TR | SET |
| 9 | GN ESA | GN | 25 | SW_BGS_TR | SET |
| 10 | GN ET / WT | GN | 26 | SW_ORO_TR | SET |
| 11 | GN SAV | GN | 27 | SW_TR_ | SET |
| 12 | GN WA / WSA | GN | 28 | TAS_BGS_TR | SET |
| 13 | GN WBS | GN | 29 | TAS_E_TR | SET |
| 14 | GAB Inshore | GAB | 30 | TAS_ORO_TR | SET |
| 15 | GAB Midshore | GAB | 31 | TAS_W_TR | SET |
| 16 | GAB Offshore | GAB | 32 | VIT | VIT |

Table 30. AL (6 clusters) - Cluster Membership for ports.

| Port Unloaded | Cluster | Port Unloaded | Cluster |
| :---: | :---: | :---: | :---: |
| CAPE JAFFA | AL-1 | PORT ALBERT | AL-4 |
| ADELAIDE | AL-1 | BEAUTY POINT | AL-4 |
| AMERICAN RIVER | AL-1 | MOOLOOLABA | AL-4 |
| ROBE | AL-1 | BRIDPORT | AL-4 |
| PORT ADELAIDE | AL-1 | COFFS HARBOUR | AL-4 |
| NORTH ARMS | AL-1 | NELSON BAY | AL-4 |
| KINGSCOTE | AL-1 | DEVONPORT | AL-4 |
| VIVONNE BAY | AL-1 | LAKES ENTRANCE | AL-4 |
| EMU POINT | AL-1 | EDEN | AL-4 |
| CAPE JERVIS | AL-1 | GEORGETOWN | AL-4 |
| $\begin{gathered} \hline \text { VICTOR } \\ \text { HARBOUR } \\ \hline \end{gathered}$ | AL-1 | KILLECRANKIE | AL-4 |
| MELBOURNE | AL-1 | PORT STEPHENS | AL-4 |
| WIRRINA COVE | AL-1 | PORT WELSHPOOL | AL-4 |
| APOLLO BAY | AL-2 | SYDNEY | AL-4 |
| SOUTHPORT TAS | AL-2 | MUSSEL ROE BAY | AL-4 |
| STANLEY | AL-2 | PORTLAND | AL-5 |
| GRASSY | AL-2 | SOUTHEND | AL-5 |
| STRAHAN | AL-2 | $\begin{gathered} \text { PORT } \\ \text { MACDONNELL } \end{gathered}$ | AL-5 |
| SAN REMO | AL-2 | BEACHPORT | AL-5 |
| WARRNAMBOOL | AL-2 | BICHENO | AL-6 |
| KING ISLAND | AL-2 | COLES BAY | AL-6 |
| PORT FAIRY | AL-2 | DOVER | AL-6 |
| PORT HUON | AL-2 | NUBEENA | AL-6 |
| QUEENSCLIFFE | AL-2 | ST HELENS | AL-6 |
| SMITHTON | AL-2 | PORT ARTHUR | AL-6 |
| STREAKY BAY | AL-3 | KETTERING | AL-6 |
| PORT KENNY | AL-3 | HOBART | AL-6 |
| PORT LINCOLN | AL-3 | $\begin{gathered} \hline \text { EAGLEHAWK } \\ \text { NECK } \\ \hline \end{gathered}$ | AL-6 |
| VENUS BAY | AL-3 | TRIABUNNA | AL-6 |
| COFFIN BAY | AL-3 | ULLADULLA | AL-6 |
| THEVENARD | AL-3 | DUNALLEY | AL-6 |
|  |  | MARGATE | AL-6 |

Table 31. AL (6 clusters) - Mean Cluster Percentages. These are the mean proportion of shots falling into different strata for each cluster.

|  | AL CSA | AL ESA | AL ET / WT | AL WSA / WA | AL EBS / NSW | AL WBS / SAV |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| AL-1 | 0.923 | 0.022 | 0.000 | 0.054 | 0.000 | 0.002 |
| AL-2 | 0.019 | 0.000 | 0.021 | 0.000 | 0.012 | 0.948 |
| AL-3 | 0.147 | 0.006 | 0.000 | 0.761 | 0.000 | 0.085 |
| AL-4 | 0.006 | 0.001 | 0.061 | 0.002 | 0.916 | 0.014 |
| AL-5 | 0.182 | 0.707 | 0.000 | 0.004 | 0.001 | 0.106 |
| AL-6 | 0.001 | 0.000 | 0.884 | 0.000 | 0.067 | 0.047 |

Table 32. GN (7 clusters) - Cluster Membership.

| SOUTHEND | GN-1 | KILLECRANKIE | GN-3 |
| :---: | :---: | :---: | :---: |
| NEWHAVEN | GN-1 | MUSSEL ROE BAY | GN-3 |
| SAN REMO | GN-1 | LADY BARRON | GN-3 |
| APOLLO BAY | GN-1 | $\begin{aligned} & \hline \text { PORT } \\ & \text { WELSHPOOL } \end{aligned}$ | GN-3 |
| CURRIE | GN-1 | EDEN | GN-3 |
| QUEENSCLIFFE | GN-1 | ST HELENS | GN-3 |
| WILLIAMSTOWN | GN-1 | PORT ALBERT | GN-3 |
| STRAHAN | GN-1 | WHITEMARK | GN-3 |
| WYNYARD | GN-1 | GEORGETOWN | GN-3 |
| MELBOURNE | GN-1 | PORT FRANKLIN | GN-3 |
| STANLEY | GN-1 | EUCLA | GN-4 |
| EDINBURGH | GN-2 | CEDUNA | GN-4 |
| ADELAIDE | GN-2 | COFFIN BAY | GN-4 |
| AMERICAN RIVER | GN-2 | FOWLERS BAY | GN-4 |
| BEACHPORT | GN-2 | STREAKY BAY | GN-4 |
| VIVONNE BAY | GN-2 | THEVENARD | GN-4 |
| WIRRINA COVE | GN-2 | VENUS BAY | GN-4 |
| KINGSCOTE | GN-2 | BICHENO | GN-5 |
| MARION BAY | GN-2 | KETTERING | GN-5 |
| CAPE JAFFA | GN-2 | HOBART | GN-5 |
| KANGAROO POINT | GN-2 | CREMORNE | GN-5 |
| NORTH ARMS | GN-2 | PIRATES BAY | GN-5 |
| POINT TURTON | GN-2 | DOVER | GN-5 |
| PONDALOWIE BAY | GN-2 | DUNALLEY | GN-5 |
| PORT ADELAIDE | GN-2 | EAGLEHAWK NECK | GN-5 |
| PORT LINCOLN | GN-2 | MARGATE | GN-5 |
| ROBE | GN-2 | DODGES FERRY | GN-5 |
| SUNSET COVE | GN-2 | SOUTHPORT | GN-5 |
| VICTOR HARBOUR | GN-2 | TRIABUNNA | GN-5 |
| LAKES ENTRANCE | GN-3 | BLACKFELLOWS | GN-6 |
| BEAUTY POINT | GN-3 | PORT <br> MACDONNELL | GN-6 |
| BELL BAY | GN-3 | PORT FAIRY | GN-7 |
| BRIDPORT | GN-3 | WARRNAMBOOL | GN-7 |
| DEVONPORT | GN-3 | PORTLAND | GN-7 |

Table 33. GN (7 clusters) - Mean Cluster Percentages. These are the mean proportion of shots falling into different strata for each cluster.

|  | GN CSA | GN ESA | GN WBS | GN SAV | GN WA / <br> WSA | GN ET / <br> WT | GN EBS <br> / NSW |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| GN-1 | 0.000 | 0.000 | 0.895 | 0.005 | 0.006 | 0.004 | 0.089 |
| GN-2 | 0.936 | 0.029 | 0.000 | 0.000 | 0.034 | 0.000 | 0.000 |
| GN-3 | 0.000 | 0.000 | 0.029 | 0.000 | 0.000 | 0.063 | 0.907 |
| GN-4 | 0.061 | 0.001 | 0.000 | 0.000 | 0.937 | 0.000 | 0.000 |
| GN-5 | 0.000 | 0.000 | 0.021 | 0.000 | 0.000 | 0.973 | 0.006 |
| GN-6 | 0.006 | 0.888 | 0.005 | 0.101 | 0.000 | 0.000 | 0.000 |
| GN-7 | 0.006 | 0.034 | 0.046 | 0.816 | 0.007 | 0.003 | 0.088 |

Table 34. GAB (3 clusters) - Cluster Membership.

| ADELAIDE | GAB-1 |
| :--- | :--- |
| BANDY CREEK | GAB-1 |
| CEDUNA | GAB-1 |
| ESPERANCE | GAB-1 |
| HOBART | GAB-1 |
| LAKES <br> ENTRANCE | GAB-1 |
| MELBOURNE | GAB-1 |
| PORT ADELAIDE | GAB-1 |
| PORT ALBERT | GAB-1 |
| PORT LINCOLN | GAB-1 |
| PORT <br> WELSHPOOL | GAB-1 |
| STRAHAN | GAB-1 |
| THEVENARD | GAB-1 |
| FREMANTLE | GAB-2 |
| PORTLAND | GAB-2 |
| ALBANY | GAB-3 |

Table 35. GAB (3 clusters) - Mean Cluster Percentages. These are the mean proportion of shots falling into different strata for each cluster.

|  | GAB-Inshore | GAB- <br> Midshore | GAB- <br> offshore |
| :---: | :---: | :---: | :---: |
| GAB-1 | 0.885 | 0.058 | 0.057 |
| GAB-2 | 0.039 | 0.575 | 0.385 |
| GAB-3 | 0.101 | 0.009 | 0.890 |

Table 36. SET ( 15 clusters) - Cluster Membership.

| WOOLGOOLGA | SET-1 | PORT KEMBLA | SET-8 |
| :---: | :---: | :---: | :---: |
| ADELAIDE | SET-2 | MOOLOOLABA | SET-8 |
| PORTLAND | SET-2 | BRISBANE | SET-8 |
| PORT ADELAIDE | SET-2 | AUCKLAND | SET-8 |
| EDEN | SET-3 | NELSON | SET-9 |
| WILLIAMSTOWN | SET-3 | BURNIE | SET-9 |
| DEVONPORT | SET-4 | STANLEY | SET-9 |
| HOBART | SET-4 | NELSON NZ | SET-9 |
| BELL BAY | SET-4 | NAMBUCCA HEADS | SET-10 |
| PORT ALBERT | SET-5 | DOVER | SET-11 |
| PORT FAIRY | SET-6 | KETTERING | SET-11 |
| PORT MACDONNELL | SET-6 | MELBOURNE | SET-11 |
| SOUTHEND | SET-6 | PORT <br> MELBOURNE | SET-11 |
| ROBE | SET-6 | STRAHAN | SET-11 |
| VICTOR HARBOUR | SET-6 | BEAUTY POINT | SET-12 |
| PORT JACKSON | SET-7 | $\begin{aligned} & \hline \text { PORT } \\ & \text { WELSHPOOL } \end{aligned}$ | SET-12 |
| GREENWELL POINT | SET-7 | APOLLO BAY | SET-13 |
| BROKENDOWN | SET-7 | SORRENTO | SET-13 |
| BERMAGUI | SET-7 | SAN REMO | SET-13 |
| BATEMANS BAY | SET-7 | QUEENSCLIFFE | SET-13 |
| NEWCASTLE | SET-7 | LAKES | SET-13 |
| TERRIGAL | SET-7 | LAKES <br> ENTRANCE | SET-13 |
| ULLADULLA | SET-7 | BEACHPORT | SET-14 |
| WOLLONGONG | SET-7 | PORT LINCOLN | SET-14 |
| SYDNEY | SET-7 | ST HELENS | SET-15 |
|  |  | BICHENO | SET-15 |
|  |  | CEDUNA | SET-15 |
|  |  | PORT ARTHUR | SET-15 |
|  |  | PIRATES BAY | SET-15 |
|  |  | TRIABUNNA | SET-15 |

Table 37. SET ( 15 clusters) - Main Strata per Port Cluster.

|  | Main Stratum |  |
| :--- | :---: | :---: |
| SET-1 | NSW_RRP TR |  |
| SET-2 | SW_TR |  |
| SET-3 | EDL_OFF TR |  |
| SET-4 | TAS_E_TR |  |
| SET-5 | BS_IN_TR | EDL_DS |
| SET-6 | SW_TR |  |
| SET-7 | NSW_IN_TR |  |
| SET-8 | ECDW_TR |  |
| SET-9 | TAS_BGS_TR |  |
| SET-10 | NSW_OFF_TR |  |
| SET-11 | TAS_W_TR |  |
| SET-12 | EDL_IN_TR |  |
| SET-13 | EDL_DS |  |
| SET-14 | SW_TR |  |
| SET-15 | TAS_E_TR |  |

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Table 38. SET (15 port clusters) - Mean Cluster Percentages. These are the mean proportion of shots falling into different strata for each cluster.

|  | $\begin{gathered} \mathrm{SW} \\ \mathrm{TR} \end{gathered}$ | $\begin{gathered} \text { EDL } \\ \text { DS } \end{gathered}$ | $\underset{\text { TR }}{\text { ECDW }}$ | $\begin{aligned} & \text { TAS } \\ & \text { E_TR } \\ & \hline \end{aligned}$ | $\begin{gathered} \text { BS } \\ \text { IN_TR } \end{gathered}$ | $\begin{aligned} & \text { TAS } \\ & \text { W_TR } \end{aligned}$ | $\begin{aligned} & \text { EDL } \\ & \text { IN_TR } \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { NSW } \\ & \text { IN_TR } \\ & \hline \end{aligned}$ | $\begin{gathered} \text { SW } \\ \text { ORO_TR } \end{gathered}$ | $\begin{gathered} \text { SW } \\ \text { BGS TR } \end{gathered}$ | $\begin{aligned} & \text { EDL } \\ & \text { OFF_TR } \end{aligned}$ | $\begin{aligned} & \text { NSW } \\ & \text { OFF_TR } \end{aligned}$ | $\begin{gathered} \text { TAS } \\ \text { ORO_TR } \end{gathered}$ | $\begin{gathered} \text { NSW } \\ \text { RRP TR } \end{gathered}$ | $\begin{gathered} \text { TAS } \\ \text { BGS_TR } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SET-1 |  |  |  |  |  |  |  |  |  |  |  |  |  | 1.000 |  |
| SET-2 | 0.716 |  |  | 0.023 |  | 0.041 |  |  | 0.051 | 0.156 |  |  | 0.001 |  | 0.011 |
| SET-3 |  | 0.002 | 0.001 | 0.021 | 0.001 |  | 0.402 | 0.091 |  |  | 0.457 | 0.024 |  |  |  |
| SET-4 | 0.015 |  | 0.005 | 0.333 | 0.004 | 0.179 | 0.051 |  |  | 0.006 | 0.050 | 0.011 | 0.125 |  | 0.222 |
| SET-5 |  | 0.500 |  |  | 0.500 |  |  |  |  |  |  |  |  |  |  |
| SET-6 | 1.000 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SET-7 |  | 0.005 | 0.034 |  |  |  | 0.001 | 0.720 |  |  |  | 0.150 |  | 0.089 |  |
| SET-8 |  |  | 1.000 |  |  |  |  |  |  |  |  |  |  |  |  |
| SET-9 |  |  |  |  |  | 0.116 |  |  |  |  |  |  |  |  | 0.884 |
| $\begin{array}{\|l} \hline \text { SET- } \\ 10 \\ \hline \end{array}$ |  |  |  |  |  |  |  |  |  |  |  | 1.000 |  |  |  |
| $\begin{array}{\|l} \hline \text { SET- } \\ 11 \\ \hline \end{array}$ | 0.052 | 0.042 |  | 0.020 | 0.001 | 0.782 |  |  | 0.004 | 0.004 |  |  | 0.008 |  | 0.087 |
| $\begin{aligned} & \hline \text { SET- } \\ & 12 \\ & \hline \end{aligned}$ | 0.010 | 0.126 |  | 0.203 | 0.006 | 0.064 | 0.466 |  |  |  | 0.080 | 0.001 | 0.001 |  | 0.043 |
| $\begin{array}{\|l\|} \hline \text { SET- } \\ 13 \\ \hline \end{array}$ | 0.004 | 0.936 |  | 0.005 | 0.009 | 0.001 | 0.030 |  |  |  | 0.014 |  |  |  |  |
| $\begin{aligned} & \hline \text { SET- } \\ & 14 \\ & \hline \end{aligned}$ | 0.860 |  |  | 0.004 |  | 0.047 |  |  | 0.002 | 0.088 |  |  |  |  |  |
| $\begin{array}{\|l\|} \hline \text { SET- } \\ 15 \\ \hline \end{array}$ | 0.024 | 0.010 |  | 0.858 | 0.008 | 0.029 | 0.030 |  |  |  | 0.035 |  | 0.001 |  | 0.005 |

Table 39-M17. VIT PORTS

| VIT' PORTS |
| :--- |
| APOLLO BAY |
| EDEN |
| HOBART |
| LAKES ENTRANCE |
| PORT ALBERT |
| PORT |
| MELBOURNE |
| PORT WELSHPOOL |
| PORTLAND |
| QUEENSCLIFFE |
| SAN REMO |
| SORRENTO |
| ULLADULLA |

Table 40. A breakdown, by species and year, of the MWCVs of the length frequencies obtained from the sea based ISMP sampling data, for the retained component of the LF distribution.

| SPECIES | Retained |  |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ |
| Alfonsino | 24.6 | 65.5 | 25.2 | 64.4 | 84.5 | 80.2 | 40.8 |
| Bight redfish | 20.7 | 18.1 | 19.7 | 14.9 | 15.4 | 28.4 | 24.7 |
| Blue grenadier | 16.8 | 25.5 | 12.7 | 36.6 | 25.4 | 60.9 | 29.8 |
| Blue warehou | 36.4 | 58.5 | 29.5 | 15.2 | 27.0 | 59.1 | 45.2 |
| Blue-eye trevalla | 147.1 | 23.5 | 20.1 | 32.9 | 40.7 | 31.1 | 46.5 |
| Deepwater flathead | 13.6 | 16.8 | 15.7 | 11.4 | 17.9 | 26.1 | 20.5 |
| Deepwater shark basket |  | 167.1 | 47.9 | 7.9 | 90.2 |  |  |
| Eastern school whiting | 14.2 | 16.1 | 21.5 | 16.5 | 27.1 | 22.4 | 30.9 |
| Gemfish | 32.8 | 28.1 | 23.9 | 52.7 | 28.9 | 75.2 | 50.3 |
| Gummy shark | 70.2 | 74.8 | 69.2 | 87.5 |  | 83.7 | 112.9 |
| Inshore ocean perch | 33.3 | 51.3 | 26.7 | 56.9 | 33.4 | 39.7 | 26.9 |
| Jackass morwong | 13.6 | 26.7 | 16.5 | 10.7 | 14.9 | 31.4 | 19.2 |
| John dory | 46.7 | 17.3 | 16.5 | 28.3 | 27.7 | 37.1 | 25.5 |
| Mirror dory | 18.3 | 25.1 | 12.7 | 15.0 | 22.3 | 62.7 | 35.5 |
| Orange roughy | 14.4 | 12.0 | 11.3 | 10.7 | 10.4 | 34.1 | 28.6 |
| Oreo basket | 99.0 | 54.1 | 55.8 | 23.0 |  | 88.7 |  |
| Pink ling | 19.7 | 14.4 | 41.3 | 18.1 | 20.2 | 42.1 | 34.9 |
| Redfish | 15.1 | 13.4 | 14.8 | 14.2 | 10.4 | 29.9 | 32.4 |
| Ribaldo |  | 38.6 | 28.6 | 52.4 | 45.6 | 78.8 | 56.2 |
| Royal red prawn | 10.7 | 19.0 | 12.4 | 21.2 | 23.5 | 27.2 |  |
| Sawshark basket | 64.0 | 90.3 | 73.0 |  | 2.9 | 74.5 | 143.6 |
| School shark | 4.5 | 48.4 | 61.7 | 41.7 | 28.3 | 83.7 | 32.1 |
| Silver trevally | 18.8 | 21.1 | 15.3 | 19.3 | 24.9 | 90.0 | 29.4 |
| Silver warehou | 24.5 | 21.8 | 27.2 | 17.0 | 21.2 | 33.6 | 34.4 |
| Tiger flathead | 27.3 | 7.4 | 12.3 | 9.5 | 6.0 | 22.5 | 15.6 |
| Yellowspotted boarfish | 25.9 | 27.8 | 20.4 | 26.6 | 37.3 | 25.7 |  |

