Australian Government
Australian Fisheries Management Authority

## Stock Assessment for the Southern and Eastern Scalefish and Shark Fishery: 2012



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## Cover photographs

Front cover, blue grenadier, ocean perch, flathead and orange roughy.

## Report structure

Part 1 of this report describes the Tier 1 assessments of 2012. Part 2 describes the Tier 3 and Tier 4 assessments, catch rate standardisations and other general work contributing to the assessment and management of SESSF stocks in 2012.

# Stock Assessment for the Southern and Eastern Scalefish and Shark Fishery: 2012 

Part 2: Tier 3 and Tier 4, catch rate standardisations and other work contributing to the assessment and management of SESSF stocks in 2012

G.N. Tuck<br>June 2013<br>Report 2010/0818

Australian Fisheries Management Authority

## Stock Assessment for the Southern and Eastern Scalefish and Shark Fishery: 2012 Part 2

## Table of Contents

12. MAINTAINING RISK EQUIVALENCY AMONG FISHERY HARVEST CONTROL
RULES IN THE SESSF
12.1 SUMMARY 1
12.2 InTRODUCTION 2
12.3 METHODS 4
12.4 ReSULTS 7
12.5 DISCUSSION 9
12.6 ACKNOWLEDGEMENTS 12
12.7 REFERENCES 13
12.8 TABLES 16
12.9 Figures 22
13. CATCH RATE STANDARDIZATIONS FOR SELECTED SPECIES FROM THE
SESSF (DATA 1986 - 2011)
13.1 SUMMARY 30
13.2 InTRODUCTION 31
13.3 METHODS 32
13.4 Results 35
13.5 School Whiting (WHS - 37330014) SilLago FLINDERSI 38
13.6 EAStERN GEmFISH (GEM - 37439002 - REXEA SOLANDRI) SPAwning FISHERY 42
13.7 EASTERN GEMFISH NON-SpAWNING (GEM - 37439002 - REXEA SOLANDRI) 46
13.8 JACKASS MORWONG Z10-50 (MOR - 37377003 NEMADACTYLUS MACROPTERUS) 50
13.9 JACKASS MORWONG Z1020 (MOR-37377003 N. MACROPTERUS) 57
13.10 JACKASS MORWONG Z30 (MOR - 37377003 N. MACROPTERUS) 61
13.11 JACKASS MORWONG Z4050 OT (MOR - 37377003 N. MACROPTERUS) 65
13.12 Flathead Trawl (FLT - 37296001 - Neoplatycephalus Richardsoni) 69
13.13 Flathead Trawl Z1020(FLT - 37296001 - N. RIChardsoni) 70
13.14 Flathead Trawl Z30 (FLT - 37296001 - N. RIChardsoni) 74
13.15 Flathead Danish Seine (FLT - 37296001 - N. RIChardsoni) 79
13.16 RedFish Zone 10 (RED - 37258003 - CENTROBERYX AFFINIS) 84
13.17 REDFISH ZONE 20 (RED - 37258003 - CENTROBERYX AFFINIS) 88
13.18 Silver Trevally (TRE - 37337062 - PSEUDOCARANX DENTEX) 92
13.19 Royal Red Prawn (PRR - 28714005 - HALIPOROIDES Sibogat) 96
13.20 Blue Eye, Z2030 (TBE - 37445001 - Hyperoglyphe antarctica) 102
13.21 Blue Eye, Z4050 (TBE - 37445001 - H. ANTARCTICA) 106
13.22 Blue Eye, AL (TBE - 37445001 - H. antarctica) 110
13.23 Blue Eye, DL (TBE - 37445001 - H. antarctica) 114
13.24 Blue Eye, AL \& DL (TBE - 37445001 - H. ANTARCTICA) 117
13.25 Blue Grenadier Spawning (GRE - 37227001 - Macruronus novaezelandiae) 123
13.26 Blue Grenadier Non-Spawning (GRE - 37227001 - M. novaezelandiae) 127
13.27 Silver Warehou (TRS - 37445006 - Seriolella punctata) 131
13.28 Blue Warehou Zones 10, 20, 30 (TRT - 37445005 - SERIOLELLA BRAMA) 136
13.29 BLUE WAREHOU Z4050 (TRT - 37445005 - S. BRAMA) 140
13.30 Blue Warehou Z10-50 (TRT - 37445005 - S. BRAMA) 144
13.31 Pink Ling TW (LIG - 37228002 - GENYPTERUS BLACODES) 149
13.32 Pink Ling, Z102030 (LIG-37228002 - G. BLACODES) 150
13.33 Pink Ling, Z4050 (LIG - 37228002 - G. BLACODES) 154
13.34 Pink Ling, Z10 (LIG - 37228002 - G. BLACODES) ..... 158
13.35 Pink Ling, Z20 (LIG - 37228002 - G. BLACODES) ..... 162
13.36 Pink Ling, Z30 (LIG - 37228002 - G. BLACODES) ..... 166
13.37 Pink Ling, Z40 (LIG-37228002 - G. BLACODES) ..... 170
13.38 Pink Ling, Z50 (LIG - 37228002 - G. BLACODES) ..... 174
13.39 Western Gemfish and GAB (GEM - 37439002 - REXEA SOLANDRI) ..... 178
13.40 Western Gemfish Z4050 (GEM - 37439002 - R. SOLANDRI) ..... 182
13.41 WESTERN GEMFISH GAB (GEM - 37439002 - R. SOLANDRI) ..... 186
13.42 Offshore Ocean Perch, Z1020 (REG - 37287001 - H. Percoides) 200M ..... 190
13.43 Inshore Ocean Perch, Z1020 (REG - 37287001 - H. PERCOIDES) 0-200M ..... 195
13.44 JOhn Dory (DOJ - 37264004) Zeus faber ..... 199
13.45 Mirror Dory (DOM - 37264003 ZENOPSIS NEBULOSUS) ..... 203
13.46 Mirror Dory East (DOM - 37264003 Zenopsis nebulosus) ..... 208
13.47 Mirror Dory West (DOM - 37264003 Zenopsis nebulosus) ..... 212
13.48 Ribaldo (RBD - 37224002 - MORA MORO) ..... 216
13.49 Ribaldo (RBD - 37224002 - MORA MORO) AUTOLINE ..... 222
13.50 OCEAN JACKETS (LTC - 37465006 - NELUSETTA AYRAUDI) ..... 224
13.51 OCEAN JACKETS - GAB (LTC - 37465006 - N. AYRAUDI) ..... 230
13.52 Bibliography ..... 234
14. CATCH RATE STANDARDIZATION UPDATES WITH DATA TO OCT 2012 ..... 237
14.1 SUMMARY ..... 237
14.2 Introduction ..... 238
14.3 Methods ..... 239
14.4 Results ..... 241
14.5 School Whiting (WHS - 37330014) Sillago FLindersi ..... 242
14.6 JACKASS MORWONG SUMMARY (MOR-37377003 NEMADACTYLUS MACROPTERUS) ..... 243
14.7 JACKASS MORWONG Z1020 (MOR-37377003 N. MACROPTERUS) ..... 244
14.8 JACKASS MORWONG Z30 (MOR - 37377003 N. MACROPTERUS) ..... 245
14.9 JACKASS MORWONG Z4050 (MOR - 37377003 N. MACROPTERUS) ..... 246
14.10 Flathead Summary (FLT - 37296001 - NEOPLATYCEPHALUS RICHARDSONI) ..... 247
14.11 RedFish Zone 10 (RED - 37258003 - CENTROBERYX AFFINIS) ..... 251
14.12 Silver Trevally (TRE - 37337062 - PseudocaranX dentex) ..... 252
14.13 Royal Red Prawn (PRR - 28714005 - Haliporoides sibogae) ..... 254
14.14 Blue Eye, AL \& DL (TBE - 37445001 - H. antarctica) ..... 255
14.15 BLUE GRENADIER SUMMARY (GRE - 37227001 - MACRURONUS NOVAEZELANDIAE) ..... 256
14.16 Silver Warehou (TRS - 37445006 - Seriolella punctata) ..... 259
14.17 Pink Ling Summary (LIG - 37228002 - Genypterus blacodes) ..... 260
14.18 Western Gemfish Z4050 (GEM - 37439002 - R. SOLANDRI) ..... 264
14.19 Offshore Ocean Perch, Z1020 (REG - 37287001 - H. Percoides) 200M ..... 265
14.20 Inshore Ocean Perch, Z1020 (REG - 37287001 - H. PERCOIDES) 0-200m ..... 266
14.21 John Dory (DOJ - 37264004 ) Zeus Faber ..... 267
14.22 Mirror Dory Summary (DOM - 37264003 Zenopsis nebulosus) ..... 268
14.23 Mirror Dory East (DOM - 37264003 Z. nebulosus) ..... 269
14.24 Mirror Dory West (DOM - 37264003 Z. nebulosus) ..... 270
14.25 Ribaldo (RBD - 37224002 - MORA MORO) ..... 271
14.26 Bibliography ..... 272
15. STANDARDIZATION OF BIGHT REDFISH IN THE GAB 2000/2001 - FEB 2011/2012. CATCH RATE UPDATE ..... 274
15.1 SUMMARY ..... 274
15.2 METHODS ..... 274
15.3 Results ..... 275
15.4 CONCLUSION ..... 276
15.5 AcKNOWLEDGEMENTS ..... 276
16. STANDARDIZATION OF DEEPWATER FLATHEAD IN THE GAB 2000/2001 - FEB 2011/2012. CATCH RATE UPDATE. ..... 282
Malcolm Haddon ..... 282
16.1 SUMMARY ..... 282
16.2 Methods ..... 282
16.3 Results ..... 283
16.4 CONCLUSION ..... 284
16.5 ACKNOWLEDGEMENTS ..... 284
17. STANDARDIZED CATCH RATES FOR THE SESSF GUMMY SHARK FISHERY: DATA FROM 1976-2011 ..... 290
Malcolm Haddon ..... 290
17.1 SUMMARY ..... 290
17.2 InTRODUCTION ..... 290
17.3 Methods ..... 291
17.4 Results ..... 295
17.5 South Australia ..... 298
17.6 BASS STRAIT ..... 301
17.7 TASMANIA ..... 305
17.8 Extra Tables ..... 308
17.9 Bibliography ..... 309
18. STANDARDIZED CATCH RATES FOR THE SESSF SAW SHARK AND ELEPHANT FISH FISHERIES. DATA FROM 1980-2011 ..... 311
Malcolm Haddon ..... 311
18.1 SUMMARY ..... 311
18.2 Introduction ..... 312
18.3 Methods ..... 313
18.4 Results ..... 317
18.5 DISCUSSION ..... 333
18.6 REFERENCES ..... 334
18.7 ADDITIONAL GRAPHS ..... 335
18.8 TABLES ..... 339
19. YIELD, TOTAL MORTALITY VALUES AND TIER 3 ESTIMATES FOR SELECTED SHELF AND SLOPE SPECIES IN THE SESSF 2012 ..... 354
19.1 Summary ..... 354
19.2 Methods ..... 355
19.3 Results ..... 364
19.4 RBC CALCULATIONS ..... 396
19.5 References ..... 398
20. TIER 4 ANALYSES IN THE SESSF, INCLUDING DEEP WATER SPECIES. DATA FROM 1986-2011 ..... 407
Malcolm Haddon ..... 407
20.1 SUMMARY ..... 407
20.2 SUMMARY OF RBCs AND DISCARDS ..... 408
20.3 Introduction ..... 411
20.4 Methods ..... 411
20.5 ReSUlts ..... 419
20.6 DEEP-WATER ..... 447
20.7 Non-Tier 4 Species ..... 485
20.8 Bibliography ..... 514
21. SAW SHARK AND ELEPHANT FISH TIER 4 ANALYSES (DATA 1980-2011) ..... 515
21.1 SUMMARY ..... 515
21.2 Introduction ..... 516
21.3 METHODS ..... 517
21.4 ReSUlts ..... 520
21.5 DISCUSSION ..... 529
21.6 TABLES ..... 530
21.7 AcKNOWLEDGEMENTS ..... 532
21.8 BiBLIOGRAPHY ..... 532
22. PROJECTING THE SCHOOL SHARK MODEL INTO THE FUTURE: REBUILDING TIMEFRAMES AND AUTO-LONGLINING IN SOUTH AUSTRALIA ..... 533
22.1 SUMMARY ..... 533
22.2 BACKGROUND ..... 533
22.3 Methods ..... 534
22.4 Results and conclusions ..... 535
22.5 References ..... 536
22.6 Figures ..... 537
22.7 TABLES ..... 541
23. INCIDENTAL BYCATCH RATIOS OF SCHOOL SHARK (GALEORHINUS GALEUS) TO GUMMY SHARK (MUSTELUS ANTARCTICUS) OFF SOUTH AUSTRALIA WHEN USING AUTOMATIC LONGLINES COMPARED WITH GILLNETS ..... 543
23.1 SUMMARY ..... 543
23.2 InTRODUCTION ..... 543
23.3 Methods ..... 544
23.4 Results ..... 545
23.5 CONCLUSIONS ..... 546
23.6 AcKNOWLEDGEMENTS ..... 546
23.7 REFERENCES ..... 546
24. PREDICTED PUP PRODUCTION OF GUMMY SHARK (MUSTELUS ANTARTICTCUS) ASSUMING AN AUTOMATIC LONGLINE FISHERY OFF SOUTH AUSTRALIA ..... 547
24.1 Summary ..... 547
24.2 Introduction ..... 547
24.3 Methods ..... 548
24.4 RESULTS AND DISCUSSION ..... 550
24.5 Conclusions ..... 553
24.6 AcKNOWLEDGEMENTS ..... 554
24.7 ReFERENCES ..... 554
25. BENEFITS ..... 555
26. CONCLUSION ..... 556
27. APPENDIX: INTELLECTUAL PROPERTY ..... 559
28. APPENDIX: PROJECT STAFF ..... 560

# 12. Maintaining risk equivalency among fishery harvest control rules in the SESSF 

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### 12.1 Summary

This chapter presents results from using management strategy evaluation (MSE) to test alternative ways of implementing precaution among the tier system of harvest control rules (HCRs) in the Southern and Eastern Scalefish and Shark fishery (SESSF). The work considers the relative performance of two data-poor harvest control rules (Tier 3, Tier 4) when compared to that obtained under Tier 1 (the most data-rich case).

We tested a range of discount factors and compared the risk of failing to achieve the fishery objectives under different management strategies, with that from the Tier 1 HCR. The perceived risk to the stock was defined as:

- the probability of estimating the stock to be below the limit reference point more than $10 \%$ of the time during the MSE simulations.

The discount factors were compared to three other approaches to achieving risk equivalency that relied on catch rate stability, reference point adjustment, and observation error. Three species were considered, tiger flathead, school whiting, and jackass morwong. Simulations made use of available information from recent stock assessments for these data-rich species.

Scenarios that only applied discount factors when catch rates were deemed to be stable (at least in the manner defined in these analyses) were found to be inadequate for implementing precaution. In these scenarios, larger discount factors were required to obtain risk equivalency.
As the outcomes were variable across the species, the harvest strategies, and the methods used to implement precaution, it is not possible to provide a simple conclusion that a single optimum method exists for balancing risk against uncertainty for each Tier level of assessment. Further work could include analyses utilizing different initial stock status, and alternative levels of observation error, to help explain the impact of these factors.

A critical assumption underlying the Tier 4 HCR is a linear relationship between CPUE and biomass. The analyses did not account for the effect of possible nonlinearity in this relationship. It can be expected that, if the true relationship is non-linear, then further precaution would be required in order to maintain the same perceived risk to the stock.

The analyses considered risk equivalency in terms of the perceived risk to the stock (as determined by the applied stock assessment method), as opposed to the actual risk to the stock (as determined by the underlying MSE operating model). In practice the perceived risk is what would be observed and used to judge the ability to meet management objectives. Risk could also be defined in terms of the probability and magnitude of the methods giving very low (and/or variable) catch quota advice.

While the analyses focused on risk equivalency with Tier 1, it may be more relevant to judge performance against a metric that quantifies an acceptable level of not meeting the harvest strategy objectives; such as staying above the limit reference point more than $10 \%$ of the time with a probability of 0.9 . Exactly how to implement such metrics, along with the time periods over which to calculate them, still needs to be prescribed.

### 12.2 Introduction

Marine fisheries are increasingly being managed using formal harvest strategies that include specification of harvest control rules (HCRs). Management agencies are also recognizing the need to incorporate estimates of uncertainty into the decision-making process, with the precautionary approach to management calling for more conservative action as the uncertainty regarding estimates of resource state increases (e.g. Caddy and McGarvey 1996). Specification of management action through the use of HCRs in datapoor situations can respond to this uncertainty by explicitly adjusting the output (e.g. recommended catch) as a function of uncertainty. Many fisheries management agencies have chosen to adopt precautionary targets within harvest control rule specifications (e.g. Deroba and Bence 2008, Froese et al. 2011). These methods assume that the selection and estimation of the appropriate reference points are sufficiently robust to uncertainty, and advice is not based on the amount of information available. In many cases these reference points will not all be equally estimable. Uncertainty associated with estimation and implementation error can impact the performance of such harvest control rules (Deroba and Bence 2008).

Methods for incorporating uncertainty into the management advice arising from HCRs have varied. The Revised Management Procedure of the International Whaling Commission directly penalizes catch limits based on data availability and also sets catch limits based on a percentile of a posterior distribution less than the mean estimate (e.g. Punt and Donovan 2007). Similarly the Potential Biological Removals (PBR) method, applied to set catch limits for U.S. marine mammal populations, uses an abundance estimate based on a percentile of the current estimate of abundance (Wade 1998). Control rules for toothfish stocks in the Antarctic adopted by the Commission for the Conservation of Antarctic Marine Living Resources (CCAMLR) are based on model projections that include both process and observation error (Constable et al. 2000). In South Africa, the management procedure for anchovy includes a scale-down factor whereby the Total Allowable Catch (TAC) is reduced prior to information on recruitment strength becoming available (de Moor et al. 2011). In the U.S., buffers to account for scientific uncertainty are applied to overfishing level estimates to calculate allowable catch limits for managed fished species (e.g. Punt et al 2012). A similar approach has been adopted in the Southern and Eastern Scalefish and Shark fishery (SESSF) in Australia, where a discount factor is applied to the TAC for quota species that are assessed and managed under data-poor HCRs (Smith et al. 2008, Smith et al. 2009).

The SESSF has operated in some form off Southeast Australia since the early 1900s, and is the leading source of fresh fish for Sydney and Melbourne. The fishery employs a diverse range of gears exploiting over one hundred species. The primary management tool within the SESSF for the major commercial species is an annual TAC implemented under an Individual Transferable Quota (ITQ) system. Additional management measures include gear restrictions and area closures, the latter of which are used mainly for conserving protected or at risk species (Smith and Wayte 2002, Smith et al. 2008). The TAC for each species is based on the calculation of a Recommended Biological Catch (RBC). RBCs are determined using a tier framework of HCRs, with the choice of tier rule for each species determined by the information available (Table 12.1). The Tier 1 HCR is used with the most information rich species, and the HCR relies on current estimates of biomass and fishing mortality rate. The data-poor Tier 3 and Tier 4 analyses rely on either equilibrium-based estimates of fishing mortality (Tier 3), or an empirical rule based on the current catch-per-unit-effort (cpue) relative to a selected reference level (Tier 4).

As the results of data-poor (Tier 3 and Tier 4) analyses are expected to be more uncertain than those from the data-rich Tier 1 (because the estimates of stock status are based on less information), this uncertainty is factored into the TAC calculation by application of a discount factor (currently $5 \%$ for Tier 3 and $15 \%$ for Tier 4 species), which reduces the RBC from the Tier 3 and 4 analyses (Smith et al. 2008, Smith et al 2009). Such a measure is important as the target and limit reference points used in the Tier 3 and Tier 4 HCRs are equivalent (at least in intent) to those used for Tier 1 stocks, and do not in themselves prescribe additional precaution.

In theory, if uncertainty associated with the data-poor HCRs is to be buffered, then management action should operate such that the risk (to the stock or fishery) of managing under the data-poor control rules should be at least the same as the data-rich case. That is, discount factors should operate so that they maintain risk equivalency among HCRs (Figure 12.1). Currently, perceived risk is defined as the estimated probability of failing to meet the objective that the stock should not fall below $20 \%$ of unfished spawning biomass more than $10 \%$ of the time (DAFF 2007).

While the method of discounting catch recommendations from less certain assessments in the SESSF is both transparent and consistent with a precautionary approach, there has been considerable discussion over the conditions under which these discount factors should be applied. Arguments for not applying the discount factors have included alternative management actions (e.g. the presence of closed areas), or the stability of indicators such as time series of catches and catch rates. The use of a discount factor is not the only possible method of implementing precaution among Tier levels, and it is not clear whether the current method actually achieves its intent of equalizing relative risk among Tier levels. Alternative methods could include the use of different reference points (e.g. alternative targets and limits), incorporating estimation uncertainty into the estimate of stock status used to calculate the RBC (as per the PBR method of Wade (1998)), and stability (or otherwise) in other stock indicators (the more unstable the indicator, the greater the reduction in the recommended catch). While the application of discount factors is designed to result in similar behaviour to these other approaches, it is not clear whether a) their application is appropriate, b) whether the current values
applied are the 4'best' ones, and c) whether uniform application across species is desirable.

This paper uses Management Strategy Evaluation (MSE, c.f. Bunnefeld et al. 2011) to investigate the performance of both the discount factors and alternative methods of implementing precaution when applying the SESSF data-poor harvest strategies. The MSE approach enables incorporation of many of the uncertainties associated with management of fish stocks, including process error in the stock dynamics (annual recruitments), observation error (sampling variability), estimation error (from stock assessment models and HCR application), and implementation error (actual catches may differ from the recommended management action).

We first use MSE to compare the performance of the data-poor harvest strategies both with and without the currently applied discount factors. These comparisons are made for three quota species in the SESSF for which data rich (Tier 1) assessments are available. We then determine the discount factors required for the data-poor harvest strategies so that these strategies are perceived to be risk equivalent with the Tier 1 HCR. Finally, we identify and explore tradeoffs among risk, catch, and stock size for alternative approaches of including precaution when implementing the harvest strategies. The analyses focus on using the perceived performance of the Tier 1 HCR as a baseline for judging the relative performance of the data-poor control rules, with respect to HCR performance against the risk-based objective of Australia's harvest strategy policy.

### 12.3 Methods

### 12.3.1 Simulation framework

The general MSE framework is similar to that previously conducted for SESSF species (e.g. Fay et al. 2011, Little et al. 2011, Wayte and Klaer 2010). An operating model, representing the true state of the system, is conditioned using the most recent stock assessment for each species, with uncertainty in true stock status characterised by the uncertainty in parameter estimates obtained using Bayesian methods (e.g. Fay and Tuck 2011). A set of scenarios are run for three species, with each scenario being characterised by the combination of a particular HCR (Tiers 1, 3, or 4), and for Tiers 3 and 4 , a method for implementing precaution in the harvest strategy. The operating model is projected forward under the application of a specified harvest strategy for thirty years, with data generated annually from the operating model representing the quantity and quality of data currently available, estimation using the appropriate stock assessment method for the HCR, application of the HCR, and updating of the TAC. This TAC is then converted to an actual catch that is used to update the dynamics of the operating model. A set of summary statistics are calculated after the thirty year projection period.
Table 12.1 describes the specifications for the three harvest strategies.

### 12.3.2 Species

The analyses were applied to three SESSF target species, using information from recent stock assessments: tiger flathead (Klaer 2010), school whiting (Day 2009), and jackass morwong (Wayte 2010). These species have biological characteristics representative of species typically caught in the SESSF on the continental shelf. Maximum age varies among the species (school whiting: 9, tiger flathead: 20, jackass morwong: 25). The
fishable school whiting population size is mostly determined by annual recruitment strength. The current assessment for jackass morwong indicates that the spawning stock biomass is close to the limit reference point and that the stock is depleted. Tiger flathead is currently assessed to be near the biomass target that is a proxy for the level needed to achieve maximum economic yield.

### 12.3.3 Harvest strategies

### 12.3.3.1 Tier 1

Tier 1 (T1) stock assessments are conducted annually using Stock Synthesis (SS; e.g. Methot 2007), according to the specifications for the base-case model of the most recent stock assessment (Table 12.2 and Table 12.3). Data available for the Tier 1 stock assessments within the MSE are: the historical data on catches, generated catch rates, and for some years the age structure and length structure of the catch. In addition, data are generated by the operating model according to the effective sample sizes of the contemporary data, as calculated in the published stock assessment reports (Table 12.2).

RBC calculation under T1 is obtained from the SS output using a F-based harvest control rule (
Table 12.1) with a target fishing mortality rate equivalent to that which on average would achieve a spawning biomass of $0.48 \mathrm{~B}_{0}$ using the current year's fishing pattern by fleet.

### 12.3.3.2 Tier 3

The Tier 3 (T3) harvest strategy uses recent age composition data to obtain an estimate of the current rate of fishing mortality (Wayte and Klaer 2010). Target and limit reference points are the values for the fishing mortality rate that result in the equilibrium stock depletion (ESD) values given in Table 12.1 (Cordue 2012). ESD is akin to the spawning potential ratio (SPR) except that the stock-recruitment relationship is taken into account.

### 12.3.3.3 Tier 4

The Tier 4 (T4) harvest strategy compares an estimate of current catch per unit effort (cpue) to that during a pre-specified reference period and adjusts a target catch from this reference period accordingly (Little et al. 2011). To account for uncertainty in the calculation of reference points when implementing $T 4$, the reference catch ( $\mathrm{C}^{*}$ ) and reference cpue ( $\mathrm{CPUE}_{\text {ref }}$ ) for each simulation are randomly determined from the historical data during the default reference period (1986-1995, Table 12.3). T4 can then be biased with respect to the T1 target biomass of $0.48 \mathrm{~B}_{0}$, and it is possible that $\mathrm{C}^{*}$ and $\mathrm{CPUE}_{\text {ref }}$ may not be in equilibrium with each other, as is assumed to be the case when applying T4 (Smith et al. 2008; Little et al. 2011). Data available to T4 are the historical cpue data, and data generated from the operating model according to the observation error ascribed to the cpue data in the most recent Tier 1 stock assessment (Table 12.2). RBC calculation is achieved by comparing the recent average cpue in the simulation to the selected value for the reference period, as described in Haddon (2011).

### 12.3.4 TAC calculation

The TAC is calculated by subtracting estimates of the discards (Table 12.2) from the RBC, and applying the discount factor (depending on the scenario). The TAC is also constrained so as not to change by more than $50 \%$ from one year to the next. Implementation error, such that the catch taken in the next year from the operating model may not be equal to the TAC, is applied using estimates of the relationship between TAC and catch obtained from data on SESSF species (L. Richard Little, unpublished data).

### 12.3.5 Application of discount factor

Discount factors were applied to the TAC by reducing the prescribed TAC by a fixed percentage in each year that the harvest strategies were applied. Values currently applied in the SESSF are $5 \%$ for T3, and $15 \%$ for T4. In addition, simulations also considered discount factor values ranging from 0 (i.e. no discount factor) to $50 \%$.

The resulting estimates of catch were then input into the operating model based on the current proportional allocation by fishing fleet (averaged over the final 5 years of the stock assessment; Table 12.4).

### 12.3.6 Scenarios for implementing precaution

Four scenarios for implementing precaution were tested:

1. Discount factor. TACs were reduced by a fixed percentage (Section 3.5).
2. CPUE stability. This scenario included a discount factor, but also includes a rule where the discount factor was not applied if catch rates have historically been stable. The discount factor (risk premium) is not applied in a given year if the CV of the cpue data over a time period is less than $20 \%$. The time period examined was set at the estimated mean generation time for the species (Table 12.3), the values for which were obtained using the base-case life history parameters in the most recent stock assessment.
3. Conservative target. The target reference point was adjusted. No discount factor applied. For T3, the target fishing mortality was adjusted to different values of ESD (e.g. $F_{60}$ rather than $F_{48}$ ). For T4, the target cpue was increased by a fixed percentage (ranges from $101 \%$ to $150 \%$ ).
4. Estimation error. A more conservative value for the stock indicator was used in the HCR when calculating the RBC, based on the estimation error of the stock indicator. No discount factor was applied. A fixed percentage of the cumulative error distribution of the stock indicator was used, with this ordinarily being $50 \%$ such that values less than $50 \%$ are more conservative (e.g. Figure 12.2). For T3, the error associated with the estimate of the current rate of fishing mortality is calculated using asymptotic methods by the estimation procedure (e.g. Wayte and Klaer 2010, Fay et al. 2011). For T4, the error distribution is obtained by calculating the standard error of the mean current cpue (Figure 12.2).

### 12.3.7 Performance measures

Performance was primarily assessed by using the simulation results to calculate the perceived risk to the stock. Risk was defined according to the harvest strategy policy as:

- the probability of the stock declining below the estimated limit biomass reference point more than $10 \%$ of the time (in this case estimated over the thirty year projection period).

Note that the performance metric is the perceived probability of going below the limit based on the results of stock assessments, not the true probability according to the dynamics of the operating model.

A set of additional performance measures were calculated, which included the mean and variability of estimated and true final stock biomass, the magnitudes of TAC and catch obtained under the harvest strategy, and the true value for the risk to the stock. The true risk was calculated in the same way as the perceived risk, except that the actual stock status from the operating model was used rather than the estimated stock status.

### 12.4 Results

### 12.4.1 Tier 3 and Tier 4 performance with current discount factors

Differences in the MSE performance when the current discount factors (5\% T3 and $15 \% \mathrm{~T} 4)$ were applied compared to when these factors were not applied were primarily associated with the risk to the stock. There was no apparent difference in summaries of the distribution of relative spawning biomass (perceived or true) under T3 when the 5\% discount factor was applied (Figure 12.3). TAC variability was however constrained with the $5 \%$ discount factor compared to no discount factor, particularly with tiger flathead (Figure 12.3, 3f, 3i). Under T4, both perceived and true relative spawning biomasses were higher with the $15 \%$ discount factor than when no discount factor was applied (Figure 12.4, Figure 12.5). For tiger flathead under T4, the average TAC in the first seven years of harvest strategy application was lower with the $15 \%$ discount, but this difference in the average TAC was minimal towards the end of the projected period (Figure 12.4 c ). True risk for tiger flathead was slightly higher than perceived risk under T 4 , but lower than that estimated under T 3 (Table 12.5).

### 12.4.2 Tiger flathead

The perceived risk that the management objective would not be met (i.e. the probability that the stock would go below the limit reference point more than $10 \%$ of the time) under T1 was very low ( $<1 \%$, Table 12.5). Perceived risks under T3 and T4 were high and did not approach the T 1 value even with very high discount factors (Figure 12.5).

For the conservative target scenario, the T4 CPUE target needed to be just $2 \%$ larger to meet the same risk as implementing a $15 \%$ discount factor (Table 12.6). When accounting for uncertainty in the estimate of the T 4 multiplier on catch, the value associated with the 42 nd percentile of the error distribution around the multiplier gave the same level of risk as a $15 \%$ discount factor (Figure 12.5; Table 12.6)

When stability in catch rates was used to determine whether or not to apply a discount factor to T 4 results, a discount factor of only $6 \%$ was needed to match the perceived
risk obtained when using a $15 \%$ discount factor every year (Table 12.6). However, the perceived risk using just a discount factor of $6 \%$ was the same as that for a $15 \%$ discount factor (Figure 12.6), meaning that the catch rate stability scenario was not actually precautionary. Average TACs were $2,798 \mathrm{t}$ in the stable catch rate scenario, compared to $2,768 t$ when the $15 \%$ discount factor was applied every year.

T3 had higher perceived (and true) risks to the stock than T4 for flathead (Table 12.6, Figure 12.5). In general, similar performance (in terms of perceived risk) could be achieved either by use of the current $5 \%$ discount factor, or using a slightly more conservative target ( $\mathrm{F}_{51}$ vs $\mathrm{F}_{48}$, Table 12.6). Under the assessment uncertainty scenario, the average TAC was lower for the percentile of estimate consistent with the same perceived risk than when applying the $5 \%$ discount factor (Table 12.6).

Figure 12.5 shows that the perceived risk under T3 was much more sensitive to the different scenarios than that under T4. The catch rate stability scenario was not conservative, as the same (or higher) discount factors were required in this scenario to achieve the same perceived risk as when just using a discount factor every year (Figure 12.5a), the catch rate stability line has the same or increased risk for a given $\%$ discount factor. If CPUE stability was more conservative than using the discount factor alone, then the risk should be lower for this scenario for the same $\%$ discount factor).

### 12.4.3 School whiting

Under T1, the perceived risk was zero (Figure 12.6). This was also the case for T3 and T4 without any need for a discount factor (results not shown). Implementation error on average tends to set catches lower than the prescribed TAC, except when the stock is at low levels relative to the target (L. Richard Little, unpublished data). The current (2009) stock status for school whiting is estimated to be above the biomass target, and so the final biomass tends to be above the target. This is exacerbated under T4 by the default reference period (Table 12.3) in fact representing a stock above the $0.48 \mathrm{~B}_{0}$ target (Little et al. 2011).

The simulations were also conducted with no implementation error (i.e. the annual retained catch was equal to the TAC). With no implementation error, the perceived risk to the stock under T1 was still very low (although non-zero, Table 12.5). However, perceived risks under T3 and T4 were below the T1 risk level without need for a discount factor (Table 12.5, Figure 12.6). Average TACs under T3 and T4 were substantially lower than those under T 1 (Table 12.5).

The T3 HCR appears to under-estimate stock status, with the estimated final depletion being around the target (Figure 12.3, Table 12.4), but with the true value for depletion being close to $70 \%$ unfished biomass (Table 12.5). Consequently, the mean average TACs associated with the T3 scenarios were low (Table 12.5). This is unsurprising, as the T3 HCR is not thought to be appropriate to school whiting because of its potential to have a rapidly changing age structure and with few age classes being fully selected by the fishery (Smith et al. 2008, Wayte and Klaer 2010).

### 12.4.4 Jackass morwong

In contrast to tiger flathead, T 4 had higher perceived risk than T3 for jackass morwong (Table 12.5, Figure 12.7). The current T1 assessment for jackass morwong estimates relative spawning biomass to be close to the limit biomass reference point (e.g. Figure
12.3). However, the recent fishing mortality rate for morwong has been low (Wayte 2010). Consequently, T3 over-estimates stock status (in terms of ESD, Figure 12.3). The SESSF $50 \%$ change rule (whereby the TAC cannot change by more than $50 \%$ from year to year) constrains increases in the TAC, and so the perceived risk does not increase as to increase risk (results of simulations that did not include this constraint showed higher levels of risk, results not shown).

It was not possible to obtain risk equivalency of T 4 with T 1 over the entire projection period, even with very large discounts (Figure 12.7). This is because there is a high probability that T4 estimates CPUE to be below the limit (given the reference period CPUE in Table 12.3) at the start of the projection period, particularly given the uncertainty in the target catch rate during the default reference period. This can be partially attributed to the high CV of CPUE during the default reference period (used to determine possible values for the target CPUE), which can result in high (possibly unattainable given the reference period catches) target CPUE values. Despite this, the true risk to the stock associated with a $15 \%$ discount factor under T4 is less than that under T 1 (Table 12.5), even though the perceived risk under T 4 with the $15 \%$ discount is 0.18 (Table 12.5). The Catch rate stability scenario required a $25 \%$ discount factor to obtain risk equivalency with the Discount factor scenario with the discount factor of $15 \%$ applied every year (Table 12.7, Figure 12.7). The T4 reference catch rate only needed to be increased by $5 \%$ to match the risk with a $15 \%$ discount under the conservative target scenario (Table 12.7), whereas the assessment uncertainty scenario needed to be very conservative for risk equivalency, with the $16^{\text {th }}$ percentile of the T4 result needed (Table 12.7).

### 12.5 Discussion

Implementing precaution in fisheries management advice based on uncertain estimates of resource status is clearly warranted. This is perhaps even more important when stocks are assessed using data-poor methods, because these methods frequently make assumptions that can result in biased estimates of stock status, in addition to providing uncertain estimates. Additional uncertainties associated with both the management process, and the methods used to determine the scientific advice, will tend to increase the need for further precautionary advice (Ralston et al. 2011).