Table 41. A breakdown, by species and year, of the MWCVs of the length frequencies obtained from the sea based ISMP sampling data, for the discarded component of the LF distribution.

| SPECIES | Discarded |  |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ |
| Alfonsino |  |  |  |  |  | 100.0 | 72.3 |
| Bight redfish |  | 59.6 |  |  |  | 7.3 |  |
| Blue grenadier | 64.7 | 45.5 | 34.9 | 25.8 | 17.3 | 100.0 | 71.7 |
| Blue warehou |  |  | 83.3 | 28.8 | 34.9 |  | 41.0 |
| Blue-eye trevalla | 30.0 | 22.1 | 55.7 | 59.4 | 17.9 | 78.6 |  |
| Deepwater flathead | 19.5 | 27.0 | 58.5 | 46.4 |  |  |  |
| Deepwater shark basket | 52.4 | 35.9 | 32.9 | 47.9 | 98.3 | 81.1 | 57.2 |
| Eastern school whiting | 66.8 | 71.8 | 52.9 |  |  | 10.5 | 30.6 |
| Gemfish | 36.7 | 28.0 | 31.5 | 20.4 | 35.4 | 38.1 | 66.6 |
| Gummy shark | 49.5 | 36.8 | 61.8 | 19.9 | 23.5 | 6.4 | 63.8 |
| Inshore ocean perch | 89.4 | 100.0 | 2.4 |  |  |  | 54.0 |
| Jackass morwong | 30.0 | 22.5 | 27.1 | 86.8 | 48.7 | 1.4 | 41.2 |
| John dory | 52.6 | 31.9 | 14.6 | 37.8 | 96.6 | 100.0 | 39.1 |
| Mirror dory | 123.3 | 87.5 |  |  | 96.0 | 62.3 |  |
| Orange roughy |  | 47.1 |  |  |  | 0.0 | 42.2 |
| Oreo basket | 16.4 | 16.5 | 18.8 | 29.8 | 55.9 |  | 81.8 |
| Pink ling |  |  |  |  | 12.9 |  | 104.1 |
| Redfish |  |  |  |  |  |  |  |
| Ribaldo |  |  |  |  |  | 100.0 | 34.2 |
| Royal red prawn | 75.0 | 0.0 | 16.7 |  | 0.0 | 0.0 |  |
| Sawshark basket | 0.0 |  | 84.0 | 50.9 |  |  | 19.8 |
| School shark | 29.8 | 40.3 | 31.4 | 44.2 | 42.1 |  | 32.9 |
| Silver trevally | 20.8 | 29.9 | 36.4 | 21.3 | 71.9 | 19.1 |  |
| Silver warehou |  |  |  |  |  |  |  |
| Tiger flathead |  |  |  |  |  |  |  |
| Yellowspotted boarfish |  |  |  |  |  |  |  |

Table 42 - Data sets required for design of ISMP sampling regime.

| Data set | Brief description | Custodian |
| :---: | :---: | :---: |
| Catch and Effort logbook data for all SESSF sectors and all years until December 2008. | Records of catch by species, effort and gear used for each fishing event. | AFMA |
| Catch Disposal Record data from all SESSF sectors and all years until December 2008. | Records of landed catch by species for each fishing event trip. | AFMA |
| AFMA Observer data for all SESSF sectors and all years until December 2008. | Independent on-board observations of retained and discarded catch weights and biological (length and otoliths) sampling. | AFMA |
| Fishery Independent survey data for all SESSF sectors where available. | Statistically robust estimates of relative abundance of species caught and biological sampling. | AFMA |
| Ecological Risk Assessment database. | List of species caught by each sector and the relative risks to those species by each sector. | CSIRO |
| ISMP onboard observer data for all SESSF sectors and all years until December 2008. | Independent on-board observations of retained and discarded catch weights, fishing effort and biological (length and otoliths) sampling. | AFMA/Fisheries Victoria |
| ISMP port-based observer data for all SESSF sectors and all years until December 2008. | Independent port-based observations of biological (length and otoliths) sampling of catches. | AFMA/Fisheries Victoria |
| ISMP PET interactions data for all SESSF sectors and all years until December 2008. | Independent on-board observations of PET interactions. | AFMA/Fisheries Victoria |
| CAF aging database for all SESSF species and all years until December 2008. | Biological data (e.g. age and length) from sampled catches. | AFMA/Fisheries Victoria |
| Shark database. | Historical records of shark catch by species, effort and gear used for each fishing event. | Fisheries Victoria |

Table 43 - The breakdown of the SESSF into fisheries and gears.

| Fishery | Sub-Fishery | Gear | Strata |
| :---: | :---: | :--- | :--- |
| SESSF | SET | Danish Seine (DS) | 1 |
|  | GIT | Board Trawl (TW) | 14 |
|  | GAB | Trawl | 1 |
|  | GHAT | TrawlLine. Includes <br> Automatic longline (ALL); other <br> longline (LL); <br> drop line (DL); <br> bottom line (BL) | 6 |
|  |  | Gillnet (GN) |  |
|  |  | 7 |  |

Table 44. Quota species in the SESSF.

| Project Key | CAAB | Common Name | Scientific Name |
| :---: | :---: | :---: | :---: |
| ALFONSINO | 37258002 | Alfonsino | Bervx splendens |
| BIGHT REDFISH | 37258004 | Bight redfish | Centrobervx gerrardi |
| BLUE GRENADIER | 37227001 | Blue Grenadier | Macruronus novaezelandiae |
| BLUE WAREHOU | 37445005 | Blue Warehou | Seriolella brama |
| BLUE-EYE TREVALLA | 37445001 | Blue-eve Trevalla | Hvperoglvphe antarctica |
| BOARFISH (UNSPECIFIED) | 37367000 | Boarfishes | Pentacerotidae |
| DEEPWATER FLATHEAD | 37296002 | Deepwater flathead | Neoplatvcephalus conatus |
| DEEPWATER SHARK BASKET | 37020905 | Platypus shark (mixed) | Deania calcea \& quadrispinosa |
| DEEPWATER SHARK BASKET | 37020003 | Brier Shark | Deania calcea |
| DEEPWATER SHARK BASKET | 37020004 | Longsnout Dogfish | Deania quadrispinosa |
| DEEPWATER SHARK BASKET | 37020013 | Plunket's Dogfish | Centroscvmnus plunketi |
| DEEPWATER SHARK BASKET | 37020021 | Southern lanternshark | Etmopterus granulosus |
| DEEPWATER SHARK BASKET | 37020904 | Roughskin dogfish | Centroscvmnus \& Deania spp |
| DEEPWATER SHARK BASKET | 37020906 | Deepwater dogfish | Centroscymnus spp. |
| DEEPWATER SHARK BASKET | 37020907 | Lantern shark (mixed) | Etmopterus spp |
| EASTERN SCHOOL WHITING | 37330014 | Eastern School Whiting | Sillago flindersi |
| ELEPHANTFISH | 37043001 | Elephantfish | Callorhinchus milii |
| GEMFISH | 37439002 | Gemfish | Rexea solandri |
| GUMMY SHARK | 37017001 | Gummy shark | Mustelus antarcticus |
| INSHORE OCEAN PERCH | 37287001 | Reef Ocean Perch | Helicolenus Percoides |
| JACKASS MORWONG | 37377003 | Jackass Morwong | Nemadactvlus macropterus |
| JOHN DORY | 37264004 | John Dory | Zeus faber |
| MIRROR DORY | 37264003 | Mirror Dory | Zenopsis nebulosus |
| ORANGE ROUGHY | 37255009 | Orange roughy | Hoplostethus atlanticus |
| OREO BASKET | 37266001 | Spikey oreo dory | Neocvttus rhomboidalis |
| OREO BASKET | 37266004 | Warty oreo | Allocyttus verrucosus |
| OREO BASKET | 37266005 | Black oreo dory | Allocvttus niger |
| PINK LING | 37228002 | Pink ling | Genvpterus blacodes |
| REDFISH | 37258003 | Redfish | Centrobervx affinis |
| RIBALDO | 37224002 | Ribaldo | Mora moro |
| ROYAL RED PRAWN | 28714005 | Roval red prawn | Haliporoides sibogae |
| SAWSHARK BASKET | 37023000 | Sawsharks | Pristiophoridae |
| SAWSHARK BASKET | 37023001 | Southern Sawshark | Pristiophorus nudipinnis |
| SAWSHARK BASKET | 37023002 | Common Sawshark | Pristiophorus cirratus |
| SCHOOL SHARK | 37017008 | School shark | Galeorhinus galeus |
| SILVER TREVALLY | 37337062 | Silver Trevally | Pseudocaranx dentex |
| SILVER WAREHOU | 37445006 | Silver warehou | Seriolella punctata |
| SMOOTH OREODORY | 37266003 | Smooth oreo dory | Pseudocvttus Maculatus |
| TIGER FLATHEAD | 37296001 | Tiger flathead | Neoplatvcephalus richardsoni |

Table 45. Major Non-Quota species in the SESSF.

| Project Key | Caab Cd | Common Name | Scientific Name |
| :---: | :---: | :---: | :---: |
| AUSTRALIAN ANGELSHARK | 37024001 | Australian Angelshark | Squatina australis |
| BARRACOUTA | 37439001 | Barracouta | Thyrsites atun |
| BLUE MORWONG | 37377004 | Blue Morwong | Nemadactylus valenciennesi |
| BLUESTRIPED GOATFISH | 37355001 | Bluestriped Goatfish | Upeneichthys lineatus |
| BROADNOSE SHARK | 37005002 | Broadnose Shark | Notorynchus cepedianus |
| BRONZE WHALER | 37018001 | Bronze Whaler | Carcharhinus brachyurus |
| CUTTLEFISH (UNSPECIFIED) | 23607000 | Cuttlefishes | Sepiidae |
| DRAUGHTBOARD SHARK | 37015001 | Draughtboard Shark | Cephaloscyllium laticeps |
| FROSTFISH | 37440002 | Frostfish | Lepidopus caudatus |
| GOULD SQUID | 23636004 | Gould's squid | Nototodarus gouldi |
| GUITARFISH (UNSPECIFIED) | 37026000 | Guitarfishes unspecified | Rhynchobatidae |
| GUITARFISH (UNSPECIFIED) | 37026001 | Giant guitarfish | Rhynchobatus djiddensis |
| HAPUKU | 37311006 | Hapuku | Polyprion oxygeneios |
| HAPUKU (UNSPECIFIED) | 37311902 | Hapuku and Bass Groper | Polyprion spp |
| KING DORY | 37264001 | King Dory | Cyttus traversi |
| KNIFEJAW | 37369002 | Knifejaw | Oplegnathus woodwardi |
| LATCHET | 37288006 | Latchet | Pterygotrigla polyommata |
| OCEAN BLUE-EYE TREVALLA | 37445014 | Ocean Blue-eye Trevalla | Schedophilus labyrinthica |
| OCEAN JACKET | 37465006 | Ocean Jacket | Nelusetta ayraudi |
| OCTOPUS | 23650000 | Octopoda | Octopoda |
| OCTOPUS | 23659000 | Octopuses | Octopodidae |
| ORNATE ANGELSHARK | 37024002 | Ornate Angelshark | Squatina tergocellata |
| RED GURNARD | 37288001 | Red Gurnard | Chelidonichthys Kumи |
| SILVER DORY | 37264002 | Silver Dory | Cyttus australis |
| SMOOTH HAMMERHEAD | 37019004 | Smooth Hammerhead | Sphyrna zygaena |
| SNAPPER | 37353001 | Snapper | Pagrus auratus |
| SOUTHERN EAGLE RAY | 37039001 | Southern Eagle Ray | Myliobatis australis |
| SQUID (GENERAL) | 23615000 | Squids | Teuthoidea |
| STARGAZER (UNSPECIFIED) | 37400000 | Stargazers | Uranoscopidae |
| STRIPED TRUMPETER | 37378001 | Striped Trumpeter | Latris lineata |
| TRIGGERFISH \& LEATHERJACKET (UNSPECIFIED) | 37465000 | Leatherjackets | Balistidae Monacanthidae |
| WHISKERY SHARK | 37017003 | Whiskery Shark | Furgaleus macki |
| YELLOWSPOTTED BOARFISH | 37367001 | Yellowspotted Boarfish | Paristiopterus gallipavo |

Table 46. Alphabetical List of the 75 Project Keys used for discard rate analyses. Included are the $\mathbf{7}$ HIGH RISK project keys groupings.

| 1 | ALFONSINO |  |  |
| :---: | :---: | :---: | :---: |
| 2 | AUSTRALIAN ANGELSHARK | 39 | OCEAN JACKET |
| 3 | BARRACOUTA | 40 | OCTOPUS |
| 4 | BIGHT REDFISH | 41 | ORANGE ROUGHY |
| 5 | BLUE GRENADIER | 42 | OREO BASKET |
| 6 | BLUE MORWONG | 43 | ORNATE ANGELSHARK |
| 7 | BLUE WAREHOU | 44 | PINK LING |
| 8 | BLUE-EYE TREVALLA | 45 | RAYS |
| 9 | BLUESTRIPED GOATFISH | 46 | RED GURNARD |
| 10 | BOARFISH (UNSPECIFIED) | 47 | REDFISH |
| 11 | BROADNOSE SHARK | 48 | RIBALDO |
| 12 | BRONZE WHALER | 49 | ROYAL RED PRAWN |
| 13 | CHIMAERAS | 50 | SAWSHARK BASKET |
| 14 | CRUSTACEANS | 51 | SAWSHARKS |
| 15 | DEEPWATER FLATHEAD | 52 | SCHOOL SHARK |
| 16 | DEEPWATER SHARK BASKET | 53 | SHARKS |
| 17 | DOGFISH | 54 | SILVER DORY |
| 18 | DRAUGHTBOARD SHARK | 55 | SILVER TREVALLY |
| 19 | EASTERN SCHOOL WHITING | 56 | SILVER WAREHOU |
| 20 | ECHINODERMS | 57 | SMOOTH HAMMERHEAD |
| 21 | ELEPHANTFISH | 58 | SMOOTH OREODORY |
| 22 | FISH | 59 | SNAPPER |
| 23 | FROSTFISH | 60 | SOUTHERN EAGLE RAY |
| 24 | GEMFISH | 61 | SQUID (GENERAL) |
| 25 | GOULD SQUID | 62 | STARGAZER (UNSPECIFIED) |
| 26 | GUITARFISH (UNSPECIFIED) | 63 | STINGAREES |
| 27 | GUMMY SHARK | 64 | STRIPED TRUMPETER |
| 28 | HAGFISH | 65 | TIGER FLATHEAD |
| 29 | HAPUKU | 66 | TRIGGERFISH \& LEATHERJACKET |
| 30 | INSHORE OCEAN PERCH | 67 | WHISKERY SHARK |
| 31 | JACKASS MORWONG | 68 | YELLOWSPOTTED BOARFISH |
| 32 | JOHN DORY | 69 | HIGH RISK DOGFISH, OTHER |
| 33 | KING DORY | 70 | HIGH RISK MOLLUSCS |
| 34 | KNIFEJAW | 71 | HIGH RISH HAGFISH |
| 35 | LATCHET | 72 | HIGH RISK SHARKS |
| 36 | MIRROR DORY | 73 | HIGH RISK SKATES/RAYS |
| 37 | MOLLUSCS | 74 | HIGH RISK TELEOSTS |
| 38 | OCEAN BLUE-EYE TREVALLA | 75 | HIGH RISK UPPER SLOPE DOGFISH |

Table 47. List of high risk and very high risk species available in the either the CDR, GENLOG or ISMP data supplied for the redesign of the ISMP for the SESSF, either by CAAB_code, or some other species indicator.

| CAAB Code | Species Name | Scientific Name | Group | Status | Project Key |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 23607000 | Cuttlefishes | Sepiidae | Molluscs | Major <br> NonQuota | Molluscs |
| 23607001 | Giant cuttlefish | Sepia apama | Molluscs | Minor | Molluscs |
| 23607901 | Cuttlefish (mixed) | Sepia spp | Molluscs | Minor | Molluscs |
| 23608003 | Southern <br> squid | Settletail | Molluscs | Minor | Molluscs |
| 23610000 | Pygmy squids | Idiosepiidae austrinum | Octopoda | Molluscs | Minor |

Note: The column 'Project Key' indicates the level of grouping that was used for the analysis of these species to meet the project requirements.

Table 48. List of threatened, endangered and protected species available in either the CDR, GENLOG or ISMP data supplied for the redesign of the ISMP for the SESSF, either by CAAB_code, or some other species indicator.