The analyses detailed in this paper clearly demonstrate that application of a discount factor can work to buffer additional perceived risk associated with managing under data-poor harvest strategies (T3 and T4) in the SESSF relative to the data-rich harvest strategy (T1). However, for some of the data-rich cases presented here, these discounts were not necessary as perceived risks were already low. The analyses also demonstrated that the values for the discount factor that resulted (where possible) in the same risk equivalency as estimated under T1 varied by species. Furthermore, these values were different from those currently applied in the SESSF. That these values are not identical among stocks is unsurprising as the uncertainty associated with the stock status estimates differs among particular assessments, and the degree of observation error differs among species (Table 12.2). This suggests that not all stocks should be managed with the same level of discount factor to achieve the same outcome. However, choosing appropriate unique values for each stock is problematic. In principle, these values could be based on knowledge of the level of observation error and life history, given that the role of these factors in determining risk can be more formally assessed.

The T3 harvest strategy was shown to have higher perceived risk than T4 for tiger flathead, but the reverse was true for jackass morwong. The precision of T3 estimates of stock status through the MSE projections for jackass morwong compared to flathead (Figure 12.3) was due to the fact that the annual sample sizes for the age data were 200 for morwong compared to 100 for flathead (Table 12.2), and that more years are included in the reference catch calculations for morwong than for flathead, resulting in TACs that respond more slowly to changes in current fishing mortality rate. Conversely, the CPUE data used in T4 were more precise for flathead than for morwong (Table 12.2). Conducting the simulations with different levels of observation error (sample size) would enable exploration of the effects of such factors. Data-limited species would likely have lower sample sizes.

Critical assumptions for T4 include that the reference catch and target CPUE be chosen appropriately, and that a linear relationship exists between CPUE and biomass. We attempted to capture uncertainty in the choice of reference values for T4 by sampling from the distribution of the historical data during the 'default' reference period (Table 12.3). This increased uncertainty did not however translate into higher risk, either because the historical data were not variable (tiger flathead) or the combination of reference catch and CPUE were associated with spawning biomass above target levels (school whiting). No accounting was made in the analyses for possible failure in the assumption of linearity in the relationship between CPUE and biomass. It can be expected that, if the true relationship is non-linear and T4 is applied assuming linearity, then additional precaution would be required to maintain the same perceived risk to the stock. Including a non-linear relationship between CPUE and biomass is recommended as an important next step for this work. This will require incorporating estimates of non-linearity into the operating model, and recalculating the T 1 scenario, as this will also be sensitive to this assumption. In theory, T1 should be more robust to this assumption as the associated stock assessments also include additional data streams (e.g. length and age composition, Table 12.1) that T 4 does not.

Alternative methods of implementing precaution were capable of producing similar results as applying the discount factor, with performance measures for biomass and catch approximating those for the discount factor scenario following risk equalization. Using stability in catch rates (at least in the manner defined in these analyses) as rationale for not applying a discount factor does not seem to be an adequate approach for implementing precaution. Either larger or the same level of discount factors were required in concert with this approach to obtain the same perceived risk as when applying discount factors every year (e.g. Figure 12.5a). The conservative target and assessment uncertainty scenarios both were able to match risks with the discount factor scenario. However, as with the values for the discount factor, the values for the adjustments needed for these scenarios varied by species.

The time period used to evaluate the performance of the harvest strategy can be important; stocks that are initially at lower stock levels will tend to have higher probabilities of going below the limit biomass purely due to variability in stock dynamics. Indeed, the perceived risk obtained under T4 for jackass morwong was largely due to the stock initially being at low levels relative to the (variable) target. Perceived risk would be reduced if the initial years were not included in the risk calculation.

The analyses considered risk equivalency in terms of the perceived risk to the stock, as this is based on what is observed from the resource, and is closest to that used to determine if management objectives are met. Alternatively, the scenarios could be tuned to the true risk to the stock from the operating model. The results in Table 12.5Table 12.7 suggest that the true risk in the operating model was generally lower than the perceived risk, across species and harvest strategies, although this was not the case for T3 for jackass morwong and T4 for tiger flathead. Risk was defined in terms of the probability of the estimate of the stock not meeting the limit biomass-related objective of Australia's harvest strategy policy. It is important to note that this is not the only risk associated with managing a fishery. Risk may also need to be managed with respect to catch (e.g. Sethi 2010), such as the probability that TACs will be low, and/or undesirableexcessive variability in the prescribed catch levels. For example, Figure 12.3 c and Figure 12.4 c show that low TACs were prescribed frequently during the projections for tiger flathead. Managing for risk associated with a quantity other than the limit biomass objective will result in different values for the discount factors (or chosen precautionary method). Figure 8 shows the values of the discount factor for tiger flathead required for the different scenarios to obtain the same (or lower) mean average TAC as under T1. T4 would require either a discount factor of $12 \%(23 \%$ under the catch rate stability scenario), an increase in the target of $7 \%$, or using the $20^{\text {th }}$ percentile of the assessment uncertainty distribution (Figure 12.8) lower panels). The results can also be viewed in terms of the estimates for relative spawning biomass at the end of the projection period, namely, what discount factor is required to result in the same average stock status (Figure 12.9). In both cases (for tiger flathead), the discount factor required in the catch rate stability scenario results in greater risks to achieving the particular management objective than solely applying a discount factor.

Management agencies may augment HCRs with additional 'meta-rules' that serve to constrain catches to reduce the variability of advice (e.g. catch quota), or otherwise incorporate additional information. One such rule in the SESSF specifies that the TAC for a species cannot change by more than $50 \%$ from one year to the next. Including this ' $50 \%$ change rule' in the analyses led to changes in the performance of the harvest strategies (particularly for T3). This rule (included in the analyses presented in this paper) resulted in reduced perceived risk for T3, a result that is compatible with findings of Wayte and Klaer (2010) with regard to the performance of this harvest strategy. Other meta-rules in place in the SESSF such as a CPUE multiplier rule and a minimum required TAC change of greater than $10 \%$ were not included.

Implementation of discount factors for SESSF species/stocks managed under T3 and T4 has been controversial. Arguments for not using discount factors have included stability in catch rates and the presence of closed areas. The analyses in this paper demonstrate that catch rate stability cannot be expected to increase precaution, even when there is a linear relationship between CPUE and abundance. The presence of closed areas being used as an uncertainty buffer was not addressed in this paper. Arguments for using the presence of areas closed to fishing rely on the closed area being 'on the table' as part of the fishable stock, and when assessment methods do not account for these closures when calculating indicators of stock status (for example, when using solely fishery dependent data that by necessity, is from areas open to fishing). The presence of closed areas has the potential to complicate stock assessment and the reliability of stock status indicators, and the effect on management outcomes is not necessarily straightforward (e.g. Field et al. 2006).

This chapter shows that it is possible to achieve the same end (i.e. same perceived risk to the stock) with multiple methods for accounting for uncertainty when determining TACs. However, using a conservative target, or specifying a percentile of the estimation error may be more attractive to stakeholders as accounting for uncertainty is then part of the RBC calculation, removing the potential for obtaining a high (or low) result from the HCR, only to then further take something away (via the discount factor). Using the uncertainty of the estimate is intuitive as it implies that the impact of the uncertainty on management advice can be lessened through more intensive monitoring. However this method only considers the error associated with the estimation of the stock indicator given the method used, and does not easily accommodate the uncertainty associated with assumptions related to the chosen assessment method or harvest strategy. Ralston et al. (2011) show that often only a fraction of the uncertainty can be attributed to the error quantified within stock assessments. This was clearly shown for T 4 , in which the harvest strategy is fairly precise, although with potential for large bias and subsequent deviation from desired target stock state if the values chosen for the reference period are incorrect / poorly chosen. However the need to accommodate for 'known unknowns' also applies to the scope of our analyses, which limited the quantification of operating model uncertainty to the parametric uncertainty associated with the current base-case stock assessment for each species. In this respect, simple methods (such as the discount factor or conservative target) that are able to account for additional uncertainty, may in fact be more easily implemented.

We focused on obtaining risk equivalency with T 1 , the data-rich harvest strategy. However, a perhaps more relevant assessment would be to determine the values in the different scenarios needed that meet the harvest strategy objectives. Unfortunately, the harvest policy does not prescribe the acceptable tolerance level for the risk objective. That is, there is no definition for an acceptable level of failing to meet the limit-based objective. Explicit management for risk must be based with respect to some level that is deemed appropriate. In the US, policy dictates that the risk of exceeding the overfishing level (catch that results in a rate of fishing mortality greater than prescribed under a HCR) must not exceed 0.5 . However, individual regional management councils have had to determine how best to implement this when setting catch limits (e.g. Punt et al. 2012). Indeed, although T 1 is the most information-rich case, it must also carry some risk. Our analyses provide a means for comparing risk among harvest strategies, and can be used to identify tradeoffs related to biomass and catch-based fishery objectives given a specific level of risk. When determining an acceptable level of risk, both the magnitude of risk, and (for the current limit reference point objective) the time period over which to calculate the risk (i.e. over how many years, and which years) need to be defined in order to properly apply the results of analyses such as those described here.

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12.8 Tables

Table 12.1. Specifications for the harvest strategies, and data used in them.

| Harvest strategy | Assessment <br> method | Stock Indicator | Response <br> Indicator | Target reference <br> point | Limit reference <br> point | Data input |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Table 12.2. Observation error for the data generated from the operating model, and key parameters used in the harvest strategies.

| Species | Assessment used to condition operating model | Harvest strategy | Assumed natural mortality rate | Assumed steepness | $\begin{gathered} \text { CPUE } \\ \text { CV } \end{gathered}$ | Discard rate CV | Length composition sample size | Age composition sample size (for T1 this is age-at-length) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Tiger flathead | Klaer 2010 | T1 | 0.27 | 0.62 | 0.21 | 0.27 | 100 | 100 |
|  |  |  |  |  | 0.17 | 0.75 | (retained) |  |
|  |  |  |  |  | 0.37 | 0.75 | 50 (discard) |  |
|  |  | T3 | 0.27 | 0.62 | 0.17 | 0.27 | NA | 100 |
|  |  |  |  |  |  | 0.75 |  |  |
|  |  |  |  |  |  | 0.75 |  |  |
|  |  | T4 | NA | NA | 0.17 | 0.27 | NA | NA |
|  |  |  |  |  |  | 0.75 |  |  |
|  |  |  |  |  |  | 0.75 |  |  |
| School whiting | Day 2009 | T1 | Estimated | 0.75 | 0.26 | 0.25 | 100 | 200 |
|  |  |  |  |  |  |  | (retained) |  |
|  |  |  |  |  |  |  | 100 (discard) |  |
|  |  | T3 | 0.6 | 0.75 | 0.26 | 0.25 |  | 200 |
|  |  | T4 | NA | NA | 0.26 | 0.25 | NA | NA |
| Jackass morwong | Wayte 2010 | T1 | 0.15 | 0.7 | $0.20$ | 0.25 |  | 200 |
|  |  |  |  |  | 0.34 |  | (retained) |  |
|  |  |  |  |  |  |  | 100 (discard) |  |
|  |  | T3 | 0.15 | 0.7 | 0.20 | 0.25 | NA | 200 |
|  |  | T4 | NA | NA | 0.20 | 0.25 | NA | NA |

Table 12.3. Specifications for the target reference period catches and CPUE for the Tier 4 HCR (T4). In each simulation, a reference catch and target CPUE are randomly drawn from normal distributions with the given mean and standard deviation (sd).

| Species | Reference period |  | Reference catch |  | Reference CPUE |  | Mean | Yr 1 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | sd | mean | sd | generation <br> time (yrs) |  |  |  |
| Tiger <br> flathead | 1986 | 1995 | 2871 | 458 | 1.06 | 0.217 | 9 |  |
| School <br> whiting | 1986 | 1995 | 1896 | 382 | 1.24 | 0.270 | 5 |  |
| Jackass <br> morwong | 1986 | 1995 | 1377 | 365 | 1.52 | 0.406 | 12 |  |

Table 12.4 : Proportional allocation of catches to fleets during implementation of the harvest strategies.

| Species | Fleet | Proportion annual catch |
| :--- | :--- | :--- |
| Tiger flathead | Danish seine | 0.466 |
|  | Eastern trawl | 0.478 |
|  | Tasmanian trawl | 0.056 |
| School whiting | Danish seine | 0.324 |
|  | Trawl | 0.676 |
| Jackass morwong | Eastern trawl | 0.761 |
|  | Danish seine | 0.057 |
|  | Tasmanian trawl | 0.182 |

Table 12.5. Performance measures for the three harvest strategies, and performance of T 3 and T 4 with the currently applied values for the discount factors.

| Species | Harvest strategy | Discount factor | Perceived risk | True risk in operating model | Median estimated final depletion | CV <br> estimated final depletion | Median true final depletion | $\begin{aligned} & \hline \mathrm{CV} \text { true } \\ & \text { final } \\ & \text { depletion } \end{aligned}$ | Mean ave TAC | Mean CV of annual TACs |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Tiger flathead | T1 | 0 | 0.005 | 0.01 | 0.55 | 0.15 | 0.54 | 0.21 | 2,825 | 0.13 |
|  | T3 | 0 | 0.504 | 0.34 | 0.50 | 0.74 | 0.45 | 0.63 | 2,207 | 1.01 |
|  | T3 | 5\% | 0.473 | 0.32 | 0.50 | 0.70 | 0.47 | 0.60 | 2,151 | 0.99 |
|  | T4 | 0 | 0.127 | 0.18 | 0.49 | 0.44 | 0.43 | 0.46 | 2,960 | 0.66 |
|  | T4 | 15\% | 0.124 | 0.15 | 0.53 | 0.44 | 0.46 | 0.45 | 2,774 | 0.64 |
| School whiting | T1 | 0 | 0.003 | 0.00 | 0.47 | 0.29 | 0.44 | 0.29 | 1,940 | 0.37 |
|  | T3 | 0 | 0.001 | 0.00 | 0.47 | 0.28 | 0.67 | 0.26 | 1,145 | 0.41 |
|  | T3 | 5\% | 0.000 | 0.00 | 0.49 | 0.28 | 0.68 | 0.25 | 1,069 | 0.40 |
|  | T4 | 0 | 0.000 | 0.00 | 0.44 | 0.28 | 0.51 | 0.29 | 1,691 | 0.38 |
|  | T4 | 15\% | 0.000 | 0.00 | 0.46 | 0.27 | 0.54 | 0.28 | 1,552 | 0.37 |
| Jackass morwong | T1 | 0 | 0.006 | 0.07 | 0.46 | 0.10 | 0.45 | 0.32 | 780 | 0.30 |
|  | T3 | 0 | 0.005 | 0.05 | 0.63 | 0.21 | 0.66 | 0.28 | 564 | 0.35 |
|  | T3 | 5\% | 0.002 | 0.05 | 0.65 | 0.20 | 0.67 | 0.28 | 532 | 0.34 |
|  | T4 | 0 | 0.191 | 0.09 | 0.42 | 0.28 | 0.42 | 0.33 | 964 | 0.55 |
|  | T4 | 15\% | 0.179 | 0.05 | 0.45 | 0.27 | 0.45 | 0.31 | 900 | 0.54 |

Table 12.6. Performance measures for optimized results for tiger flathead with implementation error. It was not possible to achieve risk equivalency with T1 (
Figure 12.5). Entries in 'Value needed for risk equivalency' are the values for the adjusting factors (e.g. discount factor, conservative target) needed to obtain the same risk as for T3 and T4 with current SESSF discount factors. Remaining columns show performance statistics associated with the optimized HCR parameters.

| Harvest strategy / Scenario | Value needed for risk equivalenc y | Perceived risk | True risk in operating model | Median estimated final depletion | CV of estimated final depletion | Median true final depletion | CV of <br> true final depletio n |  | Mean CV of TACs |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| T1 | NA | 0.005 | 0.013 | 0.55 | 0.15 | 0.54 | 0.21 | 2,825 | 0.13 |
| T3 |  |  |  |  |  |  |  |  |  |
| Discount factor | 5\% | 0.473 | 0.32 | 0.50 | 0.70 | 0.47 | 0.60 | 2,151 | 0.99 |
| Catch rate stability | 5\% | 0.464 | 0.33 | 0.49 | 0.77 | 0.45 | 0.64 | 2,192 | 1.00 |
| Conservative target | $F_{51}$ | 0.471 | 0.32 | 0.50 | 0.73 | 0.45 | 0.62 | 2,176 | 1.00 |
| Assessment uncertainty | 0.54 | 0.448 | 0.31 | 0.51 | 0.72 | 0.48 | 0.61 | 2,095 | 0.97 |
| T4 |  |  |  |  |  |  |  |  |  |
| Discount factor | 15\% | 0.124 | 0.15 | 0.53 | 0.44 | 0.46 | 0.45 | 2,774 | 0.64 |
| Catch rate stability | 6\% | 0.124 | 0.17 | 0.50 | 0.44 | 0.43 | 0.46 | 2,927 | 0.66 |
| Conservative target | +2\% | 0.124 | 0.17 | 0.50 | 0.44 | 0.43 | 0.46 | 2,917 | 0.64 |
| Assessment uncertainty | 0.42 | 0.124 | 0.17 | 0.50 | 0.44 | 0.43 | 0.46 | 2,929 | 0.68 |

Table 12.7. Performance measures for optimized results for jackass morwong.

| Scenario | Value needed for risk equivalenc y | Perceived risk | True risk in operating model | Median estimated final depletion | CV of estimated final depletion | Median True final depletion | CV of <br> True final depletio n | Mean <br> Averag e TAC | Mean CV of TACs |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| T1 | NA | 0.007 | 0.075 | 0.46 | 0.10 | 0.45 | 0.32 | 766 | 0.30 |
| T3 |  |  |  |  |  |  |  |  |  |
| Discount factor | 5\% | 0.002 | 0.05 | 0.65 | 0.20 | 0.67 | 0.28 | 532 | 0.34 |
| Catch rate stability | 7\% | 0.002 | 0.04 | 0.66 | 0.20 | 0.68 | 0.28 | 523 | 0.35 |
| Conservative target | $\mathrm{F}_{52}$ | 0.001 | 0.04 | 0.66 | 0.19 | 0.68 | 0.27 | 514 | 0.34 |
| Assessment uncertainty | 0.52 | 0.002 | 0.05 | 0.64 | 0.20 | 0.66 | 0.28 | 559 | 0.34 |
| T4 |  |  |  |  |  |  |  |  |  |
| Discount factor | 15\% | 0.179 | 0.05 | 0.45 | 0.27 | 0.45 | 0.31 | 900 | 0.54 |
| Catch rate stability | 25\% | 0.179 | 0.04 | 0.46 | 0.27 | 0.45 | 0.31 | 880 | 0.55 |
| Conservative target | +5\% | 0.177 | 0.06 | 0.42 | 0.27 | 0.43 | 0.32 | 930 | 0.54 |
| Assessment uncertainty | 0.16 | 0.179 | 0.03 | 0.46 | 0.26 | 0.45 | 0.32 | 914 | 0.60 |

### 12.9 Figures



Figure 12.1. Example of risk equivalency among harvest control rules. All three control rules provide the same point estimate of stock status (black points), but with increasing level of uncertainty (vertical error bars). To obtain risk equivalency with $\mathrm{T} 1, \mathrm{~T} 3$ and T 4 must be conservative (gray points/bars) so that the probability of being below some level is the same as that for T 1 (grey dashed horizontal line).


Figure 12.2. Assessment uncertainty scenario: example for Tier 4 HCR. Rather than use point estimate of current CPUE to determine status relative to reference CPUE, some percentile (here the $5^{\text {th }}$ ) of the error distribution (based on CV of current mean CPUE) is used.

## perceived status

true stock status
TAC
Tiger flathead







Jackass morwong




Figure 12.3. T3 summaries of perceived status (equilibrium stock depletion), true stock status (relative spawning biomass), and Total Allowable Catch (TAC, t ) when no discount factor is applied (solid line [median], shading [inter-quartile range], dotted lines [central $95 \%$ interval]), and when a $5 \%$ discount factor is applied (boxplots).


Figure 12.4. T4 summaries of perceived status (relative to cpue target), true stock status (relative spawning biomass), and Total Allowable Catch (TAC, t ) when no discount factor is applied (solid line [median], shading [inter-quartile range], dotted lines [central $95 \%$ interval]), and when a $15 \%$ discount factor is applied (boxplots).


Figure 12.5. Perceived risk for tiger flathead for the four scenarios. Results of CPUE stability are shown in grey in panels (a) and (d). Dashed grey horizontal line represents perceived risk under T1. Black points indicate values for perceived risk given no discount factor. Blue points indicate perceived risk given currently applied values for the discount factor.


Figure 12.6. Perceived risk for school whiting for the four scenarios. Results of CPUE stability are shown in grey in panels (a) and (d). Dashed horizontal line represents perceived risk under T1. Black points indicate values for perceived risk given no discount factor. Blue points indicate perceived risk given currently applied values for the discount factor.


Figure 12.7. Perceived risk for jackass morwong for the four scenarios. Results of CPUE stability are shown in grey in panels (a) and (d). Dashed horizontal line represents perceived risk under T1. Black points indicate values for perceived risk given no discount factor. Blue points indicate perceived risk given currently applied values for the discount factor.

## Tiger flathead



Figure 12.8. Mean average TACs for tiger flathead. Results for T3 and T4 are compared to mean avg TAC obtained under T1 (dashed line). 'discount only’ corresponds to discount factor scenario,' stable CPUE' corresponds to CPUE stability scenario.

## Tiger flathead

Discount Conservative Target Assessment uncertainty


Figure 12.9. Median estimated final depletion for tiger flathead. Results for T3 and T4 are compared to the median estimated final depletion under T1 (dashed line). Scenarios in left-hand panels as for Figure 12.8.

# 13. Catch Rate Standardizations for Selected Species from the SESSF (data 1986 - 2011) 

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### 13.1 Summary

Catch-per-unit-effort (CPUE) data is an important input to many of the stock assessments conducted within the South East and Southern Shark Fishery (SESSF), where it is used as an index of relative abundance through time. The catch and effort log-book data from the SESSF, which is the source of CPUE data, constitutes shot by shot data derived from a wide range of vessels, areas (zones), months, depths, and fishing gears. The catch rates used in the assessments are standardized to reduce the effects of factors such as which vessel fished, where and when fishing occurred, what gear was used, at what depths fishing was conducted, and whether fishing occurred during the day or night. The intent is to focus on any changes in catch rates that occurred between years as a result of changes in stock size rather than changes that occur in any of these other factors. This intent is not always realized when there are unknown influential factors or factors for which we have no data so interpretation of the catch rate trends should not necessarily be taken at face value. This is especially the case when there have been major management changes, such as the introduction of quotas or the more recent structural adjustment. Such large events can greatly influence fishing behaviour, which in turn influences catch rates. Because these changes affected the whole fleet at the same time it is not possible to standardize for their effects.

Catch rates, generally as kilograms per hour fished (though sometimes as catch per shot e.g. Danish Seine, or non-trawl methods), were natural log-transformed to normalize the data and stabilize the variance before standardization. A General Linear Model was used rather than using a Generalized Linear Model with a log-link. This relatively simple analytical approach means that the exact same methods can be applied to all species/stock combinations in a relatively robust manner. The statistical models were variants on the form: LnCE $=$ Year + Vessel + Month + DepthCategory + Zone + Daynight. For some fisheries weeknumber or gear type was also included. In addition, there were interaction terms which could sometimes be fitted, such as Month:Zone or Month:DepthCategory. The data from all vessels reporting catches of a species were included although a preliminary data selection was made on a given depth range for each species for the zones of interest to focus attention on those depths contributing significantly to the fishery for each assumed stock and to reduce the number of empty categories within the statistical models.

This document reports the statistical standardization of the commercial catch and effort data for 19 species, distributed across 44 different combinations of stocks and fisheries ready for inclusion in the annual round of stock assessments. These included School Whiting, Eastern Gemfish, Jackass Morwong, Flathead, Redfish, Silver Trevally, Royal Red Prawn, Blue Eye, Blue Grenadier, Spotted/Silver Warehou, Blue Warehou, Pink Ling, Western Gemfish, Ocean Perch, John Dory, Mirror Dory, Ribaldo, and Ocean Jackets. The statistical package R was used, with especial use being made of the biglm library, which was necessary because of the large amount of data available for some
species. Despite the large numbers of observations available in most analyses, the use of the AIC was able to discriminate between the more complex models. In fact, the visual difference between the CPUE trends exhibited by the top few models tends to be only minor.

Summary graphs are provided across all species (Figure 13.2 and Figure 13.3), as well as more detailed information for each stock. Out of 36 stocks there were 10 whose catch rates have increased over the last 10 years, there were 13 stocks where catch rates were stable (two of which were stable and low; Blue Warehou 102030 and Jackass Morwong 30 ), and there were 7 stocks whose catch rates have declined over the last 10 years. Many of the species are also examined for trends in catches and geometric catch rates between zones; this was to provide a check that there were only minor year x zone interactions (differences in catch rate trends between zones).

### 13.2 Introduction

Commercial catch and effort data are used in in very many fishery stock assessments in Australia as an index of relative abundance through time. The assumption is made that there is a direct relationship between catch rates and the amount of exploitable biomass. However, many factors can influence catch rates, including who was fishing with what gear in what depth, in what season, in what area, and whether it was day or night (plus other factors). The use of catch rates as an index of relative abundance means that it would be best to remove the effects of variation due to changes in these other factors on the assumption that what remains will provide a better estimate of the dynamics of the underlying stock biomass. This process of adjusting the time series for the effects of other influential factors is known as standardization and the accepted way of doing this is to use some statistical modelling procedure that focuses attention onto the annual average catch rates adjusted for the variation in the averages brought about by all the other factors identified.

The diversity of species and methods in the SESSF fishery means that each fishery/stock for which standardized catch rates are required entails its own set of conditions and selection of data. The Resource Assessment Groups (RAGs) have direct input on what combinations of depths and area need to be used in the standardization of each species/stock.

### 13.2.1 The Limits of Standardization

The assumption behind using commercial catch rates in stock assessments is that they reflect the relative abundance of the exploitable biomass through time. The legitimacy behind using commercial catch rates can be questioned when there are factors significantly influencing catch rates which cannot be included in any standardization. Over the last two decades there have been a number of major management interventions in the South East Scalefish and Shark Fishery (SESSF) including the introduction of the quota management system in 1992 and that of the Harvest Strategy Policy (HSP) and associated structural adjustment in 2005 - 2007. The combination of limited quotas and the HSP is now controlling catches in such a way that many fishers have been altering their fishing behaviour to take into account the availability of quota and their own access to quota needed to land the species taken in the mixed species SESSF.

Some stocks, such as flathead, are near or around their target stock size and catch rates are at historically good levels. As a result of this success, some fishers report having to avoid catching species, such as flathead, so as to avoid having to discard and to stay within the bounds of their own quota holdings. Such influences on catch rates tend to bias the catch rates downwards, or at very least add noise to any CPUE signal, which could lead to misinformation passing to any assessment. Currently, there is no way to handle this issue but care needs to be taken not to provide incorrectly conservative advice or inappropriately high catch targets. Included in the management changes is the on-going introduction of numerous area closures imposed for a range of different reasons.

Another example of catch rates not necessarily reflecting the stock dynamics can be found with BlueEye Autoline catch rates. Some of the closures (e.g. the gulper closures north east of Flinders Island) cover areas where auto-line catch rates were previously relatively high. Fishing continues mostly along the western edge of the St Helens Hill closure (even though this closure is open to Autoline vessels but the catch rates on the periphery are only about $2 / 3$ the catch rates on the St Helens Hill itself. The geographical scale of these changes is much finer than that already included in the analyses and so the impression gained is that catch rates in general have declined whereas this may be much more about exactly where the fishing is occurring than what the stock is doing. A FRDC funded research project has only recently begun to examine the influence of closures on stock assessments and this exploration is on-going. The preliminary findings findings, however, indicate that again, great care needs to be taken when trying to interpret the outcomes of the catch rate standardization.

### 13.3 Methods

### 13.3.1 Catch Rate Standardization

### 13.3.1.1 Preliminary Data Selection

The precise methods used when standardizing commercial catch and effort data in the SESSF continue to be discussed in the Commonwealth stock assessment RAGs. This discussion continues because the catch rate time series are very influential in many of the assessments. Previously, various filters were placed on the data available in a preliminary attempt to focus on those vessels that actively target a species. These data filters involved only using vessels that had taken the species for more than two years and those that had taken some minimum annual catch level. The objective of these selections was to remove noise from whatever signal was present in the available data. After examining the effects of these data selections they appear to have only very minor influences on the catch rate trends because the number of records involved was only minor (often differences were not apparent in the graphs, i.e. less effect than the thickness of the lines) and so such selections are again not used this year. Far more influential were restrictions based upon depth of operation. In recognition that there are records which report activity in unlikely depths, there are usually restrictions placed on the depth range from which records could be validly reported. This is necessary as depth tends to be one of the most influential factors used in the statistical standardizations and rare outlying depths only served to confuse the analysis by introducing many combinations of factors that contained no data. In addition the choice of which particular reporting zones or areas are to be examined also leads to a prior selection of data.

Briefly, initial data selection for a particular species consists of using those data relating to a specific fishery (e.g. SET, GHT, GAB, etc), those data within a specified depth range and taken with a specified method in specified statistical zones within the years specified for the analysis.

The graphical representation of results includes the depiction of the unstandardized geometric mean catch rate along with the optimum statistical model representing the standardized time series. This provides a visual indication of whether the standardization changes any trend away from the nominal catch rate. To avoid visual distortions introduced by scaling the standardization relative to a particular year, the time series have all been scaled relative to the average of each time series of yearly indices, which means that the overall average in each case equates to one; this centres the vertical location of each series but does not change the relative trends through time. In all cases the differences between this year's analysis and last years' were minimal; both are illustrated in the individual stock graphs. In addition, for most analyses there is a graph of the relative contribution made by the different factors considered to the changes in the trend between the geometric mean and the optimum model. The scale of the changes introduced by a factor is not always in the same order as the relative proportion of the variation accounted for by a particular factor.

### 13.3.1.2 General Linear Modelling

In each case, catch rates, generally as kilograms per hour fished (though sometimes as catch per shot e.g. School Whiting caught by Danish Seine), were natural logtransformed. A General Linear Model was used rather than using a Generalized Linear Model with a log-link; this has advantages in terms of normalizing the data while stabilizing the variance, which the Generalized Linear Model approach does not always achieve appropriately (Venables \& Dichmont, 2004). This relatively simple analytical approach means that the exact same methods can be applied to all species in a relatively robust manner. The statistical models were variants on the form: LnCE = Year + Vessel + Month + DepthCategory + Zone + Daynight. For some fisheries weeknumber or gear type was also included. In addition, there were interaction terms which could sometimes be fitted, such as Month:Zone or Month:DepthCategory. Thus, the CPUE, conditioned on positive catches of the species of interest, was statistically modelled with a normal GLM on log-transformed CPUE data:

$$
\begin{equation*}
\operatorname{Ln}\left(C P U E_{i}\right)=\alpha_{0}+\alpha_{1} x_{i, 1}+\alpha_{2} x_{i, 2}+\sum_{j=3}^{N} \alpha_{j} x_{i j}+\varepsilon_{i} \tag{1}
\end{equation*}
$$

where $\operatorname{Ln}\left(C P U E_{i}\right)$ is the natural logarithm of the catch rate (usually $\mathrm{kg} / \mathrm{h}$, but sometimes $\mathrm{kg} / \mathrm{shot}$ ) for the $i$-th shot, $x_{i j}$ are the values of the explanatory variables $j$ for the $i$-th shot and the $\alpha_{j}$ are the coefficients for the $N$ factors $j$ to be estimated ( $\alpha_{0}$ is the intercept, $\alpha_{l}$ is the coefficient for the first factor, etc.).

### 13.3.1.3 The Overall Year Effect

For the lognormal model the expected back-transformed year effect involves a biascorrection to account for the log-normality; this then focuses on the mean of the distribution rather than the median:

$$
\begin{equation*}
C P U E_{t}=e^{\left(\gamma_{t}+\sigma_{t}^{2} / 2\right)} \tag{2}
\end{equation*}
$$

where $\gamma_{\mathrm{t}}$ is the Year coefficient for year $t$ and $\sigma_{t}$ is the standard deviation of the $\log$ transformed data (obtained from the analysis). The year coefficients were all divided by the average of the year coefficients to simplify the visual comparison of catch rate changes:

$$
\begin{equation*}
C E_{t}=\frac{C P U E_{t}}{\left(\sum C P U E_{t}\right) / n} \tag{3}
\end{equation*}
$$

where CPUE $_{t}$ is the yearly coefficients from the standardization, $\left(\Sigma \mathrm{CPUE}_{t}\right) / \mathrm{n}$ is the arithmetic average of the yearly coefficients, $n$ is the number of years of observations, and $\mathrm{CE}_{\mathrm{t}}$ is the final time series of yearly index of relative abundance.


Figure 13.1. A schematic diagram depicting the statistical reporting zones in the SESSF, as used in this document. The GAB fishery is to the west of Zone 50. The main SESSF trawl zones are zones $10-50$. Each zone extends out to the boundary of the EEZ, except for zones 50 and 60, and for zones 92 and 91 , which are bounded by zone 70 .

### 13.3.1.4 Data Manipulations

A standard set of database extracts were designed to identify positive shots containing the species of interest in each case. For each species the analyses were restricted to particular zones and depth ranges within a particular fishery and using a particular method (Table 13.1).

The statistical software $R$ was used in all analyses (R Development Core Team, 2009), which, because of the large size of the datasets, required the use of the library "biglm".

### 13.4 Results

Table 13.1. Data characteristics for each analysis. Records show the number of records, the depths, the zones, and other details used in the data selection for the analyses.

|  | Species | Zone | Depths | Comment | Records |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | School Whiting | 60 | 0-100 | Danish Seine, catch per shot. | 78728 |
| 2 | Eastern Gemfish | 10-30,40/2 | 200-500 | June-Sept 93 onwards, Spawning | 14402 |
| 3 | Eastern Gemfish | 10-30,40/2 | 0-600 | Oct-May 86-09 0-600m, Jun-Sep | 38323 |
| 4 | Jackass Morwong | 10-50 | 70-360 |  | 143899 |
| 5 | Jackass Morwong | 10,20 | 70-300 |  | 110061 |
| 6 | Jackass Morwong | 30 | 70-300 |  | 18557 |
| 7 | Jackass Morwong | 40,50 | 70-360 |  | 12525 |
| 8 | Flathead | 10,20 | 0-400 | Trawl | 245841 |
| 9 | Flathead | 30 | 0-400 |  | 19485 |
| 10 | Flathead DS | 20,60 | 0-200 | Danish Seine, catch per shot | 174461 |
| 11 | RedFish | 10 | 0-400 |  | 69927 |
| 12 | RedFish | 20 | 0-400 |  | 26195 |
| 13 | Silver Trevally | 10,20 | 0-200 | Remove State waters and MPAs | 32456 |
| 14 | Royal Red Prawn | 10 | 200-700 |  | 23586 |
| 15 | Blue Eye | 20,30 | 0-1000 |  | 12048 |
| 16 | Blue Eye | 40,50 | 0-1000 |  | 12317 |
| 17 | Blue Eye | 10-50,83- | 200-600 | Autolining and Droplining 1997 | 13855 |
| 18 | Blue Grenadier | 40 | 100- | Spawning Jun-Aug, Combined big | 12238 |
| 19 | Blue Grenadier | 10-60 | 0-1000 | Except Zone 40 Jun-Aug | 126964 |
| 20 | Silver Warehou | 10-50 | 0-600 |  | 122577 |
| 21 | Blue Warehou | 10-30 | 0-400 |  | 36668 |
| 22 | Blue Warehou | 40,50 | 0-600 |  | 12576 |
| 23 | Blue Warehou | 10-50 | 0-600 |  | 49703 |
| 24 | Pink Ling | 10-30 | 0-600 |  | 148537 |
| 25 | Pink Ling | 40,50 | 200-800 |  | 71064 |
| 26 | Pink Ling | 10 | 250-600 | For use in disaggregated analyses | 43615 |
| 27 | Pink Ling | 20 | 250-600 | " | 42339 |
|  | Pink Ling | 30 | 250-600 | " | 7800 |
| 29 | Pink Ling | 40 | 350-800 | " | 29435 |
| 30 | Pink Ling | 50 | 200-800 | " | 41427 |
| 31 | Western Gemfish | 40,50,GAB | 100-600 |  | 41220 |
| 32 | Western Gemfish | 40,50 | 200-600 |  | 30604 |
| 33 | Western Gemfish | GAB | 100-600 | Only 1995 onwards | 9045 |
| 34 | Off-Ocean Perch | 10,20 | 200-700 |  | 75911 |
| 35 | In-Ocean Perch | 10,20 | 0-200 |  | 16044 |
| 36 | Total Ocean | 10,20 | 0-700 |  | 91409 |
| 37 | John Dory | 10,20 | 0-200 |  | 130809 |
| 38 | Mirror Dory | 10-50 | 0-600 |  | 117422 |
| 39 | Mirror Dory East | 10-30 | 0-600 |  | 88365 |
| 40 | Mirror Dory West | 40,50 | 0-600 |  | 29025 |
| 41 | Ribaldo (RBD) | 10-50 | 0-1000 |  | 19029 |
| 42 | Ribaldo | 20-50,81- | 0-1000 |  | 4249 |
| 43 | Ocean Jackets | 10-50 | 0-300 |  | 75344 |
| 44 | Ocean Jackets | 82-83 | 0-300 |  | 43965 |


|  | BlueEye ALDL | $\overbrace{-4}^{-0.0366}$ <br> BlueEye AL |  <br> BlueEyeDL |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |

(
WGemGAB
STA


Figure 13.2. Summary graph of the optimum standardizations for 19 species and 36 different stocks, each with a linear regression across the last ten years (2002-2011). The gradient is given in the top right corner of each graph and the colour of the standardized CPUE line relates to the gradient: Green indicates a positive gradient $>0.015$, blue a flat line with a gradient between 0.0149 and -0.0149 , while red indicates a negative gradient $<-0.015$. There were 11 stocks with a positive gradient, 14 stocks with a flat gradient, and 11 stocks with a negative gradient. Composite stocks, such as MirrorDory10-50 and TotalOceanPerch are omitted.