| CAAB code | Species Name | Scientific Name | Group | status | Project Key |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 37008000 | Grey nurse sharks | Odontaspididae | Sharks | Minor | Sharks |
| 37008001 | Grey nurse shark | Carcharias taurus | Sharks | Minor | Sharks |
| 37010003 | White Shark | Carcharodon carcharias | Sharks | Minor | Sharks |
| 37014001 | Whale shark | Rhincodon typus | Sharks | Minor | Sharks |
| 39001001 | Turtles | Testudines | Omit | Minor | Reptiles |
| 39012001 | Pignose turtle | Carettochelys insculpta | Omit | Minor | Reptiles |
| 39020001 | Loggerhead turtle | Caretta caretta | Omit | Minor | Reptiles |
| 39020002 | Green turtle | Chelonia mydas | Omit | Minor | Reptiles |
| 39020003 | Hawksbill turtle | Eretmochelys imbricata | Omit | Minor | Reptiles |
| 39020004 | Pacific (Olive) Ridely turtle | Lepidochelys olivacea | Omit | Minor | Reptiles |
| 39020005 | Flatback turtle | Natator depressus | Omit | Minor | Reptiles |
| 39021001 | Leatherback turtle | Dermochelys coriacea | Omit | Minor | Reptiles |
| 39125000 | Seasnakes | Hydrophiidae | Omit | Minor | Reptiles |
| 40000000 | Birds | Avians | Omit | Minor | Birds |
| 40000006 | Red Cormorant | Red Cormorant | Omit | Minor | Birds |
| 40001000 | Penguins | Spheniscidae | Omit | Minor | Birds |
| 40040000 | Albatrosses | Diomedeidae | Omit | Minor | Birds |
| 40040001 | Bullers Albatross | Thalassarche bulleri | Omit | Minor | Birds |
| 40040002 | Shy Albatross | Thalassarche cauta | Omit | Minor | Birds |
| 40040003 | Yellow Nosed Albatross | Thalassarche chlororhynchos | Omit | Minor | Birds |
| 40040004 | Grey Headed Albatross | Thalassarche chrysostoma | Omit | Minor | Birds |
| 40040005 | Southern Royal Albatross | Diomedea epomophora | Omit | Minor | Birds |
| 40040006 | Wandering Albatross | Diomedea exulans | Omit | Minor | Birds |
| 40040007 | Black Browed Albatross | Thalassarche melanophrys | Omit | Minor | Birds |
| 40040008 | Sooty Albatross | Phoebetria fusca | Omit | Minor | Birds |
| 40040009 | Light Mantled Sooty Albatross | Phoebetria palpebrata | Omit | Minor | Birds |
| 40040010 | Gibsons Albatross | Diomedea gibsoni | Omit | Minor | Birds |
| 40040012 | Northern Royal Albatross | Diomedea sanfordi | Omit | Minor | Birds |
| 40040013 | Campbell Albatross | Thalassarche impavida | Omit | Minor | Birds |
| 40040014 | Indian Yellow Nosed Albatross | Thalassarche carteri | Omit | Minor | Birds |
| 40041000 | Petrels Prions and Shearwaters | Procellariidae | Omit | Minor | Birds |
| 40041003 | Cape Petrel | Daption capense | Omit | Minor | Birds |
| 40041007 | Southern Giant Petrel | Macronectes giganteus | Omit | Minor | Birds |
| 40041008 | Northern Giant Petrel | Macronectes halli | Omit | Minor | Birds |
| 40041013 | Fairy Prion | Pachyptila turtur | Omit | Minor | Birds |
| 40041015 | Snow petrel | Pagodroma nivea | Omit | Minor | Birds |
| 40041018 | White Chinned Petrel | Procellaria aequinictialis | Omit | Minor | Birds |
| 40041019 | Grey Petrel | Procellaria cinerea | Omit | Minor | Birds |
| 40041020 | Parkinsons petrel | Procellaria parkinsoni | Omit | Minor | Birds |
| 40041021 | Westland Petrel | Procellaria westlandica | Omit | Minor | Birds |
| 40041022 | Tahiti Petrel | Pseudobulweria rostrata | Omit | Minor | Birds |
| 40041025 | White Necked Petrel | Pterodroma cervicalis | Omit | Minor | Birds |
| 40041030 | Coulds Petrel | Pterodroma leucoptera | Omit | Minor | Birds |
| 40041031 | Great Winged Petrel | Pterodroma macroptera | Omit | Minor | Birds |
| 40041035 | Providence Petrel | Pterodroma solandri | Omit | Minor | Birds |
| 40041038 | Flesh Footed Shearwater | Puffinus carneipes | Omit | Minor | Birds |
| 40041040 | Fluttering Shearwater | Puffinus gavia | Omit | Minor | Birds |
| 40041042 | Sooty Shearwater | Puffinus griseus | Omit | Minor | Birds |
| 40041043 | Huttons Shearwater | Puffinus huttoni | Omit | Minor | Birds |
| 40041045 | Wedge Tailed Shearwater | Puffinus pacificus | Omit | Minor | Birds |
| 40041047 | Short Tailed Shearwater | Puffinus tenuirostris | Omit | Minor | Birds |
| 40041999 | Shearwaters | Puffinus spp | Omit | Minor | Birds |
| 40042004 | Wilsons Storm Petrel | Oceanites oceanus | Omit | Minor | Birds |
| 40042007 | White Faced Storm Petrel | Pelagodroma marina | Omit | Minor | Birds |
| 40047002 | Australian gannet | Morus serrator | Omit | Minor | Birds |
| 40047004 | Masked booby | Sula dactylatra | Omit | Minor | Birds |
| 40128000 | Gulls skuas noddys and terns | Laridae | Omit | Minor | Birds |
| 40128004 | South polar skua | Catharacta maccormicki | Omit | Minor | Birds |
| 40128005 | Great Skua | Catharacta skua | Omit | Minor | Birds |
| 40128013 | Silver gull | Larus novaehollandiae | Omit | Minor | Birds |
| 40128025 | Crested Tern | Sterna bergii | Omit | Minor | Birds |
| 40128028 | Sooty tern | Sterna fuscata | Omit | Minor | Birds |
| 40128031 | Gull-billed tern | Sterna nilotica | Omit | Minor | Birds |
| 40128999 | Terns | Terns | Omit | Minor | Birds |
| 41000001 | Whales | Whales (order Cetacea in part) | Omit | Minor | Whales/Dolphins |
| 41000002 | Toothed whales | Toothed whales (suborder Odontoceti in part) | Omit | Minor | Whales/Dolphins |
| 41000003 | Baleen whales | Baleen whales (suborder Mysticeti) | Omit | Minor | Whales/Dolphins |
| 41110000 | Baleen whales | Mysticeti | Omit | Minor | Whales/Dolphins |


| CAAB code | Species Name | Scientific Name | Group | status | Project Key |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 41110001 | Southern right whale | Eubalaena australis | Omit | Minor | Whales/Dolphins |
| 41110002 | Pygmy right whale | Caperea marginata | Omit | Minor | Whales/Dolphins |
| 41112001 | Minke whale | Balaenoptera acutorostrata | Omit | Minor | Whales/Dolphins |
| 41112002 | Sei whale | Balaenoptera borealis | Omit | Minor | Whales/Dolphins |
| 41112003 | Brydes whale | Balaenoptera edeni | Omit | Minor | Whales/Dolphins |
| 41112004 | Blue whale | Balaenoptera musculus | Omit | Minor | Whales/Dolphins |
| 41112005 | Fin whale | Balaenoptera physalus | Omit | Minor | Whales/Dolphins |
| 41112006 | Humpback whale | Megaptera novaeangliae | Omit | Minor | Whales/Dolphins |
| 41116000 | Dolphins | Delphinidae | Omit | Minor | Whales/Dolphins |
| 41116001 | Common dolphin | Delphinus delphis | Omit | Minor | Whales/Dolphins |
| 41116002 | Pygmy killer whale | Feresa attenuata | Omit | Minor | Whales/Dolphins |
| 41116003 | Short-finned pilot whale | Globicephala macrorhynchus | Omit | Minor | Whales/Dolphins |
| 41116004 | Long-finned pilot whale | Globicephala melas | Omit | Minor | Whales/Dolphins |
| 41116006 | Frasers dolphin | Lagenodelphis hosei | Omit | Minor | Whales/Dolphins |
| 41116008 | Dusky dolphin | Lagenorhynchus obscurus | Omit | Minor | Whales/Dolphins |
| 41116009 | Southern right whale dolphin | Lissodelphis peronii | Omit | Minor | Whales/Dolphins |
| 41116011 | Killer whales | Orcinus orca | Omit | Minor | Whales/Dolphins |
| 41116012 | Melon-headed whale | Peponocephala electra | Omit | Minor | Whales/Dolphins |
| 41116013 | False killer whale | Pseudorca crassidens | Omit | Minor | Whales/Dolphins |
| 41116014 | Indo-Pac. hump-backed dolphin | Sousa chinensis | Omit | Minor | Whales/Dolphins |
| 41116015 | Pantropical spotted dolphin | Stenella attenuata | Omit | Minor | Whales/Dolphins |
| 41116016 | Striped dolphin | Stenella coeruleoalba | Omit | Minor | Whales/Dolphins |
| 41116017 | Spinner dolphin | Stenella longirostris | Omit | Minor | Whales/Dolphins |
| 41116018 | Rough-toothed dolphin | Steno bredanensis | Omit | Minor | Whales/Dolphins |
| 41116019 | Bottlenose dolphin | Tursiops truncatus | Omit | Minor | Whales/Dolphins |
| 41117001 | Spectacled porpoise | Australophocaena dioptrica | Omit | Minor | Whales/Dolphins |
| 41119001 | Pygmy sperm whale | Kogia breviceps | Omit | Minor | Whales/Dolphins |
| 41119002 | Dwarf sperm whale | Kogia simus | Omit | Minor | Whales/Dolphins |
| 41119003 | Sperm Whale | Physeter catodon | Omit | Minor | Whales/Dolphins |
| 41120000 | Beaked whales | Mesoplodon spp | Omit | Minor | Whales/Dolphins |
| 41120001 | Arnouxs beaked whale | Berardius arnuxii | Omit | Minor | Whales/Dolphins |
| 41120002 | Southern bottlenose whale | Hyperoodon planifrons | Omit | Minor | Whales/Dolphins |
| 41120003 | Longmans beaked whale | Mesoplodon pacificus | Omit | Minor | Whales/Dolphins |
| 41120004 | Andrews beaked whale | Mesoplodon bowdoini | Omit | Minor | Whales/Dolphins |
| 41120005 | Blainvilles beaked whale | Mesoplodon densirostris | Omit | Minor | Whales/Dolphins |
| 41120006 | Ginkgo-toothed beaked whale | Mesoplodon ginkgodens | Omit | Minor | Whales/Dolphins |
| 41120007 | Grays beaked whale | Mesoplodon grayi | Omit | Minor | Whales/Dolphins |
| 41120008 | Hectors beaked whale | Mesoplodon hectori | Omit | Minor | Whales/Dolphins |
| 41120009 | Strap-toothed whale | Mesoplodon layardii | Omit | Minor | Whales/Dolphins |
| 41120010 | Trues beaked whale | Mesoplodon mirus | Omit | Minor | Whales/Dolphins |
| 41120011 | Sherpherds beaked whale | Tasmacetus shepherdi | Omit | Minor | Whales/Dolphins |
| 41120012 | Cuviers beaked whale | Ziphius cavirostris | Omit | Minor | Whales/Dolphins |
| 41120999 | Whales | Whales | Omit | Minor | Whales/Dolphins |
| 41131000 | Eared seals | Otariidae | Omit | Minor | Seals |
| 41131001 | New Zealand fur seal | Arctocephalus forsteri | Omit | Minor | Seals |
| 41131002 | Antarctic fur seal | Arctocephalus gazella | Omit | Minor | Seals |
| 41131003 | South African fur seal | Arctocephalus pusillus | Omit | Minor | Seals |
| 41131004 | Subantarctic fur seal | Arctocephalus tropicalis | Omit | Minor | Seals |
| 41131005 | Australian sea lion | Neophoca cinerea | Omit | Minor | Seals |
| 41131006 | New Zealand sea lion | Phocarctos hookeri | Omit | Minor | Seals |
| 41131999 | Sealions | Sealions | Omit | Minor | Seals |
| 41132999 | Seals | Otariidae and Phocidae | Omit | Minor | Seals |
| 41136001 | Leopard seal | Hydrurga leptonyx | Omit | Minor | Seals |
| 41136002 | Weddell seal | Leptonychotes weddellii | Omit | Minor | Seals |
| 41136003 | Crabeater seal | Lobodon carcinophagus | Omit | Minor | Seals |
| 41136004 | Southern elephant seal | Mirounga leonina | Omit | Minor | Seals |
| 41206001 | Dugong | Dugong dugon | Omit | Minor | Omit |

Note: The column 'Project Key' indicates the level of grouping that was used for the analysis of these species to meet the project requirements.

Table 49 - A summary of the GENLOG and ISMP shots on record, for all years considered, and for the years 2004 to 2006.

|  | ALL YRS | $\begin{aligned} & \hline \text { ALL } \\ & \text { YRS } \end{aligned}$ | 2004 |  | 2005 |  | 2006 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Stratum | GENLOG | ISMP | GENLOG | ISMP | GENLOG | ISMP | GENLOG | ISMP |
| AL 10 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| AL 20 | 805 | 52 | 396 | 38 | 229 | 10 | 180 | 4 |
| AL 30 | 707 | 80 | 342 | 53 | 199 | 14 | 166 | 13 |
| AL 40 | 977 | 89 | 514 | 53 | 279 | 5 | 184 | 31 |
| AL 50 | 1978 | 11 | 598 | 5 | 852 | 6 | 528 | 0 |
| AL 60 | 345 | 14 | 147 | 14 | 101 | 0 | 97 | 0 |
| AL 70 | 15 | 0 | 2 | 0 | 4 | 0 | 9 | 0 |
| AL GAB | 1598 | 206 | 382 | 51 | 440 | 95 | 776 | 60 |
| BS_IN_TR | 1762 | 7 | 411 | 6 | 749 | 1 | 602 | 0 |
| DL 20 | 40 | 0 | 36 | 0 | 2 | 0 | 2 | 0 |
| DL 30 | 491 | 46 | 141 | 14 | 171 | 9 | 179 | 23 |
| DL 40 | 45 | 0 | 28 | 0 | 7 | 0 | 10 | 0 |
| DL 50 | 183 | 0 | 116 | 0 | 56 | 0 | 11 | 0 |
| DL 60 | 2 | 0 | 2 | 0 | 0 | 0 | 0 | 0 |
| DL 70 | 63 | 0 | 17 | 0 | 23 | 0 | 23 | 0 |
| DL GAB | 435 | 0 | 174 | 0 | 129 | 0 | 132 | 0 |
| ECDW_TR | 291 | 25 | 230 | 25 | 60 | 0 | 1 | 0 |
| EDL_DS | 22493 | 249 | 8323 | 48 | 7681 | 71 | 6489 | 130 |
| EDL_IN_TR | 14695 | 297 | 4853 | 47 | 5588 | 133 | 4254 | 117 |
| EDL_OFF_TR | 7497 | 216 | 2816 | 76 | 2606 | 74 | 2075 | 66 |
| GN 10 | 2 | 0 | 2 | 0 | 0 | 0 | 0 | 0 |
| GN 20 | 3019 | 0 | 1024 | 0 | 920 | 0 | 1075 | 0 |
| GN 30 | 2313 | 0 | 990 | 0 | 798 | 0 | 525 | 0 |
| GN 40 | 1079 | 0 | 382 | 0 | 398 | 0 | 299 | 0 |
| GN 50 | 6866 | 0 | 2427 | 0 | 2327 | 0 | 2112 | 0 |
| GN 60 | 10646 | 0 | 4357 | 0 | 3473 | 0 | 2816 | 0 |
| GN GAB | 10264 | 0 | 3389 | 0 | 2929 | 0 | 3946 | 0 |
| GAB Inshore | 15186 | 459 | 4696 | 123 | 5325 | 203 | 5165 | 133 |
| GAB Midshore | 1611 | 55 | 474 | 29 | 649 | 1 | 488 | 25 |
| NSW_IN_TR | 18843 | 782 | 7849 | 297 | 6151 | 272 | 4843 | 213 |
| NSW_OFF_TR | 5671 | 297 | 2132 | 128 | 2024 | 81 | 1515 | 88 |
| NSW_RRP_TR | 1201 | 13 | 441 | 4 | 437 | 5 | 323 | 4 |
| GAB Offshore | 615 | 47 | 246 | 21 | 177 | 11 | 192 | 15 |
| SW_BGS_TR | 1350 | 76 | 501 | 31 | 414 | 30 | 435 | 15 |
| SW_ORO_TR | 227 | 3 | 93 | 1 | 79 | 2 | 55 | 0 |
| SW_TR | 12528 | 404 | 4966 | 97 | 3902 | 187 | 3660 | 120 |
| TAS_BGS_TR | 1758 | 32 | 831 | 8 | 445 | 19 | 482 | 5 |
| TAS_E_TR | 7320 | 122 | 2978 | 35 | 2391 | 21 | 1951 | 66 |
| TAS_ORO_TR | 760 | 72 | 245 | 25 | 270 | 25 | 245 | 22 |
| TAS_W_TR | 4452 | 38 | 1944 | 1 | 1451 | 28 | 1057 | 9 |
| VIT | 1474 | 0 | 314 | 0 | 739 | 0 | 421 | 0 |

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Table 50. Results of analyses on the seasonal patterns of discard ratios for clusters identified in the GAB fishery. One out of 7 tests indicate a significant difference in discard ratios for a cluster across seasons (the grenadier cluster in the offshore depth zone). The sample size for this particular comparison is relatively small, however the trend in discard ratio from summer to autumn to winter is consistently increasing.

| Depth Category | Cluster | Autumn |  |  | Spring |  |  | Summer |  |  | Winter |  |  | F | Sig. | Kruskal Wallace Chi Square | Sig. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Sample | Mean Discard | Variance | Sample | Mean Discard | Variance | Sample | Mean Discard | Variance | Sample | Mean <br> Discard | Variance |  |  |  |  |
| Inshore | DEEP <br> WATER <br> FLATHEAD | 140 | 0.4518 | 0.0367 | 135 | 0.4466 | 0.0312 | 129 | 0.4351 | 0.0349 | 110 | 0.4912 | 0.0262 | 2.109 | . 098 | 6.737 | 0.081 |
|  | GEMFISH | 0 |  |  | 6 | 0.4656 | 0.0290 | 21 | 0.3107 | 0.0262 | 0 |  |  | 4.182 | . 052 | 3.704 | 0.054 |
|  | GRENADIER | 0 |  |  | 0 |  |  | 1 | 0.7837 |  | 0 |  |  |  |  |  |  |
|  | ORANGE ROUGHY | 1 | 0.4950 |  | 0 |  |  | 0 |  |  | 0 |  |  |  |  |  |  |
|  | REDFISH | 60 | 0.3591 | 0.0301 | 4 | 0.4209 | 0.0568 | 48 | 0.3268 | 0.0304 | 6 | 0.2441 | 0.0045 | 1.240 | . 299 | 3.488 | 0.322 |
| Midshore | DEEP <br> WATER <br> FLATHEAD |  |  |  |  |  |  |  |  |  | 5 | 0.5458 | 0.0047 |  |  |  |  |
|  | GEMFISH | 1 | 0.3527 |  | 4 | 0.4416 | 0.0554 | 5 | 0.4069 | 0.0272 | 1 | 0.5114 |  | . 131 | . 939 | 1.291 | 0.731 |
|  | GRENADIER | 27 | 0.2435 | 0.0234 | 3 | 0.5274 | 0.0588 | 22 | 0.3290 | 0.0384 | 9 | 0.3492 | 0.0327 | 2.973 | . 039 | 6.978 | 0.073 |
|  | ORANGE ROUGHY | 1 | 0.0127 |  | 0 |  |  | 0 |  |  | 6 | 0.0343 | 0.0035 | . 114 | . 749 | 0 | 1 |
|  | REDFISH | 0 |  |  | 0 |  |  | 0 |  |  | 0 |  |  |  |  |  |  |
| Offshore | DEEP WATER FLATHEAD | 0 |  |  | 0 |  |  | 0 |  |  | 0 |  |  |  |  |  |  |
|  | GEMFISH | 0 |  |  | 0 |  |  | 0 |  |  | 0 |  |  |  |  |  |  |
|  | GRENADIER | 4 | 0.4147 | 0.1546 | 0 |  |  | 4 | 0.1964 | 0.0048 | 10 | 0.7751 | 0.1043 | 5.692 | . 014 | 6.335 | 0.042 |
|  | ORANGE ROUGHY | 7 | 0.2935 | 0.2102 | 0 |  |  | 0 |  |  | 62 | 0.1371 | 0.0526 | 2.307 | . 133 | 0.095 | 0.757 |
|  | REDFISH | 0 |  |  | 0 |  |  | 0 |  |  | 0 |  |  |  |  |  |  |

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Table 51. Results of analyses on the longitudinal patterns of discard ratios for clusters identified in the GAB fishery. One out of 9 tests indicate a significant difference in discard ratios for a cluster across longitude groupings (the grenadier cluster in the offshore depth zone). Discard rates for
this cluster are larger in the far eastern
group

| Depth Category | Cluster | long <= 127 |  |  | 127<long<=130 |  |  | long > 130 |  |  | F | Sig. | Kruskal Wallace Chi Square | Sig. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Sample | Mean Discard | Variance | Sample | Mean Discard | Variance | Sample | Mean Discard | Variance |  |  |  |  |
| Inshore | DEEP <br> WATER <br> FLATHEAD | 188 | 0.4563 | 0.0324 | 206 | 0.4530 | 0.0270 | 120 | 0.4551 | 0.0438 | . 017 | . 983 | 0.161 | 0.922 |
|  | GEMFISH | 14 | 0.3030 | 0.0119 | 13 | 0.3904 | 0.0480 |  |  |  | 1.765 | . 196 | 1.592 | 0.207 |
|  | GRENADIER |  |  |  | 1 | 0.7837 |  |  |  |  |  |  |  |  |
|  | ORANGE ROUGHY | 1 | 0.4950 |  |  |  |  |  |  |  |  |  |  |  |
|  | REDFISH | 13 | 0.3490 | 0.0276 | 39 | 0.3281 | 0.0377 | 66 | 0.3492 | 0.0267 | . 191 | . 827 | 1.291 | 0.524 |
| Midshore | DEEP WATER FLATHEAD | 5 | 0.5458 | 0.0047 |  |  |  |  |  |  |  |  |  |  |
|  | GEMFISH | 3 | 0.2727 | 0.0197 | 7 | 0.4991 | 0.0229 | 1 | 0.3527 |  | 2.559 | . 138 | 4.39 | 0.111 |
|  | GRENADIER | 1 | 0.5872 |  | 15 | 0.3767 | 0.0363 | 45 | 0.2733 | 0.0310 | 3.126 | . 051 | 5.653 | 0.059 |
|  | ORANGE ROUGHY | 6 | 0.0109 | 0.0002 |  |  |  | 1 | 0.1525 |  | 113.748 | . 000 | 2.291 | 0.13 |
|  | REDFISH |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Offshore | DEEP <br> WATER <br> FLATHEAD |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | GEMFISH |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | GRENADIER | 5 | 0.8537 | 0.1070 |  |  |  | 13 | 0.4559 | 0.1244 | 4.760 | . 044 | 3.921 | 0.048 |
|  | ORANGE ROUGHY | 42 | 0.1346 | 0.0361 |  |  |  | 27 | 0.1816 | 0.1194 | . 532 | . 468 | 0.237 | 0.626 |
|  | REDFISH |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 52. Results of analyses on the zonal patterns of discard ratios for clusters identified in the GHAT fishery. 6 out of 8 tests indicate a significant difference in discard ratios for a cluster across seasons. On this basis it is recommended that the ISMP design for the GHAT fishery comprises a zonal split.

| Gear <br> Code | Cluster |  | Zone |  |  |  |  |  |  | F | Sig. | Kruskal <br> Wallace Chi <br> Square | Sig. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | GAB | 20 | 30 | 40 | 50 | 60 | $\begin{aligned} & \text { NORTH } \\ & \text { EAST } \\ & \hline \end{aligned}$ |  |  |  |  |
| $\begin{array}{ll} \mathbf{A L} & \boldsymbol{\&} \\ \mathbf{B L} & \end{array}$ | AL BLUE-EYE TREVALLA | Sample | 66 | 8 | 58 | 11 | 2 |  | 3 | 2.440 | 0.037 | 8.285 | 0.141 |
|  |  | Mean Discard | 0.2142 | 0.1804 | 0.1176 | 0.1423 | 0.1291 |  | 0.0000 |  |  |  |  |
|  |  | Variance | 0.0510 | 0.0169 | 0.0124 | 0.0296 | 0.0012 |  | 0.0000 |  |  |  |  |
|  | AL GUMMY SHARK | Sample |  | 2 | 1 |  |  | 14 |  | 2.998 | 0.082 | 6.050 | 0.049 |
|  |  | Mean Discard |  | 0.6080 | 0.6982 |  |  | 0.3542 |  |  |  |  |  |
|  |  | Variance |  | 0.0003 |  |  |  | 0.0515 |  |  |  |  |  |
|  | AL HAPUKU | Sample | 41 | 1 | 2 |  |  |  | 3 | 0.532 | 0.663 | 7.260 | 0.867 |
|  |  | Mean Discard | 0.2363 | 0.0310 | 0.1564 |  |  |  | 0.0326 |  |  |  |  |
|  |  | Variance | 0.1073 |  | 0.0067 |  |  |  | 0.0015 |  |  |  |  |
|  | AL MIXED | Sample | 3 | 3 | 2 | 1 | 6 |  | 16 | 17.346 | 0.000 | 17.017 | 0.004 |
|  |  | Mean Discard | 0.0000 | 0.4806 | 0.7500 | 0.0000 | 0.1565 |  | 0.0313 |  |  |  |  |
|  |  | Variance | 0.0000 | 0.0705 | 0.1250 |  | 0.0194 |  | 0.0019 |  |  |  |  |
|  | AL PINK LING | Sample | 96 | 62 | 124 | 152 | 3 |  | 2 | 4.401 | 0.001 | 56.417 | 0.000 |
|  |  | Mean Discard | 0.2087 | 0.1331 | 0.2233 | 0.1231 | 0.1920 |  | 0.0023 |  |  |  |  |
|  |  | Variance | 0.1076 | 0.0140 | 0.0322 | 0.0255 | 0.1106 |  | 0.0000 |  |  |  |  |
| DL | $\begin{aligned} & \text { DL BLUE-EYE } \\ & \text { TREVALLA } \end{aligned}$ | Sample |  | 87 | 121 | 116 | 96 | 1 | 122 | 8.495 | 0.000 | 60.819 | 0.000 |
|  |  | Mean Discard |  | 0.0144 | 0.0215 | 0.0173 | 0.1056 | 0.0000 | 0.0739 |  |  |  |  |
|  |  | Variance |  | 0.0040 | 0.0081 | 0.0066 | 0.0362 |  | 0.0269 |  |  |  |  |
|  | DL HAPUKU | Sample |  | 60 | 38 | 31 | 36 |  | 42 | 22.431 | 0.000 | 62.120 | 0.000 |
|  |  | Mean Discard |  | 0.0355 | 0.3542 | 0.1395 | 0.4323 |  | 0.6733 |  |  |  |  |
|  |  | Variance |  | 0.0122 | 0.2061 | 0.1167 | 0.1904 |  | 0.1842 |  |  |  |  |
| GN | GN GUMMY MIXED | Sample |  | 99 |  |  |  | 4 |  | 55.162 | 0.000 | 10.415 | 0.001 |
|  |  | Mean Discard |  | 0.1454 |  |  |  | 0.7999 |  |  |  |  |  |
|  |  | Variance |  | 0.0287 |  |  |  | 0.0692 |  |  |  |  |  |
|  | GN GUMMY SHARK | Sample |  |  |  |  |  | 1 |  |  |  |  |  |
|  |  | Mean Discard |  |  |  |  |  | 0.4143 |  |  |  |  |  |
|  |  | Variance |  |  |  |  |  |  |  |  |  |  |  |
|  | GN SCHOOL SHARK | Sample |  | 1 |  |  |  |  |  |  |  |  |  |
|  |  | Mean Discard |  | 0.3723 |  |  |  |  |  |  |  |  |  |
|  |  | Variance |  |  |  |  |  |  |  |  |  |  |  |

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Table 53. Results of analyses on the seasonal patterns of discard ratios for clusters identified in the GHAT fishery. One out of 8 tests indicate a significant difference in discard ratios for a cluster across seasons (the Blue-Eye Trevalla Cluster for automatic longlining AL, longlining LL and bottom longline BL).

| Gear Code | Cluster | Autumn |  |  | Spring |  |  | Summer |  |  | Winter |  |  | F | Sig. | Kruskal Wallace Chi Square | Sig. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Sample | Mean Discard | Variance | Sample | Mean Discard | Variance | Sample | Mean Discard | Variance | Sample | Mean Discard | Variance |  |  |  |  |
| $\begin{array}{\|l} \mathbf{A L} \\ \boldsymbol{\&} \\ \mathbf{B L} \end{array}$ | AL BLUE-EYE TREVALLA | 41 | 0.120 | 0.020 | 15 | 0.174 | 0.019 | 48 | 0.091 | 0.014 | 44 | 0.280 | 0.050 | 11.348 | . 000 | 23.576 | 0 |
|  | AL GUMMY <br> SHARK  |  |  |  |  |  |  | 16 | 0.318 | 0.055 | 1 | 0.698 |  | 2.490 | . 134 | 1.5 | 0.221 |
|  | AL HAPUKU | 3 | 0.010 | 0.000 | 2 | 0.155 | 0.007 | 5 | 0.130 | 0.049 | 37 | 0.247 | 0.113 | 0.695 | 0.56 | 1.929 | 0.587 |
|  | AL MIXED | 4 | 0.360 | 0.105 |  |  |  | 9 | 0.271 | 0.002 | 18 | 0.028 | 0.058 | 6.835 | 0.004 | 5.104 | 0.078 |
|  | AL PINK LING | 117 | 0.150 | 0.020 | 109 | 0.126 | 0.019 | 135 | 0.132 | 0.030 | 78 | 0.335 | 0.118 | 21.342 | . 000 | 6.003 | 0.111 |
| DL | DL BLUE-EYE <br> TREVALLA | 98 | 0.022 | 0.007 | 189 | 0.046 | 0.020 | 122 | 0.052 | 0.018 | 134 | 0.059 | 0.021 | 1.548 | 0.201 | 6.564 | 0.087 |
|  | DL HAPUKU | 33 | 0.276 | 0.181 | 74 | 0.381 | 0.220 | 42 | 0.344 | 0.171 | 58 | 0.207 | 0.135 | 1.985 | . 117 | 5.351 | 0.148 |
| GN | GN GUMMY <br> MIXED  | 63 | 0.145 | 0.037 |  |  |  |  |  |  | 40 | 0.212 | 0.059 | 2.428 | 0.122 | 1.059 | 0.304 |
|  | GN GUMMY <br> SHARK  | 1 | 0.414 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | $\substack{\text { GN } \\ \text { SHARK }}$  | 1 | 0.372 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

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Table 54. A breakdown, by species and year, of the MWCVs of the length frequencies obtained from the sea based ISMP sampling data, by retained and discarded Retained
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| Yellowspotted boarfish | 27.78 | 20.42 | 26.58 | 37.28 |
| :---: | :---: | :---: | :---: | :---: |

## Appendix A - Data Pre-Processing

The following broad categories of data were supplied:

1. ISMP - data gathered from the independent monitoring - see A-Table 1.
2. GENLOG - shot level catch data (referred to as SEF1 in earlier ISMP reports and design studies) - see A-Table 2.
3. CDR- Catch Disposal Records (referred to as SEF2 in earlier ISMP reports and design studies) - see A-Table 3.

Initially, four data pre-processing activities were carried out to prepare the data for higher level analyses to consider the merits of different experimental designs for future ISMP programs for the SET, GAB and GHAT fisheries. These are as follows:

1. Analysis 1. Merging the CDR and GENLOG data.
2. Analysis 2. Combine ISMP data sources and allocation of ISMP shots to ISMP strata.
3. Analysis 3. Allocation of the GENLOG shots to ISMP strata.

The reason for carrying out Analysis 1 was twofold,
a) To link GENLOG shots to the PORT information in the CDR data - since there is no record of PORT in the GENLOG data. The reason for requiring PORT information at the level PORT (preferably port of departure) is that, since observers are required to select vessels for ISMP sampling purposes, they need to be given some indication of which vessels to choose in order to achieve particular sample size by stratum targets. Thus it is important to be able to build a profile of the typical stratum breakdown of vessels operating from particular ports and this can only be done if port of departure is linked in the data to actual shots.
b) To correct the GENLOG catch weight estimates at the level of shots. The rationale behind this part of the exercise is that the shot estimates of catch are regarded as being imprecise estimates, while the PORT estimates of catch are regarded as accurate estimates of the actual catch for the trip. This situation pertains to many fisheries around the world, and it is typical to correct shots such that the sum of catch over all shots for a trip is equal to the landing recorded for the trip as a whole. This exercise is carried out separately for each species.

The reason for carrying out Analysis 2 is to be able to allocate ISMP shots to strata so as to be able to calculate stratified means and variances for the existing ISMP design in order to be able to draw comparisons with alternative ISMP designs.

The need for Analysis 3 is to provide an estimate of the size of the strata used for the recent ISMP. This statement is only applicable to the SET fishery, since the ISMP's for the GAB and GHAT fisheries have not been carried out according to an explicit stratified design.

These stratum proportions are required in order to be able to calculate stratified means and variances for the existing ISMP design, as well as for alternative ISMP designs that may be considered.

## Appendix A-Analysis 1 - Merging the CDR and GENLOG data

The GENLOG data covers the year range 1985 - 2009. The CDR data supplied to the study spans the year range 1996-2008. Therefore, if a merge to correct shot level catches in the GENLOG data on the basis of trip landings catch records is obligatory, then the GENLOG data from 1985 - 1995 and 2009 will have to be excluded from the analysis. Although a request for CDR data prior to 1996 was considered, on subsequent reanalysis of the data it was felt that merging the $1992-1995$ CDR data with the $1992-1995$ GENLOG data was either (a) probably not a feasible exercise, or (b) feasible computationally but insufficiently reliable for the purpose of the calculations reported in this document. Thus the earliest year for which a CDR - GENLOG merge was contemplated was 1996. Please see, A-Table 4, Table 5 and A-Table 6 for a detailed breakdown.

We note that the CDR - GENLOG merge is not simple to achieve since there is no single key which links shots in the GENLOG data to landings in the CDR data. The reasons for this are historic. We would however strongly advise, if this has not already been done, that such a key be established in the data recording process, since the absence of this key compromises the quality and amount of data available for a wide range of analyses (not just that which is presently under consideration).

In the absence of the key just mentioned, the basis for a merge of the CDR and the GENLOG data therefore has to be a combination of a unique vessel identifier and a logical test on shot and trip dates, along the lines that
"a shot belongs to a designated trip if the vessel for the shot matches the vessel of the trip AND the shot date lies in between or on the start and/or end dates of the trip".
Merging of the CDR and GENLOG data was carried out separately for the SET, GAB and GHAT fisheries.

## Appendix A - Merging the CDR and GENLOG data

## i) Shot level dataset merge issues

A preliminary analysis showed that, restricting the GENLOG data to the period after and including 1996, about 100 of the Vessel_IDs in the GENLOG data do not appear in the CDR data and in the order of $30-40$ of the Vessel_IDs in the CDR data do not appear in the GENLOG data. On the basis of this preliminary analysis, Vessel_ID did not appear to be a good basis for a merge of the CDR and GENLOG data post-1996. Vessel_Names offer an alternative basis for such a merge but a preliminary analysis of the Vessel_Names which are common between the CDR data and the GENLOG data suggested a similar situation existed as for Vessel_IDs.

Two other tables in the GENLOG dataset which contained what appear to be a range of legacy variables related to vessel were considered as providing ancillary information which might assist with the CDR - GENLOG merge for the SET fishery. These tables are VESSELS_CSIRO and VESSELS_CSIRO_ALLVESSELS.

A-Table 7 summarises the results of the investigative work conducted to determine the possibility of using either Vessel_ID or Vessel_Name as the basis for the merge.

A speculative investigation was then launched on the assumption that the match of Vessel_Ids between the CDR and the GENLOG data might perhaps have improved over time.

The results of this investigation are presented in A-Table 8 below. They clearly show that the number of matched vessels increases significantly from the year 2000 onwards. This bears out the assumption that the quality of the matching based on Vessel_Ids will improve with time. The quality of the resulting dataset can be further assessed by examining the number of GENLOG shots and CDR landings which would be lost as a result of a merge based on Vessel_Ids - the results of this analysis are presented in A-Table 9.

The above analytical work demonstrates that using Vessel_ID as a basis for merging the CDR and GENLOG datasets becomes increasing reliable from 2000 onwards. The authors of this document chose to limit the data included in the analysis to records of shots occurring during or after 2002 on this basis.

Despite the above, it is important to record that the following data issues and problems exist in the in the CDR data:

- Trips for which end dates occur after the start date of the next trip for the same vessel overlapping trips
- Trips in which the end date is equal to the start date of the next trip for the same vessel special kind of overlapping trip
- Trips in which the end dates occur before start date - inverted trips
- Situations where shot dates lie in-between the end-date and start date of the next closest trips for the same vessel - crack situation
- Situations in which a trip is nested within another trip - nested trips
- Situations where two trips have the same start and/or end-dates - really a special combination of all of the above

The approach that is taken for this report for dealing with these problems relies on the basic assumptions that the end dates are more reliable and that erroneous start dates are causing many, but not all, of the problems mentioned above. The generic method developed here to deal with this issue is as follows:

- Step 1. Merge all trips which have the same end date.
- Step 2. Merge all trips for which end dates are equal to the start date of the next trip overlapping trips - the merged 'super trip' will have a start date equal to the earlier trip.
- Step 3. For trips for which end dates occur after the start date of the next trip for the same vessel - make the start data of the second trip equal to the end date of the previous trip plus 1 day.

The rationale behind Step 3 is that it seems likely that an overlapping trip is the result of an error, and that the error is more likely to lie in the start date - hence we keep those trips separate. In the case of Step 2, when the end date is exactly equal to the start date of the next trip, it seems that this coincidence is less likely to be due to an error and hence the two trips are merged.

- Step 4. For all other trips, start dates are set equal to [the end date of the previous trip + 1 day].

The method was tested (a) in terms of its logic, (b) for specific examples in the data which correspond to the situations described as well as various combinations of those situations, and (c) by checking that all shots were allocated to trips provided that a matching vessel existed, and (d) checking that no shots were allocated to more than one trip.

## ii) Shot-species level dataset merge issues

Having merged the CDR data to the GENLOG catch data, an analysis of the quality of record matching based on species was launched. The aim of this analysis was to establish the methodology required to calculate per shot catch weight correction factors for GENLOG records. During the course of this analysis, it became evident that a number of landed species in the CDR landings table had no matching GENLOG catch record. The authors theorise that this is due to imprecise species identification in the GENLOG catch records meaning that species were either not identified as part of a catch by the vessel skipper, or were incorrectly identified. The landed catch is, of course subjected to a more detailed scrutiny and would provide a more accurate breakdown of the species caught.

This species mismatching presents two problems to the calculation of corrected catch weights:

1) When calculating correction factors for GENLOG retained weight values, unreasonable large, or small catch correction factors are derived
2) Species identified in CDR landing records but not found in GENLOG catch records for that trip.

The first of these problems is illustrated in A-Figure 1 which shows histogram of the distribution of the calculated catch correction factor. The catch correction factors are calculated in such a way that the summed retained catch weight per species from the GENLOG catch data can be adjusted to match the landed weight for that species in the CDR data. In order to deal with the excessively large correction factors, a pragmatic approach was adopted to set a maximum correction factor of 4 . In this way, unrealistic values are truncated down to a more reasonable level based on the argument that a recorded retained catch weight from the GENLOG data will never be less than $1 / 4$ of the landed catch (correction factors greater than 0 imply that the vessel skipper underestimated the catch weight). The distribution of the resulting correction factors is shown in A-Figure 2.

The second problem results in small differences in the summed strata weights (corrected by landing weights) when compared to the landed totals on a species by species basis. Therefore the summed retained catch weight from the GENLOG tables is less than the total species catch for a particular species.

Table 10 illustrates this problem by listing the total landed weight by quota and major nonquota species against the summed GENLOG catch weights.

In order to correct the final computed stratum weights for this error an additional weighting factor was calculated after determining the stratum weights. This factor was used to assign the landed weight from the CDR records not accounted for in the GENLOG catch data to the strata on a pro rata basis. In this way the summed strata weights match the summed landing weights.

## Appendix A-Analysis 2 - Combine ISMP datasets

ISMP observer data was supplied by AFMA in two different formats, a Microsoft Access Database: ISMP_ARCHIVE_DATA and 3 delimited text files: onbl.txt, onbw.txt and portlf.txt. The analysis described below was initially built on data sourced from the ISMP_ARCHIVE_DATA, since this was the data initially provided - this source provided usable data for the period 2002 to 2006. During the course of the project, additional data was provided by AFMA which contained observer data for 2007 and 2008. It should be noted that the two datasets, while containing similar data, were not in the same format. This resulted in
additional work to enable the data pre-processing tasks to run correctly against the final ISMP dataset for the period 2002-2008.

## Appendix A-Analysis 2a - Allocation of ISMP shots to ISMP strata, SET Fishery

There are two main aspects to the allocation of ISMP shots to the strata definitions used for the design of the existing SET ISMP. The first is the allocation of shots to ISMP zones.

## Appendix A-Zones

This involves the application of an algorithm to decide whether a shot lies in one of 7 different zones being:
70 - East Coast Deep Water Zone
60 - Bass Strait
50 - Western Victoria
40 - Western Tasmania
30 - Eastern Tasmania
20 - Eastern Victoria
10 - New South Wales
The allocation of ISMP shots to Zones is achieved by rendering each of the areas described above into a collage of one or more trapezoids with parallel eastern and western boundary lines running north - south. Each trapezoid can thus be described by six values (two longitudes and four latitudes), i.e.

LongW, LongE, LatNE, LatNW, LatSE, LatSW.

A shot lies within a trapezoid if its longitude value lies between LongW and LongE, and if its latitude value lies in between the latitude values obtained by the intersection of the northsouth line passing through the location of the shot and the following two lines

A northern boundary line passing between the two points (LongW; LatNW), (LongE ; LatNE) And
A southern boundary line passing between the two points (LongW; LatSW), (longE; LatSE).
A shot lies within a zone if it falls inside one of the trapezoids corresponding to that zone. The trapezoid values for the 7 zones described for the SET fishery are given in A-Table 11.
This includes values for the East Coast Deep Water zone not previously included in the SET ISMP.
A-Figure 3 shows the ISMP shots per zone as a scatter plot using the longitude and latitude of the shot.