Figure 13.3. Summary graph of the optimum standardizations for 19 species and 36 different stocks, each with a linear regression across the last five years (2007-2011). The gradient is given in the top right corner of each graph and the colour of the standardized CPUE line relates to the gradient: Green indicates a positive gradient $>0.015$, blue a flat line with a gradient between 0.0149 and -0.0149 , while red indicates a negative gradient $<-0.015$. There were 9 stocks with a positive gradient, 3 stocks with a flat gradient, and 24 stocks with a negative gradient. In many instances five years provides a false impression, for example, Pink Ling 10-30 appears negative here but positive across ten years. While declines can be serious, responding to noise would just make management variable over short time periods.

### 13.5 School Whiting (WHS - 37330014) Sillago flindersi

School Whiting are taken primarily by Danish Seine (and within State waters). In Commonwealth waters the catches are primarily within Zone 60, and in depths less than or equal to 100 m . All vessels and all records were included in the analysis. Catch rates were expressed as the natural log of catch per shot. There were a total of 78,728 records used.

Table 13.2. School Whiting from Zone 60 in depths 0 to 100 m by Danish Seine. Total Catch is the total reported in the database, Records are the number used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr}$ ). Mth:DepC is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Mth:DepC | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 1302.410 | 5667 | 1181.583 | 26 | 112.3054 | 1.1573 | 0.0000 |
| 1987 | 995.965 | 4119 | 920.495 | 23 | 131.1624 | 1.2735 | 0.0292 |
| 1988 | 1255.688 | 3815 | 1177.456 | 25 | 168.5490 | 1.6496 | 0.0298 |
| 1989 | 1061.513 | 4440 | 994.408 | 27 | 127.0438 | 1.0919 | 0.0288 |
| 1990 | 1930.368 | 6263 | 1859.923 | 24 | 165.2959 | 1.6853 | 0.0269 |
| 1991 | 1630.255 | 4871 | 1517.794 | 26 | 164.1905 | 1.4283 | 0.0285 |
| 1992 | 854.106 | 2980 | 777.524 | 23 | 124.7066 | 1.0180 | 0.0327 |
| 1993 | 1694.896 | 4696 | 1471.559 | 23 | 152.4819 | 1.4500 | 0.0288 |
| 1994 | 946.201 | 4503 | 879.162 | 24 | 93.9314 | 0.8477 | 0.0290 |
| 1995 | 1212.561 | 4270 | 1065.934 | 21 | 122.4731 | 1.0666 | 0.0294 |
| 1996 | 898.213 | 4297 | 718.814 | 22 | 81.4339 | 0.6978 | 0.0296 |
| 1997 | 697.380 | 3314 | 481.660 | 20 | 64.5619 | 0.5411 | 0.0318 |
| 1998 | 594.153 | 2988 | 464.154 | 20 | 66.0158 | 0.5219 | 0.0327 |
| 1999 | 681.252 | 2044 | 452.215 | 21 | 84.3634 | 0.5966 | 0.0375 |
| 2000 | 700.880 | 1913 | 335.075 | 17 | 65.1233 | 0.6006 | 0.0380 |
| 2001 | 890.925 | 1980 | 425.095 | 18 | 93.2089 | 0.8457 | 0.0391 |
| 2002 | 788.331 | 2192 | 429.218 | 20 | 90.8874 | 0.8697 | 0.0374 |
| 2003 | 866.808 | 2352 | 463.528 | 20 | 87.1013 | 0.8898 | 0.0368 |
| 2004 | 604.886 | 1771 | 334.631 | 20 | 79.7648 | 0.8435 | 0.0395 |
| 2005 | 662.684 | 1750 | 311.428 | 20 | 77.2502 | 0.9513 | 0.0412 |
| 2006 | 667.505 | 1428 | 270.272 | 18 | 76.2250 | 0.8220 | 0.0430 |
| 2007 | 535.358 | 1488 | 347.049 | 14 | 89.2381 | 1.0799 | 0.0420 |
| 2008 | 502.245 | 1260 | 317.058 | 15 | 92.3448 | 1.0860 | 0.0451 |
| 2009 | 461.891 | 1569 | 350.723 | 15 | 93.6200 | 1.1350 | 0.0418 |
| 2010 | 409.501 | 1179 | 273.470 | 15 | 88.7190 | 1.0156 | 0.0461 |
| 2011 | 373.911 | 1579 | 260.300 | 14 | 72.0269 | 0.8354 | 0.0415 |



Figure 13.4. School Whiting in zone 60 in depths 0 to 100 m taken by Danish Seine. The top left is the depth distribution of all records reporting School Whiting, the top right graph depicts the depth distribution of shots containing School Whiting in Zone 60 and depths $0-100 \mathrm{~m}$. The middle left diagram depicts the distribution of catch by depth within zone 60 across all years, the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the School Whiting catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.5. School Whiting in zone 60 in depths 0 to 100 m by Danish Seine. The dashed black line represents the geometric mean catch rate, the solid black line the standardized catch rates, and the blue line is last year's analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.3. School Whiting from Zone 60 in depths 0 to 100 m by Danish Seine. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories. DN is DayNight

Model 1 LnCE~Year
Model 2 LnCE~Year+Vessel
Model 3 LnCE~Year+Vessel+DN
Model 4 LnCE $\sim$ Year+Vessel+DN+Month
Model 5 LnCE $\sim$ Year + Vessel + DN + Month + DepCat
Model 6 LnCE~Year+Vessel+DN+Month+DepCat+DN:DepCat
Model 7 LnCE~Year+Vessel+DN+Month+DepCat+DepCat:Month
Model $8 \quad$ LnCE $\sim$ Year+Vessel + DN + Month + DepCat + DN:Month

Table 13.4. School Whiting from Zone 60 in depths 0 to 100 m by Danish Seine. Model selection criteria, including the AIC, the adjusted $\mathrm{r}^{2}$ and the change in adjusted $\mathrm{r}^{2}$. The optimum was model 7. DN is DayNight, DepC is depth category and Mth is Month.

|  | Year | Vessel | DayNight | Month | DepCat | DN:DepC | DepC:Mth | DN:Mth |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 55414 | 53351 | 51068 | 49087 | 47636 | 47460 | 47127 | 47488 |
| RSS | 159048 | 154753 | 150319 | 146543 | 142760 | 142392 | 141662 | 142365 |
| MSS | 7531 | 11826 | 16261 | 20036 | 23819 | 24187 | 24917 | 24214 |
| Nobs | 78728 | 78728 | 78728 | 78728 | 77211 | 77211 | 77211 | 77211 |
| Npars | 26 | 72 | 75 | 86 | 90 | 102 | 134 | 123 |
| adj_r2 | 4.491 | 7.015 | 9.677 | 11.933 | 14.200 | 14.408 | 14.811 | 14.401 |
| \%Change | 0.000 | 2.525 | 2.661 | 2.256 | 2.267 | 0.208 | 0.404 | -0.410 |



Figure 13.6. The relative influence of each factor used on the final trend in the optimal standardization for School Whiting in Zone 60. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.6 Eastern Gemfish (GEM - 37439002 - Rexea solandri) Spawning Fishery

Only use June through September from 1993 - 2010, 300-500m depth, Catch effort > 0.0 , Zones $10-30$ plus below 42 degrees on the west coast of Tasmania (zone 40).

Eastern Gemfish are taken by trawl in the spawning season from June to September in Zones 10, 20, 30, in the bottom half of 40 and between depths of 300 to 500 m . There were 13,270 records used. The spawning run of Eastern Gemfish is considered to be a bycatch fishery. Particular records related to the Eastern Gemfish surveys in 2007 and 2008 are removed from the data set prior to the analysis.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 353.410 | 825 | 143.241 | 50 | 17.7557 | 1.9638 | 0.0000 |
| 1994 | 232.179 | 825 | 49.142 | 48 | 11.8113 | 1.2980 | 0.0616 |
| 1995 | 181.746 | 661 | 21.946 | 50 | 7.3422 | 0.8735 | 0.0650 |
| 1996 | 382.196 | 773 | 135.175 | 50 | 10.8392 | 1.0971 | 0.0626 |
| 1997 | 571.976 | 1239 | 268.914 | 50 | 18.7012 | 1.5816 | 0.0580 |
| 1998 | 404.817 | 887 | 144.689 | 47 | 11.4499 | 1.0604 | 0.0621 |
| 1999 | 448.677 | 1092 | 88.365 | 48 | 8.0762 | 0.8808 | 0.0600 |
| 2000 | 336.464 | 1235 | 37.978 | 48 | 4.6672 | 0.6138 | 0.0604 |
| 2001 | 331.486 | 902 | 34.567 | 50 | 4.4410 | 0.6324 | 0.0638 |
| 2002 | 196.526 | 977 | 22.949 | 47 | 3.3317 | 0.4550 | 0.0632 |
| 2003 | 269.227 | 976 | 31.653 | 49 | 4.4807 | 0.6497 | 0.0624 |
| 2004 | 525.201 | 694 | 20.563 | 50 | 3.9729 | 0.6155 | 0.0688 |
| 2005 | 498.511 | 666 | 20.076 | 40 | 4.4130 | 0.5452 | 0.0684 |
| 2006 | 509.019 | 580 | 35.198 | 36 | 7.5671 | 0.8669 | 0.0709 |
| 2007 | 542.778 | 313 | 25.424 | 21 | 8.7558 | 1.0657 | 0.0858 |
| 2008 | 252.302 | 450 | 35.264 | 25 | 10.2842 | 1.3209 | 0.0782 |
| 2009 | 194.843 | 426 | 37.246 | 26 | 9.0364 | 1.2096 | 0.0789 |
| 2010 | 220.639 | 422 | 42.549 | 26 | 9.6316 | 1.3126 | 0.0791 |
| 2011 | 147.321 | 459 | 28.479 | 26 | 6.8540 | 0.9574 | 0.0771 |



Figure 13.7. Eastern Gemfish, spawning fishery in depths between $300-500 \mathrm{~m}$, taken by trawl. The top left is the depth distribution of all records reporting Eastern Gemfish, the top right graph depicts the depth distribution of shots containing Eastern Gemfish, spawning fishery in depths between $300-500 \mathrm{~m}$, taken by trawl. The middle left diagram depicts the distribution of catch by depth by SESSF zone, the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Eastern Gemfish catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.8. Eastern Gemfish, spawning fishery in depths between $300-500 \mathrm{~m}$, taken by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.6. Eastern Gemfish, spawning fishery in depths between $300-500 \mathrm{~m}$, taken by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel+Month |
| Model 4 | LnCE $\sim$ Year+Vessel+Month +DepCat |
| Model 5 | LnCE $\sim$ Year+Vessel+Month +DepCat +DayNight |
| Model 6 | LnCE $\sim$ Year+Vessel+Month +DepCat +DayNight+Zone |
| Model 7 | LnCE $\sim$ Year+Vessel+Month +DepCat +DayNight+Zone+Zone:Month |
| Model 8 | LnCE $\sim$ Year+Vessel+Month +DepCat +DayNight+Zone+Zone:DepCat |

Table 13.7. Eastern Gemfish, spawning fishery in depths between $300-500 \mathrm{~m}$, taken by trawl. Model selection criteria, including the AIC, the deviance and the change in deviance. The optimum is model Zone:Month.

|  | Year | Vessel | DepCat | Month | DayNight | Zone | Zone:Mth Zone:DepC |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 8690 | 6694 | 5940 | 5436 | 5396 | 5284 | 5026 | 5265 |
| RSS | 26261 | 22532 | 21374 | 20518 | 20452 | 20283 | 19896 | 20172 |
| MSS | 4079 | 7807 | 8965 | 9822 | 9888 | 10056 | 10443 | 10168 |
| Nobs | 14402 | 14402 | 14402 | 14301 | 14301 | 14301 | 14301 | 14301 |
| Npars | 19 | 124 | 127 | 137 | 140 | 143 | 152 | 173 |
| adj_r2 | 13.335 | 25.093 | 28.928 | 31.724 | 31.929 | 32.476 | 33.722 | 32.704 |
| \%Change | 0.000 | 11.759 | 3.835 | 2.796 | 0.206 | 0.546 | 1.246 | -1.018 |



Figure 13.9. The relative influence of each factor used on the final trend in the optimal standardization for the eastern gemfish spawning fishery. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.7 Eastern Gemfish Non-Spawning (GEM - 37439002 - Rexea solandri)

Use October to May 1986-2010, all depths to 600 m , June to September, $<300 \mathrm{~m}$ depth, Zones $10-30$ plus below 42 on the west coast of Tasmania (zone 40).

[^0]| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:DepC | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 3639.955 | 2203 | 414.317 | 89 | 15.3241 | 2.3896 | 0.0000 |
| 1987 | 4660.447 | 2019 | 784.860 | 80 | 25.2674 | 2.9835 | 0.0413 |
| 1988 | 3515.819 | 2372 | 532.525 | 84 | 20.3738 | 2.6781 | 0.0411 |
| 1989 | 1778.325 | 1742 | 189.588 | 78 | 12.8697 | 1.9458 | 0.0444 |
| 1990 | 1206.897 | 858 | 107.120 | 73 | 12.0080 | 1.7081 | 0.0547 |
| 1991 | 580.322 | 863 | 70.785 | 73 | 8.4919 | 1.1499 | 0.0543 |
| 1992 | 494.441 | 717 | 135.384 | 52 | 10.6133 | 1.5743 | 0.0576 |
| 1993 | 353.410 | 1543 | 94.850 | 62 | 8.9852 | 1.2954 | 0.0467 |
| 1994 | 232.179 | 1845 | 69.080 | 56 | 6.2854 | 0.8826 | 0.0448 |
| 1995 | 181.746 | 1714 | 50.937 | 54 | 5.4906 | 0.8009 | 0.0454 |
| 1996 | 382.196 | 1930 | 55.795 | 61 | 3.9966 | 0.5858 | 0.0447 |
| 1997 | 571.976 | 1829 | 67.152 | 57 | 4.1253 | 0.5934 | 0.0464 |
| 1998 | 404.817 | 1304 | 45.991 | 52 | 4.0091 | 0.5742 | 0.0491 |
| 1999 | 448.677 | 1432 | 30.842 | 54 | 2.7336 | 0.4234 | 0.0483 |
| 2000 | 336.464 | 1856 | 33.128 | 60 | 2.5299 | 0.3905 | 0.0462 |
| 2001 | 331.486 | 1680 | 31.424 | 53 | 1.9996 | 0.3369 | 0.0476 |
| 2002 | 196.526 | 1727 | 20.203 | 53 | 1.5421 | 0.2622 | 0.0475 |
| 2003 | 269.227 | 1627 | 20.423 | 51 | 1.6954 | 0.2954 | 0.0479 |
| 2004 | 525.201 | 1825 | 39.127 | 56 | 2.5873 | 0.4142 | 0.0473 |
| 2005 | 498.511 | 1772 | 44.504 | 51 | 2.7875 | 0.4370 | 0.0471 |
| 2006 | 509.019 | 1364 | 34.001 | 43 | 2.8952 | 0.4652 | 0.0501 |
| 2007 | 542.778 | 812 | 24.887 | 27 | 4.0265 | 0.6460 | 0.0570 |
| 2008 | 252.302 | 843 | 34.918 | 27 | 5.5997 | 0.8828 | 0.0567 |
| 2009 | 194.843 | 646 | 29.796 | 31 | 5.4510 | 0.9137 | 0.0620 |
| 2010 | 220.639 | 909 | 28.997 | 27 | 3.8269 | 0.7197 | 0.0559 |
| 2011 | 147.321 | 891 | 24.935 | 26 | 3.5366 | 0.6515 | 0.0566 |



Figure 13.10. Non-spawning Eastern Gemfish from the SET in depths between $0-600 \mathrm{~m}$, taken by trawl. The top left is the depth distribution of all records reporting Eastern Gemfish, the top right graph depicts the depth distribution of shots containing Non-spawning Eastern Gemfish from the SET in depths between $0-600 \mathrm{~m}$, taken by trawl. The middle left diagram depicts the distribution of catch by depth by SESSF zone, the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Eastern Gemfish catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<$ 30 Kg ).


Figure 13.11. Non-spawning Eastern Gemfish from the SET in depths between $0-600 \mathrm{~m}$, taken by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The blue line is last year's optimum standardization. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.9. Non-spawning Eastern Gemfish from the SET in depths between $0-600 \mathrm{~m}$, taken by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel+DepCat |
| Model 4 | LnCE $\sim$ Year+Vessel+DepCat+Month |
| Model 5 | LnCE $\sim$ Year+Vessel+DepCat+Month+Zone |
| Model 6 | LnCE $\sim$ Year+Vessel+DepCat+Month+Zone+DayNight |
| Model 7 | LnCE $\sim$ Year+Vessel+DepCat+Month+Zone+DayNight+Zone:Month |
| Model 8 | LnCE $\sim$ Year+Vessel+DepCat+Month+Zone+DayNight+Zone:DepCat |

Table 13.10. Nonspawning Eastern Gemfish from the SET in depths between $0-600 \mathrm{~m}$, taken by trawl. Model selection criteria, including the AIC, the deviance and the change in deviance. The optimum is model Zone:DepCat.

|  | Year | Vessel | DepCat | Month | Zone |  |  | DayNight |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | Zone:Mth Zone:DepC .



Figure 13.12. The relative influence of each factor used on the final trend in the optimal standardization for Non-spawning Eastern Gemfish. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.8 Jackass Morwong Z10-50 (MOR - 37377003 Nemadactylus macropterus)

Only data from Zones 10 to 50 in depths $70-360 \mathrm{~m}$ taken by trawl.

> Table 13.11. Jackass Morwong from zones 10 to 50 in depths $70-360 \mathrm{~m}$ by trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr} \mathrm{)}. \mathrm{Zone:Mth} \mathrm{is} \mathrm{the} \mathrm{optimum} \mathrm{model} \mathrm{and} \mathrm{StDev} \mathrm{is} \mathrm{the}$ standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 982.811 | 5772 | 873.211 | 106 | 22.5592 | 1.8326 | 0.0000 |
| 1987 | 1087.690 | 4948 | 1000.054 | 104 | 26.1917 | 2.0862 | 0.0265 |
| 1988 | 1483.512 | 5984 | 1314.397 | 102 | 29.1554 | 2.0565 | 0.0258 |
| 1989 | 1667.373 | 5434 | 1500.604 | 89 | 33.9001 | 1.9994 | 0.0266 |
| 1990 | 1001.414 | 5022 | 837.357 | 86 | 24.2137 | 1.6572 | 0.0276 |
| 1991 | 1138.070 | 5233 | 899.685 | 85 | 21.1181 | 1.4684 | 0.0274 |
| 1992 | 758.254 | 3483 | 523.779 | 63 | 19.1937 | 1.2157 | 0.0306 |
| 1993 | 1014.985 | 4732 | 821.881 | 73 | 21.3530 | 1.2320 | 0.0287 |
| 1994 | 818.418 | 5660 | 684.800 | 71 | 18.0744 | 1.0502 | 0.0274 |
| 1995 | 789.528 | 5852 | 705.409 | 63 | 16.3623 | 0.9875 | 0.0271 |
| 1996 | 827.191 | 7535 | 749.574 | 70 | 13.8607 | 0.9246 | 0.0260 |
| 1997 | 1063.363 | 7561 | 934.001 | 70 | 16.1581 | 0.9910 | 0.0264 |
| 1998 | 876.404 | 5941 | 688.705 | 65 | 13.4363 | 0.8476 | 0.0274 |
| 1999 | 961.262 | 5801 | 779.703 | 66 | 14.1587 | 0.8793 | 0.0276 |
| 2000 | 945.098 | 6902 | 732.188 | 77 | 10.1983 | 0.7263 | 0.0269 |
| 2001 | 790.188 | 6786 | 644.178 | 71 | 8.3295 | 0.5431 | 0.0271 |
| 2002 | 811.136 | 7761 | 691.282 | 65 | 8.3275 | 0.5693 | 0.0267 |
| 2003 | 775.123 | 6538 | 601.484 | 64 | 7.9077 | 0.4899 | 0.0273 |
| 2004 | 765.506 | 6483 | 604.476 | 70 | 8.6153 | 0.4905 | 0.0276 |
| 2005 | 784.128 | 6376 | 597.416 | 58 | 8.9785 | 0.5259 | 0.0276 |
| 2006 | 811.298 | 5446 | 616.102 | 49 | 11.5427 | 0.6003 | 0.0285 |
| 2007 | 607.870 | 3812 | 443.366 | 30 | 12.2504 | 0.5987 | 0.0309 |
| 2008 | 700.439 | 4491 | 546.640 | 33 | 13.7889 | 0.7042 | 0.0299 |
| 2009 | 454.352 | 3383 | 344.429 | 27 | 11.4713 | 0.6212 | 0.0318 |
| 2010 | 380.248 | 3438 | 292.104 | 30 | 8.5497 | 0.4635 | 0.0319 |
| 2011 | 422.130 | 3525 | 303.284 | 28 | 8.5254 | 0.4389 | 0.0318 |



Figure 13.13. Jackass Morwong from zones 10 to 50 in depths $70-360 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Jackass Morwong, the top right graph depicts the depth distribution of shots containing Jackass Morwong from zones 10 to 50 in depths $70-360 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zones 10 to 50 , the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Jackass Morwong catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.14. Jackass Morwong from zones 10 to 50 in depths $70-360 \mathrm{~m}$ by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.12. Jackass Morwong from zones 10 to 50 in depths $70-360 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE~Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel+Month |
| Model 4 | LnCE $\sim$ Year+Vessel+Month+Zone |
| Model 5 | LnCE $\sim$ Year+Vessel+Month+Zone+DepCat |
| Model 6 | LnCE $\sim$ Year+Vessel+Month+Zone+DepCat+DayNight |
| Model 7 | LnCE $\sim$ Year+Vessel+Month+Zone+DepCat+DayNight+Zone:Month |
| Model 8 | LnCE $\sim$ Year+Vessel+Month+Zone+DepCat+DayNight+Zone:DepCat |

Table 13.13. Jackass Morwong from zones 10 to 50 in depths $70-360 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the adjusted r2 and the change in adjusted r2. The optimum was model Zone:Month.

|  | Year | Vessel | Month | DepCat | Zone |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | DayNight Zone:Month Zone:DepCat



Figure 13.15. The relative influence of each factor used on the final trend in the optimal standardization for Jackass Morwong in Zones $10-50$. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.


Figure 13.16. The trends in catch and geometric mean catch rates for Jackass Morwong taken by trawl across SESSF zones $10-50$. The catch rate trends across zones $10-30$ are very similar, whilst those for zones 40 to 50 are noisy due to low catches until after 1996.


Figure 13.17. The standardized trends for the Month and DepCat factors for Jackass Morwong taken by trawl across SESSF zones $10-50$.


Figure 13.18. The standardized trends for the Month and DepCat factors for Jackass Morwong taken by trawl across SESSF zones $10-20$.


Figure 13.19. The standardized trends for the Month and DepCat factors for Jackass Morwong taken by trawl across SESSF zone 30.



Figure 13.20. The standardized trends for the Month and DepCat factors for Jackass Morwong taken by trawl across SESSF zones 40-50.

Table 13.14. The split of reported catches in tonnes by zone as taken by trawl in the identified depths. GAB includes zones $82,83,84$, and 85 .

| Year | $\mathbf{1 0}$ | $\mathbf{2 0}$ | $\mathbf{3 0}$ | $\mathbf{4 0}$ | $\mathbf{5 0}$ | $\mathbf{6 0}$ | $\mathbf{G A B}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 153.290 | 597.906 | 32.287 | 0.400 | 152.246 | 27.077 | 16.565 |
| 1987 | 142.674 | 770.594 | 80.446 | 13.775 | 46.426 | 19.748 | 12.960 |
| 1988 | 177.971 | 922.634 | 213.955 | 16.700 | 51.072 | 56.980 | 41.625 |
| 1989 | 80.174 | 896.639 | 505.097 | 50.770 | 34.226 | 39.482 | 51.348 |
| 1990 | 82.706 | 606.652 | 158.494 | 14.701 | 68.417 | 22.015 | 45.800 |
| 1991 | 107.642 | 690.990 | 225.715 | 14.382 | 33.105 | 22.191 | 32.935 |
| 1992 | 56.005 | 444.369 | 132.726 | 27.490 | 34.501 | 7.577 | 45.160 |
| 1993 | 104.483 | 431.220 | 344.380 | 4.474 | 21.107 | 20.498 | 46.599 |
| 1994 | 105.480 | 436.446 | 185.204 | 4.641 | 18.665 | 18.064 | 46.813 |
| 1995 | 77.205 | 388.259 | 187.464 | 67.835 | 10.855 | 3.854 | 52.929 |
| 1996 | 97.641 | 475.605 | 162.715 | 10.917 | 27.350 | 6.793 | 45.263 |
| 1997 | 62.813 | 652.029 | 205.295 | 29.995 | 27.213 | 13.946 | 66.733 |
| 1998 | 58.295 | 441.898 | 193.305 | 45.258 | 12.960 | 13.458 | 72.596 |
| 1999 | 44.685 | 445.380 | 249.027 | 64.502 | 16.404 | 8.962 | 102.751 |
| 2000 | 49.760 | 475.166 | 126.249 | 107.740 | 13.703 | 20.428 | 73.115 |
| 2001 | 37.154 | 273.619 | 112.989 | 137.773 | 149.603 | 17.559 | 52.075 |
| 2002 | 76.130 | 291.396 | 110.840 | 98.844 | 156.460 | 15.729 | 48.200 |
| 2003 | 32.855 | 240.440 | 196.687 | 62.151 | 114.646 | 12.053 | 98.563 |
| 2004 | 31.203 | 223.494 | 205.915 | 48.383 | 141.841 | 7.189 | 104.330 |
| 2005 | 37.108 | 288.939 | 151.947 | 36.915 | 162.915 | 8.309 | 96.830 |
| 2006 | 30.714 | 289.117 | 166.045 | 24.665 | 167.622 | 6.735 | 121.021 |
| 2007 | 14.548 | 230.969 | 118.917 | 25.839 | 96.708 | 5.620 | 109.069 |
| 2008 | 38.791 | 327.492 | 122.652 | 29.875 | 74.678 | 6.366 | 91.719 |
| 2009 | 27.405 | 230.783 | 55.928 | 20.819 | 45.113 | 3.843 | 64.330 |
| 2010 | 21.984 | 190.938 | 59.890 | 13.603 | 27.382 | 3.445 | 39.384 |
| 2011 | 17.680 | 184.592 | 51.259 | 35.147 | 51.226 | 11.685 | 24.997 |

### 13.9 Jackass Morwong Z1020 (MOR-37377003 N. macropterus)

Only data from zone 10 and 20 were used for trawl vessels only (i.e. exclude Danish Seine vessels), and depths between 70 and 300 m .

Table 13.15. Jackass Morwong from zones 10 and 20 in depths $70-300 \mathrm{~m}$ by trawl. Total Catch is the total reported in the database, Records is the number of reported records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr}$ ). Zone:Mth is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 982.811 | 5045 | 686.225 | 87 | 21.2677 | 1.8013 | 0.0000 |
| 1987 | 1087.690 | 4266 | 858.475 | 79 | 26.2295 | 2.1904 | 0.0291 |
| 1988 | 1483.512 | 5147 | 1025.256 | 79 | 27.6740 | 2.0603 | 0.0283 |
| 1989 | 1667.373 | 4325 | 929.409 | 65 | 27.9306 | 1.9409 | 0.0293 |
| 1990 | 1001.414 | 4127 | 600.553 | 59 | 21.9897 | 1.6146 | 0.0303 |
| 1991 | 1138.070 | 4436 | 661.796 | 55 | 19.4037 | 1.5285 | 0.0301 |
| 1992 | 758.254 | 2842 | 378.592 | 46 | 17.3690 | 1.2047 | 0.0339 |
| 1993 | 1014.985 | 3363 | 464.955 | 49 | 17.0123 | 1.2640 | 0.0325 |
| 1994 | 818.418 | 4470 | 473.423 | 49 | 16.1919 | 1.1081 | 0.0304 |
| 1995 | 789.528 | 4600 | 435.209 | 47 | 14.0323 | 1.0326 | 0.0301 |
| 1996 | 827.191 | 6218 | 544.828 | 51 | 12.3880 | 0.9524 | 0.0286 |
| 1997 | 1063.363 | 6031 | 672.142 | 53 | 14.8970 | 1.0413 | 0.0294 |
| 1998 | 876.404 | 4790 | 435.779 | 46 | 11.3605 | 0.8423 | 0.0304 |
| 1999 | 961.262 | 4429 | 447.847 | 50 | 11.3334 | 0.8527 | 0.0309 |
| 2000 | 945.098 | 5719 | 479.565 | 54 | 8.7637 | 0.6997 | 0.0296 |
| 2001 | 790.188 | 4930 | 258.551 | 48 | 5.8826 | 0.4890 | 0.0306 |
| 2002 | 811.136 | 5702 | 328.002 | 44 | 6.3660 | 0.5411 | 0.0300 |
| 2003 | 775.123 | 4585 | 237.585 | 47 | 5.3371 | 0.4243 | 0.0310 |
| 2004 | 765.506 | 4196 | 220.279 | 52 | 5.4124 | 0.4215 | 0.0319 |
| 2005 | 784.128 | 4378 | 262.616 | 39 | 6.8948 | 0.5102 | 0.0315 |
| 2006 | 811.298 | 3417 | 275.501 | 36 | 8.8173 | 0.6092 | 0.0332 |
| 2007 | 607.870 | 2437 | 212.373 | 20 | 9.2385 | 0.5636 | 0.0366 |
| 2008 | 700.439 | 3167 | 321.578 | 25 | 11.2739 | 0.7233 | 0.0345 |
| 2009 | 454.352 | 2447 | 228.460 | 19 | 10.4057 | 0.6625 | 0.0367 |
| 2010 | 380.248 | 2593 | 193.811 | 19 | 7.6433 | 0.4697 | 0.0364 |
| 2011 | 422.130 | 2401 | 170.945 | 18 | 7.3903 | 0.4519 | 0.0374 |



Figure 13.21. Jackass Morwong from zones 10 and 20 in depths $70-300 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Jackass Morwong, the top right graph depicts the depth distribution of shots containing Jackass Morwong from zones 10 and 20 in depths $70-300 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zones 10 and 20 ( 20 is top red line), the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Jackass Morwong catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).

Table 13.16. Jackass Morwong from zones 10 and 20 in depths $70-300 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.
Model $1 \quad$ LnCE $\sim$ Year

Model 2 LnCE~Year+Vessel
Model 3 LnCE~Year+Vessel+Month
Model $4 \quad$ LnCE $\sim$ Year+Vessel+Month+DepCat
Model 5 LnCE~Year+Vessel+Month+DepCat+Zone
Model 6 LnCE~Year+Vessel+Month+DepCat+Zone+DayNight
Model 7 LnCE~Year+Vessel+Month+DepCat+Zone+DayNight+Zone:Month
Model 8 LnCE $\sim$ Year+Vessel+Month+DepCat+Zone+DayNight+Zone:DepCat


Figure 13.22. Jackass Morwong from zones 10 and 20 in depths $70-300 \mathrm{~m}$ by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.17. Jackass Morwong from zones 10 and 20 in depths $70-300 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the adjusted $\mathrm{r}^{2}$ and the change in adjusted $\mathrm{r}^{2}$. The optimum was model Zone:Month.

|  | Year | Vessel | Month | DepCat | Zone | DayNight | Zone:Mth | Zone:DepC |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 78210 | 63936 | 61270 | 59207 | 57411 | 56112 | 55278 | 55810 |
| RSS | 223893 | 196049 | 191319 | 186963 | 183908 | 181721 | 180301 | 181179 |
| MSS | 29119 | 56963 | 61693 | 66049 | 69105 | 71292 | 72712 | 71833 |
| Nobs | 110061 | 110061 | 110061 | 109099 | 109099 | 109099 | 109099 | 109099 |
| Npars | 26 | 197 | 208 | 220 | 221 | 224 | 235 | 236 |
| adj_r2 | 1.489 | 22.376 | 24.241 | 25.957 | 27.166 | 28.030 | 28.585 | 28.237 |
| \%Change | 0.000 | 10.887 | 1.865 | 1.715 | 1.209 | 0.864 | 0.555 | -0.349 |



Figure 13.23. The relative influence of each factor used on the final trend in the optimal standardization for Jackass Morwong in Zones $10-20$. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.10 Jackass Morwong Z30 (MOR - 37377003 N. macropterus)

Only data from zone 30 were used, depths between 70 and 300 m taken by trawl.

Table 13.18. Jackass Morwong from zone 30 in depths $70-300 \mathrm{~m}$ by trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr)}$. Month:DepCat is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Month:Dep | CtDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 982.811 | 69 | 29.887 | 6 | 52.3193 | 1.7155 | 0.0000 |
| 1987 | 1087.690 | 210 | 57.476 | 13 | 45.8807 | 1.8206 | 0.1762 |
| 1988 | 1483.512 | 283 | 207.935 | 13 | 90.9064 | 2.4624 | 0.1710 |
| 1989 | 1667.373 | 687 | 475.039 | 19 | 125.0173 | 3.0771 | 0.1642 |
| 1990 | 1001.414 | 386 | 148.857 | 26 | 64.6762 | 2.2018 | 0.1649 |
| 1991 | 1138.070 | 427 | 189.534 | 29 | 68.3860 | 1.4168 | 0.1633 |
| 1992 | 758.254 | 335 | 106.819 | 18 | 50.3448 | 1.5066 | 0.1680 |
| 1993 | 1014.985 | 1042 | 325.873 | 27 | 49.6567 | 1.2045 | 0.1580 |
| 1994 | 818.418 | 762 | 180.185 | 22 | 40.3412 | 0.8326 | 0.1590 |
| 1995 | 789.528 | 826 | 185.282 | 19 | 36.4017 | 0.8047 | 0.1599 |
| 1996 | 827.191 | 890 | 161.402 | 19 | 29.4500 | 0.8239 | 0.1588 |
| 1997 | 1063.363 | 940 | 202.389 | 15 | 32.4284 | 0.9798 | 0.1573 |
| 1998 | 876.404 | 772 | 191.733 | 15 | 38.4649 | 0.9081 | 0.1587 |
| 1999 | 961.262 | 855 | 246.913 | 17 | 46.7614 | 1.0608 | 0.1592 |
| 2000 | 945.098 | 552 | 123.785 | 22 | 30.7755 | 0.7116 | 0.1611 |
| 2001 | 790.188 | 796 | 108.097 | 19 | 16.1559 | 0.4402 | 0.1583 |
| 2002 | 811.136 | 1044 | 108.944 | 15 | 13.9509 | 0.3855 | 0.1579 |
| 2003 | 775.123 | 1126 | 187.053 | 19 | 20.4814 | 0.5377 | 0.1569 |
| 2004 | 765.506 | 1500 | 201.278 | 15 | 18.1516 | 0.4021 | 0.1561 |
| 2005 | 784.128 | 1159 | 137.710 | 17 | 12.3142 | 0.2886 | 0.1574 |
| 2006 | 811.298 | 1127 | 154.482 | 14 | 17.6164 | 0.3623 | 0.1579 |
| 2007 | 607.870 | 714 | 111.625 | 8 | 22.5650 | 0.5004 | 0.1600 |
| 2008 | 700.439 | 768 | 119.020 | 9 | 24.1797 | 0.5271 | 0.1599 |
| 2009 | 454.352 | 463 | 54.343 | 10 | 16.5669 | 0.3751 | 0.1634 |
| 2010 | 380.248 | 372 | 58.189 | 9 | 19.1085 | 0.3938 | 0.1664 |
| 2011 | 422.130 | 452 | 48.260 | 8 | 11.9546 | 0.2606 | 0.1640 |



Figure 13.24. Jackass Morwong from zone 30 in depths $70-300 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Jackass Morwong, the top right graph depicts the depth distribution of shots containing Jackass Morwong from zone 30 in depths $70-300 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zone 30 , the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Jackass Morwong catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.25. Jackass Morwong from zone 30 in depths $70-300 \mathrm{~m}$ by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.19. Jackass Morwong from zone 30 in depths $70-300 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Month |
| Model 3 | LnCE $\sim$ Year+Month+Vessel |
| Model 4 | LnCE $\sim$ Year+Month+Vessel+DepCat |
| Model 5 | LnCE $\sim$ Year+Month+Vessel+DepCat+DayNight |
| Model 6 | LnCE $\sim$ Year+Month+Vessel+DepCat+DayNight+DayNight:Month |
| Model 7 | LnCE $\sim$ Year+Month+Vessel+DepCat+DayNight+Month:DepCat |
| Model 8 | LnCE $\sim$ Year+Month+Vessel+DepCat+DayNight+DayNight:DepCat |

Table 13.20. Jackass Morwong from zone 30 in depths $70-300 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the adjusted r 2 and the change in adjusted r2. The optimum was model Month:DepCat.

|  | Year | Month | Vessel | DepC | DN | DN:Mth | Mth:DepC | DN:DepC |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 9058 | 7165 | 6093 | 5614 | 5492 | 5449 | 5398 | 5533 |
| RSS | 30150 | 27193 | 25418 | 24554 | 24383 | 24239 | 23912 | 24342 |
| MSS | 6284 | 9241 | 11016 | 11880 | 12051 | 12196 | 12522 | 12092 |
| Nobs | 18557 | 18557 | 18557 | 18361 | 18361 | 18361 | 18361 | 18361 |
| Npars | 26 | 37 | 127 | 139 | 142 | 175 | 274 | 178 |
| adj_r2 | 17.137 | 25.219 | 29.758 | 32.098 | 32.558 | 32.837 | 33.379 | 32.539 |
| \%Change | 0.000 | 8.081 | 4.539 | 2.340 | 0.461 | 0.278 | 0.542 | -0.840 |



Figure 13.26. The relative influence of each factor used on the final trend in the optimal standardization for Jackass Morwong in Zone 30. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.11 Jackass Morwong Z4050 OT (MOR - 37377003 N. macropterus)

The data restrictions used in selecting the data for analysis were, depths between 70 and 360 m .

Table 13.21. Jackass Morwong from zones 40 and 50 in depths $70-360 \mathrm{~m}$ by trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates (kg/hr). Zone:Mth is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 982.811 | 551 | 149.261 | 19 | 40.7569 | 1.8476 | 0.0000 |
| 1987 | 1087.690 | 350 | 58.464 | 21 | 24.4475 | 1.4499 | 0.0873 |
| 1988 | 1483.512 | 402 | 65.444 | 19 | 32.2567 | 2.1629 | 0.0876 |
| 1989 | 1667.373 | 346 | 83.203 | 21 | 32.2213 | 1.5649 | 0.0924 |
| 1990 | 1001.414 | 412 | 80.657 | 22 | 28.9610 | 1.5788 | 0.0938 |
| 1991 | 1138.070 | 281 | 40.380 | 26 | 18.6097 | 1.0830 | 0.0980 |
| 1992 | 758.254 | 252 | 28.878 | 14 | 15.3915 | 0.8735 | 0.1009 |
| 1993 | 1014.985 | 248 | 24.971 | 17 | 15.5454 | 0.8514 | 0.1021 |
| 1994 | 818.418 | 312 | 22.679 | 16 | 14.6606 | 0.8207 | 0.0954 |
| 1995 | 789.528 | 295 | 77.615 | 17 | 21.5262 | 0.8724 | 0.0964 |
| 1996 | 827.191 | 346 | 37.071 | 17 | 15.3414 | 0.9493 | 0.0937 |
| 1997 | 1063.363 | 489 | 53.851 | 20 | 12.8372 | 0.7580 | 0.0870 |
| 1998 | 876.404 | 267 | 54.630 | 19 | 14.8359 | 0.7991 | 0.0990 |
| 1999 | 961.262 | 383 | 77.235 | 17 | 15.5951 | 0.7257 | 0.0917 |
| 2000 | 945.098 | 429 | 118.868 | 25 | 22.5254 | 1.0187 | 0.0919 |
| 2001 | 790.188 | 914 | 273.953 | 25 | 34.2135 | 1.0783 | 0.0810 |
| 2002 | 811.136 | 860 | 251.749 | 22 | 33.1596 | 1.0560 | 0.0814 |
| 2003 | 775.123 | 655 | 171.726 | 24 | 30.9832 | 0.9012 | 0.0847 |
| 2004 | 765.506 | 681 | 176.677 | 25 | 30.6678 | 0.9677 | 0.0837 |
| 2005 | 784.128 | 722 | 190.703 | 21 | 28.0502 | 1.0456 | 0.0832 |
| 2006 | 811.298 | 818 | 183.204 | 19 | 21.6176 | 0.8384 | 0.0823 |
| 2007 | 607.870 | 594 | 115.405 | 15 | 19.7196 | 0.6831 | 0.0851 |
| 2008 | 700.439 | 473 | 101.945 | 16 | 24.9534 | 0.6936 | 0.0883 |
| 2009 | 454.352 | 413 | 59.154 | 13 | 14.8023 | 0.5479 | 0.0913 |
| 2010 | 380.248 | 411 | 38.336 | 13 | 10.0135 | 0.4047 | 0.0908 |
| 2011 | 422.130 | 621 | 82.817 | 14 | 12.6335 | 0.4278 | 0.0864 |



Figure 13.27. Jackass Morwong from zones 40 and 50 in depths $70-360 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Jackass Morwong, the top right graph depicts the depth distribution of shots containing Jackass Morwong from zone 40 and 50 in depths $70-360 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zone 40 and 50 ( 50 is top red line), the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Jackass Morwong catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.28. Jackass Morwong from zones 40 and 50 in depths $70-360 \mathrm{~m}$ by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.22. Jackass Morwong from zones 40 and 50 in depths $70-360 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE~Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+DepCat |
| Model 3 | LnCE $\sim$ Year+DepCat + Month |
| Model 4 | LnCE $\sim$ Year+DepCat + Month+Vessel |
| Model 5 | LnCE $\sim$ Year+DepCat + Month+Vessel+DayNight |
| Model 6 | LnCE $\sim$ Year+DepCat + Month+Vessel+DayNight+Zone |
| Model 7 | LnCE $\sim$ Year+DepCat + Month+Vessel+DayNight+Zone + Zone:Month |
| Model 8 | LnCE $\sim$ Year+DepCat + Month + Vessel + DayNight + Zone + Zone:DepCat |

Table 13.23. Jackass Morwong from zones 40 and 50 in depths $70-360 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the adjusted r2 and the change in adjusted r2.The optimum was model 7 .

|  | Year | DepCat | Month | Vessel | DayNight | Zone | Zone:Mth | Zone:DepC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AIC | 7635 | 5345 | 4160 | 3556 | 3478 | 3389 | 3246 | 3285 |
| RSS | 22946 | 18985 | 17229 | 16194 | 16086 | 15968 | 15758 | 15797 |
| MSS | 1783 | 5744 | 7500 | 8535 | 8643 | 8761 | 8971 | 8932 |
| Nobs | 12525 | 12433 | 12433 | 12433 | 12433 | 12433 | 12433 | 12433 |
| Npars | 26 | 41 | 52 | 135 | 138 | 139 | 150 | 154 |
| adj_r2 | 7.025 | 22.979 | 30.041 | 33.800 | 34.228 | 34.702 | 35.505 | 35.323 |
| \%Change | 0.000 | 15.954 | 7.062 | 3.759 | 0.428 | 0.474 | 0.803 | -0.182 |



Figure 13.29. The relative influence of each factor used on the final trend in the optimal standardization for Jackass Morwong in Zones $40-50$. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.12 Flathead Trawl (FLT - 37296001 - Neoplatycephalus richardsoni)





Figure 13.30. The trends in catches and geometric mean catch rates for flathead as taken by trawl in Zones 10 to 30 . The catch rate trends in 10 and 20 are similar to each other but are different from that expressed in zone 30 . For this reason, zones 10 and 20 are standardized separately from Zone 30.


Figure 13.31. The standardized trends for the Month and DepCat factors for Flathead taken by trawl across SESSF zones 10-20.

### 13.13 Flathead Trawl Z1020(FLT - 37296001 - N. richardsoni)

Only data from zones 10 and 20 were used, depths less than 400 m .

Table 13.24 . Flathead from zones 10 and 20 in depths $0-400 \mathrm{~m}$ by trawl. Total Catch is the total reported
in the database, Records was the number of records used in the analysis, CatchT is the reported catch in
the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is
the geometric mean of catch rates $(\mathrm{kg} / \mathrm{hr})$. Zone:DepCat is the optimum model and StDev is the standard
deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:DepC | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 1892.183 | 10196 | 963.031 | 95 | 16.7357 | 0.8010 | 0.0000 |
| 1987 | 2461.337 | 8104 | 1008.332 | 86 | 20.4621 | 1.0713 | 0.0160 |
| 1988 | 2469.526 | 9175 | 1171.699 | 86 | 23.7988 | 1.1721 | 0.0158 |
| 1989 | 2599.063 | 8841 | 1210.472 | 74 | 23.9908 | 1.1672 | 0.0159 |
| 1990 | 2032.323 | 7765 | 1221.459 | 64 | 30.1854 | 1.3861 | 0.0168 |
| 1991 | 2230.185 | 7797 | 1145.652 | 57 | 28.7154 | 1.3155 | 0.0168 |
| 1992 | 2375.366 | 6810 | 871.934 | 53 | 23.8898 | 1.0282 | 0.0175 |
| 1993 | 1879.140 | 8782 | 998.146 | 58 | 23.8001 | 1.0500 | 0.0167 |
| 1994 | 1710.404 | 10280 | 902.906 | 56 | 17.9798 | 0.7610 | 0.0160 |
| 1995 | 1800.616 | 10305 | 994.134 | 54 | 18.0790 | 0.8067 | 0.0159 |
| 1996 | 1879.872 | 11089 | 958.779 | 59 | 16.4549 | 0.7138 | 0.0158 |
| 1997 | 2355.987 | 10395 | 997.137 | 60 | 16.8264 | 0.7166 | 0.0162 |
| 1998 | 2306.407 | 9986 | 999.535 | 52 | 17.7430 | 0.7588 | 0.0162 |
| 1999 | 3117.465 | 10377 | 1129.356 | 57 | 20.4344 | 0.9100 | 0.0160 |
| 2000 | 2945.581 | 13110 | 1696.814 | 59 | 24.4338 | 1.0130 | 0.0155 |
| 2001 | 2599.512 | 11957 | 1375.379 | 53 | 22.3118 | 0.9759 | 0.0158 |
| 2002 | 2876.253 | 12357 | 1444.049 | 49 | 22.8273 | 1.0657 | 0.0157 |
| 2003 | 3230.066 | 12879 | 1593.850 | 52 | 22.5536 | 1.0530 | 0.0156 |
| 2004 | 3222.611 | 12220 | 1343.072 | 52 | 19.7879 | 0.9091 | 0.0158 |
| 2005 | 2844.045 | 10703 | 1154.986 | 49 | 17.7159 | 0.7770 | 0.0162 |
| 2006 | 2585.823 | 9137 | 1148.779 | 46 | 22.2550 | 0.9429 | 0.0167 |
| 2007 | 2648.311 | 6337 | 1076.563 | 25 | 31.3544 | 1.1537 | 0.0184 |
| 2008 | 2910.286 | 7292 | 1330.559 | 27 | 31.6602 | 1.2088 | 0.0178 |
| 2009 | 2460.393 | 6311 | 1060.713 | 26 | 30.0219 | 1.1115 | 0.0185 |
| 2010 | 2501.518 | 6872 | 1124.212 | 25 | 29.4565 | 1.0698 | 0.0181 |
| 2011 | 2465.166 | 6764 | 1095.324 | 24 | 28.3798 | 1.0611 | 0.0182 |



Figure 13.32. Flathead from zones 10 and 20 in depths $0-400 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Jackass Morwong, the top right graph depicts the depth distribution of shots containing Flathead from zones 10 and 20 in depths $0-400 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zones 10 and 20 ( 20 is top red line), the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Flathead catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.33. Flathead from zones 10 and 20 in depths $0-400 \mathrm{~m}$ by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.25. Flathead from zones 10 and 20 in depths $0-400 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel+DepCat |
| Model 4 | LnCE $\sim$ Year+Vessel+DepCat+Month |
| Model 5 | LnCE $\sim$ Year+Vessel+DepCat+Month+DayNight |
| Model 6 | LnCE $\sim$ Year+Vessel+DepCat + Month+DayNight + Zone |
| Model 7 | LnCE $\sim$ Year+Vessel+DepCat + Month+DayNight+Zone+Zone:Month |
| Model 8 | LnCE $\sim$ Year+Vessel+DepCat + Month+DayNight + Zone + Zone:DepCat |

Table 13.26. Flathead from zones 10 and 20 in depths $0-400 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the adjusted r 2 and the change in adjusted r 2 . The optimum was model Zone:DepCat.

|  | Year | Vessel | DepCat | Month | DayNight | Zone | Zone:Month |  | Zone:DepCat |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 44640 | 17349 | 9494 | 8636 | 8503 | 8456 | 6606 | 5504 |  |
| RSS | 294728 | 263379 | 253108 | 252197 | 252053 | 252002 | 250075 | 248929 |  |
| MSS | 9583 | 40932 | 51203 | 52114 | 52258 | 52309 | 54236 | 55382 |  |
| Nobs | 245841 | 245841 | 243892 | 243892 | 243892 | 243892 | 243892 | 243892 |  |
| Npars | 26 | 204 | 224 | 235 | 238 | 239 | 250 | 259 |  |
| adj_r2 | 3.139 | 13.379 | 16.750 | 17.046 | 17.092 | 17.108 | 17.738 | 18.112 |  |
| \%Change | 0.000 | 10.240 | 3.371 | 0.296 | 0.046 | 0.016 | 0.630 | 0.374 |  |



Figure 13.34. The relative influence of each factor used on the final trend in the optimal standardization for Flathead in Zones $10-20$. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.14 Flathead Trawl Z30 (FLT - 37296001 - N. richardsoni)

Only data from zone 30 were used, depths less than 400 m .

Table 13.27. Flathead from zone 30 in depths $0-400 \mathrm{~m}$ by trawl. Total Catch is the total reported in the
database, Records was the number of records used in the analysis, CatchT is the reported catch in the area
and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the
geometric mean of catch rates $(\mathrm{kg} / \mathrm{hr)}. \mathrm{Month:DepCat} \mathrm{is} \mathrm{the} \mathrm{optimum} \mathrm{model} \mathrm{and} \mathrm{StDev} \mathrm{is} \mathrm{the} \mathrm{standard}$
deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Mth:DepC | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 1892.183 | 71 | 16.754 | 6 | 23.1157 | 0.9649 | 0.0000 |
| 1987 | 2461.337 | 90 | 5.155 | 9 | 11.1912 | 0.5830 | 0.1898 |
| 1988 | 2469.526 | 193 | 39.976 | 9 | 21.2587 | 1.0107 | 0.1701 |
| 1989 | 2599.063 | 516 | 48.443 | 19 | 20.5177 | 0.7583 | 0.1625 |
| 1990 | 2032.323 | 253 | 24.619 | 27 | 20.3187 | 0.7837 | 0.1647 |
| 1991 | 2230.185 | 314 | 33.353 | 29 | 15.9189 | 0.7143 | 0.1608 |
| 1992 | 2375.366 | 272 | 33.897 | 15 | 22.4408 | 0.6725 | 0.1649 |
| 1993 | 1879.140 | 902 | 92.079 | 24 | 17.1065 | 0.6436 | 0.1563 |
| 1994 | 1710.404 | 612 | 64.487 | 17 | 18.5289 | 0.6696 | 0.1573 |
| 1995 | 1800.616 | 694 | 71.349 | 17 | 19.8905 | 0.7323 | 0.1576 |
| 1996 | 1879.872 | 714 | 61.425 | 17 | 15.7596 | 0.6761 | 0.1571 |
| 1997 | 2355.987 | 885 | 104.875 | 14 | 20.7052 | 0.8488 | 0.1554 |
| 1998 | 2306.407 | 707 | 118.552 | 14 | 28.8666 | 1.0027 | 0.1563 |
| 1999 | 3117.465 | 770 | 175.052 | 17 | 31.0992 | 1.1046 | 0.1567 |
| 2000 | 2945.581 | 520 | 83.664 | 20 | 25.4446 | 0.8942 | 0.1580 |
| 2001 | 2599.512 | 916 | 101.308 | 17 | 18.0579 | 0.7669 | 0.1553 |
| 2002 | 2876.253 | 1367 | 212.158 | 15 | 30.1174 | 1.4329 | 0.1544 |
| 2003 | 3230.066 | 1454 | 240.110 | 21 | 30.0485 | 1.4655 | 0.1538 |
| 2004 | 3222.611 | 1923 | 477.416 | 15 | 47.0053 | 1.9274 | 0.1534 |
| 2005 | 2844.045 | 1540 | 388.325 | 18 | 43.4956 | 1.7224 | 0.1539 |
| 2006 | 2585.823 | 1315 | 287.968 | 13 | 37.5195 | 1.3838 | 0.1547 |
| 2007 | 2648.311 | 823 | 173.155 | 8 | 33.0381 | 1.1403 | 0.1562 |
| 2008 | 2910.286 | 874 | 173.739 | 11 | 29.3148 | 1.0596 | 0.1560 |
| 2009 | 2460.393 | 600 | 100.225 | 10 | 29.0939 | 1.0262 | 0.1576 |
| 2010 | 2501.518 | 537 | 104.186 | 10 | 28.3260 | 1.0402 | 0.1586 |
| 2011 | 2465.166 | 623 | 131.274 | 9 | 29.1229 | 0.9756 | 0.1577 |



Figure 13.35. Flathead from zone 30 in depths $0-400 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Flathead, the top right graph depicts the depth distribution of shots containing Flathead from zone 30 in depths $0-400 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zone 30 , the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Flathead catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.36. Flathead from zone 30 in depths $0-400 \mathrm{~m}$ by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.28. Flathead from zone 30 in depths $0-400 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel+DepCat |
| Model 4 | LnCE $\sim$ Year+Vessel+DepCat+DayNight |
| Model 5 | LnCE $\sim$ Year+Vessel+DepCat+DayNight+Month |
| Model 6 | LnCE $\sim$ Year+Vessel+DepCat+DayNight+Month+DayNight:DepCat |
| Model 7 | LnCE $\sim$ Year+Vessel+DepCat+DayNight+Month+Month:DepCat |
| Model 8 | LnCE $\sim$ Year+Vessel+DepCat+DayNight+Month+DayNight:DepCat |

Table 13.29. Flathead from zone 30 in depths $0-400 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the adjusted r 2 and the change in adjusted r 2 . The optimum was model 7.

|  | Year | Vessel | DepCat | DN | Mth | DN:Mth | Mth:DepC | DN:DepC |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 2767 | 1076 | 41 | -304 | -613 | -664 | -987 | -695 |
| RSS | 22398 | 20352 | 19052 | 18708 | 18390 | 18279 | 17629 | 18198 |
| MSS | 2161 | 4208 | 5508 | 5851 | 6169 | 6280 | 6930 | 6362 |
| Nobs | 19485 | 19485 | 19278 | 19278 | 19278 | 19278 | 19278 | 19278 |
| Npars | 26 | 114 | 134 | 137 | 148 | 181 | 368 | 208 |
| adj_r2 | 8.683 | 16.651 | 21.887 | 23.283 | 24.544 | 24.871 | 26.825 | 25.098 |
| \%Change | 0.000 | 7.968 | 5.236 | 1.396 | 1.260 | 0.327 | 1.954 | -1.727 |



Figure 13.37. The relative influence of each factor used on the final trend in the optimal standardization for Flathead from zone 30. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.


Figure 13.38. The standardized trends for the Month and DepCat factors for Flathead taken by trawl across SESSF zone 30 .
13.15 Flathead Danish Seine (FLT - 37296001 - N. richardsoni)

Only data from zones 20, and 60 were used, for Danish Seine vessels only (i.e. exclude Otter Trawl vessels), and depths less than 200 m .

Table 13.30. Flathead from zones 20 and 60 in depths $0-200 \mathrm{~m}$ by Danish Seine. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates (kg/hr). Zone:Month is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 1892.183 | 5501 | 763.945 | 26 | 45.0535 | 1.0196 | 0.0000 |
| 1987 | 2461.337 | 5651 | 1366.944 | 23 | 88.6187 | 1.4382 | 0.0227 |
| 1988 | 2469.526 | 5823 | 1097.541 | 25 | 88.9194 | 1.5679 | 0.0225 |
| 1989 | 2599.063 | 5412 | 1142.708 | 27 | 78.4955 | 1.3677 | 0.0229 |
| 1990 | 2032.323 | 4653 | 586.018 | 25 | 48.3882 | 0.8956 | 0.0241 |
| 1991 | 2230.185 | 4670 | 775.768 | 28 | 69.8580 | 1.2546 | 0.0242 |
| 1992 | 2375.366 | 6643 | 1218.041 | 24 | 85.5977 | 1.3526 | 0.0223 |
| 1993 | 1879.140 | 5859 | 539.588 | 24 | 39.0251 | 0.8576 | 0.0230 |
| 1994 | 1710.404 | 7332 | 649.481 | 25 | 37.6721 | 0.7210 | 0.0218 |
| 1995 | 1800.616 | 5505 | 656.665 | 21 | 36.2337 | 0.7409 | 0.0232 |
| 1996 | 1879.872 | 7679 | 755.670 | 22 | 33.6052 | 0.6993 | 0.0218 |
| 1997 | 2355.987 | 8480 | 1150.436 | 21 | 60.3446 | 0.9063 | 0.0215 |
| 1998 | 2306.407 | 9904 | 1134.732 | 21 | 60.5323 | 0.7587 | 0.0210 |
| 1999 | 3117.465 | 8818 | 1702.605 | 23 | 98.4160 | 1.0917 | 0.0214 |
| 2000 | 2945.581 | 7092 | 1037.689 | 19 | 64.0436 | 0.8036 | 0.0225 |
| 2001 | 2599.512 | 7457 | 1004.507 | 18 | 62.0182 | 0.7550 | 0.0226 |
| 2002 | 2876.253 | 8218 | 1144.075 | 22 | 75.2709 | 0.8971 | 0.0222 |
| 2003 | 3230.066 | 9006 | 1210.597 | 23 | 80.7627 | 0.9580 | 0.0219 |
| 2004 | 3222.611 | 7784 | 1253.026 | 22 | 83.7818 | 0.9388 | 0.0224 |
| 2005 | 2844.045 | 7212 | 1125.753 | 22 | 87.7421 | 0.9572 | 0.0228 |
| 2006 | 2585.823 | 5563 | 968.051 | 21 | 89.1577 | 0.9501 | 0.0239 |
| 2007 | 2648.311 | 5551 | 1182.067 | 15 | 104.4620 | 1.1512 | 0.0238 |
| 2008 | 2910.286 | 6214 | 1283.489 | 15 | 103.2936 | 1.0314 | 0.0234 |
| 2009 | 2460.393 | 5499 | 1168.928 | 15 | 91.4234 | 1.0605 | 0.0239 |
| 2010 | 2501.518 | 6048 | 1166.861 | 15 | 101.4483 | 0.9437 | 0.0235 |
| 2011 | 2465.166 | 6887 | 1121.755 | 14 | 85.7656 | 0.8817 | 0.0230 |



Figure 13.39. Flathead from zones 20 and 60 in depths $0-200 \mathrm{~m}$ by Danish Seine. The top left is the depth distribution of all records reporting Flathead, the top right graph depicts the depth distribution of shots containing Flathead from zones 20 and 60 in depths $0-200 \mathrm{~m}$ by Danish Seine. The middle left diagram depicts the distribution of catch by depth within zones 20 and 60 , the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Flathead catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.40. Flathead from zones 20 and 60 in depths $0-200 \mathrm{~m}$ by Danish Seine. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.31. Flathead from zones 20 and 60 in depths $0-200 \mathrm{~m}$ by Danish Seine. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE~Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Zone |
| Model 3 | LnCE $\sim$ Year+Zone+DepCat |
| Model 4 | LnCE $\sim$ Year+Zone+DepCat+Vessel |
| Model 5 | LnCE $\sim$ Year+Zone+DepCat+Vessel+Month |
| Model 6 | LnCE $\sim$ Year+Zone+DepCat+Vessel+Month+DayNight |
| Model 7 | LnCE $\sim$ Year+Zone+DepCat+Vessel+Month+DayNight+Zone:Month |
| Model 8 | LnCE $\sim$ Year+Zone+DepCat+Vessel+Month+DayNight+Zone:DepCat |

Table 13.32. Flathead from zones 20 and 60 in depths $0-200 \mathrm{~m}$ by Danish Seine. Model selection criteria, including the AIC, the adjusted $r^{2}$ and the change in adjusted $r^{2}$. The optimum was model Zone:Month.

|  | Year | Zone | DepCat | Vessel | Month | DayNight |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | Zone:Mth Zone:DepC



Figure 13.41. The relative influence of each factor used on the final trend in the optimal standardization for Flathead by Danish Seine in Zones $20 \& 60$. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.


Figure 13.42. Standardized trends for Month and DepCat factors for Flathead taken by trawl Danish Seine.

### 13.16 RedFish Zone 10 (RED - 37258003 - Centroberyx affinis)

Only data from zone 10 were used, depths less than 400 m .

Table 13.33. Redfish from zone 10 in depths $0-400 \mathrm{~m}$ by trawl. Total Catch is the total reported in the
database, Records is the number of records used in the analysis, CatchT is the reported catch in the area
and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the
geometric mean of catch rates $(\mathrm{kg} / \mathrm{hr})$. Month:DepCat is the optimum model and StDev is the standard
deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Mth:DepC | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 1687.471 | 4503 | 1528.926 | 81 | 38.3044 | 1.5835 | 0.0000 |
| 1987 | 1252.658 | 3383 | 1114.805 | 73 | 35.9993 | 1.2675 | 0.0371 |
| 1988 | 1125.492 | 2966 | 904.361 | 70 | 37.3114 | 1.3164 | 0.0390 |
| 1989 | 714.316 | 2156 | 586.942 | 64 | 29.4122 | 1.0928 | 0.0431 |
| 1990 | 931.370 | 1894 | 699.754 | 49 | 37.2522 | 1.4854 | 0.0453 |
| 1991 | 1570.607 | 2467 | 1056.996 | 44 | 39.9367 | 1.5398 | 0.0421 |
| 1992 | 1636.687 | 2428 | 1393.725 | 41 | 50.0990 | 1.9882 | 0.0430 |
| 1993 | 1921.347 | 2960 | 1611.795 | 47 | 56.0385 | 2.5202 | 0.0407 |
| 1994 | 1487.717 | 4208 | 1140.891 | 49 | 35.8972 | 1.7668 | 0.0378 |
| 1995 | 1240.617 | 4397 | 1027.576 | 46 | 27.8589 | 1.1458 | 0.0368 |
| 1996 | 1344.049 | 4063 | 1094.993 | 50 | 26.2588 | 0.9344 | 0.0375 |
| 1997 | 1397.328 | 2952 | 1157.743 | 50 | 33.5183 | 1.0959 | 0.0406 |
| 1998 | 1553.718 | 3072 | 1363.404 | 43 | 43.1196 | 1.3762 | 0.0402 |
| 1999 | 1116.403 | 2998 | 969.424 | 44 | 32.7876 | 1.0812 | 0.0402 |
| 2000 | 758.275 | 3300 | 642.137 | 48 | 22.7760 | 0.7288 | 0.0398 |
| 2001 | 742.268 | 3209 | 607.215 | 41 | 17.8301 | 0.7123 | 0.0398 |
| 2002 | 807.133 | 3481 | 601.823 | 44 | 16.4201 | 0.6089 | 0.0396 |
| 2003 | 615.183 | 2690 | 478.879 | 43 | 17.0122 | 0.5826 | 0.0417 |
| 2004 | 476.009 | 2717 | 390.967 | 44 | 15.2541 | 0.4941 | 0.0416 |
| 2005 | 483.516 | 2443 | 360.961 | 41 | 16.1484 | 0.5057 | 0.0429 |
| 2006 | 325.092 | 1768 | 256.212 | 34 | 15.6812 | 0.4745 | 0.0472 |
| 2007 | 216.279 | 1207 | 149.288 | 18 | 15.4678 | 0.4237 | 0.0547 |
| 2008 | 183.757 | 1396 | 155.290 | 22 | 13.9780 | 0.3993 | 0.0524 |
| 2009 | 160.525 | 1171 | 123.810 | 20 | 11.3207 | 0.3253 | 0.0558 |
| 2010 | 152.816 | 1228 | 112.793 | 19 | 10.4815 | 0.3059 | 0.0547 |
| 2011 | 87.305 | 870 | 63.806 | 17 | 8.5118 | 0.2447 | 0.0615 |



Figure 13.43. Redfish from zone 10 in depths $0-400 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Redfish, the top right graph depicts the depth distribution of shots containing Redfish from zone 10 in depths $0-400 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zone 10 , the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Redfish catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<$ 30 Kg ).


Figure 13.44. Redfish from zone 10 in depths $0-400 \mathrm{~m}$ by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.34. Redfish from zone 10 in depths $0-400 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel+DepCat |
| Model 4 | LnCE $\sim$ Year+Vessel+DepCat+DayNight |
| Model 5 | LnCE $\sim$ Year+Vessel+DepCat+DayNight + Month |
| Model 6 | LnCE $\sim$ Year+Vessel+DepCat+DayNight+Month+DayNight:DepCat |
| Model 7 | LnCE $\sim$ Year+Vessel+DepCat+DayNight+Month+Month:DepCat |
| Model 8 | LnCE $\sim$ Year+Vessel+DepCat+DayNight+Month+DayNight:DepCat |

Table 13.35. Redfish from zone 10 in depths $0-400 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the adjusted $r^{2}$ and the change in adjusted $r^{2}$. The optimum was model Month:DepCat.

|  | Year | Vessel | DepC | DN | Mth | DN:Mth | Mth:DepC | DN:DepC |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 74101 | 66157 | 61252 | 60659 | 60164 | 60033 | 58865 | 59387 |
| RSS | 201623 | 179231 | 166879 | 165448 | 164223 | 163758 | 160168 | 162119 |
| MSS | 13765 | 36156 | 48508 | 49939 | 51165 | 51630 | 55220 | 53268 |
| Nobs | 69927 | 69927 | 69550 | 69550 | 69550 | 69550 | 69550 | 69550 |
| Npars | 26 | 170 | 190 | 193 | 204 | 237 | 424 | 264 |
| adj_r2 | 6.357 | 16.585 | 22.310 | 22.973 | 23.532 | 23.712 | 25.182 | 24.446 |
| \%Change | 0.000 | 10.228 | 5.725 | 0.663 | 0.558 | 0.180 | 1.471 | -0.737 |



Figure 13.45. The relative influence of each factor used on the final trend in the optimal standardization for Redfish in Zone 10. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.17 RedFish Zone 20 (RED - 37258003 - Centroberyx affinis)

Only data from zone 20 were used, depths less than 400 m .

Table 13.36. Redfish from zone 20 in depths $0-400 \mathrm{~m}$ by trawl. Total Catch is the total reported in the
database, Records is the number of records used in the analysis, CatchT is the reported catch in the area
and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the
geometric mean of catch rates $(\mathrm{kg} / \mathrm{hr})$. Month:DepCat is the optimum model and StDev is the standard
deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Mth:DepC | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 1687.471 | 838 | 69.648 | 34 | 12.7888 | 1.2276 | 0.0000 |
| 1987 | 1252.658 | 548 | 70.567 | 28 | 16.3056 | 1.5256 | 0.0867 |
| 1988 | 1125.492 | 1008 | 174.671 | 35 | 22.5742 | 2.1550 | 0.0784 |
| 1989 | 714.316 | 567 | 57.490 | 32 | 13.8221 | 1.3220 | 0.0880 |
| 1990 | 931.370 | 699 | 95.090 | 34 | 16.4273 | 1.5491 | 0.0864 |
| 1991 | 1570.607 | 886 | 181.397 | 27 | 20.9240 | 1.9209 | 0.0851 |
| 1992 | 1636.687 | 691 | 100.149 | 25 | 18.2135 | 1.5608 | 0.0902 |
| 1993 | 1921.347 | 836 | 175.486 | 25 | 23.8774 | 1.9120 | 0.0871 |
| 1994 | 1487.717 | 1291 | 212.848 | 26 | 22.1556 | 1.7884 | 0.0820 |
| 1995 | 1240.617 | 1316 | 169.079 | 24 | 14.7891 | 1.1242 | 0.0805 |
| 1996 | 1344.049 | 1751 | 210.919 | 26 | 11.8255 | 1.1110 | 0.0787 |
| 1997 | 1397.328 | 1456 | 196.332 | 28 | 10.9003 | 0.9231 | 0.0811 |
| 1998 | 1553.718 | 1237 | 164.642 | 24 | 11.9357 | 0.9827 | 0.0822 |
| 1999 | 1116.403 | 947 | 122.433 | 25 | 9.4628 | 0.8459 | 0.0853 |
| 2000 | 758.275 | 1364 | 92.988 | 27 | 5.0564 | 0.5684 | 0.0825 |
| 2001 | 742.268 | 1345 | 113.456 | 24 | 5.9658 | 0.5810 | 0.0831 |
| 2002 | 807.133 | 1725 | 172.165 | 24 | 6.7628 | 0.6660 | 0.0818 |
| 2003 | 615.183 | 1428 | 77.081 | 26 | 4.5183 | 0.4330 | 0.0831 |
| 2004 | 476.009 | 1248 | 59.212 | 22 | 4.2622 | 0.4417 | 0.0855 |
| 2005 | 483.516 | 1353 | 92.209 | 20 | 5.5759 | 0.5611 | 0.0840 |
| 2006 | 325.092 | 821 | 46.469 | 21 | 4.7612 | 0.4929 | 0.0895 |
| 2007 | 216.279 | 673 | 59.701 | 11 | 5.6299 | 0.5908 | 0.0934 |
| 2008 | 183.757 | 536 | 24.505 | 17 | 4.1887 | 0.4893 | 0.0978 |
| 2009 | 160.525 | 448 | 30.527 | 12 | 4.9795 | 0.5023 | 0.1016 |
| 2010 | 152.816 | 644 | 34.686 | 15 | 4.4782 | 0.4597 | 0.0971 |
| 2011 | 87.305 | 539 | 20.314 | 12 | 2.6852 | 0.2652 | 0.0996 |



Figure 13.46. Redfish from zone 20 in depths $0-400 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Redfish, the top right graph depicts the depth distribution of shots containing Redfish from zone 20 in depths $0-400 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zone 20 , the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Redfish catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches < 30 Kg ).