## Appendix A-Remainder of Stratum allocation

The next phase of the allocation of strata as per the most recent design and implementation (see A-Figure 3) involves a combination of species (presence/absence), depth, gear and seasonal considerations. There are 15 strata in the final design for the SET ISMP, as shown in A-Table 12. Note that A-Table 12 also shows the stratum definitions. The column headed 'Precedence' means that rules take precedence in the order 1-15. For example, if there is any ambiguity and a shots falls into more than one stratum, then the precedence order will remove that ambiguity.

The procedure described above still left a number of unknown shots pre-1999 as shown in ATable 13. As a result an additional set of rules was used to deal with these unallocated shots:

Designate the shots inshore or offshore. This is largely based on the known depth distributions of the target species - refer to Table 9 in Smith et al 1997 (Design of an ISMP for the SEF) which reports depth distributions for main species (also see the attached table of depth distributions). This can be backed up by looking at the depths of targeted shots for these species when depth is known. This worked for 16 of the target species which together make up $70 \%$ of the troublesome shots.

Unfortunately 2 of the main species straddle the 200 m depth: redfish and spotted warehou, as does Goulds squid. They make up $23 \%$ of the shots. The following algorithm was used to sort out most of the unallocated shots based on the species that appear to be characteristic for each depth category (inshore/offshore).

To allocate inshore/offshore the following rule is used:

1. If target_sp $=37264004$ or 37296001 or 37337062 or 37377003 or 37445005 then depth_strata=inshore;
2. Else if target_sp=37020000 or 37020003 or 37020905 or 37227001 or 37228002 or $372550093 \overline{7} 264003$ or 37287093 or 37439002 then depth_strata $=$ offshore;
3. Else if target_sp= 28714005 then depth_strata=RRP; Else do; if the shot contains (John dory 37264004 and/or grey morwong 37377002 and/or blackspot boarfish 37367005 ) and does NOT contain (offshore ocean perch 37287093 and/or deepsea flathead 37297001), then depth_strata=inshore;
4. else if the shot does NOT contain (John dory 37264004 and/or grey morwong 37377002 and/or blackspot boarfish 37367005 ) and does contain (offshore ocean perch 37287093 and/or deepsea flathead 37297001), then depth_strata=offshore;
5. else depth_strata="unknown.

Application of this rule in conjunction with the other rules applied leads to the shot allocation results presented in A-Table 14 which are regarded as satisfactory to proceed.

Note: In the data supplied for this study the Royal Red Prawn code has been changed to the species code 28714005.

## Appendix A - Analysis 2b - Allocation of ISMP shots to ISMP strata, GAB Fishery

As previously discussed, stratification of the GAB Fishery is based on depth criteria which were developed by a close examination of the species content per shot with the shot depth overlaid. As previously discussed, the utility of a spatial component was also assessed.

The analysis was conducted in two stages:

1. Group shots by similar species content using the K-Means clustering algorithm and score each shot record with its assigned cluster name/description.
2. Overlay the resulting cluster names against the depth and location of the shot.

Step 1 of the analysis required the calculation of a new column of data for each species caught in a shot. The percentage of the retained catch weight for this species was then stored in this new field. The resulting dataset contained, for each shot, the species caught and the percentage of the catch for that species. This data was then passed through the clustering algorithm and the resulting dataset was used to determine the rules for stratifying the GAB fishery.

The analytical work briefly outlined above resulted in the determination of 3 depth bands as the strata for the GAT fishery.

## Appendix A-Analysis 2c - Allocation of ISMP shots to ISMP strata, GHAT Fishery

The stratification of the GHAT fishery followed the same basic process as the GAB fishery this led to the selection of GEAR as. Since the GHAT fishery spans both the SET and GAB fisheries spatially, it was decided to further subdivide the GHAT Fishery into the SET Zones and one additional zone representing the are covered by the GAB fishery (since the SET fishery and the GAB fishery are spatially distinct). The zoning of the GHAT fishery data required the re-use of the processes developed for the SET fishery (as discussed in Analysis 2 a above), amended to include the GAB area. These zones, in conjunction with the GEAR type were used to determine the possible strata for the GHAT fishery.

## Appendix A - Analysis 3a. Allocation of GENLOG shots to ISMP strata for the SET Fishery

In order to allocate GENLOG shots to the ISMP strata, it was necessary to create a variable mapping which allowed the GENLOG shots to be fed through the existing strata allocation process. Using this approach we were able to re-use much of the work completed for the ISMP strata allocation process, it should be noted, however that a number additional modifications in order to ensure that the process functioned correctly. Of particular note was a systematic error which appeared when an analysis of the resulting strata was conducted - the results of which are shown in A-Table 15. As can be seen from this result, the strata BS_IN_TR, EDL_IN_TR and NSW_IN_TR were not represented in the allocated strata dataset. Further analysis of the data revealed that the underlying cause of the missing strata was a result of missing or "null" values in the depth field for catch records which should have been allocated to these strata (see A-Table 8 for the stratum rules)

In order to resolve the missing depth values, a process needed to be developed which would accurately impute a depth value for those records with null or missing depths. An examination of the strata allocation rules (A-Table 8) revealed that it was not necessary to impute the actual value of the missing depth, the model need only determine if the correct depth should be greater or less than 200 m .

In order to impute the missing depth, a dataset containing a dichotomous variable representing the depth of the shot was created such that "Y" = Depth $>200 \mathrm{~m}$ and "N" $<200 \mathrm{~m}$. Acting on the premise that the catch content would be a strong indicator of shot depth, additional variables were derived to represent all possible species which could have been caught. These derived variables were then populated with the percentage of the retained catch for each species of a shot (a value of zero, in this case would indicate that the species was not caught in this shot). The resulting fields were then presented to a decision tree algorithm as independent variables along with fields describing the vessel and gear used with the dichotomous depth variable as the dependent variable.

A "hold out" sample of $50 \%$ of the available data was used to test and validate the resulting model. Various diagnostic tools revealed that the decision tree had achieved an accuracy of $\sim 92 \%$ on both the testing and training datasets and showed no evidence of over fitting when the model fit was checked against the training dataset and the validation dataset. A-Figure 5 and A-Figure 6 display "Gains" plots, panelled by the dataset of the decision tree's accuracy in assigning a shot to a depth $>200 \mathrm{~m}$ or to a depth $<200 \mathrm{~m}$.

These diagnostics indicate a stable model which was deemed to be acceptable for the purposes of the stratum allocations. Using the results of the decision tree model, the missing strata could be filled, resulting in A-Table 16.

Table 1 - Appendix A: ISMP Dataset Supplied by AFMA.

| TABLE: Catch | TABLE: Cruise | TABLE: Lengthfreqs |
| :---: | :---: | :---: |
| RECORD COUNT: 271939 | RECORD COUNT: 2413 | RECORD COUNT: 306078 |
| FIELD NAMES: | FIELD NAMES: | FIELD NAMES: |
| ID | ID | ID |
| Db | Db | db |
| fishery | Fishery | fishery |
| CALLSIGN | CALLSIGN | CALLSIGN |
| depart_date | depart_date | depart_date |
| shot_date | STATE | shot_date |
| shot_number | OBSERVER | shot_number |
| retained_number | Portdepart | CAAB_code |
| retained_kgs | Portreturn | lf_type |
| discard_number | Boardingdate | lf_percentage |
| discard_kgs | Boardingtime | length |
| CAAB_code | Disembarkingdate | lengthcode |
| common_name | Disembarkingtime | retained |
| scientific_name | COMMENTS | discarded |
| lf_type |  | sorted |
| If_percentage |  | total_number |
| COMMENTS |  | males |
|  |  | females |
|  |  |  |
| TABLE: PORTLENFREQ | TABLE: TbIPetabundance | TABLE: TbIPetinteraction |
| RECORD COUNT: 210641 | RECORD COUNT: 5898 | RECORD COUNT: 420 |
| FIELD NAMES: | FIELD NAMES: | FIELD NAMES: |
| ID | ID | ID |
| CALLSIGN | CALLSIGN | CALLSIGN |
| Vessel_name | depart_date | depart_date |
| Fisher | shot_date | shot_date |
| sample_date | shot_number | shot_number |
| gear | Observer | observer |
| observer | start_time | start_time |
| port | Endtime | CAAB_code |
| state | CAAB_code | contactcode |
| zone | IRR | contactcount |
| CAAB_code | Int | contactpoint |
| catch_kgs | ROM | earsample |
| catch_no | Count | latitude |
| sample_kgs | Countmethod | longitude |
| sample_no | Total | length |
| grade | Comments | Number Dead |
| length_code |  | sex |
| length |  | sightcount |
| frequency |  | sightdistancecode |
| females |  | sightmethodcount |
| males |  | vesselactivity |
|  |  | contactmortality |
|  |  | age |
|  |  | comments |
|  |  |  |

Table 1 - Appendix A: ISMP Dataset Supplied by AFMA Continued....

| FILE NAME: portlf.txt | FILE NAME: onbw.txt | FILE NAME: portlf.txt |
| :---: | :---: | :---: |
| RECORD COUNT: | RECORD COUNT: | RECORD COUNT: |
| FIELD NAMES: | FIELD NAMES: | FIELD NAMES: |
| CAAB species code | CSIROcode | CSIROcode |
| Calendar year | Year | Year |
| Month | Month | Month |
| Day | Day | Day |
| Gear code | Latitude | Latitude |
| Zone code | Longitude | Longitude |
| Grade | Gear code | Gear code |
| Lencode | RetWhole | Length |
| CatchWeight | RetNumber | Lencode |
| CatchNum | DisWhole | NuMale |
| SampleWeight | DisNumber | NuFemale |
| SampleNum | DepthMin | TotNum |
| Length | DepthMax | DepthMin |
| TotalNum | Process | DepthMax |
| CallSign | CallSign | 1 fret |
| Port |  | 1fdis |
|  |  | Process |
|  |  | Sorted |
|  |  | Retained |
|  |  | Discarded |
|  |  | TempSST |
|  |  | CallSign |
|  |  |  |
| TABLE: Trawls |  |  |
| RECORD COUNT: 14499 |  |  |
| FIELD NAMES: |  |  |
| ID | Dimension1 | TRAWLTYPE |
| db | Dimension2 | TVI1 |
| fishery | Dimension3 | TVI2 |
| CALLSIGN | Dimension4 | DISCARD |
| depart_date | Dimension5 | Discardsobserved |
| shot date | Dimension6 | ESTCATCH |
| SHOTNU | Dimension7 | REASONOTH |
| dtstart | Gearloss | RETCATCH |
| endlatitude | Haulfinishtime | BENTHOS |
| endlongitude | Haulfrom | BOTTOMSUB |
| startlatitude | Haulstarttime | BOTTOMTOP |
| startlongitude | Longlinetype | SUBOTHER |
| DEPTHMAX | Material1 | TEMPNETS |
| DEPTHMIN | Material2 | TEMPSST |
| hauldate | Material3 | TOPOTHER |
| searchtime | Material4 | Tidedirection |
| CONTACT | Material5 | VALOTHER |
| DSEM1 | Material6 | WINDDIR |
| DSEM2 | NETNU | WINDSPEED |
| DSEM2OTHER | NETOPEN | CLOUD |
| GRIDFIN | Numberoffloaters | Moonphase |
| GRIDSTART | Setdamage | SEAHT |
| LATERR | Shotfinishtime | SWELLHT |
| LONGERR | Sinkrate | INFOCOMP |
| TIMEZONE | TOWDIR | Targetfish |
| gear | TOWSPEED | target1 |
| Baitefficiency | TRAWLDES1 | target2 |
| Baiter | TRAWLDES3 | target3 |
|  |  | targetshot |
|  |  | COMMENTS |

## Table 2 - Appendix A: GENLOG Dataset Supplied by AFMA.

| TABLE: Ms_san_operator_landing | TABLE: Ms_san_receiver_landing | TABLE: Ms_species_code |
| :---: | :---: | :---: |
| RECORD COUNT: 723366 | RECORD COUNT: 859780 | RECORD COUNT: 631 |
| FIELD NAMES: | FIELD NAMES: | FIELD NAMES: |
| log_book type code | log_book type code | species code |
| log_book_serial_no | log_book_serial_no | csiro_code |
| log book page_no | log book page no | effective date |
| species code | species code | fao species code |
| process_form_code | process_form_code | animal_type |
| state catch kg | state catch kg | caab version |
| commonwealth catch_kg | Commonwealth catch_kg | common name |
| high_seas_catch_kg | high_seas_catch_kg | converted_wt_type |
| area of waters | area of waters | quota species code |
| permissible activity | permissible activity | redundant date |
| container_type_code | container_type_code | species_name |
| csiro code | csiro code | default_process form_code |
| number of carcasses | number_of carcasses | epbc_act_defined |
| number_of_containers | number_of_containers |  |
| process_conversion factor | process conversion_factor |  |
| record_no | record no |  |
| trip_limit_trigger_flag |  |  |
|  |  |  |
|  |  |  |
| TABLE: Ms_vessels |  | TABLE: OPERATION COLLECTION |
| RECORD COUNT: 12604 |  | RECORD COUNT: 646 |
| FIELD NAMES: | ... | FIELD NAMES: |
| vessel_id | vessel_contact_phone | ID_CSIRO |
| vessel name | vessel depth | record no |
| carrying capacity | vessel depth type | dive method |
| home_port | vessel_length | depth_maximum |
| hull material | vessel length type | depth_measure |
| hull units | vessel photo | depth_minimum |
| last_hull_survey | vessel_tonnage | no_of_divers |
| nationality | year built | no of tenders |
| other dist symbol | Vcode | total dive hours |
| other_dist_state | vms_required | trip_length |
| owner_id | Vtype |  |
| place built | port of registry |  |
| power_units | previous_flag |  |
| primary dist symbol | australian flag_status |  |
| primary dist state |  |  |
| radio_call_sign |  |  |
| int_radio_call_sign |  |  |
| registered_user_id |  |  |
| registration_date |  |  |
| ships register number |  |  |
| vessel breadth |  |  |
| vessel_breadth_type |  |  |
| vessel colour |  |  |
| TABLE: OPERATION | TABLE: OPERATION_CSIRO | TABLE: OPERATION_DREDGE |
| RECORD COUNT: 2264962 | RECORD COUNT: 2264962 | RECORD COUNT: 3051 |
| FIELD NAMES: | FIELD NAMES: | FIELD NAMES: |
| ID CSIRO | ID CSIRO | ID CSIRO |
| log_book_type_code | ERA_SUB_FISHERY_ID | record_no |
| log book_serial_no | FISHERY ID | depth maximum |
| log_book page no | LOG BOOK TYPE CODE | depth measure |
| shot_number | GEAR_CODE_CSIRO | depth_minimum |
| activity code | TRAWL STRATA | grids traversed |
| authorised representative | ACTIVITY CODE | hours fished |
| client_id | AVG_TRAWL_DEPTH_METRES | number_of_hauls |
| concession_id | AVG_TRAWL_DEPTH_METRES ID DATA QUALITY | operation_activity |
| distinguishing symbol | AVG_TRAWL DEPTH_METRES_GIS |  |
| end grid | LENGTH_KM | TABLE: |
| Fishwell Consulting | 163 | AFMA Project F2008/0627 |


|  |  | OPERATION_LONGLINE |
| :---: | :---: | :---: |
| end latitude | TRAWL_SPEED_KMHR | RECORD COUNT: 490270 |
| end_latitude_degrees | START_TIME_DEC | FIELD NAMES: |
| end latitude minutes | END TIME DEC | ID CSIRO |
| end location | EFFORT TIME DECIMAL | ID HISTORICAL |
| end_longitude | UTC_OFFSET | record_no |
| end longitude degrees | DAY NIGHT | avg branch line length |
| end longitude minutes | SEF ZONE | avg_branch_ln_len_meas |
| end_time | OR_ZONE | avg_bubble_line_length |
| fishery id | SEF ZONE GIS | avg bubble ln len_meas |
| fishing_method | STATE WATERS_GIS | depth code |
| fishing_method_historical | SHARK_AREA | depth_maximum |
| gear_code | SHARK REGION | depth measure |
| gear code historical | VESSEL ID CSIRO | depth minimum |
| licence_no | VESSEL_NAME_CSIRO | direction_of_set |
| moon_phase | VESSEL_CALLSIGN_CSIRO | line shooter used |
| record no | VESSELNAME_CALLSIGN | main line length |
| season | DANISHSEINE | main_line_length_meas |
| shot date | SHOT NUMBER_CSIRO | no hooks btwn bubbles |
| start grid | SHOT DATE | no_light_sticks_used |
| start_latitude | YEAR | number_of_hooks |
| start latitude degrees | MONTH | number of lines |
| start latitude _minutes | DAY | set direction |
| start_location | LONGITUDE_START | gear_lost_line_length |
| start longitude | LATITUDE_START | gear_lost no hooks |
| start longitude degrees | LONGITUDE END | time start haul |
| start_longitude_minutes | LATITUDE_END | time_end_haul |
| start time | QUARDEG_CODE | start haul latitude degrees |
| time zone | HALFDEG CODE | start haul latitude minutes |
| vessel_id | HALFDEG_CODE_AFMA | start haul_latitude |
| waters fished | C SQ CODE | start haul longitude degrees |
| zone marker | INDEX1KM ID | start haul longitude minutes |
| vessel_name | INDEX1KM_ID_TRAWLEND | start_haul_longitude |
| entered_by | DATA_SOURCE | end_haul_latitude_degrees |
| entered_on | DATA SOURCE ID | end haul latitude minutes |
| changed_by | LENGTH_KM_GREATCIRCLE | end_haul_latitude |
| changed_on | RECORD_NO | end_haul_longitude_degrees |
| DATA SOURCE | ID HISTORICAL | end haul longitude minutes |
| ID_CSIRO_2008 | ID_CSIRO_2007 | end_haul_longitude |
| ID CSIRO 2008 AUTONUMBER | VESSEL NAME FROMID | vessel shooting_speed |
| ID_HISTORICAL | VESSEL_SYMBOL_FROMID |  |
|  | ID_CSIRO_AUTONUMBER |  |
|  | ID_CSIRO_2008A |  |
| TABLE: OPERATION MINOR LINE | TARGET_SPECIES | TABLE: ZONES_SHARK_AREA |
| RECORD COUNT: 31114 | net_mesh_size | RECORD COUNT: 58 |
| FIELD NAMES: | total net length | FIELD NAMES: |
| ID_CSIRO | net height | ZONENAME |
| record no | ID Historical, total net length lost | ZONENO |
| fishing method |  | LATITUDE |
| am_hours_fished | TABLE: VESSELS_CSIRO | LONGITUD |
| depth_maximum | RECORD COUNT: 1631 | wlong |
| depth_measure | FIELD NAMES: | elong |
| depth_minimum | VESSEL_ID_CSIRO | nwlat |
| hours fished | VESSEL NAME | nelat |
| hours searched | DISTINGUISHING_SYMBOL | swlat |
| no_of_crew | BUYOUT | selat |
| no of tenders | VESSEL NAME ORIGINAL |  |
| number_of_hooks | DISTINGUISHING_SYMBOL_ORIGINA | TABLE: ZONES SHARK REGIONS |
| number_of_lines | VESSEL_ID_CSIRO_20070214 | RECORD COUNT: 13 |
| pm hours fished |  | FIELD NAMES: |
| trip_length | TABLE: <br> VESSELS_CSIRO_ALLVESSELS | ZONENAME |
| port of_departure | RECORD COUNT: 2257 | ZONENO |