Figure 13.47. Redfish from zone 20 in depths $0-400 \mathrm{~m}$ by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.37. Redfish from zone 20 in depths $0-400 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel+DepCat |
| Model 4 | LnCE $\sim$ Year+Vessel+DepCat+DayNight |
| Model 5 | LnCE $\sim$ Year+Vessel+DepCat+DayNight+Month |
| Model 6 | LnCE $\sim$ Year+Vessel+DepCat+DayNight+Month+DayNight:DepCat |
| Model 7 | LnCE $\sim$ Year+Vessel+DepCat+DayNight+Month+Month:DepCat |
| Model 8 | LnCE $\sim$ Year+Vessel+DepCat+DayNight+Month+DayNight:DepCat |

Table 13.38. Redfish from zone 20 in depths $0-400 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the adjusted $r^{2}$ and the change in adjusted $r^{2}$. The optimum was model Month:DepCat.

|  | Year | Vessel | DepCat | Mth | DayNight | DN:Mth | Mth:DepCat | DN:DepCat |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AIC | 24240 | 20932 | 19183 | 18735 | 18700 | 18668 | 18404 | 18616 |
| RSS | 65955 | 57674 | 53783 | 52822 | 52737 | 52540 | 51269 | 52326 |
| MSS | 9493 | 17773 | 21665 | 22626 | 22710 | 22908 | 24179 | 23122 |
| Nobs | 26195 | 26195 | 26049 | 26049 | 26049 | 26049 | 26049 | 26049 |
| Npars | 26 | 129 | 149 | 160 | 163 | 196 | 383 | 223 |
| adj_r2 | 12.499 | 23.182 | 28.308 | 29.558 | 29.663 | 29.837 | 31.036 | 30.050 |
| \%Change | 0.000 | 10.683 | 5.126 | 1.251 | 0.105 | 0.174 | 1.198 | -0.986 |



Figure 13.48. The relative influence of each factor used on the final trend in the optimal standardization for Redfish in Zone 20. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.18 Silver Trevally (TRE - 37337062 - Pseudocaranx dentex)

Only data from zones 10 and 20 combined were used, depths less than 200 m . In order to discount the influence of catches taken within the Batemans Bay MPA, all data in Commonwealth waters within the MPA have been excluded from the analysis. The selection of which records to exclude is improved over last year's analysis through the use of improved GIS.

Table 13.39. Silver Trevally from Zones 10 and 20 in depths 0 to 200 m , excluding data taken in State waters (Bateman's Bay MPA). Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr)}$ ). Zone:Month is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 469.508 | 1765 | 278.628 | 74 | 17.0086 | 1.1373 | 0.0000 |
| 1987 | 198.490 | 1090 | 116.317 | 63 | 17.5072 | 1.3565 | 0.0593 |
| 1988 | 278.541 | 1299 | 226.620 | 52 | 23.7642 | 1.7779 | 0.0546 |
| 1989 | 376.196 | 1838 | 278.037 | 62 | 23.0657 | 1.8873 | 0.0501 |
| 1990 | 450.391 | 1841 | 288.809 | 52 | 23.2975 | 2.2336 | 0.0518 |
| 1991 | 340.683 | 1909 | 213.903 | 49 | 18.1137 | 2.0038 | 0.0522 |
| 1992 | 296.493 | 1194 | 108.366 | 44 | 12.0774 | 1.1420 | 0.0586 |
| 1993 | 377.673 | 1262 | 132.861 | 47 | 13.4863 | 1.2467 | 0.0575 |
| 1994 | 392.828 | 1839 | 139.154 | 46 | 9.4912 | 0.9557 | 0.0531 |
| 1995 | 413.439 | 1570 | 136.637 | 43 | 10.2789 | 1.0853 | 0.0551 |
| 1996 | 340.616 | 1883 | 129.536 | 47 | 7.5806 | 0.8718 | 0.0534 |
| 1997 | 328.839 | 1450 | 88.499 | 48 | 6.2012 | 0.8265 | 0.0572 |
| 1998 | 210.136 | 1023 | 48.972 | 40 | 5.2414 | 0.6064 | 0.0610 |
| 1999 | 166.018 | 882 | 41.568 | 39 | 4.9696 | 0.6057 | 0.0642 |
| 2000 | 154.753 | 1020 | 43.620 | 42 | 3.6777 | 0.4529 | 0.0615 |
| 2001 | 270.175 | 1536 | 82.085 | 43 | 4.1345 | 0.5290 | 0.0554 |
| 2002 | 232.787 | 1474 | 67.852 | 40 | 3.0864 | 0.4299 | 0.0572 |
| 2003 | 337.667 | 1124 | 57.733 | 45 | 3.3755 | 0.4218 | 0.0595 |
| 2004 | 458.075 | 1345 | 84.499 | 42 | 4.5401 | 0.5836 | 0.0579 |
| 2005 | 290.940 | 673 | 59.560 | 40 | 4.7971 | 0.5154 | 0.0691 |
| 2006 | 247.284 | 493 | 48.824 | 32 | 5.7178 | 0.7212 | 0.0764 |
| 2007 | 172.720 | 463 | 47.115 | 20 | 7.4274 | 0.8211 | 0.0791 |
| 2008 | 128.386 | 818 | 69.665 | 23 | 8.0833 | 0.8476 | 0.0658 |
| 2009 | 164.752 | 838 | 94.881 | 23 | 9.2632 | 0.8553 | 0.0651 |
| 2010 | 240.308 | 967 | 135.510 | 24 | 11.7000 | 1.0954 | 0.0634 |
| 2011 | 192.047 | 860 | 139.299 | 20 | 11.0945 | 0.9902 | 0.0654 |
|  |  |  |  |  |  |  |  |



Figure 13.49. Silver Trevally from Zones 10 and 20 in depths 0 to 200 m, excluding data taken in State waters (Bateman's Bay MPA). The top left is the depth distribution of all records reporting Silver Trevally, the top right graph depicts the depth distribution of shots containing Silver Trevally from Zones 10 and 20 in depths 0 to 200 m by trawl, excluding data taken in State waters (Bateman's Bay MPA). The middle left diagram depicts the distribution of catch by depth within zones 10 and 20 ( 20 is top red line), the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Silver Trevally catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.50. Silver Trevally from Zones 10 and 20 in depths 0 to 200 m , excluding data taken in State waters (Bateman's Bay MPA). The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.40. Silver Trevally from Zones 10 and 20 in depths 0 to 200 m , excluding data taken in State waters (Bateman's Bay MPA). Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.



Figure 13.51. The relative influence of each factor used on the final trend in the optimal standardization for Silver Trevally in Zones 10 and 20. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.19 Royal Red Prawn (PRR - 28714005 - Haliporoides sibogae)

Only data from Zone 10 were used, depths between 200 - 700 m .

Table 13.42. Royal Red Prawn from zone 10 in depths $200-700 \mathrm{~m}$ by trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates (kg/hr). Month:DepCat is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Mth:DepC | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 277.717 | 1592 | 231.844 | 47 | 27.7627 | 0.6928 | 0.0000 |
| 1987 | 351.294 | 1764 | 324.716 | 47 | 41.9857 | 0.8818 | 0.0380 |
| 1988 | 362.505 | 1395 | 344.457 | 41 | 49.1496 | 0.9761 | 0.0409 |
| 1989 | 329.254 | 1143 | 310.760 | 39 | 45.8268 | 0.8322 | 0.0428 |
| 1990 | 337.134 | 727 | 311.118 | 25 | 95.1525 | 1.5596 | 0.0491 |
| 1991 | 334.134 | 734 | 299.370 | 29 | 79.4866 | 1.3942 | 0.0495 |
| 1992 | 166.860 | 434 | 146.081 | 19 | 70.3817 | 1.0429 | 0.0579 |
| 1993 | 298.797 | 673 | 232.774 | 21 | 68.5216 | 1.1944 | 0.0493 |
| 1994 | 359.830 | 661 | 240.363 | 26 | 77.7193 | 1.1363 | 0.0496 |
| 1995 | 335.592 | 1070 | 252.905 | 25 | 58.4998 | 0.9009 | 0.0435 |
| 1996 | 360.776 | 1216 | 272.675 | 25 | 60.5827 | 0.8126 | 0.0420 |
| 1997 | 252.693 | 855 | 166.703 | 21 | 51.9861 | 0.7661 | 0.0463 |
| 1998 | 233.298 | 1234 | 190.732 | 23 | 39.1713 | 0.8258 | 0.0427 |
| 1999 | 367.042 | 1607 | 348.804 | 25 | 49.7799 | 0.8163 | 0.0404 |
| 2000 | 434.931 | 1538 | 398.474 | 27 | 49.6136 | 1.0256 | 0.0408 |
| 2001 | 276.786 | 1307 | 228.699 | 22 | 35.9685 | 0.8744 | 0.0431 |
| 2002 | 484.209 | 1740 | 417.370 | 23 | 47.9208 | 1.0510 | 0.0402 |
| 2003 | 230.805 | 801 | 163.184 | 26 | 39.7063 | 1.0958 | 0.0491 |
| 2004 | 193.801 | 579 | 170.681 | 22 | 50.4687 | 1.1216 | 0.0535 |
| 2005 | 173.896 | 601 | 159.805 | 21 | 47.1225 | 1.0226 | 0.0536 |
| 2006 | 192.262 | 455 | 178.579 | 17 | 55.0038 | 1.2309 | 0.0580 |
| 2007 | 121.545 | 324 | 116.430 | 9 | 48.8072 | 0.8506 | 0.0663 |
| 2008 | 75.799 | 252 | 70.605 | 8 | 39.0864 | 0.7329 | 0.0751 |
| 2009 | 68.785 | 250 | 67.607 | 9 | 59.2670 | 0.9328 | 0.0788 |
| 2010 | 96.765 | 343 | 82.821 | 9 | 40.3732 | 0.8928 | 0.0662 |
| 2011 | 110.923 | 291 | 108.960 | 8 | 82.0762 | 1.3368 | 0.0706 |



Figure 13.52. Royal Red Prawn from zone 10 in depths $200-700 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Royal Red Prawn, the top right graph depicts the depth distribution of shots containing Royal Red Prawn from zone 10 in depths $200-700 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zone 10 , the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Royal Red Prawn catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.53. Royal Red Prawn from zone 10 in depths $200-700 \mathrm{~m}$ by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.43. Royal Red Prawn from zone 10 in depths 200 - 700m by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+DepCat |
| Model 3 | LnCE $\sim$ Year+DepCat+Vessel |
| Model 4 | LnCE $\sim$ Year+DepCat+Vessel+Month |
| Model 5 | LnCE $\sim$ Year+DepCat+Vessel+Month+DayNight |
| Model 6 | LnCE $\sim$ Year+DepCat+Vessel+Month+DayNight+DayNight:DepCat |
| Model 7 | LnCE $\sim$ Year+DepCat+Vessel+Month+DayNight+Month:DepCat |
| Model 8 | LnCE $\sim$ Year+DepCat+Vessel+Month+DayNight+DayNight:DepCat |

Table 13.44. Royal Red Prawn from zone 10 in depths 200 - 700m by trawl. Model selection criteria, including the AIC, the adjusted $r^{2}$ and the change in adjusted $r^{2}$. The optimum was model Month:DepCat.

|  | Year |  |  | DepCat | Vessel | Mth DayNight |  | DN:Mth |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mth:DepCat | DN:DepCat |  |  |  |  |  |  |  |
| AIC | 12994 | 8188 | 2947 | 1341 | 1250 | 1218 | 762 | 1205 |
| RSS | 40828 | 33110 | 26293 | 24530 | 24429 | 24327 | 23393 | 24232 |
| MSS | 1761 | 9479 | 16295 | 18059 | 18160 | 18262 | 19196 | 18356 |
| Nobs | 23586 | 23453 | 23453 | 23453 | 23453 | 23453 | 23453 | 23453 |
| Npars | 26 | 50 | 133 | 144 | 147 | 180 | 411 | 219 |
| adj_r2 | 4.033 | 22.094 | 37.913 | 42.050 | 42.281 | 42.440 | 44.096 | 42.568 |
| \%Change | 0.000 | 18.061 | 15.819 | 4.137 | 0.231 | 0.159 | 1.656 | -1.528 |



Figure 13.54. The relative influence of each factor used on the final trend in the optimal standardization for Royal Red Prawn in Zone 10. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.19.1 Comparison between Different Mesh Sizes

Royal Red Prawns are targeted with so-called prawn nets that are significantly smaller meshed than usual trawl nets (Figure 13.55). The smaller mesh nets, $<60 \mathrm{~mm}$, have significantly higher catch rates than larger mesh nets with rates between 4 and 6 times higher than meshes $>80 \mathrm{~mm}$ (Figure 13.56).


Figure 13.55. A scatter plot depicting the size distribution of reported mesh sizes used when Royal Red Prawn were landed. Each individual operation has been varied slightly (jittered) so as to illustrate the concentrations of mesh size and related catch rates. Thus, there are concentrations around $40-45 \mathrm{~mm}$, another at around $60-65 \mathrm{~mm}$, and another around 90 mm . The three red lines depict the average log catch rates for the three clusters of data points, with the larger values relating to the clusters left to right.


Figure 13.56. The geometric mean catch rates ( $\mathrm{kg} / \mathrm{hr)}$ ) of shots using nets with meshes less than 60 mm and those with meshes greater than 80 mm .

While it is undoubtedly true that the absolute catch rates of Royal Red Prawn are much higher than when fishing with a normal trawl net the important aspect for the assessment of the relative abundance through time is any trend in the catch rates through time. By re-scaling all catch rates to a mean of 1.0 the trends in the catch rates from the different data sets can be directly compared Figure 13.57.

The trends exhibited by the different data selections are noisy but essentially track a similar path. The optimum model, that uses all available data but doesn't distinguish between mesh sizes, is less variable than the smaller and larger mesh categories. Nevertheless, all series exhibit a rise between 2006 and 2007, and a further rise from
average in 2011. These trend lines are clearly noisy about the average in each case, with the variation being greater from $>=80 \mathrm{~mm},<60 \mathrm{~mm}$, and the Optimum series, which reflects the number of records in each data set. The conclusion remains that the use of the total dataset provides a good representation of the changes in the catch rates and can continue to be used in the Tier 4 assessment.


Figure 13.57. A comparison of the standardized catch rates, scaled to a mean of 1.0 , for the optimum model using all data, and the separate statistical models for those data from meshes $<60 \mathrm{~mm}$ and those $>$ 80 mm . The CV for the optimum series is $17.8 \%$, for $<60 \mathrm{~mm}$ it is $26.9 \%$ and for the $>=80 \mathrm{~mm}$ it is $37.8 \%$. The CVs in each case also reflect the relative amount of data available in each analysis.

Table 13.45. The scaled standardized catch rate data for different data selections. The Geomean and Optimum relate to all available data, $<60$ relates to data from net meshes $<60 \mathrm{~mm},>=80$ relate to data from meshes $>=80 \mathrm{~mm}$, and BothMeshes relates to a standardization that includes data from both $<60$ and $>=80 \mathrm{~mm}$, using mesh size as a factor in the standardization, where it is, not surprisingly, highly influential. Columns 3-5 contain the data plotted in Figure 13.57.

| Year | GeoMean | Optimum | $<60$ | $>=80$ | BothMeshes |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2002 | 0.93993 | 1.02359 | 0.64274 | 0.94438 | 1.03085 |
| 2003 | 0.77881 | 1.06722 | 0.67969 | 0.79677 | 0.98518 |
| 2004 | 0.98991 | 1.09235 | 0.99474 | 0.49541 | 1.03903 |
| 2005 | 0.92427 | 0.99593 | 0.76834 | 0.86704 | 0.94001 |
| 2006 | 1.07886 | 1.19880 | 1.08025 | 1.55634 | 1.15983 |
| 2007 | 0.95732 | 0.82842 | 1.53417 | 1.54000 | 1.02956 |
| 2008 | 0.76665 | 0.71378 | 1.12281 | 0.76311 | 0.82392 |
| 2009 | 1.16248 | 0.90847 | 0.91818 | 0.93033 | 0.80293 |
| 2010 | 0.79189 | 0.86951 | 1.02308 | 0.66411 | 0.88371 |
| 2011 | 1.60987 | 1.30193 | 1.23600 | 1.44252 | 1.30499 |

### 13.20 Blue Eye, Z2030 (TBE - 37445001 - Hyperoglyphe antarctica)

Trawling data from zones 20 and 30, depths less than 1000 m .
Table 13.46. BlueEye from zones 20 and 30 in depths $0-1000 \mathrm{~m}$ by trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr)}$ ). Zone:DepCat is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:DepC | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 37.962 | 166 | 9.117 | 17 | 10.0553 | 2.0191 | 0.0000 |
| 1987 | 15.495 | 190 | 10.026 | 14 | 9.8390 | 1.8963 | 0.1371 |
| 1988 | 105.177 | 307 | 19.433 | 21 | 14.4132 | 2.3535 | 0.1297 |
| 1989 | 88.066 | 315 | 33.371 | 32 | 14.6333 | 2.6517 | 0.1321 |
| 1990 | 79.298 | 264 | 39.845 | 36 | 24.1892 | 3.3877 | 0.1347 |
| 1991 | 76.024 | 474 | 29.189 | 37 | 9.3594 | 1.8116 | 0.1268 |
| 1992 | 49.305 | 313 | 14.232 | 23 | 8.3976 | 1.3516 | 0.1339 |
| 1993 | 59.654 | 736 | 37.789 | 31 | 7.9893 | 1.0762 | 0.1239 |
| 1994 | 109.975 | 855 | 89.033 | 33 | 10.7324 | 1.2447 | 0.1232 |
| 1995 | 58.572 | 489 | 28.335 | 29 | 5.8281 | 0.8326 | 0.1279 |
| 1996 | 71.684 | 648 | 35.518 | 29 | 5.7645 | 0.6522 | 0.1254 |
| 1997 | 470.716 | 604 | 19.921 | 31 | 4.6731 | 0.5711 | 0.1269 |
| 1998 | 475.965 | 475 | 18.704 | 24 | 4.1103 | 0.6687 | 0.1295 |
| 1999 | 574.484 | 633 | 41.733 | 27 | 3.5948 | 0.6835 | 0.1260 |
| 2000 | 667.056 | 657 | 37.661 | 33 | 2.7104 | 0.4728 | 0.1245 |
| 2001 | 612.354 | 692 | 25.038 | 24 | 2.2460 | 0.4228 | 0.1248 |
| 2002 | 758.103 | 700 | 33.732 | 28 | 3.0245 | 0.4238 | 0.1266 |
| 2003 | 592.295 | 723 | 14.094 | 25 | 2.2565 | 0.4333 | 0.1258 |
| 2004 | 598.119 | 623 | 15.172 | 28 | 2.7233 | 0.4238 | 0.1275 |
| 2005 | 455.408 | 502 | 17.920 | 26 | 2.6096 | 0.4175 | 0.1307 |
| 2006 | 573.719 | 327 | 36.782 | 17 | 3.9462 | 0.5129 | 0.1349 |
| 2007 | 631.172 | 248 | 10.641 | 11 | 3.1268 | 0.4084 | 0.1404 |
| 2008 | 337.335 | 434 | 13.654 | 15 | 5.6341 | 0.3804 | 0.1344 |
| 2009 | 443.095 | 246 | 22.849 | 14 | 5.4891 | 0.3759 | 0.1420 |
| 2010 | 385.706 | 199 | 11.939 | 13 | 3.5048 | 0.2615 | 0.1473 |
| 2011 | 517.919 | 228 | 7.870 | 12 | 2.2147 | 0.2667 | 0.1442 |



Figure 13.58. BlueEye from zones 20 and 30 in depths $0-1000 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting BlueEye, the top right graph depicts the depth distribution of shots containing BlueEye from zones 20 and 30 in depths $0-1000 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zones 20 and 30 ( 30 is top red line), the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the BlueEye catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.59. BlueEye from zones 20 and 30 in depths $0-1000 \mathrm{~m}$ by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.47. BlueEye from zones 20 and 30 in depths $0-1000 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel+Zone |
| Model 4 | LnCE $\sim$ Year+Vessel+Zone+DepCat |
| Model 5 | LnCE $\sim$ Year+Vessel+Zone+DepCat+DayNight |
| Model 6 | LnCE $\sim$ Year+Vessel+Zone+DepCat+DayNight+Month |
| Model 7 | LnCE $\sim$ Year+Vessel+Zone+DepCat+DayNight+Month+Zone:Month |
| Model 8 | LnCE $\sim$ Year+Vessel+Zone+DepCat+DayNight+Month+Zone:DepCat |

Table 13.48. BlueEye from zones 20 and 30 in depths $0-1000 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the deviance and the change in deviance. The optimum was model Zone:DepCat.

|  | Year |  | Vessel | Zone | DepCat |  | DayNight | Month | Zone:Mth Zone:DepC |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: |
| AIC | 10592 | 4594 | 4201 | 4083 | 3971 | 3956 | 3938 | 3765 |  |  |
| RSS | 28898 | 17227 | 16671 | 16308 | 16149 | 16099 | 16045 | 15718 |  |  |
| MSS | 4590 | 16261 | 16817 | 17180 | 17339 | 17389 | 17443 | 17770 |  |  |
| Nobs | 12048 | 12048 | 12048 | 11975 | 11975 | 11975 | 11975 | 11975 |  |  |
| Npars | 26 | 143 | 144 | 192 | 195 | 206 | 217 | 254 |  |  |
| adj_r2 | 13.528 | 47.944 | 49.619 | 50.512 | 50.982 | 51.089 | 51.207 | 52.050 |  |  |
| \%Change | 0.000 | 34.416 | 1.675 | 0.893 | 0.470 | 0.107 | 0.118 | 0.843 |  |  |



Figure 13.60. The relative influence of each factor used on the final trend in the optimal standardization for BlueEye in Zones $20-30$. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.21 Blue Eye, Z4050 (TBE - 37445001 - H. antarctica)

Data from zones 40 and 50, depths less than 1000 m .
Table 13.49. BlueEye from Zones 40 and 50 in depths 0 to 1000 m by Trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates (kg/hr). Zone:Month is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 37.962 | 194 | 15.955 | 18 | 13.1296 | 0.9050 | 0.0000 |
| 1987 | 15.495 | 56 | 3.145 | 14 | 11.6895 | 0.7678 | 0.1748 |
| 1988 | 105.177 | 142 | 76.410 | 15 | 41.5696 | 2.2576 | 0.1554 |
| 1989 | 88.066 | 238 | 43.985 | 24 | 25.5841 | 1.8566 | 0.1371 |
| 1990 | 79.298 | 157 | 30.910 | 16 | 13.0702 | 1.9495 | 0.1575 |
| 1991 | 76.024 | 129 | 18.954 | 18 | 17.4424 | 1.6212 | 0.1558 |
| 1992 | 49.305 | 129 | 28.643 | 15 | 21.8842 | 1.8823 | 0.1556 |
| 1993 | 59.654 | 289 | 18.109 | 19 | 8.5334 | 0.8648 | 0.1394 |
| 1994 | 109.975 | 348 | 16.282 | 19 | 8.8991 | 0.9188 | 0.1360 |
| 1995 | 58.572 | 500 | 26.381 | 21 | 6.4723 | 0.8197 | 0.1323 |
| 1996 | 71.684 | 523 | 30.184 | 24 | 8.0361 | 0.8539 | 0.1328 |
| 1997 | 470.716 | 788 | 82.371 | 18 | 6.5139 | 0.8725 | 0.1295 |
| 1998 | 475.965 | 780 | 58.946 | 19 | 5.3540 | 1.0375 | 0.1307 |
| 1999 | 574.484 | 877 | 46.303 | 19 | 6.4046 | 1.0684 | 0.1298 |
| 2000 | 667.056 | 1109 | 44.729 | 22 | 5.2927 | 0.9274 | 0.1290 |
| 2001 | 612.354 | 955 | 42.188 | 26 | 5.7866 | 0.8756 | 0.1306 |
| 2002 | 758.103 | 802 | 32.268 | 26 | 5.0532 | 0.7333 | 0.1307 |
| 2003 | 592.295 | 392 | 11.023 | 25 | 3.1895 | 0.6717 | 0.1372 |
| 2004 | 598.119 | 852 | 31.296 | 24 | 4.2166 | 0.5914 | 0.1309 |
| 2005 | 455.408 | 508 | 12.750 | 22 | 3.6280 | 0.5420 | 0.1340 |
| 2006 | 573.719 | 533 | 16.279 | 17 | 3.6218 | 0.5591 | 0.1337 |
| 2007 | 631.172 | 538 | 26.188 | 16 | 4.4303 | 0.5942 | 0.1337 |
| 2008 | 337.335 | 324 | 16.371 | 14 | 4.9605 | 0.7664 | 0.1388 |
| 2009 | 443.095 | 343 | 15.751 | 13 | 4.0530 | 0.7055 | 0.1385 |
| 2010 | 385.706 | 430 | 31.436 | 14 | 5.5190 | 0.7466 | 0.1358 |
| 2011 | 517.919 | 381 | 14.696 | 14 | 2.8213 | 0.6110 | 0.1370 |



Figure 13.61. BlueEye from Zones 40 and 50 in depths 0 to 1000 m by Trawl. The top left is the depth distribution of all records reporting BlueEye, the top right graph depicts the depth distribution of shots containing BlueEye from Zones 40 and 50 in depths 0 to 1000 m by Trawl. The middle left diagram depicts the distribution of catch by depth within zones 40 and 50 ( 50 is top red line), the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the BlueEye catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.62. BlueEye from Zones 40 and 50 in depths 0 to 1000 m by Trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.50. BlueEye from Zones 40 and 50 in depths 0 to 1000 m by Trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel+DepCat |
| Model 4 | LnCE $\sim$ Year+Vessel+DepCat + DayNight |
| Model 5 | LnCE $\sim$ Year+Vessel+DepCat + DayNight + Month |
| Model 6 | LnCE $\sim$ Year+Vessel+DepCat + DayNight + Month + Zone |
| Model 7 | LnCE $\sim$ Year+Vessel+DepCat + DayNight + Month + Zone + Zone: Month |
| Model 8 | LnCE $\sim$ Year+Vessel+DepCat + DayNight + Month + Zone + Zone: DepCat |

Table 13.51. BlueEye from Zones 40 and 50 in depths 0 to 1000 m by Trawl. Model selection criteria, including the AIC, the adjusted $\mathrm{r}^{2}$ and the change in adjusted $\mathrm{r}^{2}$. The optimum was model Zone:DepCat.

|  | Year | Vessel | DepCat | DayNight | Month | Zone | Zone:Mth |  | Zone:DepC |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 8034 | 2689 | 2278 | 2017 | 1974 | 1927 | 1911 | 1903 |  |
| RSS | 23547 | 15061 | 14399 | 14090 | 14015 | 13959 | 13917 | 13821 |  |
| MSS | 2683 | 11169 | 11831 | 12141 | 12216 | 12272 | 12314 | 12409 |  |
| Nobs | 12317 | 12317 | 12265 | 12265 | 12265 | 12265 | 12265 | 12265 |  |
| Npars | 26 | 106 | 155 | 158 | 169 | 170 | 181 | 219 |  |
| adj_r2 | 10.047 | 42.088 | 44.406 | 45.587 | 45.829 | 46.040 | 46.155 | 46.354 |  |
| \%Change | 0.000 | 32.040 | 2.319 | 1.181 | 0.242 | 0.211 | 0.115 | 0.199 |  |



Figure 13.63. The relative influence of each factor used on the final trend in the optimal standardization for BlueEye in Zones $40-50$. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.22 Blue Eye, AL (TBE - 37445001 - H. antarctica)

Depths between $200-600 \mathrm{~m}$. All data from auto-longlining. 1997 was omitted as being unrepresentative due to very lower numbers of records.

Table 13.52. BlueEye from the SESSF in depths $200-600 \mathrm{~m}$ by AutoLongLine. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr)}$ ) Zone:DepCat is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:DepC | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1998 | 475.965 | 28 | 14.989 | 2 | 249.6862 | 0.5615 | 0.0000 |
| 1999 | 574.484 | 50 | 47.670 | 2 | 536.1933 | 1.9718 | 0.3284 |
| 2000 | 667.056 | 29 | 28.299 | 2 | 608.0267 | 1.5945 | 0.3624 |
| 2001 | 612.354 | 65 | 40.232 | 2 | 246.5002 | 0.8418 | 0.3151 |
| 2002 | 758.103 | 228 | 131.686 | 4 | 162.2961 | 0.7989 | 0.2880 |
| 2003 | 592.295 | 434 | 157.016 | 7 | 133.4303 | 1.1368 | 0.2818 |
| 2004 | 598.119 | 1147 | 269.120 | 11 | 72.0019 | 1.1079 | 0.2776 |
| 2005 | 455.408 | 1137 | 300.462 | 7 | 77.8010 | 0.8835 | 0.2778 |
| 2006 | 573.719 | 1067 | 345.481 | 9 | 102.2372 | 0.9318 | 0.2767 |
| 2007 | 631.172 | 658 | 453.819 | 6 | 364.8943 | 1.1963 | 0.2787 |
| 2008 | 337.335 | 604 | 277.917 | 6 | 232.1695 | 0.8064 | 0.2789 |
| 2009 | 443.095 | 550 | 313.987 | 6 | 289.6046 | 0.9263 | 0.2784 |
| 2010 | 385.706 | 483 | 230.042 | 5 | 184.8051 | 0.5961 | 0.2795 |
| 2011 | 517.919 | 526 | 225.716 | 5 | 209.8939 | 0.6464 | 0.2792 |



Figure 13.64. BlueEye from the SESSF in depths $200-600 \mathrm{~m}$ by Auto-LongLine. The top left is the depth distribution of all records reporting BlueEye, the top right graph depicts the depth distribution of shots containing BlueEye from the SESSF in depths $200-600 \mathrm{~m}$ by Auto-LongLine. The middle left diagram depicts the distribution of catch by depth across the zones of the SESSF, the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the BlueEye catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.65. BlueEye from the SESSF in depths $200-600 \mathrm{~m}$ by Auto-LongLine. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.53. BlueEye from the SESSF in depths 200 - 600 m by Auto-LongLine. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE~Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel+Month |
| Model 4 | LnCE $\sim$ Year+Vessel+Month+Zone |
| Model 5 | LnCE $\sim$ Year+Vessel+Month+Zone+DayNight |
| Model 6 | LnCE $\sim$ Year+Vessel+Month+Zone+DayNight+DepCat |
| Model 7 | LnCE $\sim$ Year+Vessel+Month+Zone+DayNight+DepCat+Zone:Month |
| Model 8 | LnCE $\sim$ Year+Vessel+Month+Zone+DayNight+DepCat+Zone:DepCat |

Table 13.54. BlueEye from the SESSF in depths $200-600 \mathrm{~m}$ by Auto-LongLine. Model selection criteria, including the AIC, the adjusted $\mathrm{r}^{2}$ and the change in adjusted $\mathrm{r}^{2}$. The optimum was model Zone:Month.

|  | Year | Vessel | Month | Zone | DayNight | DepCat | Zone:Mth |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |



Figure 13.66. The relative influence of each factor used on the final trend in the optimal standardization for BlueEye in by Auto-Longline. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.23 Blue Eye, DL (TBE - 37445001 - H. antarctica)

Depths between 200-600m. All data from Drop-lining. All vessels reporting blue eye by drop line are included. There are records following 2005 but there does not appear to be any depth data associated with Drop Line records.

[^1]| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1997 | 470.091 | 515 | 248.463 | 38 | 269.5548 | 1.6934 | 0.0000 |
| 1998 | 475.965 | 700 | 320.649 | 30 | 234.7678 | 1.2383 | 0.0769 |
| 1999 | 574.484 | 868 | 338.923 | 29 | 182.2847 | 1.0864 | 0.0794 |
| 2000 | 667.056 | 1056 | 377.828 | 33 | 172.0296 | 1.0734 | 0.0817 |
| 2001 | 612.354 | 739 | 318.120 | 26 | 200.3223 | 1.1806 | 0.0856 |
| 2002 | 758.103 | 570 | 180.454 | 22 | 164.7123 | 0.9831 | 0.0908 |
| 2003 | 592.295 | 533 | 167.639 | 22 | 162.1210 | 0.8273 | 0.0945 |
| 2004 | 598.119 | 484 | 148.266 | 23 | 161.8513 | 0.9684 | 0.0972 |
| 2005 | 455.408 | 338 | 79.885 | 16 | 133.7709 | 0.7652 | 0.1064 |
| 2006 | 573.719 | 303 | 104.599 | 13 | 224.9012 | 1.0556 | 0.1143 |
| 2007 | 631.172 | 125 | 45.301 | 10 | 213.7543 | 1.3748 | 0.1411 |
| 2008 | 337.335 | 77 | 15.679 | 7 | 137.5913 | 0.8192 | 0.1612 |
| 2009 | 443.095 | 81 | 17.818 | 9 | 124.4663 | 0.5704 | 0.1729 |
| 2010 | 385.706 | 191 | 28.218 | 9 | 77.7373 | 0.5393 | 0.1405 |
| 2011 | 517.919 | 166 | 32.368 | 9 | 104.9216 | 0.8247 | 0.1513 |



Figure 13.67. BlueEye catches by zone from the SESSF in depths $200-600 \mathrm{~m}$ by DropLine.


Figure 13.68. BlueEye from the SET and GHT fishery in depths between $200-600 \mathrm{~m}$, taken by Drop Line. The top left is the depth distribution of all records reporting BlueEye, the top right graph depicts the depth distribution of shots containing BlueEye from the SET and GHT fishery in depths between 200 600 m , taken by Drop Line. The middle left diagram depicts the distribution of catch by depth by SESSF zone, the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the BlueEye catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.69. BlueEye from the SEN and GHT fishery in depths between $200-600 \mathrm{~m}$, taken by Drop Line. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.56. BlueEye from the SET and GHT fishery in depths between $200-600 \mathrm{~m}$, taken by Drop Line. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE~Year |
| :--- | :--- |
| Model 2 | LnCE~Year+Vessel |
| Model 3 | LnCE~Year+Vessel+Month |
| Model 4 | LnCE~Year+Vessel+Month+Zone |
| Model 5 | LnCE $\sim$ Year+Vessel+Month+Zone+DayNight |
| Model 6 | LnCE $\sim$ Year+Vessel+Month+Zone+DayNight+ Zone:Month |

Table 13.57. BlueEye from the SET and GHT fishery in depths between $200-600 \mathrm{~m}$, taken by Drop Line. Model selection criteria, including the AIC, the adjusted $\mathrm{r}^{2}$ and the change in adjusted $\mathrm{r}^{2}$. The optimum is model 7.

|  | Year | Vessel | Month | Zone | DayNight | Zone:Mth |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 3949 | 2904 | 2533 | 2513 | 2502 | 2482 |
| RSS | 12060 | 10063 | 9494 | 9457 | 9438 | 9381 |
| MSS | 408 | 2405 | 2973 | 3010 | 3029 | 3087 |
| Nobs | 6746 | 6746 | 6746 | 6746 | 6746 | 6746 |
| Npars | 15 | 103 | 114 | 117 | 118 | 129 |
| adj_r2 | 3.070 | 18.047 | 22.550 | 22.816 | 22.959 | 23.302 |
| \%Change | 0.000 | 14.977 | 4.502 | 0.266 | 0.143 | 0.344 |

### 13.24 Blue Eye, AL \& DL (TBE - 37445001 - H. antarctica)

Depths between $200-600 \mathrm{~m} \mathrm{~m}$. All data from auto-longlining and droplining combined. Zones $20-50$, and $83-85$ included ( $83-85$ are in the GAB ).

Table 13.58. BlueEye from the SEN and GHT in depths $200-600 \mathrm{~m}$ by Auto Long Line and Drop Line. Total Catch is the total reported in the database, Records is the number of records used in the analysis, Catch T is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr}$ ). Zone:Month is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1997 | 470.716 | 547 | 254.786 | 39 | 258.2795 | 1.8553 | 0.0000 |
| 1998 | 475.965 | 758 | 340.896 | 30 | 226.1524 | 1.2439 | 0.0795 |
| 1999 | 574.484 | 927 | 387.055 | 30 | 189.1263 | 1.1126 | 0.0815 |
| 2000 | 667.056 | 1086 | 406.152 | 34 | 177.6127 | 1.0566 | 0.0843 |
| 2001 | 612.354 | 807 | 358.910 | 27 | 202.9873 | 1.1035 | 0.0873 |
| 2002 | 758.103 | 799 | 312.210 | 24 | 163.8436 | 0.8559 | 0.0889 |
| 2003 | 592.295 | 969 | 324.984 | 25 | 148.5823 | 0.9295 | 0.0883 |
| 2004 | 598.119 | 1638 | 418.296 | 29 | 91.4807 | 1.0215 | 0.0874 |
| 2005 | 455.408 | 1479 | 381.087 | 23 | 88.2645 | 0.8194 | 0.0899 |
| 2006 | 573.719 | 1368 | 447.130 | 19 | 121.2856 | 0.9396 | 0.0900 |
| 2007 | 631.172 | 783 | 498.943 | 15 | 333.7817 | 1.1791 | 0.0959 |
| 2008 | 337.335 | 684 | 293.497 | 13 | 214.3734 | 0.7867 | 0.0975 |
| 2009 | 443.095 | 631 | 331.806 | 15 | 259.8521 | 0.8701 | 0.0977 |
| 2010 | 385.706 | 680 | 259.006 | 14 | 142.9654 | 0.5634 | 0.0978 |
| 2011 | 517.919 | 692 | 258.084 | 14 | 177.7306 | 0.6629 | 0.0976 |



Figure 13.70. BlueEye from the SEN and GHT in depths $200-600 \mathrm{~m}$ by Auto Long Line and Drop Line. The top left is the depth distribution of all records reporting BlueEye, the top right graph depicts the depth distribution of shots containing BlueEye from the SEN and GHT in depths $200-600 \mathrm{~m}$ by Auto Long Line and Drop Line. The middle left diagram depicts the distribution of catch by depth, the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the BlueEye catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.71. BlueEye from the SEN and GHT in depths $200-600 \mathrm{~m}$ by Auto Long Line and Drop Line. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates

Table 13.59. BlueEye from the SEN and GHT in depths $200-600 \mathrm{~m}$ by Auto Long Line and Drop Line. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE~Year |
| :--- | :--- |
| Model 2 | LnCE~Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel+DepCat |
| Model 4 | LnCE $\sim$ Year+Vessel+DepCat+Month |
| Model 5 | LnCE $\sim$ Year+Vessel+DepCat+Month+DayNight |
| Model 6 | LnCE~Year+Vessel+DepCat+Month+DayNight+Zone |
| Model 7 | LnCE~Year+Vessel+DepCat+Month+DayNight+Zone+Method |
| Model 8 | LnCE $\sim$ Year+Vessel+DepCat+Month+DayNight+Zone+Method+Zone:Month |
| Model 9 | LnCE $\sim$ Year+Vessel+DepCat+Month+DayNight+Zone+Method+Zone:DepCat |
| Model 10 | LnCE $\sim$ Year+Vessel+DepCat+Month+DayNight+Zone+Method+Zone:Method |
| Model 11 | LnCE $\sim$ Year+Vessel+DepCat+Month+DayNight+Zone+Method+Month:Method |

Table 13.60. BlueEye from the SEN and GHT in depths $200-600 \mathrm{~m}$ by Auto Long Line and Drop Line. Model selection criteria, including the AIC, the adjusted $\mathrm{r}^{2}$ and the change in adjusted $\mathrm{r}^{2}$. The optimum is model Zone, though Zone:Month is very close. DepC is Depth Category, Mth is Month, DN is DayNight, Meth is Method and Zon is Zone.

|  | Year Vessel | DepC | Mth | DN | Zone | Meth | Zon:Mth Zon:DepC Zon:Meth | Mth:Meth |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 11057 | 7575 | 7133 | 6388 | 6372 | 6191 | 6221 | 6192 | 6207 | 6246 | 6285 |
| RSS | 30709 | 23556 | 21896 | 20624 | 20587 | 20281 | 20280 | 20200 | 20194 | 20272 | 19863 |
| MSS | 2086 | 9239 | 10898 | 12171 | 12207 | 12514 | 12515 | 12595 | 12600 | 12522 | 12932 |
| Nobs | 13855 | 13855 | 12802 | 12802 | 12802 | 12784 | 12784 | 12784 | 12784 | 12784 | 12784 |
| Npars | 15 | 111 | 131 | 142 | 145 | 146 | 147 | 172 | 181 | 162 | 326 |
| adj_r2 | 6.265 | 27.597 | 32.547 | 36.412 | 36.509 | 37.449 | 37.446 | 37.570 | 37.542 | 37.396 | 37.852 |
| \%Change | 0 | 21.332 | 4.950 | 3.865 | 0.096 | 0.940 | -0.003 | 0.124 | -0.028 | -0.146 | 0.456 |



Figure 13.72. The relative influence of each factor used on the final trend in the optimal standardization for BlueEye by AL and DL. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.24.1 Preliminary Consideration of Closure Effects on CPUE

One issue relating to the impacts of closures, for which a working solution has still to be found, is the effect of Industry members fishing close along the edges of closures. While there is nothing wrong with fishing in this manner the resolution of the GPS data entered in to the catch effort database is sometimes such that an array of reported shots appear to have been conducted inside the closures. Given the use of VMS in all cases these locations are assumed to be a result of small rounding errors when reporting locations.

In order to validly compare the older catch rates within a closure and the catch rates outside it is currently necessary to consider the data shot-by-shot and decide whether those on the edges and just inside the boundary of closures are to be treated as if they were made inside or outside. Owing to the scale of this only the St Helens Hill closure has been examined in the first instance (Figure 13.73).


Figure 13.73. A schematic map illustrating the location and extent of the St Helens Hill closure on the east coast of Tasmania. Since the introduction of the closure in 2007 catches have been very much reduced inside the closure although there has been some confusion brought about by what is assumed to be rounding errors on the reporting of GPS locations (Table 13.61).

While there was a surge of catches in 2007, since then catches have been reduced relative to those taken in the general area in earlier years.

Table 13.61. Catches of BlueEye in the vicinity of the St Helens Closure. Actually Outside relates to the catches reported that came from the edge or seemingly just inside the closure (these are the sum of the Apparently Outside and the Inside Edge of Closure). Actually Inside relates to the catches that are inferred to have in fact been taken inside the closure.

| Year | Actually <br> Outside | Apparently <br> Outside | Actually <br> Inside | Inside Edge <br> of Closure | Apparently <br> Inside |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1997 | 3.155 | 3.155 | 12.735 |  | 12.735 |
| 1998 | 17.179 | 16.179 | 21.428 | 1.000 | 22.428 |
| 1999 | 18.109 | 17.739 | 46.907 | 0.370 | 47.277 |
| 2000 | 4.898 | 4.573 | 61.971 | 0.325 | 62.296 |
| 2001 | 10.627 | 6.187 | 62.234 | 4.440 | 66.674 |
| 2002 | 18.415 | 5.990 | 38.205 | 12.425 | 50.630 |
| 2003 | 8.011 | 5.897 | 40.758 | 2.114 | 42.872 |
| 2004 | 5.597 | 3.329 | 37.070 | 2.268 | 39.338 |
| 2005 | 2.518 | 2.218 | 18.827 | 0.300 | 19.127 |
| 2006 | 6.948 | 6.728 | 10.120 | 0.220 | 10.340 |
| 2007 | 52.284 | 40.600 | 0.823 | 11.684 | 12.507 |
| 2008 | 15.156 | 13.764 | 1.100 | 1.392 | 2.492 |
| 2009 | 14.001 | 10.217 | 0.308 | 3.784 | 4.092 |
| 2010 | 14.036 | 9.619 | 1.793 | 4.417 | 6.210 |
| 2011 | 17.358 | 16.947 |  | 0.411 | 0.411 |



Figure 13.74. The catches and geometric mean catch-rates for both inside and outside the St Helens Closure. The average catch rate inside and outside (the grey lines) were $418 \mathrm{~kg} / \mathrm{hr}$ relative to $215 \mathrm{~kg} / \mathrm{hr}$. There were only 6 records for the years 2007 - 2011 inside the closure so those catch rates are omitted as being accidentally high.
The catch rates on the edge of the closure relative to those inside were almost $50 \%$ lower, albeit rather more variable. At least in and around the St Helens closure the closure does appear to have led to the appearance that catch rates have declined. Of course this analysis cannot be certain that catch rates will be better on the St Helens hill, as in the past, and the only solid way to determine this would be to sample within the closure.

### 13.25 Blue Grenadier Spawning (GRE - 37227001 Macruronus novaezelandiae)

Data from Zone 40 in months June to August, depths between 100 and 1000m.
Table 13.62. Blue Grenadier from Zone 40 between June and August in depths between $100-1000 \mathrm{~m}$, taken by Trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr}$ ). Model 5 is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | DayNight | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 1451.778 | 89 | 237.730 | 8 | 252.1522 | 1.0104 | 0.0000 |
| 1987 | 2244.828 | 205 | 780.225 | 12 | 444.9725 | 1.2815 | 0.3107 |
| 1988 | 1849.147 | 92 | 319.022 | 8 | 387.7497 | 2.5819 | 0.3208 |
| 1989 | 1890.855 | 31 | 36.120 | 4 | 125.3994 | 0.6686 | 0.4459 |
| 1990 | 2280.471 | 158 | 565.635 | 10 | 268.6781 | 0.6558 | 0.3078 |
| 1991 | 3669.036 | 112 | 618.606 | 14 | 745.6978 | 2.5251 | 0.3500 |
| 1992 | 2474.546 | 152 | 500.586 | 10 | 484.1498 | 1.1723 | 0.3312 |
| 1993 | 2482.270 | 181 | 789.508 | 14 | 527.5902 | 1.7533 | 0.3289 |
| 1994 | 2315.490 | 323 | 974.288 | 17 | 311.1013 | 0.9693 | 0.3155 |
| 1995 | 1931.046 | 478 | 908.891 | 15 | 65.6455 | 0.4066 | 0.3138 |
| 1996 | 2304.234 | 497 | 1198.174 | 14 | 71.7968 | 0.6127 | 0.3209 |
| 1997 | 3654.679 | 555 | 2622.788 | 11 | 114.6872 | 0.4675 | 0.3204 |
| 1998 | 4226.177 | 581 | 2704.903 | 18 | 127.2732 | 0.5976 | 0.3154 |
| 1999 | 7572.998 | 1058 | 5441.838 | 14 | 359.1395 | 0.4746 | 0.3132 |
| 2000 | 7503.140 | 945 | 5627.807 | 15 | 299.1618 | 0.5129 | 0.3133 |
| 2001 | 8370.799 | 1089 | 7281.264 | 15 | 482.7733 | 0.7799 | 0.3130 |
| 2002 | 7978.310 | 1038 | 6825.011 | 14 | 333.3040 | 0.5865 | 0.3129 |
| 2003 | 7948.324 | 1054 | 7201.361 | 17 | 620.3759 | 0.5402 | 0.3141 |
| 2004 | 6093.498 | 825 | 4614.533 | 15 | 225.8117 | 0.4352 | 0.3154 |
| 2005 | 4506.740 | 417 | 2845.032 | 11 | 488.4239 | 0.9876 | 0.3226 |
| 2006 | 3544.354 | 470 | 2034.825 | 13 | 475.8843 | 1.7716 | 0.3240 |
| 2007 | 3128.212 | 306 | 1738.986 | 8 | 270.5783 | 1.2380 | 0.3332 |
| 2008 | 4152.329 | 293 | 2808.682 | 10 | 737.4291 | 0.8791 | 0.3279 |
| 2009 | 3874.668 | 349 | 2704.191 | 9 | 620.2829 | 1.1669 | 0.3248 |
| 2010 | 4552.385 | 456 | 3362.757 | 6 | 681.3735 | 0.9120 | 0.3218 |
| 2011 | 4476.805 | 484 | 3527.195 | 10 | 387.0848 | 1.0127 | 0.3191 |



Figure 13.75. Blue Grenadier from Zone 40 between June and August in depths between $100-1000 \mathrm{~m}$, taken by Trawl. The top left is the depth distribution of all records reporting Blue Grenadier, the top right graph depicts the depth distribution of shots containing Blue Grenadier from Zone 40 between June and August in depths between $100-1000 \mathrm{~m}$, taken by Trawl. The middle left diagram depicts the distribution of catch by depth by SESSF zone, the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Blue Grenadier catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.76. Blue Grenadier from Zone 40 between June and August in depths between $100-1000 \mathrm{~m}$, taken by Trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.63. Blue Grenadier from Zone 40 between June and August in depths between $100-1000 \mathrm{~m}$, taken by Trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.
Model $1 \quad$ LnCE $\sim$ Year
Model 2 LnCE~Year+Vessel
Model 3 LnCE~Year+Vessel+DepCat
Model 4 LnCE~Year+Vessel+DepCat+Weeknum
Model 5 LnCE~Year+Vessel+DepCat+Weeknum+DayNight
Model 6 LnCE~Year+Vessel+DepCat+Weeknum+DayNight+Weeknum:DepCat

Table 13.64. Blue Grenadier from Zone 40 between June and August in depths between $100-1000 \mathrm{~m}$, taken by Trawl. Model selection criteria, including the AIC, the deviance and the change in deviance. The optimum is model DayNight (model 5).

|  | Year | Vessel | DepCat | Weeknum | DayNight | Weeknum:DepCat |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 27246 | 18472 | 16740 | 16026 | 15508 | 15733 |
| RSS | 112915 | 54485 | 47096 | 44312 | 42438 | 39430 |
| MSS | 5158 | 63588 | 70977 | 73761 | 75635 | 78643 |
| Nobs | 12238 | 12238 | 12141 | 12141 | 12141 | 12141 |
| Npars | 26 | 98 | 141 | 154 | 157 | 716 |
| adj_r2 | 4.173 | 53.486 | 59.647 | 61.992 | 63.590 | 64.515 |
| \%Change | 0.000 | 49.313 | 6.161 | 2.345 | 1.598 | 0.925 |



Figure 13.77. The relative influence of each factor used on the final trend in the optimal standardization for Blue Grenadier spawning fishery. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.26 Blue Grenadier Non-Spawning (GRE - 37227001 - M. novaezelandiae)

Data from zones 10 to 60 except Zone 40 in months June to August, depths less than 1000 m and greater than 0 m .

Table 13.65. Blue Grenadier from the SET in depths between $200-600 \mathrm{~m}$, taken by Trawl, omitting the Spawning fishery (zone 40 between June and August). Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr)}$. Zone:Month is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 1451.778 | 3189 | 1183.307 | 92 | 36.7375 | 1.4764 | 0.0000 |
| 1987 | 2244.828 | 3569 | 1437.434 | 91 | 37.3307 | 1.9414 | 0.0338 |
| 1988 | 1849.147 | 3961 | 1470.196 | 102 | 36.6778 | 2.1038 | 0.0339 |
| 1989 | 1890.855 | 4309 | 1813.501 | 99 | 45.3866 | 2.1808 | 0.0339 |
| 1990 | 2280.471 | 3577 | 1625.146 | 92 | 47.9497 | 2.1567 | 0.0358 |
| 1991 | 3669.036 | 4308 | 2392.687 | 86 | 48.2874 | 1.5519 | 0.0345 |
| 1992 | 2474.546 | 3228 | 1505.799 | 61 | 40.5408 | 1.2755 | 0.0367 |
| 1993 | 2482.270 | 4203 | 1619.049 | 63 | 33.2638 | 0.9613 | 0.0352 |
| 1994 | 2315.490 | 4491 | 1309.563 | 66 | 29.5414 | 0.8636 | 0.0347 |
| 1995 | 1931.046 | 5076 | 1015.261 | 61 | 19.4025 | 0.5978 | 0.0339 |
| 1996 | 2304.234 | 5370 | 1055.340 | 73 | 15.8910 | 0.5459 | 0.0338 |
| 1997 | 3654.679 | 6194 | 994.604 | 73 | 13.3293 | 0.5664 | 0.0333 |
| 1998 | 4226.177 | 6599 | 1452.552 | 65 | 18.8682 | 0.9286 | 0.0331 |
| 1999 | 7572.998 | 8045 | 2051.946 | 65 | 22.7820 | 0.9810 | 0.0324 |
| 2000 | 7503.140 | 7679 | 1751.230 | 69 | 16.8751 | 0.6986 | 0.0328 |
| 2001 | 8370.799 | 7279 | 1013.774 | 60 | 11.4735 | 0.3992 | 0.0332 |
| 2002 | 7978.310 | 6344 | 1125.943 | 57 | 13.3454 | 0.4006 | 0.0337 |
| 2003 | 7948.324 | 5675 | 670.745 | 56 | 10.1345 | 0.3344 | 0.0340 |
| 2004 | 6093.498 | 6393 | 1206.698 | 56 | 16.9690 | 0.5633 | 0.0338 |
| 2005 | 4506.740 | 5346 | 1174.711 | 54 | 19.8341 | 0.6724 | 0.0345 |
| 2006 | 3544.354 | 4362 | 1308.840 | 42 | 26.9839 | 0.8930 | 0.0356 |
| 2007 | 3128.212 | 3659 | 1204.518 | 27 | 25.1832 | 0.7950 | 0.0366 |
| 2008 | 4152.329 | 3407 | 1276.536 | 26 | 28.8353 | 0.8654 | 0.0372 |
| 2009 | 3874.668 | 3443 | 1128.896 | 23 | 25.9256 | 0.8035 | 0.0371 |
| 2010 | 4552.385 | 3308 | 1136.546 | 25 | 25.9279 | 0.7879 | 0.0375 |
| 2011 | 4476.805 | 3950 | 894.117 | 26 | 19.3008 | 0.6557 | 0.0367 |



Figure 13.78. Blue Grenadier from the SET in depths between $200-600 \mathrm{~m}$, taken by Trawl, omitting the Spawning fishery (zone 40 between June and August). The top left is the depth distribution of all records reporting Blue Grenadier, the top right graph depicts the depth distribution of shots containing Blue Grenadier from the SET in depths between $200-600 \mathrm{~m}$, taken by Trawl, omitting the Spawning fishery (zone 40 between June and August). The middle left diagram depicts the distribution of catch by depth by SESSF zone, the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Blue Grenadier catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<$ 30 Kg ).


Figure 13.79. Blue Grenadier from the SET in depths between $200-600 \mathrm{~m}$, taken by Trawl, omitting the Spawning fishery (zone 40 between June and August). The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.66. Blue Grenadier from the SET in depths between $200-600 \mathrm{~m}$, taken by Trawl, omitting the Spawning fishery (zone 40 between June and August). Statistical model structures used in this analysis.
DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel+DepCat |
| Model 4 | LnCE $\sim$ Year+Vessel+DepCat+Month |
| Model 5 | LnCE $\sim$ Year+Vessel+DepCat+Month+Zone |
| Model 6 | LnCE $\sim$ Year+Vessel+DepCat+Month+Zone+DayNight |
| Model 7 | LnCE $\sim$ Year+Vessel+DepCat+Month+Zone+DayNight+Zone:Month |
| Model 8 | LnCE $\sim$ Year+Vessel+DepCat+Month+Zone+DayNight+Zone:DepCat |

Table 13.67. Blue Grenadier from the SET in depths between $200-600 \mathrm{~m}$, taken by Trawl, omitting the Spawning fishery (zone 40 between June and August). Model selection criteria, including the AIC, the deviance and the change in deviance. The optimum is model 7.

|  | Year | Vessel | DepCat | Month | Zone | DayNight | Zone:Mth | Zone:DepC |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 118368 | 94176 | 80603 | 75723 | 73177 | 70855 | 67740 | 69526 |
| RSS | 322399 | 265666 | 238061 | 228994 | 224403 | 220304 | 214747 | 217220 |
| MSS | 24504 | 81237 | 108842 | 117909 | 122500 | 126599 | 132156 | 129682 |
| Nobs | 126964 | 126964 | 126240 | 126240 | 126240 | 126240 | 126240 | 126240 |
| Npars | 26 | 217 | 262 | 273 | 278 | 281 | 336 | 506 |
| adj_r2 | 7.045 | 23.287 | 31.233 | 33.847 | 35.170 | 36.353 | 37.931 | 37.131 |
| \%Change | 0.000 | 16.242 | 7.946 | 2.613 | 1.324 | 1.183 | 1.578 | -0.800 |



Figure 13.80. The relative influence of each factor used on the final trend in the optimal standardization for Blue Grenadier non-spawning fishery. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.27 Silver Warehou (TRS - 37445006 - Seriolella punctata)

Data from zones 10 to 50 , depths between $0-600 \mathrm{~m}$.


Figure 13.81. The trends in catches and catch rates for zones $10-50$, split east and west.
The catch rates in the east show approximately the same trends, though there are some differences between 2000 and 2003. In the west the same pattern of noisy but flat from 1992 to 2006 followed by a decline are exhibited. But the trends are different between the east and west.


Figure 13.82. The standardized trends for the Month and DepCat factors for Silver Warehou taken by trawl across SESSF zones 10-50.

Table 13.68. Silver Warehou from Zones 10 to 50 and depths $0-600 \mathrm{~m}$ by trawl. Total Catch is the total reported in the database, Records is the number of records use din the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates (kg/hr). Zone:Month is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 1156.533 | 2438 | 1135.296 | 86 | 32.2897 | 1.4228 | 0.0000 |
| 1987 | 782.151 | 1509 | 757.298 | 76 | 35.5040 | 1.5121 | 0.0563 |
| 1988 | 1646.187 | 2249 | 1617.240 | 87 | 42.9346 | 1.9169 | 0.0510 |
| 1989 | 926.257 | 2049 | 907.420 | 80 | 30.7291 | 1.5678 | 0.0539 |
| 1990 | 1346.585 | 1983 | 1290.959 | 81 | 40.6488 | 1.6590 | 0.0544 |
| 1991 | 1453.169 | 2289 | 1207.361 | 78 | 25.6848 | 1.1675 | 0.0532 |
| 1992 | 733.767 | 1857 | 625.074 | 55 | 27.9469 | 1.0123 | 0.0558 |
| 1993 | 1815.801 | 3866 | 1735.163 | 61 | 33.2988 | 1.1448 | 0.0487 |
| 1994 | 2309.510 | 4519 | 2300.083 | 57 | 34.7142 | 1.2249 | 0.0477 |
| 1995 | 2002.881 | 5016 | 1969.857 | 58 | 29.7825 | 1.1080 | 0.0470 |
| 1996 | 2188.244 | 6080 | 2137.373 | 67 | 22.7319 | 1.0502 | 0.0462 |
| 1997 | 2562.016 | 5765 | 2305.785 | 61 | 25.3481 | 1.0783 | 0.0469 |
| 1998 | 2166.021 | 4702 | 1976.667 | 57 | 26.6416 | 1.0390 | 0.0478 |
| 1999 | 2834.052 | 5148 | 2685.678 | 58 | 31.2330 | 0.8951 | 0.0474 |
| 2000 | 3401.563 | 6738 | 3324.009 | 63 | 26.0708 | 0.8113 | 0.0464 |
| 2001 | 2970.407 | 7293 | 2789.412 | 59 | 21.7853 | 0.6811 | 0.0462 |
| 2002 | 3841.439 | 8418 | 3656.597 | 57 | 22.9919 | 0.7357 | 0.0456 |
| 2003 | 2910.130 | 7402 | 2782.813 | 64 | 20.4815 | 0.7411 | 0.0461 |
| 2004 | 3198.195 | 7860 | 3032.860 | 58 | 23.3323 | 0.8231 | 0.0459 |
| 2005 | 2647.967 | 6920 | 2558.281 | 56 | 20.0277 | 0.8088 | 0.0464 |
| 2006 | 2191.402 | 5663 | 2076.280 | 47 | 18.2160 | 0.7116 | 0.0473 |
| 2007 | 1816.516 | 4657 | 1665.236 | 33 | 20.1239 | 0.6722 | 0.0484 |
| 2008 | 1381.159 | 4400 | 1279.929 | 32 | 16.1202 | 0.6018 | 0.0487 |
| 2009 | 1285.306 | 4387 | 1109.646 | 28 | 15.8837 | 0.6203 | 0.0488 |
| 2010 | 1189.353 | 4481 | 1082.522 | 28 | 13.2653 | 0.5126 | 0.0488 |
| 2011 | 1108.751 | 4888 | 1025.651 | 30 | 12.5782 | 0.4816 | 0.0484 |



Figure 13.83. Silver Warehou from Zones 10 to 50 and depths $0-600 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Silver Warehou, the top right graph depicts the depth distribution of shots containing Silver Warehou from Zones 10 to 50 and depths $0-600 \mathrm{~m}$ by trawl.. The middle left diagram depicts the distribution of catch by depth within zones 10 to 50 , the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Silver Warehou catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.84. Silver Warehou from Zones 10 to 50 and depths $0-600 \mathrm{~m}$ by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

[^2]Table 13.70. Silver Warehou from Zones 10 to 50 and depths $0-600 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the adjusted $\mathrm{r}^{2}$ and the change in adjusted $\mathrm{r}^{2}$. The optimum is Zone:Month (model 7).

|  | Year | Vessel | Month | Zone | DepCat | DayNight | Zone:Month | Zone:DepCat |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 147644 | 125985 | 119751 | 117595 | 115256 | 115050 | 113290 | 113744 |
| RSS | 408633 | 341361 | 324374 | 318478 | 312402 | 311858 | 307164 | 307925 |
| MSS | 9666 | 76938 | 93925 | 99820 | 105897 | 106440 | 111135 | 110374 |
| Nobs | 122577 | 122577 | 122577 | 121809 | 121809 | 121809 | 121809 | 121809 |
| Npars | 26 | 221 | 232 | 262 | 266 | 269 | 313 | 389 |
| adj_r2 | 2.291 | 18.246 | 22.308 | 23.700 | 25.153 | 25.282 | 26.380 | 26.151 |
| \%Change | 0.000 | 15.956 | 4.061 | 1.392 | 1.453 | 0.128 | 1.098 | -0.229 |



Figure 13.85. The relative influence of each factor used on the final trend in the optimal standardization for Silver Warehou in Zones $10-50$. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.28 Blue Warehou Zones 10, 20, 30 (TRT - 37445005 Seriolella brama)

Data from zones 10,20 , and 30 , depths less than or equal to 400 m .
Table 13.71. Blue Warehou from zones 10 to 30 in depths $0-400 \mathrm{~m}$ by trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr)}$ ). Zone:DepCat is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:DepC | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 211.877 | 702 | 138.822 | 40 | 22.9216 | 1.8787 | 0.0000 |
| 1987 | 405.851 | 457 | 168.152 | 40 | 23.2716 | 2.2848 | 0.1048 |
| 1988 | 543.976 | 775 | 334.047 | 33 | 34.8726 | 2.8020 | 0.0953 |
| 1989 | 776.041 | 1178 | 664.709 | 41 | 52.6588 | 3.5366 | 0.0926 |
| 1990 | 881.353 | 826 | 508.270 | 42 | 46.5510 | 3.2341 | 0.0977 |
| 1991 | 1284.194 | 1567 | 465.158 | 54 | 23.0208 | 1.7420 | 0.0920 |
| 1992 | 934.405 | 1343 | 406.749 | 39 | 24.3304 | 1.4238 | 0.0926 |
| 1993 | 829.573 | 2195 | 431.735 | 45 | 20.7054 | 1.1069 | 0.0892 |
| 1994 | 944.805 | 2449 | 473.899 | 44 | 17.5997 | 1.0674 | 0.0882 |
| 1995 | 815.384 | 2646 | 467.825 | 44 | 15.3567 | 0.9784 | 0.0880 |
| 1996 | 724.408 | 3551 | 531.223 | 49 | 14.6415 | 0.9835 | 0.0872 |
| 1997 | 935.159 | 2481 | 404.281 | 42 | 11.8760 | 0.9519 | 0.0894 |
| 1998 | 903.242 | 2556 | 457.247 | 39 | 13.8592 | 0.8996 | 0.0890 |
| 1999 | 590.975 | 1643 | 131.641 | 39 | 5.7097 | 0.4842 | 0.0918 |
| 2000 | 470.248 | 2217 | 185.083 | 40 | 5.0072 | 0.4275 | 0.0902 |
| 2001 | 285.464 | 1470 | 57.242 | 33 | 2.7867 | 0.2561 | 0.0937 |
| 2002 | 290.477 | 1856 | 62.867 | 36 | 2.2036 | 0.1966 | 0.0921 |
| 2003 | 233.998 | 1324 | 42.078 | 38 | 1.8331 | 0.1558 | 0.0950 |
| 2004 | 232.446 | 1249 | 52.051 | 38 | 2.7248 | 0.2122 | 0.0967 |
| 2005 | 289.063 | 830 | 21.286 | 33 | 1.8011 | 0.1409 | 0.1011 |
| 2006 | 379.527 | 776 | 25.720 | 28 | 2.2327 | 0.1689 | 0.1022 |
| 2007 | 177.774 | 583 | 16.757 | 14 | 1.8677 | 0.1826 | 0.1068 |
| 2008 | 163.260 | 738 | 27.441 | 18 | 2.6539 | 0.2522 | 0.1026 |
| 2009 | 135.224 | 447 | 36.884 | 15 | 3.5956 | 0.2922 | 0.1118 |
| 2010 | 130.098 | 374 | 12.266 | 15 | 2.1227 | 0.1894 | 0.1172 |
| 2011 | 103.243 | 435 | 9.812 | 13 | 1.7081 | 0.1516 | 0.1134 |



Figure 13.86. Blue Warehou from zones 10 to 30 in depths $0-400 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Blue Warehou, the top right graph depicts the depth distribution of shots containing Blue Warehou from zones 10 to 30 in depths $0-400 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zones 10 to 30 , the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Blue Warehou catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.87. Blue Warehou from zones 10 to 30 in depths $0-400 \mathrm{~m}$ by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.72. Blue Warehou from zones 10 to 30 in depths $0-400 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel+DepCat |
| Model 4 | LnCE $\sim$ Year+Vessel+DepCat+Month |
| Model 5 | LnCE $\sim$ Year+Vessel+DepCat+Month+Zone |
| Model 6 | LnCE $\sim$ Year+Vessel+DepCat+Month+Zone+DayNight |
| Model 7 | LnCE $\sim$ Year+Vessel+DepCat+Month+Zone+DayNight+Zone:Month |
| Model 8 | LnCE $\sim$ Year+Vessel+DepCat+Month+Zone+DayNight+Zone:DepCat |

Table 13.73. Blue Warehou from zones 10 to 30 in depths $0-400 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the adjusted $r^{2}$ and the change in adjusted $r^{2}$. The optimum is Zone:DSepCat (model 8).

|  | Year | Vessel | DepCat | Month | Zone | DayNight |  |  | Zone:Mth |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | Zone:DepC |  |  |  |  |  |  |  |  |
| AIC | 36763 | 32110 | 31473 | 31276 | 30791 | 30789 | 30539 | 30508 |  |
| RSS | 99792 | 87133 | 85466 | 84955 | 83823 | 83805 | 83132 | 82980 |  |
| MSS | 36422 | 49081 | 50748 | 51259 | 52391 | 52410 | 53082 | 53234 |  |
| Nobs | 36668 | 36668 | 36458 | 36458 | 36458 | 36458 | 36458 | 36458 |  |
| Npars | 26 | 186 | 206 | 217 | 219 | 222 | 244 | 262 |  |
| adj_r2 | 26.689 | 35.708 | 36.901 | 37.260 | 38.092 | 38.101 | 38.560 | 38.642 |  |
| \%Change | 0.000 | 9.019 | 1.194 | 0.358 | 0.833 | 0.008 | 0.460 | 0.082 |  |



Figure 13.88. The relative influence of each factor used on the final trend in the optimal standardization for Blue Warehou in Zone $10-30$. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.29 Blue Warehou Z4050 (TRT - 37445005 - S. brama)

Data from zones $40-50$ depths less than or equal to 600 m .

Table 13.74. Blue Warehou from zones 40 and 50 in depths $0-600 \mathrm{~m}$ by trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates (kg/hr). Zone:Month is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 211.877 | 159 | 71.389 | 14 | 34.3927 | 3.3403 | 0.0000 |
| 1987 | 405.851 | 183 | 215.645 | 10 | 153.6342 | 3.1136 | 0.2443 |
| 1988 | 543.976 | 180 | 197.989 | 12 | 104.5294 | 1.2972 | 0.2537 |
| 1989 | 776.041 | 56 | 81.343 | 13 | 91.5270 | 3.3111 | 0.3142 |
| 1990 | 881.353 | 444 | 298.296 | 14 | 55.8069 | 1.4434 | 0.2394 |
| 1991 | 1284.194 | 597 | 647.537 | 18 | 159.6429 | 2.2301 | 0.2374 |
| 1992 | 934.405 | 538 | 430.133 | 17 | 88.9759 | 1.2798 | 0.2395 |
| 1993 | 829.573 | 495 | 362.854 | 21 | 92.3447 | 0.9374 | 0.2409 |
| 1994 | 944.805 | 824 | 449.901 | 21 | 67.3117 | 1.0262 | 0.2365 |
| 1995 | 815.384 | 825 | 325.150 | 22 | 45.1964 | 0.7000 | 0.2342 |
| 1996 | 724.408 | 700 | 183.550 | 24 | 26.4215 | 0.4568 | 0.2353 |
| 1997 | 935.159 | 431 | 243.547 | 23 | 35.6095 | 0.4907 | 0.2410 |
| 1998 | 903.242 | 582 | 354.483 | 19 | 58.9967 | 0.7327 | 0.2393 |
| 1999 | 590.975 | 688 | 174.376 | 19 | 32.5226 | 0.4211 | 0.2388 |
| 2000 | 470.248 | 650 | 203.390 | 24 | 28.0473 | 0.3439 | 0.2391 |
| 2001 | 285.464 | 685 | 194.156 | 23 | 27.5825 | 0.3634 | 0.2379 |
| 2002 | 290.477 | 530 | 218.017 | 23 | 35.4216 | 0.4896 | 0.2405 |
| 2003 | 233.998 | 363 | 175.478 | 19 | 28.1023 | 0.4422 | 0.2463 |
| 2004 | 232.446 | 437 | 159.255 | 21 | 28.4995 | 0.5050 | 0.2430 |
| 2005 | 289.063 | 461 | 257.801 | 18 | 53.5991 | 0.7955 | 0.2434 |
| 2006 | 379.527 | 695 | 337.473 | 16 | 31.8482 | 0.5634 | 0.2399 |
| 2007 | 177.774 | 466 | 148.640 | 16 | 22.9820 | 0.4887 | 0.2437 |
| 2008 | 163.260 | 353 | 117.774 | 12 | 20.3955 | 0.3739 | 0.2459 |
| 2009 | 135.224 | 308 | 89.003 | 11 | 18.4388 | 0.2769 | 0.2482 |
| 2010 | 130.098 | 407 | 105.291 | 12 | 17.5511 | 0.3152 | 0.2435 |
| 2011 | 103.243 | 519 | 77.907 | 14 | 14.3658 | 0.2618 | 0.2432 |



Figure 13.89. Blue Warehou from zones 40 and 50 in depths $0-600 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Blue Warehou, the top right graph depicts the depth distribution of shots containing Blue Warehou from zones 40 and 50 in depths $0-600 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zones 40 and 50 ( 50 is top red line), the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Blue Warehou catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.90. Blue Warehou from zones 40 and 50 in depths $0-600 \mathrm{~m}$ by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.75. Blue Warehou from zones 40 and 50 in depths $0-600 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE~Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel+Month |
| Model 4 | LnCE $\sim$ Year+Vessel+Month+DepCat |
| Model 5 | LnCE $\sim$ Year+Vessel+Month+DepCat+DayNight |
| Model 6 | LnCE $\sim$ Year+Vessel+Month+DepCat+DayNight+Zone |
| Model 7 | LnCE $\sim$ Year+Vessel+Month+DepCat+DayNight+Zone+Zone:Month |
| Model 8 | LnCE $\sim$ Year+Vessel+Month+DepCat+DayNight+Zone+Zone:DepCat |

Table 13.76. Blue Warehou from zones 40 and 50 in depths $0-600 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the adjusted $\mathrm{r}^{2}$ and the change in adjusted $\mathrm{r}^{2}$. The optimum is Zone:Month (model 7).

|  | Year |  | Vessel | Month | DepCat |  | DayNight | Zone | Zone:Mth |  | Zone:DepC |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: |
| AIC | 14134 | 13013 | 12032 | 11232 | 11066 | 11060 | 11011 | 11032 |  |  |  |
| RSS | 38533 | 34807 | 32139 | 29999 | 29588 | 29570 | 29403 | 29364 |  |  |  |
| MSS | 4565 | 8291 | 10959 | 13099 | 13510 | 13528 | 13695 | 13734 |  |  |  |
| Nobs | 12576 | 12576 | 12576 | 12522 | 12522 | 12522 | 12522 | 12522 |  |  |  |
| Npars | 26 | 105 | 116 | 146 | 149 | 150 | 161 | 180 |  |  |  |
| adj_r2 | 10.414 | 18.564 | 24.739 | 29.578 | 30.525 | 30.563 | 30.893 | 30.879 |  |  |  |
| \%Change | 0.000 | 8.150 | 6.175 | 4.839 | 0.947 | 0.038 | 0.330 | -0.014 |  |  |  |



Figure 13.91. The relative influence of each factor used on the final trend in the optimal standardization for Blue Warehou in Zone $40-50$. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.30 Blue Warehou Z10-50 (TRT - 37445005 - S. brama)

Only data from Zones 10 to 50 in depths $0-600 \mathrm{~m}$. Only vessels present in the fishery for more than 2 years were included.