| target group | FIELD NAMES: | WLONG |
| :---: | :---: | :---: |
| weight conversion_code | VESSEL ID CSIRO | ELONG |
| ave_weight per_carton | VESSEL_NAME | NWLAT |
| no trays or cartons | DISTINGUISHING SYMBOL | NELAT |
| gear_lost line length | VESSEL BUYOUT | SWLAT |
| gear_lost_no_hooks | VESS_STD | SELAT |
| line length | COMMENT | shregcode |
| no of line lifts | VESSEL NAME AFMA | shreg |
| avg_hooks_per_line | DISTINGUISHING_SYMBOL_AFMA |  |
| time_start haul | VESSEL ID AFMA | TABLE: CATCH |
| time end haul | VESSEL_ID AFMA ORIGINAL | RECORD COUNT: 7652739 |
| ID_Historical | VESSEL_ID_CSIRO_20070214 | FIELD NAMES: |
|  | NAME FROMID | ID CSIRO |
| ```TABLE: OPERATION_MITIGATION_ME ASURES``` | SYMBOL_FROMID | csiro_code |
| RECORD COUNT: 200357 | CountOfID_CSIRO | csiro_code_historical |
| FIELD NAMES: |  | species code |
| ID CSIRO | TABLE: Zones_OR | species_code historical |
| record_no | RECORD COUNT: 8 | catch_kg |
| mitigation_measure code | FIELD NAMES: | fishing_method |
| number used | ZONENAME | weight_conversion_code |
| ID_Historical | ZONENO | meridian |
|  | WLONG | tender_number |
| TABLE: OPERATION_NET | ELONG | grade |
| RECORD COUNT: 32505 | NWLAT | life_status |
| FIELD NAMES: | NELAT | conversion factor used |
| ID CSIRO | SWLAT | est percent of school |
| record_no | SELAT | estimated_kgs_discarded |
| assisted by pole boat |  | estimated kgs_kept |
| assisting dist symbols | TABLE: Zones_SEF | hours fished |
| depth_maximum | RECORD COUNT: 18 | no_of_fish_kept |
| depth_measure | FIELD NAMES: | no_of fish released |
| depth minimum | ZONENAME | number of boxes |
| end_depth | ZONENO | number_of_cartons |
| end haul | WLONG | targeted flag |
| hours searched | ELONG | comments |
| net_code | NWLAT | DISCARD_CODE |
| spotter plane_used | NELAT | RECORD NO |
| start depth | SWLAT | ID Historical |
| start_haul | SELAT | ID_CSIRO_2007 |
| TABLE: CODE_Depths | $\begin{aligned} & \text { TABLE: } \\ & \text { CODE_FISHERY_SUBFISHERY } \end{aligned}$ | ```TABLE: OPERATION_CSIRO_DATAQUA LITY``` |
| RECORD COUNT: 21 | RECORD COUNT: 50 | RECORD COUNT: 794549 |
| FIELD NAMES: | FIELD NAMES: | FIELD NAMES: |
| MINDEPTH | ERA SUB FISHERY ID | ID CSIRO |
| MAXDEPTH | ERA FISHERY ID | ID DATA QUALITY |
|  | SUB_FISHERY_STATUS |  |
| TABLE: CODE_Effort_unit_code | FISHERY_ID_OLD | ```TABLE: DANISH_SEINE_YEARED``` |
| RECORD COUNT: 34 | FISHERY TEP | RECORD COUNT: 631 |
| FIELD NAMES: | ERA_SUB_FISHERY_NAME | FIELD NAMES: |
| ID | DBUPDATES ENDSTAGE2 | FISHERY ID |
| effort unit_code | CODE SPATIAL_UNIT | VESSEL ID CSIRO |
| description | EFFORT_MIN_SHOTS | YEAR |
|  | SHAPEFILE NAME | VesselName |
| TABLE: CODE_Effort_unit_sub_code | SHAPEFILE_NAME_AFMA | CallSign |
| RECORD COUNT: 31 | SPATIAL_OVERLAP | VesselName_CallSign |
| FIELD NAMES: | SPATIAL_OVERLAP HABITATS | SOURCE |
| ID |  |  |
| effort unit_sub_code | TABLE: CODE_DATA_QUALITY | TABLE: Dim_Depth |
| description | RECORD COUNT: 50 | RECORD COUNT: 2550 |
|  | FIELD NAMES: | FIELD NAMES: |
| TABLE: CODE_FISHERY | ID_DATA_QUALITY | Depth |


| RECORD COUNT: 31 | DATA QUALITY | Depth Band 10M |
| :---: | :---: | :---: |
| FIELD NAMES: | DATA QUALITY SQL | Depth_Band 50M |
| ERA_FISHERY_ID | MODULE | Depth_Band_100M |
| ERA FISHERY NAME |  | Depth_Strata |
|  | TABLE: CODE_QUOTA_SPECIES | Depth_Sub Strata |
| TABLE: <br> CODE_FISHERY_OPS_GEARTY PES | RECORD COUNT: 42 | Depth_Strata_Sort |
| RECORD COUNT: 118 | FIELD NAMES: | Depth_Sub_Strata Sort |
| FIELD NAMES: | ID |  |
| ERA_FISHERY_ID | Fishery_CSV | TABLE: Fishery |
| ERA SUB FISHERY ID | AFMA Code | RECORD COUNT: 73 |
| FISHERY ASSESSED | CAAB Code 8 8 Digit NewAFMA | FIELD NAMES: |
| NAME_AFMA | CAAB_Code_8Digit_CSIRO_lng | fishery_id |
| FISHERY ID | CAAB Code 6Digit OldAFMA | name |
| LOG_BOOK TYPE_CODE | CAAB Code_MAFRI_6Digit | previous season |
| GEAR_CODE_CSIRO | Scientific_name | current_season |
| GEAR ASSESSED | Common_Name | next season |
| EFFORT_MIN_SHOTS | Notes | description |
| code_description | VALUE_DOLLARSPERKG_MEL_2005 | compliance_contact |
| CountOfID CSIRO |  | concession renew date |
| Comment | TABLE: Ms_mss_codes_master | fish_receiver permit |
| NAME_AFMA_SHAPEFILE | RECORD COUNT: 196 | fishery_manager |
|  | FIELD NAMES: | fishery manager phone |
|  | code type | fishery_operational |
|  | code_type_name | fishery_type_id |
|  | code type units | fishing permit |
|  | code type description | licensing_contact |
|  |  | licensing_contact phone |
|  |  | monitoring_contact |
|  |  | monitoring_contact phone |
|  |  | principle_concess_type |
|  |  | principle contact |
|  |  | senior_licensing contact |
|  |  | sfr_certificate |
|  |  | sfr_extract |
| TABLE: Ms san landing | TABLE: Meta_Quota Species_Current | TABLE: OPERATION_TRAWL |
| RECORD COUNT: 895123 | RECORD COUNT: 43 | RECORD COUNT: 1534015 |
| FIELD NAMES: | FIELD NAMES: | FIELD NAMES: |
| VESSEL_NAME_FROMID | CAAB CODE | ID_CSIRO |
| VESSEL_SYMBOL_FROMID | QUOTA_FISHERY | record_no |
| VESSEL ID CSIRO | QUOTA SPECIES | depth_maximum |
| VESSELNAME_CALLSIGN |  | depth_measure |
| log_book_type_code | TABLE: Ms_logbook | depth_maximum_metres |
| log_book_serial_no | RECORD COUNT: 95 | depth_minimum |
| log book page no | FIELD NAMES: | depth_temperature |
| species_code | log_book_type_code | depth_temperature_meas |
| permissible activity | log_book_name | depth_temperature_type |
| process_form code | log_book_use_type | depth_type_code |
| fishery_id | replaces_log_book | est_wt_cod_end_meas |
| season | catch_summary_table | est_wt_cod_end_pre_sort |
| area of waters |  | ground_gear_code |
| decrement_quota_code | TABLE: Ms_mss_codes_detail | hours_searched |
| op commonwealth total kg | RECORD COUNT: 2064 | net_code |
| op high seas total kg | FIELD NAMES: | shot_valid |
| op_number_of_carcasses | code_type | trawl_speed |
| op state total kg | short_code | trawl_speed_meas |
| port unloaded | long_code | trawl_time |
| quota_holder_client_id | code_description | cod_end_mesh_size |
| rec_commonwealth total kg |  | cod_end_mesh_size_meas |
| rec high seas total kg | TABLE: OPERATION_NON_TRAWL | mesh_configuration |
| rec_number_of_carcasses | RECORD COUNT: 147167 | ground_gear_disk_height |
| rec state total kg | FIELD NAMES: | ground_gear_disk_height_meas |
| record no | ID_CSIRO | trawl_configuration |
| trip_end_date | record_no | trawl_strata |
| unit_code | depth measure | avg trawl_depth |


| vessel_id | hours_searched | avg_trawl_depth_meas |
| :--- | :--- | :--- |
| converted_wt | depth_maximum | avg_trawl_depth_temp |
| converted_wt_type | depth_minimum | avg_trawl_depth_temp_meas |
| decrement_fishery_id | target_species | target_species_code |
| decrement_licence_no | time_end_haul | total_no_shots |
| decrement_concession_id | time_start_haul | ID_Historical |
| last_updated_timestamp | ID_Historical |  |
| YEAR |  |  |
| DANISHSEINE |  |  |
|  |  |  |
| TABLE: FISHING_EFFORT |  |  |
| RECORD COUNT: 1573671 |  |  |
| FIELD NAMES: |  |  |
| ID_CSIRO |  |  |
| gear_code |  |  |
| effort_unit_code |  |  |
| effortunit sub_code |  |  |
| effort_unit_value |  |  |
| effor_unit_sub_code_value |  |  |
| record no |  |  |
| ID_Historical |  |  |

Table 3: Appendix A: CDR Dataset Supplied by AFMA.

| BLE: CDR |
| :--- |
| RECORD COUNT: 923485 |
| FIELD NAMES: |
| Vessel Name |
| Primary Dist Symbol |
| Log Book Type Code |
| Log Book Serial No |
| Log Book Page No |
| Species Code |
| Common Name |
| Species Name |
| Permissible Activity |
| Process Form Code |
| Fishery Id |
| Season |
| Area Of Waters |
| Decrement Quota Code |
| Op Commonwealth Total Kg |
| Op High Seas Total Kg |
| Op Number Of Carcasses |
| Op State Total Kg |
| Port Unloaded |
| Quota Holder Client Id |
| Rec Commonwealth Total Kg |
| Rec High Seas Total Kg |
| Rec Number Of Carcasses |
| Rec State Total Kg |
| Record No |
| Trip End Date |
| Unit Code |
| Vessel Id |
| Converted Wt |
| Converted Wt Type |
| Decrement Fishery Id |

Table 4 - Appendix A: Record count (at trip level) breakdown of the CDR data supplied by AFMA for this study by Fishery_ID.

| Fishery | Trip <br> Count |
| :--- | :--- |
| GAB | 1746 |
| GHAT | 20359 |
| SET | 89614 |

For the purpose of this study the Fishery_Ids have been allocated to Fisheries GAB, GHAT or SET on the following basis:
SET = ECD, SEN, SET or VIT
$\mathrm{GAB}=\mathrm{GAB}, \mathrm{GBQ}$
GHAT = GHT, SSF, SSG or SSH.
Fishery_Ids 'SCW', 'STR' and 'TCW' have been omitted from the redesign of the ISMPs.
Table 5-Appendix A: Breakdown of the record count in the CDR data by Fishery (as defined above) and year of Trip start date.

| YEAR | GAB | GHAT | SET |
| :--- | :---: | :---: | :---: |
| $\mathbf{1 9 9 6}$ | 0 | 0 | 1 |
| $\mathbf{1 9 9 7}$ | 0 | 14 | 636 |
| $\mathbf{1 9 9 8}$ | 0 | 18 | 10417 |
| $\mathbf{1 9 9 9}$ | 0 | 930 | 10497 |
| $\mathbf{2 0 0 0}$ | 0 | 1482 | 10212 |
| $\mathbf{2 0 0 1}$ | 167 | 1728 | 9286 |
| $\mathbf{2 0 0 2}$ | 117 | 1343 | 9518 |
| $\mathbf{2 0 0 3}$ | 250 | 2801 | 8252 |
| $\mathbf{2 0 0 4}$ | 326 | 2807 | 8062 |
| $\mathbf{2 0 0 5}$ | 308 | 2348 | 7204 |
| $\mathbf{2 0 0 6}$ | 255 | 2198 | 5726 |
| $\mathbf{2 0 0 7}$ | 215 | 2646 | 5344 |
| $\mathbf{2 0 0 8}$ | 108 | 2044 | 4459 |

Table 6 - Appendix A: A breakdown of the record count in the GENLOG data by Fishery (as defined above) and year of shot date.

| YEAR | GAB | GHAT | SET |
| :--- | :---: | :---: | :---: |
| $\mathbf{1 9 8 5}$ | 0 | 0 | 17425 |
| $\mathbf{1 9 8 6}$ | 368 | 0 | 66275 |
| $\mathbf{1 9 8 7}$ | 753 | 0 | 61883 |
| $\mathbf{1 9 8 8}$ | 4621 | 0 | 67587 |
| $\mathbf{1 9 8 9}$ | 9270 | 0 | 67970 |
| $\mathbf{1 9 9 0}$ | 4832 | 0 | 61234 |
| $\mathbf{1 9 9 1}$ | 3105 | 0 | 51453 |
| $\mathbf{1 9 9 2}$ | 2830 | 0 | 38187 |
| $\mathbf{1 9 9 3}$ | 2170 | 0 | 41362 |
| $\mathbf{1 9 9 4}$ | 1857 | 0 | 43735 |
| $\mathbf{1 9 9 5}$ | 2812 | 0 | 42187 |
| $\mathbf{1 9 9 6}$ | 3329 | 0 | 47568 |
| $\mathbf{1 9 9 7}$ | 4373 | 2 | 61920 |
| $\mathbf{1 9 9 8}$ | 3717 | 3686 | 62606 |
| $\mathbf{1 9 9 9}$ | 3715 | 11437 | 56585 |
| $\mathbf{2 0 0 0}$ | 2977 | 13377 | 49554 |
| $\mathbf{2 0 0 1}$ | 3215 | 12934 | 47138 |
| $\mathbf{2 0 0 2}$ | 2586 | 13181 | 44710 |
| $\mathbf{2 0 0 3}$ | 4601 | 15135 | 43446 |
| $\mathbf{2 0 0 4}$ | 5589 | 15967 | 40994 |
| $\mathbf{2 0 0 5}$ | 6261 | 13742 | 36858 |
| $\mathbf{2 0 0 6}$ | 5921 | 13596 | 30311 |
| $\mathbf{2 0 0 7}$ | 4537 | 10692 | 22307 |
| $\mathbf{2 0 0 8}$ | 2831 | 8428 | 21998 |

Table 7 - Appendix A: Potential for using vessel_Ids or vessel_names as the basis for the CDR - GENLOG merge.

| FISHERY = SET |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Dataset | Time Period | \# null vessel_ids | \#unique vessel_ids | \#vessel ids in GENLO $\bar{G}$ not in CDR | \#vessel_ids in CDR not in GENLOG |
| GENLOG | >= 1996 | 0 | 360 |  |  |
|  | > 2001 | 0 | 191 | 10 |  |
| CDR | $>=1996$ | 0 | 298 |  |  |
|  | $>=2001$ | 0 | 220 |  | 39 |
| FISHERY = SET |  |  |  |  |  |
| Dataset | Time Period | \# null vessel_names | \#unique vessel_names | \#vessel_names in GENLOG not in CDR | \#vessel_names in CDR not in GENLOG |
| GENLOG | >= 1996 | 0 | 393 |  |  |
|  | $>=2001$ | 0 | 184 | 16 |  |
| CDR | $>=1996$ | 0 | 271 |  |  |
|  | $>=2001$ | 0 | 206 |  | 36 |
| FISHERY = GAB |  |  |  |  |  |
| Dataset | Time Period | \# null vessel_ids | \#unique vessel_ids | \#vessel ids in GENLOG not in CDR | \#vessel_ids in CDR not in GENLOG |
| GENLOG | >= 1996 | 0 | 360 |  |  |
|  | >= 2001 | 0 | 191 | 10 |  |
| CDR | > $=1996$ | 0 | 298 |  |  |
|  | $>=2001$ | 0 | 220 |  | 39 |
| FISHERY = GAB |  |  |  |  |  |
| Dataset | Time Period | $\begin{gathered} \text { \# null } \\ \text { vessel_names } \end{gathered}$ | \#unique vessel_names | \#vessel names in GENLOG not in CDR | \#vessel_names in CDR not in GENLOG |
| GENLOG | > $=1996$ | 0 | 393 |  |  |
|  | $>=2001$ | 0 | 184 | 16 |  |
| CDR | $>=1996$ | 0 | 271 |  |  |
|  | $>=2001$ | 0 | 206 |  | 36 |
| FISHERY = GHAT |  |  |  |  |  |
| Dataset | Time Period | \# null vessel_ids | \#unique vessel_ids | \#vessel ids in GENLOG not in CDR | \#vessel_ids in CDR not in GENLOG |
| GENLOG | > $=1996$ | 0 | 360 |  |  |
|  | >= 2001 | 0 | 191 | 10 |  |
| CDR | $>=1996$ | 0 | 298 |  |  |
|  | $>=2001$ | 0 | 220 |  | 39 |
| FISHERY = GHAT |  |  |  |  |  |
| Dataset | Time Period | \# null vessel_names | \#unique vessel_names | \#vessel names in GENLOG not in CDR | \#vessel_names in CDR not in GENLOG |
| GENLOG | >= 1996 | 0 | 393 |  |  |
|  | >= 2001 | 0 | 184 | 16 |  |
| CDR | $>=1996$ | 0 | 271 |  |  |
|  | $>=2001$ | 0 | 206 |  | 36 |

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|  | GAB |  |  |  | GHAT |  |  |  | SET |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | CDR <br> Vessels | GENLO <br> G Vessels | Matched Vessels | \% <br> Matched | $\begin{aligned} & \text { CDR } \\ & \text { Vessels } \end{aligned}$ | GENLO <br> G Vessels | Matched Vessels | $\begin{aligned} & \text { \%Match } \\ & \text { ed } \end{aligned}$ | CDR <br> Vessels | GENLO <br> G Vessels | Matched Vessels | $\%$ Matched |
| 1996 | 0 | 1 | 0 | 0.00\% | 0 | 0 | 0 | 0.00\% | 1 | 1 | 0 | 0.00\% |
| 1997 | 0 | 1 | 0 | 0.00\% | 7 | 1 | 0 | 0.00\% | 65 | 158 | 54 | 34.18\% |
| 1998 | 0 | 1 | 0 | 0.00\% | 2 | 58 | 1 | 1.72\% | 174 | 138 | 66 | 47.83\% |
| 1999 | 0 | 1 | 0 | 0.00\% | 69 | 107 | 59 | 55.14\% | 179 | 78 | 67 | 85.90\% |
| 2000 | 0 | 11 | 0 | 0.00\% | 93 | 95 | 76 | 80.00\% | 179 | 156 | 144 | 92.31\% |
| 2001 | 7 | 10 | 7 | 70.00\% | 90 | 112 | 81 | 72.32\% | 173 | 144 | 137 | 95.14\% |
| 2002 | 6 | 8 | 5 | 62.50\% | 75 | 109 | 74 | 67.89\% | 171 | 151 | 143 | 94.70\% |
| 2003 | 11 | 13 | 11 | 84.62\% | 144 | 141 | 139 | 98.58\% | 105 | 104 | 103 | 99.04\% |
| 2004 | 10 | 12 | 10 | 83.33\% | 136 | 134 | 133 | 99.25\% | 97 | 99 | 96 | 96.97\% |
| 2005 | 10 | 10 | 10 | 100.00\% | 114 | 116 | 112 | 96.55\% | 99 | 98 | 97 | 98.98\% |
| 2006 | 12 | 12 | 12 | 100.00\% | 104 | 104 | 100 | 96.15\% | 86 | 87 | 85 | 97.70\% |
| 2007 | 9 | 9 | 9 | 100.00\% | 85 | 85 | 79 | 92.94\% | 56 | 54 | 52 | 96.30\% |
| 2008 | 6 | 7 | 6 | 85.71\% | 74 | 74 | 64 | 86.49\% | 54 | 52 | 51 | 98.08\% |

Table 9 - Appendix A: Percent loss of CDR Landing Records and GENLOG Shot records using Vessel_Ids as the basis for merging CDR and GENLOG records.

|  | GAB |  | GHAT |  | SET |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Percent <br> Shots <br> Lost | Percent <br> Landings <br> Lost | Percent <br> Shots <br> Lost | Percent <br> Landings <br> Lost | Percent <br> Shots <br> Lost | Percent <br> Landings <br> Lost |
| $\mathbf{1 9 9 6}$ | $100.00 \%$ | $100.00 \%$ | $0.00 \%$ | $0.00 \%$ | $100.00 \%$ | $100.00 \%$ |
| $\mathbf{1 9 9 7}$ | $100.00 \%$ | $100.00 \%$ | $100.00 \%$ | $100.00 \%$ | $88.48 \%$ | $11.01 \%$ |
| $\mathbf{1 9 9 8}$ | $100.00 \%$ | $100.00 \%$ | $98.53 \%$ | $5.56 \%$ | $88.15 \%$ | $86.25 \%$ |
| $\mathbf{1 9 9 9}$ | $100.00 \%$ | $100.00 \%$ | $22.30 \%$ | $8.82 \%$ | $89.40 \%$ | $83.55 \%$ |
| $\mathbf{2 0 0 0}$ | $100.00 \%$ | $100.00 \%$ | $7.36 \%$ | $6.01 \%$ | $27.72 \%$ | $7.06 \%$ |
| $\mathbf{2 0 0 1}$ | $2.24 \%$ | $0.00 \%$ | $25.54 \%$ | $3.94 \%$ | $1.75 \%$ | $6.89 \%$ |
| $\mathbf{2 0 0 2}$ | $10.09 \%$ | $8.55 \%$ | $35.39 \%$ | $1.04 \%$ | $0.66 \%$ | $5.89 \%$ |
| $\mathbf{2 0 0 3}$ | $6.65 \%$ | $0.00 \%$ | $0.38 \%$ | $0.61 \%$ | $0.01 \%$ | $0.11 \%$ |
| $\mathbf{2 0 0 4}$ | $0.54 \%$ | $0.00 \%$ | $0.46 \%$ | $0.18 \%$ | $0.20 \%$ | $0.15 \%$ |
| $\mathbf{2 0 0 5}$ | $0.00 \%$ | $0.00 \%$ | $0.69 \%$ | $2.13 \%$ | $0.14 \%$ | $0.03 \%$ |
| $\mathbf{2 0 0 6}$ | $0.00 \%$ | $0.00 \%$ | $1.90 \%$ | $0.73 \%$ | $1.57 \%$ | $0.03 \%$ |
| $\mathbf{2 0 0 7}$ | $0.00 \%$ | $0.00 \%$ | $0.82 \%$ | $1.70 \%$ | $0.09 \%$ | $2.25 \%$ |
| $\mathbf{2 0 0 8}$ | $3.00 \%$ | $0.00 \%$ | $1.68 \%$ | $3.72 \%$ | $0.02 \%$ | $0.56 \%$ |

Table 10 - Appendix A: Total landed weight by quota and major non-quota species against the summed GENLOG catch weights.