Figure 13.92. Trends in the catches and geometric mean catch rates for Blue Warehou across each of the zones $10-50$, split east and west. The extreme catch rates in zone 40 reflect very small catches
The severe depletion in the east is evident but in the west the catch rates are noisy then flat. They are depressed primarily because of early high values that reflect very low catches or relatively high catches. Zone 50 is the main part of the western Blue Warehou fishery.


Figure 13.93. The standardized trends for the Month and DepCat factors for Blue Warehou taken by trawl across SESSF zones 10-50.

Table 13.77. Blue Warehou from zones 10 to 50 in depths $0-600 \mathrm{~m}$ by trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr}$ ). Zone:Mth is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 211.877 | 863 | 210.321 | 54 | 24.6419 | 2.0083 | 0.0000 |
| 1987 | 405.851 | 655 | 384.556 | 51 | 38.9818 | 2.3004 | 0.0921 |
| 1988 | 543.976 | 963 | 532.358 | 45 | 42.2791 | 2.5658 | 0.0892 |
| 1989 | 776.041 | 1239 | 746.152 | 50 | 53.5132 | 3.5440 | 0.0876 |
| 1990 | 881.353 | 1284 | 822.419 | 56 | 49.3618 | 2.5233 | 0.0889 |
| 1991 | 1284.194 | 2193 | 1119.788 | 66 | 38.9026 | 1.9961 | 0.0845 |
| 1992 | 934.405 | 1902 | 840.304 | 56 | 34.9011 | 1.4626 | 0.0854 |
| 1993 | 829.573 | 2717 | 797.308 | 58 | 27.0143 | 1.1328 | 0.0832 |
| 1994 | 944.805 | 3300 | 927.228 | 58 | 24.5388 | 1.0921 | 0.0820 |
| 1995 | 815.384 | 3497 | 794.697 | 58 | 19.7435 | 0.9299 | 0.0817 |
| 1996 | 724.408 | 4278 | 715.754 | 66 | 16.0446 | 0.9250 | 0.0812 |
| 1997 | 935.159 | 2925 | 648.139 | 57 | 13.9027 | 0.9172 | 0.0834 |
| 1998 | 903.242 | 3152 | 813.727 | 50 | 18.0335 | 0.9174 | 0.0829 |
| 1999 | 590.975 | 2372 | 309.696 | 57 | 9.5323 | 0.4883 | 0.0847 |
| 2000 | 470.248 | 2899 | 389.591 | 58 | 7.2891 | 0.4352 | 0.0837 |
| 2001 | 285.464 | 2208 | 253.279 | 53 | 5.6327 | 0.2942 | 0.0857 |
| 2002 | 290.477 | 2408 | 281.036 | 53 | 4.0433 | 0.2499 | 0.0854 |
| 2003 | 233.998 | 1709 | 218.370 | 51 | 3.2843 | 0.2056 | 0.0879 |
| 2004 | 232.446 | 1700 | 211.509 | 51 | 4.9660 | 0.2815 | 0.0885 |
| 2005 | 289.063 | 1297 | 279.429 | 45 | 6.0446 | 0.2614 | 0.0909 |
| 2006 | 379.527 | 1474 | 363.242 | 36 | 7.8259 | 0.2625 | 0.0899 |
| 2007 | 177.774 | 1051 | 165.406 | 25 | 5.6784 | 0.2449 | 0.0933 |
| 2008 | 163.260 | 1100 | 145.318 | 27 | 5.0903 | 0.2749 | 0.0925 |
| 2009 | 135.224 | 766 | 126.232 | 24 | 6.9116 | 0.2721 | 0.0975 |
| 2010 | 130.098 | 785 | 117.741 | 22 | 6.3388 | 0.2177 | 0.0974 |
| 2011 | 103.243 | 966 | 91.479 | 23 | 5.5194 | 0.1969 | 0.0953 |



Figure 13.94. Blue Warehou from zones 10 to 50 in depths $0-400 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Blue Warehou, the top right graph depicts the depth distribution of shots containing Blue Warehou from zones 10 to 50 in depths $0-400 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zones 1050 , the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Blue Warehou catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ )


Figure 13.95. Blue Warehou from zones 10 to 50 in depths $0-400 \mathrm{~m}$ by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.78. Blue Warehou from zones 10 to 50 in depths $0-400 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel+DepCat |
| Model 4 | LnCE $\sim$ Year+Vessel+DepCat+Zone |
| Model 5 | LnCE $\sim$ Year+Vessel+DepCat+Zone+Month |
| Model 6 | LnCE $\sim$ Year+Vessel+DepCat+Zone+Month+DayNight |
| Model 7 | LnCE $\sim$ Year+Vessel+DepCat+Zone+Month+DayNight+Zone:Month |
| Model 8 | LnCE $\sim$ Year+Vessel+DepCat+Zone+Month+DayNight+Zone:DepCat |

Table 13.79. Blue Warehou from zones 10 to 50 in depths $0-400 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the adjusted $r^{2}$ and the change in adjusted $r^{2}$. The optimum is Zone:Month (model 7).

|  | Year | Vessel | DepCat Zone | Month | DayNight | Zone:Mth | Zone:DepC |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 61386 | 47536 | 46286 | 45197 | 44475 | 44410 | 43419 | 43729 |
| RSS | 170727 | 128225 | 124838 | 122097 | 120274 | 120101 | 117509 | 117886 |
| MSS | 30728 | 73231 | 76618 | 79358 | 81181 | 81354 | 83947 | 83569 |
| Nobs | 49703 | 49703 | 49439 | 49439 | 49439 | 49439 | 49439 | 49439 |
| Npars | 26 | 216 | 246 | 250 | 261 | 264 | 308 | 384 |
| adj_r2 | 15.210 | 36.074 | 37.723 | 39.086 | 39.982 | 40.064 | 41.306 | 41.026 |
| \%Change | 0.000 | 20.864 | 1.649 | 1.362 | 0.896 | 0.083 | 1.241 | -0.280 |



Figure 13.96. The relative influence of each factor used on the final trend in the optimal standardization for Blue Warehou in Zone $10-50$. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.
13.31 Pink Ling TW (LIG - 37228002 - Genypterus blacodes)


Figure 13.97. Trends in the catches and geometric mean catch rates for Pink Ling taken by trawler across zones $10-50$ split between east and west.

The trends in the geometric mean catch rates in the east all follow approximately the same trajectory, albeit with some noise. In the west, however, zones 40 and 50 appear to follow rather different trajectories with rates increasing since 2005 in zone 40 whilst staying flat in zone 50 . However, this may simply reflect that catches were increasing in zone 40 and were decreasing in zone 50 .

### 13.32 Pink Ling, Z102030 (LIG - 37228002 - G. blacodes)

Data from zones 10, 20 and 30, depths greater than 250 m and less than 600 m .

Table 13.80. Pink Ling from zones 10 to 30 in depths between $250-600 \mathrm{~m}$ by trawl. Total Catch is the
total reported in the database, Records is the number of records used in the analysis, CatchT is the
reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the
analysis. Geomean is the geometric mean of catch rates $(\mathrm{kg} / \mathrm{hr})$. Zone:Month is the optimum model and
StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 678.977 | 4512 | 498.298 | 80 | 20.665 | 1.102 | 0.0000 |
| 1987 | 765.066 | 4260 | 492.314 | 77 | 19.424 | 1.168 | 0.0224 |
| 1988 | 583.077 | 3613 | 400.077 | 77 | 20.259 | 1.114 | 0.0235 |
| 1989 | 678.896 | 3879 | 422.077 | 77 | 19.157 | 0.959 | 0.0233 |
| 1990 | 674.479 | 2794 | 413.082 | 68 | 26.820 | 1.411 | 0.0256 |
| 1991 | 736.803 | 2938 | 370.297 | 72 | 26.305 | 1.412 | 0.0256 |
| 1992 | 568.308 | 2417 | 324.371 | 57 | 24.850 | 1.089 | 0.0269 |
| 1993 | 892.796 | 3525 | 504.474 | 59 | 25.307 | 1.023 | 0.0245 |
| 1994 | 895.431 | 4066 | 470.265 | 63 | 23.516 | 1.037 | 0.0236 |
| 1995 | 1208.893 | 4361 | 586.686 | 57 | 25.811 | 1.307 | 0.0231 |
| 1996 | 1233.265 | 4268 | 667.583 | 63 | 27.657 | 1.300 | 0.0233 |
| 1997 | 1696.855 | 4808 | 732.654 | 62 | 27.937 | 1.323 | 0.0229 |
| 1998 | 1591.988 | 4909 | 730.458 | 57 | 26.016 | 1.323 | 0.0227 |
| 1999 | 1651.572 | 5964 | 832.655 | 59 | 25.229 | 1.209 | 0.0222 |
| 2000 | 1507.379 | 5113 | 660.280 | 62 | 22.405 | 1.077 | 0.0231 |
| 2001 | 1392.822 | 4544 | 484.022 | 52 | 19.062 | 0.839 | 0.0239 |
| 2002 | 1330.296 | 3898 | 360.465 | 52 | 15.866 | 0.733 | 0.0248 |
| 2003 | 1353.243 | 4309 | 445.759 | 57 | 18.293 | 0.752 | 0.0243 |
| 2004 | 1495.581 | 3359 | 347.369 | 54 | 16.798 | 0.672 | 0.0258 |
| 2005 | 1203.256 | 3454 | 329.969 | 51 | 16.334 | 0.628 | 0.0255 |
| 2006 | 1069.222 | 2593 | 323.101 | 38 | 21.319 | 0.750 | 0.0274 |
| 2007 | 875.926 | 1652 | 204.307 | 23 | 20.501 | 0.736 | 0.0315 |
| 2008 | 980.268 | 2382 | 329.036 | 24 | 25.151 | 0.852 | 0.0286 |
| 2009 | 775.047 | 1947 | 212.362 | 27 | 18.295 | 0.622 | 0.0303 |
| 2010 | 906.088 | 1990 | 271.121 | 23 | 20.721 | 0.761 | 0.0299 |
| 2011 | 1081.674 | 2199 | 294.797 | 22 | 23.444 | 0.803 | 0.0292 |



Figure 13.98. Pink Ling from zones 10 to 30 in depths between $250-600 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Pink Ling, the top right graph depicts the depth distribution of shots containing Pink Ling from zones 10 to 30 in depths between $250-600 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zones 10 to 30 , the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Pink Ling catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.99. Pink Ling from zones 10 to 30 in depths between $250-600 \mathrm{~m}$ by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.81. Pink Ling from zones 10 to 30 in depths between $250-600 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+DepCat |
| Model 3 | LnCE $\sim$ Year+DepCat+Vessel |
| Model 4 | LnCE $\sim$ Year+DepCat+Vessel+Month |
| Model 5 | LnCE $\sim$ Year+DepCat+Vessel+Month+Zone |
| Model 6 | LnCE $\sim$ Year+DepCat+Vessel+Month+Zone+DayNight |
| Model 7 | LnCE $\sim$ Year+DepCat+Vessel+Month+Zone+DayNight+Zone:Month |
| Model 8 | LnCE $\sim$ Year+DepCat+Vessel+Month+Zone+DayNight+Zone:DepCat |

Table 13.82. Pink Ling from zones 10 to 30 in depths between $250-600 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the adjusted $r^{2}$ and the change in adjusted $r^{2}$. The optimum is Zone:Month (model 7).

|  | Year | DepCat | Vessel | Month | Zone | DayNight | Zone:Mth | Zone:DepC |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 31726 | 24789 | 4336 | 605 | -1 | -44 | -1143 | -1015 |
| RSS | 131436 | 121237 | 96914 | 93079 | 92470 | 92420 | 91291 | 91390 |
| MSS | 2713 | 12911 | 37234 | 41070 | 41679 | 41728 | 42857 | 42759 |
| Nobs | 93754 | 92943 | 92943 | 92943 | 92943 | 92943 | 92943 | 92943 |
| Npars | 26 | 44 | 223 | 235 | 237 | 240 | 262 | 276 |
| adj_r2 | 1.996 | 9.583 | 27.583 | 30.440 | 30.894 | 30.928 | 31.756 | 31.672 |
| \%Change | 0.000 | 7.587 | 18.001 | 2.857 | 0.454 | 0.035 | 0.828 | 0.744 |



Figure 13.100. The relative influence of each factor used on the final trend in the optimal standardization for Pink Ling from zones 10 to 30 . The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.33 Pink Ling, Z4050 (LIG - 37228002 - G. blacodes)

Data from zones 40 and 50, depths greater than 200 m and less or equal to 800 m .

> Table 13.83 . Pink Ling from zones 40 and 50 in depths between $200-800 \mathrm{~m}$ by trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean in the geometric mean of catch rates (kg/hr). Zone:Mth is the optimum model and SDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 678.977 | 1265 | 112.944 | 23 | 17.1417 | 1.1680 | 0.0000 |
| 1987 | 765.066 | 1310 | 206.341 | 28 | 24.0155 | 1.3435 | 0.0376 |
| 1988 | 583.077 | 1026 | 95.703 | 32 | 17.6676 | 1.0475 | 0.0407 |
| 1989 | 678.896 | 1469 | 183.121 | 34 | 21.9840 | 1.0835 | 0.0389 |
| 1990 | 674.479 | 1524 | 147.412 | 32 | 16.9021 | 0.9750 | 0.0393 |
| 1991 | 736.803 | 1897 | 198.945 | 37 | 16.3936 | 1.0334 | 0.0375 |
| 1992 | 568.308 | 1633 | 102.164 | 24 | 11.9963 | 0.7724 | 0.0386 |
| 1993 | 892.796 | 2253 | 235.485 | 24 | 17.1332 | 1.0467 | 0.0373 |
| 1994 | 895.431 | 2110 | 247.793 | 24 | 20.5621 | 1.2583 | 0.0372 |
| 1995 | 1208.893 | 3516 | 426.907 | 25 | 20.0613 | 1.2904 | 0.0350 |
| 1996 | 1233.265 | 3403 | 448.044 | 26 | 19.9984 | 1.3731 | 0.0354 |
| 1997 | 1696.855 | 3732 | 577.434 | 24 | 21.1891 | 1.4460 | 0.0350 |
| 1998 | 1591.988 | 3710 | 558.641 | 21 | 22.4111 | 1.4405 | 0.0352 |
| 1999 | 1651.572 | 3794 | 427.920 | 24 | 18.0495 | 1.1255 | 0.0351 |
| 2000 | 1507.379 | 4655 | 509.304 | 27 | 16.3679 | 0.9974 | 0.0348 |
| 2001 | 1392.822 | 5061 | 500.022 | 28 | 14.7513 | 0.8910 | 0.0346 |
| 2002 | 1330.296 | 4631 | 429.572 | 27 | 13.4100 | 0.7697 | 0.0347 |
| 2003 | 1353.243 | 3821 | 360.388 | 27 | 12.6444 | 0.7741 | 0.0352 |
| 2004 | 1495.581 | 3901 | 306.551 | 25 | 11.7195 | 0.7207 | 0.0353 |
| 2005 | 1203.256 | 2663 | 195.741 | 23 | 9.9467 | 0.5988 | 0.0366 |
| 2006 | 1069.222 | 2322 | 209.985 | 21 | 10.6509 | 0.6418 | 0.0373 |
| 2007 | 875.926 | 2532 | 287.345 | 16 | 12.6778 | 0.7114 | 0.0369 |
| 2008 | 980.268 | 1795 | 214.232 | 17 | 14.6108 | 0.9077 | 0.0383 |
| 2009 | 775.047 | 1976 | 260.609 | 13 | 14.0039 | 0.8909 | 0.0378 |
| 2010 | 906.088 | 2337 | 272.103 | 14 | 13.1465 | 0.8572 | 0.0371 |
| 2011 | 1081.674 | 2728 | 346.774 | 16 | 13.1057 | 0.8357 | 0.0370 |



Figure 13.101. Pink Ling from zones 40 and 50 in depths between $200-800 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Pink Ling, the top right graph depicts the depth distribution of shots containing Pink Ling from zones 40 and 50 in depths between $200-800 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zones 40 and 50 ( 50 is top red line), the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Pink Ling catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.102. Pink Ling from zones 40 and 50 in depths between $200-800 \mathrm{~m}$ by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.84. Pink Ling from zones 40 and 50 in depths between $200-800 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+DepCat |
| Model 3 | LnCE $\sim$ Year+DepCat+Vessel |
| Model 4 | LnCE $\sim$ Year+DepCat+Vessel+Month |
| Model 5 | LnCE $\sim$ Year+DepCat+Vessel+Month+Zone |
| Model 6 | LnCE $\sim$ Year+DepCat+Vessel+Month+Zone+DayNight |
| Model 7 | LnCE $\sim$ Year+DepCat+Vessel+Month+Zone+DayNight+Zone:Month |
| Model 8 | LnCE $\sim$ Year+DepCat+Vessel+Month+Zone+DayNight+Zone:DepCat |

Table 13.85. Pink Ling from zones 40 and 50 in depths between $200-800 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the adjusted $r^{2}$ and the change in adjusted $r^{2}$. The optimum is Zone:Month (model 7).

|  | Year | DepCat | Vessel | Month | Zone | DayNight |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | -232 | -9783 | -15835 | -18000 | -18677 | -18698 | -19964 | -19316 |
| AIC | 70780 | 61450 | 56260 | 54546 | 54025 | 54004 | 53029 | 53488 |
| RSS | 3890 | 13220 | 18410 | 20124 | 20645 | 20666 | 21641 | 21182 |
| MSS | 26 | 56 | 148 | 159 | 160 | 163 | 174 | 193 |
| Nobs | 71064 | 70683 | 70683 | 70683 | 70683 | 70683 | 70683 | 70683 |
| Npars | 26 | 178 | 163 |  |  |  |  |  |
| adj_r2 | 5.176 | 17.641 | 24.498 | 26.787 | 27.486 | 27.510 | 28.808 | 28.172 |
| \%Change | 0.000 | 12.465 | 6.857 | 2.289 | 0.699 | 0.025 | 1.298 | 0.662 |



Figure 13.103. The relative influence of each factor used on the final trend in the optimal standardization for Pink Ling from zones 40 and 50. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.34 Pink Ling, Z10 (LIG - 37228002 - G. blacodes)

Data from zone 10, depths greater than 250 m and less or equal to 600 m .
Table 13.86. Pink Ling from zone 10 in depths between $250-600 \mathrm{~m}$ by trawl. Total Catch is the total Pink Ling catch from all zones reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in this analysis, and Vessels relates to all vessels used in the analysis. Geomean is the unstandardized geometric mean of catch rates (kg/hr). Vessel:Mth is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Vessel:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 678.977 | 3324 | 314.213 | 69 | 18.2806 | 1.1717 | 0.0000 |
| 1987 | 765.066 | 3017 | 270.907 | 65 | 15.1828 | 1.1250 | 0.0271 |
| 1988 | 583.077 | 2154 | 207.947 | 62 | 16.5795 | 1.1510 | 0.0295 |
| 1989 | 678.896 | 2356 | 177.865 | 61 | 13.2037 | 0.8811 | 0.0292 |
| 1990 | 674.479 | 1436 | 157.171 | 50 | 18.6577 | 1.2797 | 0.0332 |
| 1991 | 736.803 | 1319 | 145.022 | 39 | 23.1009 | 1.5669 | 0.0345 |
| 1992 | 568.308 | 1171 | 167.548 | 42 | 26.4272 | 1.3995 | 0.0357 |
| 1993 | 892.796 | 1613 | 224.873 | 43 | 24.5764 | 1.1595 | 0.0323 |
| 1994 | 895.431 | 1865 | 231.643 | 44 | 26.4614 | 1.4164 | 0.0308 |
| 1995 | 1208.893 | 2366 | 246.588 | 42 | 22.3982 | 1.5260 | 0.0289 |
| 1996 | 1233.265 | 2343 | 278.016 | 45 | 21.7797 | 1.3515 | 0.0291 |
| 1997 | 1696.855 | 2505 | 328.403 | 46 | 24.4094 | 1.4477 | 0.0287 |
| 1998 | 1591.988 | 2873 | 356.785 | 42 | 21.4118 | 1.4226 | 0.0280 |
| 1999 | 1651.572 | 3066 | 382.112 | 39 | 20.6881 | 1.3606 | 0.0279 |
| 2000 | 1507.379 | 2235 | 250.746 | 40 | 18.7962 | 1.1667 | 0.0305 |
| 2001 | 1392.822 | 1376 | 118.901 | 34 | 14.0899 | 0.8609 | 0.0351 |
| 2002 | 1330.296 | 1464 | 106.843 | 37 | 11.8033 | 0.7056 | 0.0343 |
| 2003 | 1353.243 | 1428 | 114.389 | 39 | 13.7771 | 0.6848 | 0.0350 |
| 2004 | 1495.581 | 1028 | 67.395 | 41 | 10.9097 | 0.4993 | 0.0382 |
| 2005 | 1203.256 | 1292 | 75.762 | 35 | 11.1472 | 0.4582 | 0.0353 |
| 2006 | 1069.222 | 795 | 63.499 | 27 | 12.5966 | 0.4686 | 0.0420 |
| 2007 | 875.926 | 397 | 31.023 | 16 | 11.4186 | 0.4864 | 0.0555 |
| 2008 | 980.268 | 559 | 48.896 | 17 | 15.1211 | 0.5929 | 0.0496 |
| 2009 | 775.047 | 421 | 39.817 | 15 | 15.9787 | 0.5246 | 0.0559 |
| 2010 | 906.088 | 636 | 72.524 | 15 | 17.9099 | 0.6973 | 0.0478 |
| 2011 | 1081.674 | 576 | 54.275 | 14 | 17.1346 | 0.5955 | 0.0489 |



Figure 13.104. Pink Ling from zone 10 in depths between $250-600 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Pink Ling from zone 10 taken in the SET down to 1000 m , the top right graph depicts the depth distribution of shots containing Pink Ling from zone 10 in depths between $250-600 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zone 10 , the middle right hand graph depicts the number of vessels reporting Pink Ling catches through time. The bottom left reflects the number of records used in analysis, and bottom right is the Pink Ling catches (top line, black is total catches, all zones, all methods, the middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.105. Pink Ling from zone 10 in depths between $250-600 \mathrm{~m}$ by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates; giving a mean for the series of 1.0.

Table 13.87. Pink Ling from zone 10 in depths between $250-600 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel +DepCat |
| Model 4 | LnCE $\sim$ Year+Vessel +DepCat +Month |
| Model 5 | LnCE $\sim$ Year+Vessel +DepCat +Month+DayNight |
| Model 6 | LnCE Year+Vessel +DepCat +Month +DayNight+ Vessel:Month |
| Model 7 | LnCE $\sim$ Year+Vessel +DepCat +Month +DayNight+ Month:DepCat |

Table 13.88. Pink Ling from zone 10 in depths between $250-600 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the adjusted $r^{2}$ and the change in adjusted $r^{2}$. The optimum is Vessel:Month (model 6).

|  | Year | Vessel | DepCat | Month | DayNight | Vessel:Month | Month:DepCat |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 19195 | 6531 | 2613 | -592 | -625 | -1429 | -1300 |
| RSS | 67647 | 50302 | 45720 | 42443 | 42406 | 38992 | 41372 |
| MSS | 2805 | 20150 | 24732 | 28009 | 28047 | 31460 | 29080 |
| Nobs | 43615 | 43615 | 43393 | 43393 | 43393 | 43393 | 43393 |
| Npars | 26 | 155 | 173 | 184 | 187 | 1606 | 385 |
| adj_r2 | 3.926 | 28.349 | 34.846 | 39.501 | 39.550 | 42.529 | 40.752 |
| \%Change | 0.000 | 24.422 | 6.498 | 4.654 | 0.050 | 2.979 | 1.202 |



Figure 13.106. The relative influence of each factor used on the final trend in the optimal standardization for Pink Ling in Zone 10. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph for Vessel has the geometric mean (grey line) and the effect of adding Year + Vessel (model 2). In the third graph, for DepCat, the grey line represents model 2 and the black line the effect of adding DepCat to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.35 Pink Ling, Z20 (LIG - 37228002 - G. blacodes)

Data from zone 20, depths greater than 250 m and less or equal to 600 m .
Table 13.89. Pink Ling from zone 20 in depths between $250-600 \mathrm{~m}$ by trawl. Total Catch is the total Pink Ling catch from all zones reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in this analysis, and Vessels relates to all vessels used in the analysis. Geomean is the unstandardized geometric mean of catch rates (kg/hr).Mth:DepCat is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Mth:DepCat | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 678.977 | 1173 | 182.189 | 38 | 29.2240 | 1.0087 | 0.0000 |
| 1987 | 765.066 | 1207 | 219.162 | 37 | 36.1063 | 1.4565 | 0.0430 |
| 1988 | 583.077 | 1409 | 187.752 | 39 | 27.1962 | 0.9857 | 0.0422 |
| 1989 | 678.896 | 1462 | 236.224 | 34 | 34.1990 | 1.0624 | 0.0420 |
| 1990 | 674.479 | 1253 | 247.526 | 33 | 40.5063 | 1.2737 | 0.0453 |
| 1991 | 736.803 | 1243 | 196.325 | 31 | 32.4251 | 1.0324 | 0.0460 |
| 1992 | 568.308 | 1112 | 151.017 | 25 | 25.4253 | 0.8487 | 0.0471 |
| 1993 | 892.796 | 1585 | 258.998 | 25 | 27.3764 | 0.8842 | 0.0444 |
| 1994 | 895.431 | 1713 | 210.108 | 24 | 22.5143 | 0.7841 | 0.0437 |
| 1995 | 1208.893 | 1584 | 303.948 | 24 | 33.0905 | 1.2571 | 0.0440 |
| 1996 | 1233.265 | 1544 | 353.759 | 26 | 41.1747 | 1.3787 | 0.0445 |
| 1997 | 1696.855 | 1860 | 358.577 | 28 | 36.3858 | 1.2880 | 0.0441 |
| 1998 | 1591.988 | 1870 | 355.885 | 23 | 35.8703 | 1.3284 | 0.0439 |
| 1999 | 1651.572 | 2421 | 409.166 | 26 | 34.3684 | 1.1772 | 0.0426 |
| 2000 | 1507.379 | 2493 | 375.436 | 31 | 27.0471 | 1.0360 | 0.0426 |
| 2001 | 1392.822 | 2427 | 304.034 | 24 | 23.7631 | 0.8142 | 0.0430 |
| 2002 | 1330.296 | 1934 | 218.025 | 24 | 20.1429 | 0.7770 | 0.0444 |
| 2003 | 1353.243 | 2473 | 301.477 | 30 | 22.0973 | 0.8344 | 0.0430 |
| 2004 | 1495.581 | 1954 | 253.007 | 25 | 22.4000 | 0.8566 | 0.0449 |
| 2005 | 1203.256 | 1768 | 212.464 | 24 | 20.8376 | 0.7881 | 0.0454 |
| 2006 | 1069.222 | 1542 | 228.071 | 20 | 27.6927 | 0.9720 | 0.0459 |
| 2007 | 875.926 | 1025 | 141.086 | 12 | 24.5067 | 0.8174 | 0.0492 |
| 2008 | 980.268 | 1458 | 235.294 | 13 | 30.6898 | 0.9655 | 0.0464 |
| 2009 | 775.047 | 1291 | 156.773 | 16 | 20.0214 | 0.6769 | 0.0474 |
| 2010 | 906.088 | 1175 | 182.205 | 13 | 22.6841 | 0.7855 | 0.0489 |
| 2011 | 1081.674 | 1363 | 212.576 | 13 | 27.4133 | 0.9104 | 0.0478 |



Figure 13.107. Pink Ling from zone 20 in depths between $250-600 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Pink Ling from zone 20 taken in the SET down to 1000 m , the top right graph depicts the depth distribution of shots containing Pink Ling from zone 20 in depths between $250-600 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zone 20, the middle right hand graph depicts the number of vessels reporting Pink Ling catches through time. The bottom left reflects the number of records used in analysis, and bottom right is the Pink Ling catches (top line, black is total catches, all zones, all methods, the middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.108. Pink Ling from zone 20 in depths between 250 - 600 m by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates; giving a mean for the series of 1.0 .

Table 13.90. Pink Ling from zone 20 in depths between $250-600 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel +DepCat |
| Model 4 | LnCE $\sim$ Year+Vessel +DepCat +Month |
| Model 5 | LnCE $\sim$ Year+Vessel +DepCat +Month+DayNight |
| Model 6 | LnCE $\sim$ Year+Vessel +DepCat +Month +DayNight+ Vessel:Month |
| Model 7 | LnCE $\sim$ Year+Vessel +DepCat +Month +DayNight+ Month:DepCat |

Table 13.91. Pink Ling from zone 20 in depths between $250-600 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the adjusted $\mathrm{r}^{2}$ and the change in adjusted $\mathrm{r}^{2}$. The optimum is Month:DepCat (model 7).

|  | Year | Vessel | DepCat | Month | DayNight | Vessel:Month | Month:DepCat |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 7614 | -937 | -2216 | -3350 | -3445 | -3523 | -3725 |
| RSS | 50619 | 40822 | 39364 | 38291 | 38199 | 35777 | 37587 |
| MSS | 1942 | 11739 | 13196 | 14270 | 14362 | 16783 | 14974 |
| Nobs | 42339 | 41834 | 41834 | 41834 | 41834 | 41834 | 41834 |
| Npars | 26 | 44 | 164 | 176 | 179 | 1510 | 377 |
| adj_r2 | 3.638 | 22.254 | 24.814 | 26.843 | 27.014 | 29.384 | 27.840 |
| \%Change | 0.000 | 18.616 | 2.560 | 2.029 | 0.171 | 2.370 | 0.827 |



Figure 13.109. The relative influence of each factor used on the final trend in the optimal standardization for Pink Ling in Zone 20. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph for Vessel has the geometric mean (grey line) and the effect of adding Year + Vessel (model 2). In the third graph, for DepCat, the grey line represents model 2 and the black line the effect of adding DepCat to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.36 Pink Ling, Z30 (LIG - 37228002 - G. blacodes)

Data from zone 30, depths greater than 250 m and less or equal to 600 m .
Table 13.92. Pink Ling from zone 30 in depths between $250-600 \mathrm{~m}$ by trawl. Total Catch is the total Pink Ling catch from all zones reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in this analysis, and Vessels relates to all vessels used in the analysis. Geomean is the unstandardized geometric mean of catch rates (kg/hr). DayNight is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | DayNight | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 678.977 | 15 | 1.896 | 3 | 22.1580 | 2.0010 | 0.0000 |
| 1987 | 765.066 | 36 | 2.245 | 5 | 16.8408 | 1.0199 | 0.2887 |
| 1988 | 583.077 | 50 | 4.378 | 4 | 28.4036 | 2.1634 | 0.2767 |
| 1989 | 678.896 | 61 | 7.988 | 11 | 31.1539 | 1.3596 | 0.2729 |
| 1990 | 674.479 | 105 | 8.385 | 17 | 27.9919 | 1.4765 | 0.2517 |
| 1991 | 736.803 | 376 | 28.950 | 27 | 20.7784 | 0.9609 | 0.2408 |
| 1992 | 568.308 | 134 | 5.806 | 14 | 12.0005 | 0.5687 | 0.2462 |
| 1993 | 892.796 | 327 | 20.603 | 17 | 19.9815 | 0.9560 | 0.2380 |
| 1994 | 895.431 | 488 | 28.514 | 22 | 17.4518 | 0.7898 | 0.2359 |
| 1995 | 1208.893 | 411 | 36.150 | 17 | 22.4107 | 1.0315 | 0.2366 |
| 1996 | 1233.265 | 381 | 35.808 | 18 | 23.9592 | 1.1644 | 0.2369 |
| 1997 | 1696.855 | 443 | 45.674 | 17 | 19.7673 | 1.0964 | 0.2359 |
| 1998 | 1591.988 | 166 | 17.788 | 16 | 20.3063 | 1.0470 | 0.2416 |
| 1999 | 1651.572 | 477 | 41.377 | 15 | 18.8073 | 1.0044 | 0.2366 |
| 2000 | 1507.379 | 385 | 34.098 | 18 | 18.3481 | 0.8927 | 0.2363 |
| 2001 | 1392.822 | 741 | 61.087 | 19 | 16.2336 | 0.7676 | 0.2328 |
| 2002 | 1330.296 | 500 | 35.598 | 17 | 14.9854 | 0.7536 | 0.2345 |
| 2003 | 1353.243 | 408 | 29.893 | 19 | 15.6988 | 0.7112 | 0.2362 |
| 2004 | 1495.581 | 377 | 26.968 | 14 | 12.2641 | 0.5961 | 0.2361 |
| 2005 | 1203.256 | 394 | 41.743 | 14 | 19.1660 | 0.8352 | 0.2371 |
| 2006 | 1069.222 | 256 | 31.531 | 11 | 22.6012 | 0.8236 | 0.2400 |
| 2007 | 875.926 | 230 | 32.198 | 8 | 25.4173 | 1.0015 | 0.2405 |
| 2008 | 980.268 | 365 | 44.846 | 8 | 24.7573 | 0.9377 | 0.2386 |
| 2009 | 775.047 | 235 | 15.772 | 10 | 14.2097 | 0.5549 | 0.2408 |
| 2010 | 906.088 | 179 | 16.392 | 8 | 19.2029 | 0.6997 | 0.2425 |
| 2011 | 1081.674 | 260 | 27.946 | 7 | 20.6797 | 0.7868 | 0.2395 |



Figure 13.110. Pink Ling from zone 30 in depths between $250-600 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Pink Ling from zone 30 taken in the SET down to 1000 m , the top right graph depicts the depth distribution of shots containing Pink Ling from zone 30 in depths between $250-600 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zone 30 , the middle right hand graph depicts the number of vessels reporting Pink Ling catches through time. The bottom left reflects the number of records used in analysis, and bottom right is the Pink Ling catches (top line, black is total catches, all zones, all methods, the middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.111. Pink Ling from zone 30 in depths between $250-600 \mathrm{~m}$ by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates; giving a mean for the series of 1.0. The confidence intervals are wider due to the relatively low number of records.