This table illustrates the species mismatching problem in the CDR-GENLOG merge for all fisheries where YEAR >= 2002

| Project Key | GENLOG Catch <br> Weight | CDR Landing <br> Weight | Percentage <br> Match |
| :--- | :---: | :---: | :---: |
| Royal red prawn | 1350009.9 | 1371434.4 | $98.44 \%$ |
| Orange roughy | 17038623.5 | 17309543.3 | $98.43 \%$ |
| Redfish | 3380943.3 | 3486894.73 | $96.96 \%$ |
| Eastern school whiting | 3871403.2 | 3997978.7 | $96.83 \%$ |
| Ornate angelshark | 1105432.1 | 1152120.1 | $95.95 \%$ |
| Blue-eye trevalla | 3989024.35 | 4175656.1 | $95.53 \%$ |
| Blue grenadier | 50861167.53 | 54177170.41 | $93.88 \%$ |
| Bight redfish | 6094551.9 | 6515329.99 | $93.54 \%$ |
| Smooth oreodory | 1777876.5 | 1906056.05 | $93.28 \%$ |
| Deepwater flathead | 13241539.36 | 14334068.37 | $92.38 \%$ |
| Mirror dory | 3648492.3 | 3992210.46 | $91.39 \%$ |
| Whiskery shark | 212147.42 | 235478.58 | $90.09 \%$ |
| Pink ling | 11719972.05 | 13050593.69 | $89.80 \%$ |
| Blue morwong | 521770.75 | 581155.42 | $89.78 \%$ |
| Silver warehou | 23328625.9 | 25987733.97 | $89.77 \%$ |
| Blue warehou | 2308993.22 | 2582293.87 | $89.42 \%$ |
| Hapuku | 815399.34 | 919319.09 | $88.70 \%$ |
| Tiger flathead | 24753102.47 | 28236611.55 | $87.66 \%$ |
| Deepwater shark basket | 2428702.56 | 2771367.36 | $87.64 \%$ |
| Inshore ocean perch | 2458553.61 | 2806436.58 | $87.60 \%$ |
| Frostfish | 852027.72 | 981601.59 | $86.80 \%$ |
| Gemfish | 4224221.95 | 4877661.41 | $86.60 \%$ |
| Knifejaw | 681818.68 | 792319.43 | $86.05 \%$ |
| Gould squid | 5680313.95 | 6634745.29 | $85.61 \%$ |
| Jackass morwong | 7260210.73 | 8487013.3 | $85.54 \%$ |
| John dory | 958326.32 | 1121145.38 | $85.48 \%$ |
| Bronze whaler | 159404.44 | 187136.41 | $85.18 \%$ |
| Alfonsino | 1019790.37 | 1201555.94 | $84.87 \%$ |
|  |  |  |  |
|  |  |  |  |


| Project Key | GENLOG Catch Weight | CDR Landing Weight | Percentage Match |
| :---: | :---: | :---: | :---: |
| King dory | 1786586.95 | 2110343.14 | 84.66\% |
| School shark | 1433681.9 | 1694569.23 | 84.60\% |
| Ribaldo | 1401461.65 | 1684531.13 | 83.20\% |
| Silver trevally | 920948.95 | 1120009.29 | 82.23\% |
| Barracouta | 762319.4 | 928758.7 | 82.08\% |
| Gummy shark | 13475799.48 | 17005277.75 | 79.24\% |
| Hapuku (unspecified) | 39845.25 | 51398.31 | 77.52\% |
| Octopus | 504855.95 | 653635.57 | 77.24\% |
| Triggerfish \& leatherjacket (unspecified) | 668825.79 | 867109.31 | 77.13\% |
| Sawshark basket | 2589056.1 | 3356834.21 | 77.13\% |
| Snapper | 501050.42 | 653897.2 | 76.63\% |
| Crustaceans | 296067.25 | 390483.75 | 75.82\% |
| Stargazer (unspecified) | 1633748.06 | 2155280.12 | 75.80\% |
| Yellowspotted boarfish | 860571.18 | 1145863.9 | 75.10\% |
| Smooth hammerhead | 55151.25 | 73790.45 | 74.74\% |
| Oreo basket | 1505866.16 | 2024952.07 | 74.37\% |
| Latchet | 1642283.12 | 2240577.83 | 73.30\% |
| Bluestriped goatfish | 155837.81 | 213983.71 | 72.83\% |
| Broadnose shark | 251877.14 | 356314.44 | 70.69\% |
| Southern eagle ray | 95799.2 | 141038.47 | 67.92\% |
| Sharks | 222740.77 | 331385.72 | 67.21\% |
| Ocean jacket | 3124600.36 | 4899550.72 | 63.77\% |
| Elephantfish | 497483.86 | 782578.41 | 63.57\% |
| Fish | 3079245.36 | 4864152.66 | 63.30\% |
| Guitarfish (unspecified) | 277948.7 | 443514.9 | 62.67\% |
| Red gurnard | 1445804.86 | 2323327.39 | 62.23\% |
| Silver dory | 408001.28 | 657278.31 | 62.07\% |
| Dogfish | 472209.76 | 762515.25 | 61.93\% |
| Boarfish (unspecified) | 245753.51 | 404982.01 | 60.68\% |
| Rays | 352575 | 607942.08 | 57.99\% |
| Striped trumpeter | 22218.29 | 39083.93 | 56.85\% |
| Chimaeras | 71489.9 | 152009.38 | 47.03\% |
| Ocean blue-eye trevalla | 711.56 | 1601 | 44.44\% |
| Australian angelshark | 480989.5 | 1165809 | 41.26\% |
| Squid (general) | 516289.76 | 1358246.38 | 38.01\% |
| Molluscs | 442672.85 | 1172277.82 | 37.76\% |
| Sawsharks | 21399.8 | 57739.3 | 37.06\% |
| Draughtboard shark | 46991.2 | 141324.55 | 33.25\% |
| Triggerfish \& leatherjacket (unspecified) | 668825.79 | 867109.31 | 77.13\% |
| Sawshark basket | 2589056.1 | 3356834.21 | 77.13\% |
| Snapper | 501050.42 | 653897.2 | 76.63\% |
| Crustaceans | 296067.25 | 390483.75 | 75.82\% |
| Stargazer (unspecified) | 1633748.06 | 2155280.12 | 75.80\% |
| Yellowspotted boarfish | 860571.18 | 1145863.9 | 75.10\% |
| Smooth hammerhead | 55151.25 | 73790.45 | 74.74\% |
| Oreo basket | 1505866.16 | 2024952.07 | 74.37\% |
| Latchet | 1642283.12 | 2240577.83 | 73.30\% |
| Bluestriped goatfish | 155837.81 | 213983.71 | 72.83\% |
| Broadnose shark | 251877.14 | 356314.44 | 70.69\% |
| Southern eagle ray | 95799.2 | 141038.47 | 67.92\% |
| Sharks | 222740.77 | 331385.72 | 67.21\% |
| Ocean jacket | 3124600.36 | 4899550.72 | 63.77\% |
| Elephantfish | 497483.86 | 782578.41 | 63.57\% |
| Fish | 3079245.36 | 4864152.66 | 63.30\% |
| Guitarfish (unspecified) | 277948.7 | 443514.9 | 62.67\% |
| Red gurnard | 1445804.86 | 2323327.39 | 62.23\% |
| Silver dory | 408001.28 | 657278.31 | 62.07\% |
| Dogfish | 472209.76 | 762515.25 | 61.93\% |
| Boarfish (unspecified) | 245753.51 | 404982.01 | 60.68\% |


| Project Key | GENLOG Catch <br> Weight | CDR Landing <br> Weight | Percentage <br> Match |
| :--- | :---: | :---: | :---: |
| Rays | 352575 | 607942.08 | $57.99 \%$ |
| Striped trumpeter | 22218.29 | 39083.93 | $56.85 \%$ |
| Chimaeras | 71489.9 | 152009.38 | $47.03 \%$ |
| Ocean blue-eye trevalla | 711.56 | 1601 | $44.44 \%$ |
| Australian angelshark | 480989.5 | 1165809 | $41.26 \%$ |
| Squid (general) | 516289.76 | 1358246.38 | $38.01 \%$ |
| Molluscs | 442672.85 | 1172277.82 | $37.76 \%$ |
| Sawsharks | 21399.8 | 57739.3 | $37.06 \%$ |
| Draughtboard shark | 46991.2 | 141324.55 | $33.25 \%$ |

Table 11 - Appendix A: The six values associated with each trapezoid describing in total the area of each of the $\mathbf{7}$ zones in the SET fishery.

| ZONENAME | ZONENO | WLONG | ELONG | NWLAT | NELAT | SWLAT | SELAT |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| NSW | 10 | 147 | 155.9543 | 33.583 | 33.583 | 37.25 | 37.25 |
| East Victoria | 20 | 148 | 148.317 | 38.4 | 38.4 | 40.75 | 40.75 |
| East Victoria | 20 | 148.317 | 148.733 | 38.4 | 37.75 | 40.75 | 40.75 |
| East Victoria | 20 | 148.733 | 155 | 37.25 | 37.25 | 40.75 | 40.75 |
| East Tasmania | 30 | 147 | 148 | 42 | 40.75 | 48 | 48 |
| East Tasmania | 30 | 148 | 154 | 40.75 | 40.75 | 48 | 48 |
| West Tasmania | 40 | 138.133 | 144 | 40 | 40 | 48 | 48 |
| West Tasmania | 40 | 144 | 146 | 40 | 42 | 48 | 48 |
| West Tasmania | 40 | 146 | 147 | 42 | 42 | 48 | 48 |
| West Victoria | 50 | 138.1 | 144 | 34 | 34 | 40 | 40 |
| Bass Strait | 60 | 144 | 146 | 37.25 | 37.25 | 40 | 42 |
| Bass Strait | 60 | 146 | 147 | 37.25 | 37.25 | 42 | 42 |
| Bass Strait | 60 | 147 | 148 | 37.25 | 37.25 | 42 | 40.75 |
| Bass Strait | 60 | 148 | 148.317 | 37.25 | 37.25 | 38.4 | 38.4 |
| Bass Strait | 60 | 148.317 | 148.733 | 37.25 | 37.25 | 38.4 | 37.75 |
| ECDW | 70 | 155.5 | 157 | 24.49833 | 24.49833 | 27 | 30 |
| ECDW | 70 | 157 | 165 | 24.49833 | 24.49833 | 35 | 35 |
| ECDW | 70 | 155.9543 | 157 | 33.58167 | 33.58167 | 34.00717 | 35 |

Table 12 - Appendix A: Stratum definitions for the SET ISMP study.
The column headed 'Precedence' means that rules take precedence in the order 1-15. For example, if there is any ambiguity and a shots falls into more than one stratum, then the precedence order will remove

| Stratum Code | Description | Precedence | ZONE / depth / month | Gear | Flag code | Species |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1. BS_IN_TR | Victoria Inshore Trawl | 14 | $60 /<200$ / all | Not D | Vicin | Any |
| 2. ECDW_TR | East Coast Deep Water Trawl | 15 | 70 | Not D | Ecdw | Any |
| 3. EDL_DS | Victoria East Coast Danish Seine | 11 | All / all / all | D | Edlds | Any |
| 4. EDL_IN_TR | Victoria East Coast Inshore Trawl | 12 | $20 /<200 /$ all | Not D | Edin | Any |
| 5. EDL_OFF_TR | Victoria East Coast Offshore Trawl | 13 | $20 />200 /$ all | Not D | Edoff | Any |
| 6. NSW_IN_TR | New South Wales Inshore Trawl | 9 | $10 /<200 /$ all | Not D | Nswin | Any |
| 7. NSW_OFF_TR | New South Wales Offshore Trawl | 10 | $10 />200 /$ all | Not D | Nswoff | Any |
| 8. NSW_RRP_TR | New South Wales Royal Red Prawn Trawl | 8 | All / all / all | All | Rrp | $\begin{array}{r} 701004 \\ >50 \mathrm{~kg} \\ \hline \end{array}$ |
| 9. SW_BGS_TR | Victoria West Orange Roughy | 7 | $50 /$ all / 6-8 | All | Bgs | $227001$ <br> present |
| 10. SW_ORO_TR | South West Orange Roughy Trawl | 2 | 50 / all / 6-9 | All | Oros | $\begin{gathered} 255009 \\ >50 \% \end{gathered}$ |
| 11. SW_TR | South West Trawl | 5 | 50 / all / all | All | Oths | Any |
| 12. TAS_BGS_TR | Tasmania Spawning Blue Grenadier | 6 | 30,40 / all / 6-8 | All | Bgt | $227001$ <br> present |
| 13. TAS_E_TR | Tasmania East Coast Trawl | 3 | 30 / all / all | All | Othe | Any |
| 14. TAS_ORO_TR | Tasmania Orange Roughy Trawl | 1 | 30,40 / all / 6-9 | All | Orot | $\begin{gathered} 255009 \\ >50 \% \\ \hline \end{gathered}$ |
| 15. TAS_W_TR | Tasmania West Coast Trawl | 4 | 40 / all / all | All | Othw | Any |

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Table 13 - Appendix A: Unknown SEF ISMP shots emerging from the zone allocation and first set of stratum allocation rules as described in the text.

|  | $\underset{\text { _TR_IN }}{\substack{\text { BS }}}$ | $\begin{gathered} \hline \text { ECD } \\ \text { W_TR } \\ \hline \end{gathered}$ | $\begin{gathered} \text { EDL } \\ \text { _DS } \end{gathered}$ | $\underset{\mathbf{T}_{\mathbf{R}}}{\text { EDLIN }}$ | $\underset{\substack{\text { EDL_OFF } \\ \mathbf{T}_{\mathbf{R}} \\ \hline}}{ }$ | $\underset{T \bar{R}}{\mathbf{N S W}_{-} \mathbf{I N}}$ | $\underset{\substack{\text { NSW_OFF } \\ \text { _TR }}}{ }$ | $\begin{gathered} \text { NSW_RRP } \\ \hline \mathbf{T R} \\ \hline \end{gathered}$ | $\underset{\overline{\text { TR }}}{\substack{\text { SW_BG_ }}}$ | $\underset{\substack{\text { SW_OR } \\ \text { O_TR }}}{ }$ | $\underset{\text { R }}{\text { SW_T }}$ | $\begin{gathered} \text { TAS_BG } \\ \mathbf{S}_{\mathbf{T}} \mathbf{T R} \end{gathered}$ | $\underset{\mathbf{T R}}{ }{ }^{\text {TAS_E}}$ | $\begin{gathered} \text { TAS_OR } \\ \text { O_TR } \\ \hline \end{gathered}$ | $\begin{aligned} & \text { TAS_ } \\ & \text { W_T_R } \end{aligned}$ | unknown | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 0 | 0 | 0 | 2 | 12 | 7 | 2 | 0 | 0 | 0 | 0 | 13 | 9 | 14 | 0 | 0 | 59 |
| 1993 | 0 | 0 | 27 | 25 | 24 | 18 | 8 | 0 | 0 | 0 | 0 | 15 | 103 | 24 | 74 | 200 | 518 |
| 1994 | 0 | 0 | 246 | 51 | 37 | 0 | 0 | 0 | 4 | 36 | 100 | 2 | 42 | 21 | 27 | 272 | 838 |
| 1995 | 13 | 0 | 180 | 42 | 25 | 0 | 0 | 0 | 12 | 0 | 125 | 18 | 0 | 0 | 57 | 122 | 594 |
| 1996 | 1 | 0 | 0 | 43 | 27 | 0 | 0 | 6 | 21 | 0 | 178 | 10 | 45 | 0 | 0 | 264 | 595 |
| 1997 | 0 | 0 | 0 | 148 | 52 | 0 | 0 | 0 | 31 | 11 | 132 | 0 | 28 | 0 | 4 | 314 | 720 |
| 1998 | 0 | 0 | 23 | 90 | 97 | 87 | 76 | 16 | 20 | 10 | 120 | 52 | 51 | 14 | 15 | 1 | 672 |
| 1999 | 2 | 0 | 47 | 93 | 65 | 224 | 131 | 54 | 24 | 2 | 120 | 40 | 89 | 40 | 12 | 0 | 943 |
| 2000 | 6 | 6 | 27 | 73 | 53 | 243 | 111 | 8 | 16 | 4 | 85 | 8 | 114 | 20 | 4 | 0 | 778 |
| 2001 | 16 | 32 | 41 | 118 | 93 | 236 | 134 | 17 | 23 | 8 | 94 | 42 | 62 | 22 | 36 | 0 | 974 |
| 2002 | 12 | 13 | 34 | 81 | 62 | 208 | 75 | 26 | 12 | 2 | 144 | 9 | 64 | 9 | 40 | 1 | 792 |
| 2003 | 30 | 0 | 114 | 100 | 73 | 296 | 94 | 11 | 14 | 6 | 97 | 33 | 22 | 9 | 31 | 1 | 931 |
| 2004 | 6 | 0 | 48 | 47 | 76 | 297 | 128 | 4 | 31 | 1 | 97 | 8 | 35 | 25 | 1 | 0 | 804 |
| 2005 | 1 | 0 | 71 | 132 | 74 | 271 | 80 | 5 | 30 | 2 | 187 | 19 | 21 | 25 | 28 | 1 | 947 |
| 2006 | 0 | 0 | 130 | 116 | 66 | 213 | 88 | 4 | 15 | 0 | 120 | 5 | 66 | 22 | 9 | 1 | 855 |
| 2007 | 0 | 0 | 24 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 24 |
| Total | 87 | 51 | 1012 | 1161 | 836 | 2100 | 927 | 151 | 253 | 82 | 1599 | 274 | 751 | 245 | 338 | 1177 | 11044 |


|  | $\begin{aligned} & \hline \mathbf{B S}_{2} \mathbf{I} \\ & \mathbf{N}_{\overline{\mathbf{R}}} \end{aligned}$ | $\begin{gathered} \text { ECD } \\ \mathbf{W}_{\mathbf{R}} \mathbf{T} \\ \hline \end{gathered}$ | $\underset{\text { EDL }}{\text { D }}$ | $\underset{-T \mathbf{R}}{\text { EDL_IN }}$ | $\underset{\text { EDR }}{\text { EDF }_{2} \mathrm{OFF}}$ | $\underset{T \bar{R}}{\mathbf{N S W}_{-} \mathbf{I N}}$ | $\underset{\text { _TR }}{\text { NSW_OFF }}$ | $\underset{\text { TR }}{\text { NSW_R }_{-}}$ | $\underset{\mathbf{T R}}{\text { SW }_{-}}$ | $\underset{\text { TR }}{\text { SW_ORO_ }_{-}}$ | SW_TR | $\begin{gathered} \text { TAS_BG } \\ \text { S_TR } \end{gathered}$ | TAS <br> E_TR | $\begin{gathered} \text { TAS_OR } \\ \text { O_TR } \end{gathered}$ | TAS W_TR | unknown and/or unzonable |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 0 | 0 | 0 | 2 | 12 | 7 | 2 | 0 | 0 | 0 | 0 | 13 | 9 | 14 | 0 | 0 |
| 1993 | 0 | 0 | 27 | 25 | 24 | 140 | 77 | 0 | 0 | 0 | 0 | 15 | 103 | 24 | 74 | 9 |
| 1994 | 0 | 0 | 246 | 51 | 37 | 170 | 87 | 0 | 4 | 36 | 100 | 2 | 42 | 21 | 27 | 15 |
| 1995 | 13 | 0 | 180 | 42 | 25 | 56 | 57 | 0 | 12 | 0 | 125 | 18 | 0 | 0 | 57 | 9 |
| 1996 | 1 | 0 | 0 | 43 | 27 | 159 | 96 | 6 | 21 | 0 | 178 | 10 | 45 | 0 | 0 | 9 |
| 1997 | 0 | 0 | 0 | 148 | 52 | 192 | 110 | 0 | 31 | 11 | 132 | 0 | 28 | 0 | 4 | 37 |
| 1998 | 0 | 0 | 23 | 91 | 97 | 87 | 76 | 16 | 20 | 10 | 120 | 52 | 51 | 14 | 15 | 20 |
| 1999 | 2 | 0 | 47 | 93 | 65 | 224 | 132 | 54 | 24 | 2 | 120 | 40 | 89 | 40 | 12 | 3 |
| 2000 | 6 | 51 | 27 | 73 | 53 | 243 | 111 | 8 | 16 | 4 | 85 | 8 | 114 | 20 | 4 | 1 |
| 2001 | 16 | 32 | 41 | 118 | 93 | 236 | 134 | 17 | 23 | 8 | 94 | 42 | 62 | 22 | 36 | 3 |
| 2002 | 12 | 69 | 34 | 81 | 62 | 208 | 75 | 26 | 12 | 2 | 144 | 9 | 64 | 9 | 40 | 1 |
| 2003 | 30 | 0 | 114 | 100 | 73 | 296 | 94 | 11 | 14 | 6 | 97 | 33 | 22 | 9 | 31 | 2 |
| 2004 | 6 | 25 | 48 | 47 | 76 | 297 | 128 | 4 | 31 | 1 | 97 | 8 | 35 | 25 | 1 | 0 |
| 2005 | 1 | 0 | 71 | 133 | 74 | 272 | 81 | 5 | 30 | 2 | 187 | 19 | 21 | 25 | 28 | 0 |
| 2006 | 0 | 0 | 130 | 117 | 66 | 213 | 88 | 4 | 15 | 0 | 120 | 5 | 66 | 22 | 9 | 0 |
| 2007 | 0 | 0 | 24 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 87 | 177 | 1012 | 1164 | 836 | 2800 | 1348 | 151 | 253 | 82 | 1599 | 274 | 751 | 245 | 338 | 109 |

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| Strata | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BS_IN_TR | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| ECDW_TR | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 53 | 62 | 421 | 113 | 169 | 230 | 60 | 1 | 75 | 1 |
| EDL_DS | 8116 | 8263 | 8874 | 8108 | 9478 | 10131 | 10981 | 10018 | 7722 | 8260 | 8948 | 10129 | 8330 | 7689 | 6495 | 6397 | 6440 |
| EDL_IN_TR | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| EDL_OFF_TR | 6046 | 7204 | 7853 | 7436 | 7910 | 10917 | 9688 | 9633 | 10277 | 8820 | 8549 | 8516 | 7677 | 8247 | 6816 | 4652 | 4911 |
| NSW_IN_TR | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| NSW_OFF_TR | 6608 | 7568 | 9169 | 8984 | 9327 | 8495 | 8189 | 8864 | 9565 | 8766 | 8432 | 9918 | 10048 | 8236 | 6401 | 3882 | 4103 |
| NSW_RRP_TR | 420 | 747 | 894 | 1319 | 1493 | 1153 | 1222 | 1430 | 1282 | 1094 | 1447 | 648 | 444 | 442 | 327 | 211 | 157 |
| SW_BGS_TR | 193 | 279 | 247 | 180 | 381 | 424 | 297 | 479 | 540 | 388 | 353 | 317 | 502 | 414 | 435 | 315 | 191 |
| SW_ORO_TR | 66 | 267 | 459 | 662 | 412 | 236 | 270 | 195 | 110 | 158 | 141 | 90 | 93 | 79 | 55 | 8 | 0 |
| SW_TR | 2704 | 2886 | 2859 | 3794 | 4778 | 7982 | 7716 | 5755 | 5574 | 5715 | 5247 | 4261 | 4966 | 3963 | 3721 | 2882 | 2243 |
| TAS_BGS_TR | 208 | 203 | 324 | 477 | 525 | 577 | 614 | 1158 | 992 | 1220 | 1095 | 1082 | 831 | 446 | 482 | 332 | 318 |
| TAS_E_TR | 3479 | 3464 | 3692 | 2712 | 2958 | 3623 | 3776 | 3718 | 3395 | 3400 | 3584 | 2846 | 2981 | 2424 | 1955 | 1268 | 1364 |
| TAS_ORO_TR | 1234 | 1690 | 1324 | 779 | 1016 | 694 | 582 | 598 | 1134 | 692 | 654 | 300 | 247 | 270 | 245 | 68 | 116 |
| TAS_W_TR | 1753 | 1994 | 1973 | 2838 | 1985 | 2583 | 3051 | 2536 | 2945 | 3857 | 3464 | 2647 | 1944 | 1459 | 1059 | 1377 | 1067 |
| VIT | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 488 | 606 | 415 | 174 | 341 | 740 | 424 | 355 | 275 |
| unknown | 1512 | 1994 | 1382 | 896 | 1325 | 8573 | 8764 | 4870 | 1915 | 1922 | 1054 | 1352 | 1644 | 1783 | 1472 | 213 | 479 |

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| Strata | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BS_IN_TR | 1236 | 1334 | 702 | 336 | 582 | 4484 | 4223 | 2689 | 728 | 414 | 159 | 166 | 412 | 749 | 608 | 47 | 178 |
| ECDW_TR | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 53 | 62 | 421 | 113 | 169 | 230 | 60 | 1 | 75 | 1 |
| EDL_DS | 8116 | 8263 | 8874 | 8108 | 9478 | 10131 | 10981 | 10018 | 7722 | 8260 | 8948 | 10129 | 8330 | 7689 | 6495 | 6397 | 6440 |
| EDL_IN_TR | 4039 | 4680 | 5064 | 4987 | 5425 | 7292 | 6512 | 5754 | 6301 | 5247 | 5417 | 4890 | 4853 | 5631 | 4605 | 3190 | 3224 |
| EDL_OFF_TR | 2007 | 2524 | 2789 | 2449 | 2485 | 3625 | 3176 | 3879 | 3976 | 3573 | 3132 | 3626 | 2824 | 2616 | 2211 | 1462 | 1687 |
| NSW_IN_TR | 4590 | 5355 | 6610 | 6583 | 6594 | 5356 | 5280 | 5718 | 6456 | 6408 | 6179 | 7513 | 7892 | 6173 | 4881 | 2837 | 2983 |
| NSW_OFF_TR | 2018 | 2213 | 2559 | 2401 | 2733 | 3139 | 2909 | 3146 | 3109 | 2358 | 2253 | 2405 | 2156 | 2063 | 1520 | 1045 | 1120 |
| NSW_RRP_TR | 420 | 747 | 894 | 1319 | 1493 | 1153 | 1222 | 1430 | 1282 | 1094 | 1447 | 648 | 444 | 442 | 327 | 211 | 157 |
| SW_BGS_TR | 193 | 279 | 247 | 180 | 381 | 424 | 297 | 479 | 540 | 388 | 353 | 317 | 502 | 414 | 435 | 315 | 191 |
| SW_ORO_TR | 66 | 267 | 459 | 662 | 412 | 236 | 270 | 195 | 110 | 158 | 141 | 90 | 93 | 79 | 55 | 8 | 0 |
| SW_TR | 2704 | 2886 | 2859 | 3794 | 4778 | 7982 | 7716 | 5755 | 5574 | 5715 | 5247 | 4261 | 4966 | 3963 | 3721 | 2882 | 2243 |
| TAS_BGS_TR | 208 | 203 | 324 | 477 | 525 | 577 | 614 | 1158 | 992 | 1220 | 1095 | 1082 | 831 | 446 | 482 | 332 | 318 |
| TAS_E_TR | 3479 | 3464 | 3692 | 2712 | 2958 | 3623 | 3776 | 3718 | 3395 | 3400 | 3584 | 2846 | 2981 | 2424 | 1955 | 1268 | 1364 |
| TAS_ORO_TR | 1234 | 1690 | 1324 | 779 | 1016 | 694 | 582 | 598 | 1134 | 692 | 654 | 300 | 247 | 270 | 245 | 68 | 116 |
| TAS_W_TR | 1753 | 1994 | 1973 | 2838 | 1985 | 2583 | 3051 | 2536 | 2945 | 3857 | 3464 | 2647 | 1944 | 1459 | 1059 | 1377 | 1067 |
| VIT | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 488 | 606 | 415 | 174 | 341 | 740 | 424 | 355 | 275 |
| unknown | 276 | 660 | 680 | 560 | 743 | 4089 | 4541 | 2181 | 1187 | 1508 | 895 | 1186 | 1232 | 1034 | 864 | 166 | 301 |



Figure 1 - Appendix A: Histogram of the distribution of computed GENLOG catch weight correction factors illustrating the unreasonably large correction values discussed in the text.


Figure 2-Appendix A: Histogram of the distribution of computer GENLOG catch weight correction factors, limiting the factor to a maximum of 4 .


Figure 3-Appendix A: A plot of the locations of the SEF ISMP shots, for shots that fall into zones 10, 20, 30, 40, 50, 60 and 70.


## SET STRATIFICATION

Figure 4 - Appendix A: Graphical representation of existing ISMP Strata rule definitions for the SET Fishery.


Figure 5 - Appendix A: "Gains" plot, panelled by the dataset partition of the decision tree's accuracy at imputing the depth category for shots with a depth greater than 200m.


Figure 6 - Appendix A: "Gains" plot, panelled by the dataset partition of the decision tree's accuracy at imputing the depth category for shots with a depth less than 200 m .

## Appendix B - Cluster Analysis Results

Appendix B - Cluster Analysis Results for GABTF


Figure 1 - Appendix B: GAB Fishery, Cluster sizes.


Figure 2 - Appendix B: GAB Fishery Deepwater Flathead Cluster species composition.


Figure 3 - Appendix B: GAB Fishery Redfish Cluster species composition.


Figure 4 - Appendix B: GAB Fishery Grenadier Cluster species composition.


Figure 5 - Appendix B: GAB Fishery Gemfish Cluster species composition.


Figure 6 - Appendix B: GAB Fishery Orange Roughy Cluster species composition.


Figure 7 - Appendix B: GAB Clusters By Year.


Figure 8 - Appendix B: GAB Clusters By Year Normalised.


Figure 9 - Appendix B: GAB Clusters, Longitude, Cluster, Year.


Figure 10 - Appendix B: GAB Clusters, Longitude, Cluster, Year (Normalised).


Figure 11 - Appendix B: GAB Clusters By Longitude.


Figure 12 - Appendix B: GAB Clusters By Longitude (Normalised).


Figure 13 - Appendix B: GAB Clusters by Depth, Panelled By Year.


Figure 14 - Appendix B: GAB Clusters By Depth, Panelled By Year - Normalised.


Figure 15 - Appendix B: GAB Clusters By Depth.


Figure 16 - Appendix B. GAB Clusters By Depth, Normalised.


Figure 17 - Appendix B: GAB Clusters by Year, Depth and Longitude.


Figure 18 - Appendix B: GAB Clusters by Year, Depth Category and Longitude, Normalised.

\$KM-GAB Clusters
$\square$ DEEP WATER FLATHEAD
$\square$ GEMFISH
$\square$ GRENADIER
ORANGE ROUGHY
$\square$ REDFISH

Figure 19 - Appendix B: GAB Clusters by Depth Category and Longitude.


Figure 20 - Appendix B: GAB Clusters by Depth Category and Longitude, Normalised.

## Appendix B - Cluster Analysis Results for GHAT (AL \& BL)



Figure 21 - Appendix B: GHaT (AL \& BL) Cluster Sizes


Figure 22 - Appendix B: GHaT (AL \& BL) Pink Ling Cluster Species Composition.


Figure 23 - Appendix B: GHaT (AL \& BL) Gummy Shark Cluster Species Composition.


Figure 24 - Appendix B: GHaT (AL \& BL) Blue-Eye Trevalla Cluster Species Composition.


Figure 25 - Appendix B: GHaT (AL \& BL) Mixed Cluster Species Composition.


Figure 26 - Appendix B: GHaT (AL \& BL) Hapuku Cluster Species Composition.


Figure 27 - Appendix B: GHaT (AL \& BL) Clusters By Year.


Figure 28 - Appendix B: GHaT (AL \& BL) Clusters By Year Normalised.


Figure 29 - Appendix B: GHaT (AL \& BL) Longitude, Cluster, Year.


Figure 30 - Appendix B: GHaT (AL \& BL) Longitude, Cluster, Year: Normalised.


Figure 31 - Appendix B: GHaT (AL \& BL) Longitude.


Figure 32 - Appendix B: GHaT (AL \& BL) Longitude, Cluster (Normalised).


Figure 33 - Appendix B: GHaT (AL \& BL) Depth, Cluster, Year.


Figure 34 - Appendix B: GHaT (AL \& BL) Depth, Cluster, Year: Normalised.


Figure 35-Appendix B: GHaT (AL \& BL) Clusters By Depth.


Figure 36 - Appendix B: GHaT (AL \& BL) Clusters By Depth: Normalised.

## Appendix B-Argument for Gear separation:



Figure 37 - Appendix B: GHaT (AL \& BL) Longitude, Cluster, Year, Gear.


Figure 38 - Appendix B: GHaT (AL \& BL) Longitude, Cluster, Year, Gear (Normalised).


Figure 39 - Appendix B: GHaT (AL \& BL) Longitude, Cluster, Gear.


Figure 40 - Appendix B: GHaT (AL \& BL) Longitude, Cluster, Gear (Normalised).


Figure 41 - Appendix B: GHaT (AL \& BL) Depth, Cluster, Year, Gear.


Figure 42 - Appendix B: GHaT (AL \& BL) Depth, Cluster, Year, Gear (Normalised).


Figure 43 - Appendix B: GHaT (AL \& BL) Depth, Cluster, Gear.


Figure 44 - Appendix B: GHaT (AL \& BL) Depth, Cluster, Year, Gear (Normalised).

## Appendix B - Cluster Analysis Results for GHAT (DL)



Figure 45 - Appendix B: GHaT (DL) Cluster Sizes


Figure 46 - Appendix B: GHaT (DL) Cluster Composition Blue-Eye Trevalla.


Figure 47 - Appendix B: GHaT (DL) Cluster Composition Hapuku.


Figure 48 - Appendix B: GHaT (DL) Clusters By Year.


Figure 49 - Appendix B: GHaT (DL) Clusters By Year Normalised.


Figure 50 - Appendix B: GHaT (DL) Longitude, Cluster, Year.


Figure 51 - Appendix B: GHaT (DL) Longitude, Cluster, Year (Normalised).

\$KM-GHAT (DL)
Пblue-eyetrevalla
$\square h A P U K U$

Figure 52 - Appendix B: GHaT (DL) Longitude, Cluster.


Figure 53 - Appendix B: GHaT (DL) Longitude Cluster (Normalised).


Figure 54 - Appendix B: GHaT (DL) Depth, Cluster, Year.


Figure 55 - Appendix B: GHaT (DL) Depth, Cluster, Year (Normalised).


Figure 56 - Appendix B: GHaT (DL) Depth, Cluster.


Figure 57 - Appendix B: GHaT (DL) Depth, Cluster (Normalised).

## Appendix B - Cluster Analysis Results for GHAT (GN)



Figure 58 - Appendix B: GHaT (GN) Cluster Sizes.


Figure 59 - Appendix B: GHaT (GN) Cluster Analysis Results Gummy Shark.


Figure 60 - Appendix B: GHaT (GN) Cluster Analysis Results Gummy Mixed.


Figure 61 - Appendix B: GHaT (GN) School Shark.


Figure 62 - Appendix B: GHaT (GN) Clusters By Year.


Figure 63 - Appendix B: GHaT (GN) Clusters By Year Normalised.

\$KM-GHAT (GN ONLY)
$\square$ GUMMY MIXED
GGUMMY SHARK
SCHOOL SHARK

Figure 64 - Appendix B: GHaT (GN) Longitude, Cluster, Year.


Figure 65 - Appendix B: GHaT (GN) Longitude, Cluster Year (Normalised).

\$KM-GHAT (GN ONLY)
$\square$ GUMMY MIXED
GUMMY SHARK
SCHOOL SHARK

Figure 66 - Appendix B: GHaT (GN) Longitude, Cluster.


Figure 67 - Appendix B: GHaT (GN) Longitude, Cluster (Normalised).


Figure 68 - Appendix B: GHaT (GN) Depth, Cluster, Year.


Figure 69 - Appendix B: GHaT (GN) Depth, Cluster, Year (Normalised).


Figure 70 - Appendix B: GHaT (GN) Depth, Cluster.


Figure 71 - Appendix B: GHaT (GN) Depth, Cluster (Normalised).

## CDR/GENLOG Species Match



Figure 72 - Appendix B: GHaT (GN) Percentage of CDR in GENLOG by species (All Fisheries).


Figure 73 - Appendix B: GHaT (GN) Percentage of CDR in GENLOG by species and year (All Fisheries).

## Appendix B - TEP Analysis



Figure 74 - Appendix B: TEP Analysis, Spatial Plot Seals.


Bird Shots
O 0
O 1


Figure 76 - Appendix B: TEP Analysis, Spatial Plot Sharks.


Figure 77 - Appendix B: TEP Analysis, Spatial Plot Whales/Dolphins.


Pipefish Shots

| $\mathrm{O}_{0}$ |
| :--- |
| O |

Figure 78 - Appendix B: TEP Analysis, Spatial Plot Pipefish.


[^0]:    ** - Note that the AL strata ultimately have to be disaggregated to the various line gear types pro rata to expected shots in those gear types ( $\mathrm{AL}+\mathrm{LL}, \mathrm{BL}, \mathrm{DL}$ ).

[^1]:    ** - Note that the AL strata ultimately have to be disaggregated to the various line gear types pro rata to expected shots in those gear types ( $A L+L L, B L, D L$ ).

[^2]:    ** - Note that the AL strata ultimately have to be disaggregated to the various line gear types pro rata to expected shots in those gear types (AL+LL, BL, DL).

[^3]:    ** - Note that the AL strata ultimately have to be disaggregated to the various line gear types pro rata to expected shots in those gear types (AL+LL, BL, DL).

[^4]:    ** - Note that the AL strata ultimately have to be disaggregated to the various line gear types pro rata to expected shots in those gear types (AL+LL, BL, DL).