Table 13.93. Pink Ling from zone 30 in depths between $250-600 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel +DepCat |
| Model 4 | LnCE $\sim$ Year+Vessel +DepCat + Month |
| Model 5 | LnCE $\sim$ Year+Vessel +DepCat +Month+DayNight |
| Model 6 | LnCE $\sim$ Year+Vessel +DepCat +Month +DayNight+ Vessel:Month |
| Model 7 | LnCE $\sim$ Year+Vessel +DepCat +Month +DayNight+ Month:DepCat |

Table 13.94. Pink Ling from zone 30 in depths between 250 - 600 m by trawl. Model selection criteria, including the AIC, the adjusted $\mathrm{r}^{2}$ and the change in adjusted $\mathrm{r}^{2}$. The optimum is DayNight (model 5).

|  | Year | Vessel | DepCat | Month | DayNight | Vessel:Month | Month:DepCat |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | -1143 | -1767 | -2163 | -2224 | -2282 | -1407 | -2184 |
| RSS | 6692 | 6063 | 5655 | 5595 | 5549 | 5047 | 5338 |
| MSS | 316 | 945 | 1352 | 1412 | 1458 | 1960 | 1669 |
| Nobs | 7800 | 7800 | 7716 | 7716 | 7716 | 7716 | 7716 |
| Npars | 26 | 99 | 117 | 128 | 131 | 934 | 329 |
| adj_r2 | 4.196 | 12.378 | 18.062 | 18.818 | 19.457 | 18.064 | 20.433 |
| \%Change | 0.000 | 8.182 | 5.684 | 0.756 | 0.639 | -1.393 | 2.368 |



Figure 13.112. The relative influence of each factor used on the final trend in the optimal standardization for Pink Ling in Zone 30. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph for Vessel has the geometric mean (grey line) and the effect of adding Year + Vessel (model 2). In the third graph, for DepCat, the grey line represents model 2 and the black line the effect of adding DepCat to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.37 Pink Ling, Z40 (LIG - 37228002 - G. blacodes)

Data from zone 40 , depths greater than 350 m and less or equal to 800 m .
Table 13.95. Pink Ling from zone 40 in depths between $350-800 \mathrm{~m}$ by trawl. Total Catch is the total Pink Ling catch from all zones reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in this analysis, and Vessels relates to all vessels used in the analysis. Geomean is the unstandardized geometric mean of catch rates (kg/hr). Mth:DepCat is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Mth:DepCat | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 678.977 | 340 | 50.622 | 12 | 24.8664 | 1.1577 | 0.0000 |
| 1987 | 765.066 | 464 | 149.303 | 17 | 61.5525 | 1.7336 | 0.0815 |
| 1988 | 583.077 | 323 | 52.147 | 20 | 26.7665 | 0.9423 | 0.0856 |
| 1989 | 678.896 | 727 | 134.342 | 20 | 31.2668 | 0.9999 | 0.0787 |
| 1990 | 674.479 | 543 | 92.429 | 22 | 29.7271 | 0.9472 | 0.0789 |
| 1991 | 736.803 | 597 | 97.883 | 29 | 23.7829 | 0.8809 | 0.0774 |
| 1992 | 568.308 | 483 | 39.702 | 17 | 14.1316 | 0.6003 | 0.0805 |
| 1993 | 892.796 | 841 | 118.853 | 19 | 20.2159 | 0.9014 | 0.0760 |
| 1994 | 895.431 | 775 | 133.541 | 21 | 27.0651 | 1.1136 | 0.0761 |
| 1995 | 1208.893 | 1564 | 211.632 | 18 | 20.1818 | 1.1006 | 0.0723 |
| 1996 | 1233.265 | 1205 | 235.651 | 17 | 26.7059 | 1.2508 | 0.0751 |
| 1997 | 1696.855 | 1419 | 340.323 | 16 | 27.8818 | 1.3665 | 0.0739 |
| 1998 | 1591.988 | 1671 | 349.366 | 16 | 26.2074 | 1.3275 | 0.0736 |
| 1999 | 1651.572 | 1628 | 241.419 | 18 | 21.1431 | 0.9963 | 0.0734 |
| 2000 | 1507.379 | 2060 | 338.192 | 23 | 23.8936 | 1.0500 | 0.0730 |
| 2001 | 1392.822 | 2531 | 359.654 | 24 | 20.5368 | 0.9436 | 0.0726 |
| 2002 | 1330.296 | 2290 | 298.182 | 21 | 17.3590 | 0.7497 | 0.0726 |
| 2003 | 1353.243 | 1814 | 251.303 | 22 | 17.1223 | 0.7740 | 0.0735 |
| 2004 | 1495.581 | 1292 | 143.083 | 20 | 14.1120 | 0.5990 | 0.0749 |
| 2005 | 1203.256 | 966 | 114.114 | 18 | 14.2226 | 0.5955 | 0.0762 |
| 2006 | 1069.222 | 826 | 129.898 | 16 | 17.2693 | 0.7420 | 0.0772 |
| 2007 | 875.926 | 1254 | 221.488 | 15 | 20.4467 | 0.8823 | 0.0750 |
| 2008 | 980.268 | 806 | 151.663 | 14 | 24.2630 | 1.2079 | 0.0768 |
| 2009 | 775.047 | 965 | 200.785 | 13 | 24.1352 | 1.1482 | 0.0755 |
| 2010 | 906.088 | 947 | 182.003 | 10 | 22.1986 | 1.0373 | 0.0757 |
| 2011 | 1081.674 | 1104 | 241.948 | 12 | 22.2744 | 0.9520 | 0.0746 |



Figure 13.113. Pink Ling from zone 40 in depths between $350-800 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Pink Ling from zone 40 taken in the SET down to 1000 m , the top right graph depicts the depth distribution of shots containing Pink Ling from zone 40 in depths between $350-800 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zone 40 , the middle right hand graph depicts the number of vessels reporting Pink Ling catches through time. The bottom left reflects the number of records used in analysis, and bottom right is the Pink Ling catches (top line, black is total catches, all zones, all methods, the middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.114. Pink Ling from zone 40 in depths between 350 - 800 m by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates; giving a mean for the series of 1.0. The confidence intervals are wider due to the relatively low number of records.

Table 13.96. Pink Ling from zone 40 in depths between $350-800 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel +DepCat |
| Model 4 | LnCE $\sim$ Year+Vessel +DepCat +Month |
| Model 5 | LnCE $\sim$ Year+Vessel +DepCat +Month+DayNight |
| Model 6 | LnCE $\sim$ Year+Vessel +DepCat +Month +DayNight+ Vessel:Month |
| Model 7 | LnCE $\sim$ Year+Vessel +DepCat +Month +DayNight+ Month:DepCat |

Table 13.97. Pink Ling from zone 40 in depths between $350-800 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the adjusted $\mathrm{r}^{2}$ and the change in adjusted $\mathrm{r}^{2}$. The optimum is Month:DepCat (model 7).

|  | Year | Vessel | DepCat | Month | DayNight | Vessel:Month | Month:DepCat |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 3976 | 559 | -1884 | -3064 | -3107 | -2790 | -4314 |
| RSS | 33633 | 29652 | 27250 | 26022 | 25978 | 24669 | 24498 |
| MSS | 1745 | 5726 | 8128 | 9356 | 9400 | 10709 | 10880 |
| Nobs | 29435 | 29187 | 29187 | 29187 | 29187 | 29187 | 29187 |
| Npars | 26 | 49 | 60 | 143 | 146 | 1059 | 399 |
| adj_r2 | 4.852 | 16.047 | 22.819 | 26.085 | 26.202 | 27.646 | 29.797 |
| \%Change | 0.000 | 11.195 | 6.771 | 3.267 | 0.117 | 1.444 | 3.594 |



Figure 13.115. The relative influence of each factor used on the final trend in the optimal standardization for Pink Ling in Zone 40. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph for Vessel has the geometric mean (grey line) and the effect of adding Year + Vessel (model 2). In the third graph, for DepCat, the grey line represents model 2 and the black line the effect of adding DepCat to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model. Note that the influence, which is simply the deviations between the two lines squared, are not always reflective of the adj-r ${ }^{2}$ for each factor.

### 13.38 Pink Ling, Z50 (LIG - 37228002 - G. blacodes)

Data from zone 50 , depths greater than 200 m and less or equal to 800 m .
Table 13.98. Pink Ling from zone 50 in depths between $200-800 \mathrm{~m}$ by trawl. Total Catch is the total Pink Ling catch from all zones reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in this analysis, and Vessels relates to all vessels used in the analysis. Geomean is the unstandardized geometric mean of catch rates (kg/hr). Vessel:Mth is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Vessel:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 678.977 | 923 | 62.212 | 17 | 14.9346 | 1.0579 | 0.0000 |
| 1987 | 765.066 | 841 | 54.428 | 23 | 14.1775 | 1.0217 | 0.0456 |
| 1988 | 583.077 | 701 | 42.836 | 25 | 14.5280 | 1.1529 | 0.0476 |
| 1989 | 678.896 | 729 | 45.389 | 25 | 15.3818 | 1.1792 | 0.0469 |
| 1990 | 674.479 | 957 | 47.873 | 18 | 11.9104 | 0.9867 | 0.0469 |
| 1991 | 736.803 | 1294 | 100.787 | 20 | 13.8309 | 1.1260 | 0.0434 |
| 1992 | 568.308 | 1150 | 62.462 | 17 | 11.1987 | 0.8732 | 0.0439 |
| 1993 | 892.796 | 1410 | 116.532 | 12 | 15.5287 | 1.1594 | 0.0429 |
| 1994 | 895.431 | 1335 | 114.252 | 14 | 17.5302 | 1.3934 | 0.0428 |
| 1995 | 1208.893 | 1950 | 214.425 | 18 | 19.9408 | 1.5556 | 0.0407 |
| 1996 | 1233.265 | 2197 | 211.853 | 23 | 17.0478 | 1.4577 | 0.0403 |
| 1997 | 1696.855 | 2311 | 236.711 | 21 | 17.8914 | 1.5065 | 0.0399 |
| 1998 | 1591.988 | 2039 | 209.275 | 18 | 19.7137 | 1.5719 | 0.0406 |
| 1999 | 1651.572 | 2159 | 186.384 | 17 | 16.0778 | 1.2620 | 0.0402 |
| 2000 | 1507.379 | 2587 | 170.657 | 19 | 12.1381 | 0.9971 | 0.0400 |
| 2001 | 1392.822 | 2504 | 138.777 | 21 | 10.5409 | 0.8417 | 0.0402 |
| 2002 | 1330.296 | 2318 | 129.610 | 20 | 10.4073 | 0.8121 | 0.0403 |
| 2003 | 1353.243 | 1991 | 108.241 | 20 | 9.6163 | 0.7802 | 0.0406 |
| 2004 | 1495.581 | 2589 | 162.033 | 20 | 10.7076 | 0.7658 | 0.0401 |
| 2005 | 1203.256 | 1689 | 80.704 | 19 | 8.0776 | 0.5846 | 0.0417 |
| 2006 | 1069.222 | 1494 | 79.938 | 17 | 8.1572 | 0.5741 | 0.0427 |
| 2007 | 875.926 | 1270 | 64.909 | 13 | 7.8759 | 0.5849 | 0.0434 |
| 2008 | 980.268 | 987 | 62.435 | 14 | 9.6601 | 0.7181 | 0.0447 |
| 2009 | 775.047 | 1009 | 58.834 | 9 | 8.3008 | 0.6471 | 0.0446 |
| 2010 | 906.088 | 1382 | 89.591 | 12 | 9.1906 | 0.7152 | 0.0430 |
| 2011 | 1081.674 | 1611 | 103.784 | 13 | 9.1207 | 0.6748 | 0.0434 |



Figure 13.116. Pink Ling from zone 50 in depths between 200 - 800 m by trawl. The top left is the depth distribution of all records reporting Pink Ling from zone 50 taken in the SET down to 1000 m , the top right graph depicts the depth distribution of shots containing Pink Ling from zone 50 in depths between $200-800 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zone 50 , the middle right hand graph depicts the number of vessels reporting Pink Ling catches through time. The bottom left reflects the number of records used in analysis, and bottom right is the Pink Ling catches (top line, black is total catches, all zones, all methods, the middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.117. Pink Ling from zone 50 in depths between 200 - 800 m by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates; giving a mean for the series of 1.0. The confidence intervals are wider due to the relatively low number of records.

Table 13.99. Pink Ling from zone 50 in depths between $200-800 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel +DepCat |
| Model 4 | LnCE $\sim$ Year+Vessel +DepCat +Month |
| Model 5 | LnCE $\sim$ Year+Vessel +DepCat +Month+DayNight |
| Model 6 | LnCE $\sim$ Year+Vessel +DepCat +Month +DayNight+ Vessel:Month |
| Model 7 | LnCE $\sim$ Year+Vessel +DepCat +Month +DayNight+ Month:DepCat |

Table 13.100. Pink Ling from zone 50 in depths between 200 - 800m by trawl. Model selection criteria, including the AIC, the adjusted $\mathrm{r}^{2}$ and the change in adjusted $\mathrm{r}^{2}$. The optimum is Vessel:Month (model 6).

|  | Year | Vessel | DepCat | Month | DayNight | Vessel:Month | Month:DepCat |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | - | -18069 | -20247 | -20460 | -20582 | -20806 | -20804 |
| RSS | 13213 | 30076 | 26587 | 25132 | 24990 | 24912 | 23832 |
| MSS | 3573 | 7062 | 8517 | 8659 | 8737 | 9817 | 24386 |
| Nobs | 41427 | 41294 | 41294 | 41294 | 41294 | 41294 | 41293 |
| Npars | 26 | 56 | 129 | 140 | 143 | 946 | 473 |
| adj_r2 | 10.564 | 20.881 | 25.079 | 25.483 | 25.710 | 27.515 | 26.690 |
| \%Change | 0.000 | 10.317 | 4.198 | 0.404 | 0.226 | 1.805 | -0.825 |



Figure 13.118. The relative influence of each factor used on the final trend in the optimal standardization for Pink Ling in Zone 50. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph for Vessel has the geometric mean (grey line) and the effect of adding Year + Vessel (model 2). In the third graph, for DepCat, the grey line represents model 2 and the black line the effect of adding DepCat to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model. Note that the influence, which is simply the deviations between the two lines squared, are not always reflective of the adj-r ${ }^{2}$ for each factor.

### 13.39 Western Gemfish and GAB (GEM - 37439002 - Rexea solandri)

Data from zones 40 and 50 with $82,83,84$, and 85 (the GAB), depths greater than 100 and less than or equal to 600 m .

Table 13.101. Western Gemfish from zones 40 and 50, and the GAB (zones $82,83,84$, and 85 ) in depths between $200-600 \mathrm{~m}$ by trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis., and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr)}$ ). Zone:Mth is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 3639.955 | 1721 | 308.061 | 25 | 28.8362 | 2.1398 | 0.0000 |
| 1987 | 4660.447 | 1284 | 262.356 | 29 | 30.7827 | 2.1151 | 0.0460 |
| 1988 | 3515.819 | 1427 | 261.309 | 36 | 25.6522 | 1.9795 | 0.0478 |
| 1989 | 1778.325 | 1405 | 184.753 | 38 | 19.0566 | 1.5062 | 0.0489 |
| 1990 | 1206.897 | 1261 | 146.900 | 38 | 14.3866 | 1.3103 | 0.0527 |
| 1991 | 580.322 | 1592 | 280.530 | 35 | 19.1105 | 1.2962 | 0.0493 |
| 1992 | 494.441 | 801 | 96.906 | 21 | 15.0886 | 0.9534 | 0.0567 |
| 1993 | 353.410 | 902 | 109.371 | 21 | 11.5160 | 0.8132 | 0.0556 |
| 1994 | 232.179 | 1053 | 110.188 | 26 | 11.3093 | 0.8341 | 0.0531 |
| 1995 | 181.746 | 1316 | 107.533 | 26 | 9.0719 | 0.7977 | 0.0506 |
| 1996 | 382.196 | 1631 | 164.827 | 32 | 9.5592 | 0.9324 | 0.0488 |
| 1997 | 571.976 | 2106 | 215.362 | 28 | 8.9766 | 0.8369 | 0.0470 |
| 1998 | 404.817 | 1967 | 206.881 | 26 | 10.1690 | 1.0076 | 0.0479 |
| 1999 | 448.677 | 2347 | 323.256 | 25 | 11.9957 | 1.0051 | 0.0467 |
| 2000 | 336.464 | 2357 | 260.267 | 31 | 9.5636 | 0.8284 | 0.0472 |
| 2001 | 331.486 | 2335 | 255.222 | 31 | 9.9454 | 0.7806 | 0.0473 |
| 2002 | 196.526 | 1770 | 129.588 | 29 | 6.4625 | 0.6004 | 0.0490 |
| 2003 | 269.227 | 1642 | 203.076 | 34 | 8.8216 | 0.6743 | 0.0497 |
| 2004 | 525.201 | 1952 | 434.958 | 32 | 10.3074 | 0.7246 | 0.0497 |
| 2005 | 498.511 | 1816 | 359.400 | 27 | 12.3888 | 0.7247 | 0.0503 |
| 2006 | 509.019 | 1599 | 399.243 | 26 | 11.5504 | 0.6696 | 0.0514 |
| 2007 | 561.238 | 1412 | 382.551 | 22 | 10.3604 | 0.6329 | 0.0523 |
| 2008 | 289.477 | 1265 | 152.175 | 21 | 6.6254 | 0.6486 | 0.0527 |
| 2009 | 194.843 | 1275 | 105.771 | 16 | 5.8778 | 0.6888 | 0.0525 |
| 2010 | 220.639 | 1703 | 129.526 | 18 | 6.0572 | 0.7331 | 0.0500 |
| 2011 | 147.321 | 1348 | 75.776 | 17 | 5.4642 | 0.7665 | 0.0528 |



Figure 13.119. Western Gemfish from zones 40 and 50 , and the GAB (zones $82,83,84$, and 85 ) in depths between $200-600 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Gemfish, the top right graph depicts the depth distribution of shots containing Western Gemfish from zones 40 and 50, and the GAB (zones $82,83,84$, and 85 ) in depths between $200-600 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zones, the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Gemfish catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.120. Western Gemfish from zones 40 and 50, and the GAB (zones 82, 83, 84, and 85) in depths between $200-600 \mathrm{~m}$ by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.102. Western Gemfish from zones 40 and 50, and the GAB (zones 82, 83, 84, and 85) in depths between $200-600 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE~Year |
| :--- | :--- |
| Model 2 | LnCE~Year+DepCat |
| Model 3 | LnCE~Year+DepCat+Vessel |
| Model 4 | LnCE~Year+DepCat+Vessel+Zone |
| Model 5 | LnCE~Year+DepCat+Vessel+Zone+DayNight |
| Model 6 | LnCE~Year+DepCat+Vessel+Zone+DayNight + Month |
| Model 7 | LnCE~Year+DepCat+Vessel+Zone+DayNight + Month+Zone:Month |
| Model 8 | LnCE $\sim$ Year+DepCat+Vessel+Zone + DayNight + Month + Zone:DepCat |

Table 13.103. Western Gemfish from zones 40 and 50 , and the GAB (zones 82, 83, 84, and 85) in depths between $200-600 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the deviance and the change in deviance. The optimum is Zone:Month (model 7).

|  | Year | DepCat | Vessel | Zone | DayNight | Month | Zone:Mt | Zone:DepC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AIC | 35504 | 21879 | 14758 | 13979 | 13312 | 12939 | 12009 | 12326 |
| RSS | 97439 | 69841 | 58432 | 57321 | 56391 | 55853 | 54459 | 54693 |
| MSS | 8043 | 35641 | 47050 | 48160 | 49091 | 49629 | 51023 | 50789 |
| Nobs | 41287 | 41130 | 41130 | 41130 | 41130 | 41130 | 41130 | 41130 |
| Npars | 26 | 51 | 158 | 163 | 166 | 177 | 232 | 302 |
| adj_r2 | 7.569 | 33.708 | 44.392 | 45.443 | 46.324 | 46.822 | 48.080 | 47.767 |
| \%Change | 0.000 | 26.139 | 10.684 | 1.051 | 0.881 | 0.498 | 1.257 | -0.313 |



Figure 13.121. The relative influence of each factor used on the final trend in the optimal standardization for Western Gemfish from zones 40 and 50 and the GAB. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.40 Western Gemfish Z4050 (GEM - 37439002 - R. solandri)

Data from zones 40 and 50, depths greater than 200 and less than or equal to 600 m .
Table 13.104. Western Gemfish from zones 40 and 50 in depths between $200-600 \mathrm{~m}$ by trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis., and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr})$. Zone:Mth is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 3639.955 | 1687 | 306.861 | 24 | 29.5835 | 2.2392 | 0.0000 |
| 1987 | 4660.447 | 1209 | 248.879 | 26 | 31.5896 | 2.2272 | 0.0451 |
| 1988 | 3515.819 | 1235 | 226.956 | 27 | 26.9924 | 2.1712 | 0.0472 |
| 1989 | 1778.325 | 1082 | 156.578 | 29 | 23.3363 | 1.8043 | 0.0495 |
| 1990 | 1206.897 | 1057 | 136.085 | 29 | 15.9031 | 1.3602 | 0.0528 |
| 1991 | 580.322 | 1384 | 249.415 | 28 | 22.0062 | 1.3188 | 0.0493 |
| 1992 | 494.441 | 665 | 80.930 | 15 | 16.7792 | 0.9226 | 0.0575 |
| 1993 | 353.410 | 718 | 102.489 | 17 | 16.5820 | 0.8868 | 0.0570 |
| 1994 | 232.179 | 839 | 95.378 | 20 | 16.2263 | 0.9566 | 0.0543 |
| 1995 | 181.746 | 990 | 84.688 | 21 | 12.0017 | 0.8385 | 0.0520 |
| 1996 | 382.196 | 1182 | 145.588 | 26 | 13.4563 | 0.9255 | 0.0499 |
| 1997 | 571.976 | 1389 | 153.589 | 21 | 13.2702 | 0.8307 | 0.0484 |
| 1998 | 404.817 | 1259 | 121.661 | 20 | 13.2167 | 0.8937 | 0.0498 |
| 1999 | 448.677 | 1694 | 176.323 | 19 | 12.8407 | 0.8486 | 0.0474 |
| 2000 | 336.464 | 1932 | 228.165 | 27 | 12.4996 | 0.8705 | 0.0475 |
| 2001 | 331.486 | 1694 | 169.890 | 27 | 12.1589 | 0.7042 | 0.0484 |
| 2002 | 196.526 | 1418 | 86.261 | 24 | 7.1243 | 0.5364 | 0.0496 |
| 2003 | 269.227 | 1077 | 123.722 | 24 | 11.3050 | 0.6520 | 0.0521 |
| 2004 | 525.201 | 1232 | 105.674 | 24 | 7.9049 | 0.6352 | 0.0522 |
| 2005 | 498.511 | 1073 | 117.678 | 18 | 10.6004 | 0.6620 | 0.0532 |
| 2006 | 509.019 | 889 | 101.417 | 18 | 8.9869 | 0.5380 | 0.0560 |
| 2007 | 561.238 | 715 | 61.053 | 16 | 7.4717 | 0.5238 | 0.0583 |
| 2008 | 289.477 | 770 | 53.096 | 16 | 7.5220 | 0.5815 | 0.0572 |
| 2009 | 194.843 | 925 | 56.810 | 12 | 6.4871 | 0.6658 | 0.0546 |
| 2010 | 220.639 | 1364 | 86.888 | 14 | 6.3681 | 0.6892 | 0.0508 |
| 2011 | 147.321 | 1125 | 54.897 | 13 | 5.5076 | 0.7175 | 0.0536 |



Figure 13.122. Western Gemfish from zones 40 and 50 in depths between $200-600 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Gemfish, the top right graph depicts the depth distribution of shots containing Western Gemfish from zones 40 and 50 in depths between $200-600 \mathrm{~m}$ by trawl.. The middle left diagram depicts the distribution of catch by depth within zones 40 and 50 ( 50 is top red line), the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Gemfish catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.123. Western Gemfish from zones 40 and 50 in depths between $200-600 \mathrm{~m}$ by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.105. Western Gemfish from zones 40 and 50 in depths between $200-600 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel+DepCat |
| Model 4 | LnCE $\sim$ Year+Vessel+DepCat+DayNight |
| Model 5 | LnCE $\sim$ Year+Vessel+DepCat+DayNight + Month |
| Model 6 | LnCE $\sim$ Year+Vessel+DepCat+DayNight + Month+Zone |
| Model 7 | LnCE $\sim$ Year+Vessel+DepCat+DayNight + Month+Zone+Zone:Month |
| Model 8 | LnCE $\sim$ Year+Vessel+DepCat + DayNight + Month+Zone+Zone:DepCat |

Table 13.106. Western Gemfish from zones 40 and 50 in depths between $200-600 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the adjusted $r^{2}$ and the change in adjusted $r^{2}$. The optimum is Zone:Month (model 7).

|  | Year | Vessel | DepCat | DayNight | Month | Zone |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | Zone:Mth Zone:DepC |  |  |  |  |  |  |  |
| AIC | 20863 | 13653 | 7176 | 6695 | 6416 | 6418 | 6246 | 6274 |
| RSS | 60408 | 47453 | 38240 | 37634 | 37264 | 37264 | 37028 | 37028 |
| MSS | 7119 | 20075 | 29287 | 29893 | 30263 | 30264 | 30500 | 30500 |
| Nobs | 30604 | 30604 | 30502 | 30502 | 30502 | 30502 | 30502 | 30502 |
| Npars | 26 | 115 | 140 | 143 | 154 | 155 | 166 | 180 |
| adj_r2 | 10.470 | 29.466 | 43.112 | 44.008 | 44.538 | 44.537 | 44.868 | 44.842 |
| \%Change | 0.000 | 18.996 | 13.646 | 0.896 | 0.531 | -0.002 | 0.332 | -0.026 |



Figure 13.124. The relative influence of each factor used on the final trend in the optimal standardization for Western Gemfish from zones 40 and 50 . The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.41 Western Gemfish GAB (GEM - 37439002 - R. solandri)

Data from zones $82,83,84$, and 85 (the GAB), depths greater than 100 and less than or equal to 600 m . All vessels included

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | 181.746 | 326 | 22.845 | 6 | 3.8779 | 0.7240 | 0.0000 |
| 1996 | 382.196 | 449 | 19.239 | 7 | 3.8858 | 0.9574 | 0.0937 |
| 1997 | 571.976 | 717 | 61.773 | 9 | 4.2096 | 0.9576 | 0.0890 |
| 1998 | 404.817 | 708 | 85.220 | 8 | 6.3801 | 1.5654 | 0.0910 |
| 1999 | 448.677 | 653 | 146.933 | 7 | 10.0539 | 1.8525 | 0.0937 |
| 2000 | 336.464 | 425 | 32.102 | 6 | 2.8318 | 0.6743 | 0.0996 |
| 2001 | 331.486 | 641 | 85.332 | 8 | 5.8477 | 1.1128 | 0.0942 |
| 2002 | 196.526 | 352 | 43.326 | 8 | 4.3633 | 0.9821 | 0.1025 |
| 2003 | 269.227 | 565 | 79.354 | 11 | 5.4980 | 0.9045 | 0.0981 |
| 2004 | 525.201 | 720 | 329.284 | 10 | 16.2315 | 1.1274 | 0.0987 |
| 2005 | 498.511 | 743 | 241.723 | 10 | 15.5168 | 0.9565 | 0.0999 |
| 2006 | 509.019 | 709 | 297.706 | 11 | 15.7716 | 0.9620 | 0.0986 |
| 2007 | 561.238 | 697 | 321.498 | 10 | 14.4877 | 0.8490 | 0.0970 |
| 2008 | 289.477 | 495 | 99.079 | 7 | 5.4384 | 0.8573 | 0.0989 |
| 2009 | 194.843 | 350 | 48.961 | 4 | 4.5291 | 0.7948 | 0.1054 |
| 2010 | 220.639 | 339 | 42.638 | 4 | 4.9524 | 0.8698 | 0.1060 |
| 2011 | 147.321 | 223 | 20.879 | 4 | 5.2504 | 0.8526 | 0.1181 |



Figure 13.125. Western Gemfish in the GAB (zones 82, 83, 84, and 85) in depths between 100 and 600 m by trawl. The top left is the depth distribution of all records reporting western gemfish, the top right graph depicts the depth distribution of shots containing Western Gemfish in the GAB (zones 82, 83, 84, and 85) in depths between 100 and 600 m by trawl. The middle left diagram depicts the distribution of catch by depth within zones 82 and 85 , the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the western Gemfish catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.126. Western Gemfish in the GAB (zones 82, 83, 84, and 85) in depths between 100 and 600 m by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.108. Western Gemfish in the GAB (zones 82, 83, 84, and 85) in depths between 100 and 600 m by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.
Model 1 LnCE~Year
Model 2 LnCE $\sim$ Year + DepCat
Model 3 LnCE $\sim$ Year + DepCat+Vessel
Model $4 \quad$ LnCE~Year+DepCat+Vessel+Month
Model 5 LnCE $\sim$ Year + DepCat + Vessel + Month + DayNight
Model 6 LnCE~Year+DepCat+Vessel+Month+DayNight+Zone
Model 7 LnCE $\sim$ Year+DepCat+Vessel+Month+DayNight+Zone+Zone:Month
Model $8 \quad$ LnCE $\sim$ Year + DepCat+Vessel+Month+DayNight+Zone+Zone:DepCat

Table 13.109. Western Gemfish in the GAB (zones 82, 83, 84, and 85) in depths between 100 and 600 m by trawl. Model selection criteria, including the AIC, the adjusted $r^{2}$ and the change in adjusted $r^{2}$. The optimum is Zone:Month (model 7).

|  | Year | DepCat | Vessel | Month | DayNight | Zone | Zone:Mth | Zone:DepC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AIC | 10454 | 6747 | 5422 | 4823 | 4572 | 4338 | 4054 | 4268 |
| RSS | 28592 | 18909 | 16253 | 15178 | 14753 | 14368 | 13824 | 14024 |
| MSS | 2986 | 12668 | 15324 | 16399 | 16824 | 17210 | 17753 | 17553 |
| Nobs | 9112 | 9071 | 9071 | 9071 | 9071 | 9071 | 9071 | 9071 |
| Npars | 17 | 42 | 66 | 77 | 80 | 83 | 116 | 158 |
| adj_r2 | 9.296 | 39.847 | 48.156 | 51.528 | 52.869 | 54.085 | 55.660 | 54.806 |
| \%Change | 0.000 | 30.551 | 8.310 | 3.372 | 1.341 | 1.216 | 1.575 | -0.854 |



Figure 13.127. The relative influence of each factor used on the final trend in the optimal standardization for Western Gemfish in the GAB (zones $82,83,84$, and 85 ). The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.42 Offshore Ocean Perch, Z1020 (REG - 37287001 - H. percoides) 200 m

In the November 2009 Slope RAG meeting the depth distribution of offshore Ocean Perch was revised to $300-700 \mathrm{~m}$ to avoid overlap with inshore Ocean Perch; however, this decision was reversed in 2010 and so the analysis was repeated using 200-700 m.

Table 13.110. Offshore Ocean Perch from zones 10 and 20 in depths $200-700 \mathrm{~m}$ by trawl. Total Catch is
the total reported in the database, Records is the number of records used in the analysis, CatchT is the
reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the
analysis. Geomean is the geometric mean of catch rates $(\mathrm{kg} / \mathrm{hr})$. Zone:Month is the optimum model and
StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 262.446 | 3479 | 207.363 | 77 | 12.1440 | 1.0298 | 0.0000 |
| 1987 | 198.347 | 3140 | 132.797 | 70 | 8.9237 | 0.9538 | 0.0254 |
| 1988 | 186.712 | 2808 | 150.765 | 73 | 10.5074 | 1.0669 | 0.0264 |
| 1989 | 206.258 | 3036 | 160.004 | 67 | 10.6494 | 1.0257 | 0.0263 |
| 1990 | 180.560 | 1970 | 115.943 | 57 | 12.0207 | 1.3644 | 0.0295 |
| 1991 | 223.188 | 2093 | 138.991 | 53 | 13.4339 | 1.4423 | 0.0292 |
| 1992 | 169.669 | 1845 | 114.079 | 47 | 11.9264 | 1.2143 | 0.0301 |
| 1993 | 259.310 | 2924 | 199.186 | 53 | 12.9555 | 1.2142 | 0.0268 |
| 1994 | 257.241 | 3014 | 180.955 | 49 | 11.8001 | 1.1325 | 0.0265 |
| 1995 | 239.951 | 3146 | 150.341 | 50 | 10.4874 | 1.0249 | 0.0262 |
| 1996 | 263.235 | 3411 | 176.808 | 53 | 9.8364 | 0.9240 | 0.0258 |
| 1997 | 296.334 | 3725 | 193.773 | 54 | 9.7119 | 0.9739 | 0.0256 |
| 1998 | 292.098 | 3850 | 194.629 | 49 | 9.4285 | 0.8662 | 0.0253 |
| 1999 | 290.643 | 4406 | 219.065 | 52 | 9.7566 | 0.9802 | 0.0250 |
| 2000 | 269.827 | 4178 | 180.750 | 52 | 7.5464 | 0.7702 | 0.0255 |
| 2001 | 281.541 | 4038 | 183.911 | 43 | 8.3956 | 0.8632 | 0.0257 |
| 2002 | 255.307 | 3646 | 150.622 | 45 | 7.3709 | 0.8206 | 0.0264 |
| 2003 | 322.581 | 3960 | 185.006 | 53 | 7.6242 | 0.8719 | 0.0261 |
| 2004 | 315.869 | 3129 | 150.459 | 46 | 8.0648 | 0.8707 | 0.0275 |
| 2005 | 316.769 | 3089 | 170.080 | 46 | 9.3641 | 0.9783 | 0.0273 |
| 2006 | 237.601 | 2326 | 113.168 | 39 | 7.8433 | 0.8351 | 0.0292 |
| 2007 | 180.579 | 1528 | 94.900 | 22 | 9.9183 | 1.0332 | 0.0329 |
| 2008 | 184.267 | 1843 | 101.836 | 23 | 9.1917 | 0.9554 | 0.0315 |
| 2009 | 173.879 | 1694 | 99.608 | 23 | 9.0355 | 0.9499 | 0.0324 |
| 2010 | 195.594 | 1759 | 118.107 | 21 | 9.8647 | 0.9792 | 0.0319 |
| 2011 | 186.639 | 1874 | 116.696 | 22 | 9.0998 | 0.8592 | 0.0314 |



Figure 13.128. Offshore Ocean Perch from zones 10 and 20 in depths $200-700 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Ocean perch, the top right graph depicts the depth distribution of shots containing Offshore Ocean Perch from zones 10 and 20 in depths $200-700 \mathrm{~m}$ by trawl. . The middle left diagram depicts the distribution of catch by depth within zones 10 and 20 ( 20 is top red line), the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Ocean Perch catches (top line, black is total catches, middle line, red, are those used in the analysis).


Figure 13.129. Offshore Ocean Perch from zones 10 and 20 in depths 200 - 700m by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.111. Offshore Ocean Perch from zones 10 and 20 in depths $200-700 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+DepCat |
| Model 3 | LnCE $\sim$ Year+DepCat+Vessel |
| Model 4 | LnCE $\sim$ Year+DepCat+Vessel+Month |
| Model 5 | LnCE $\sim$ Year+DepCat+Vessel+Month+DayNight |
| Model 6 | LnCE $\sim$ Year+DepCat+Vessel+Month+DayNight+Zone |
| Model 7 | LnCE $\sim$ Year+DepCat+Vessel+Month+DayNight+Zone+Zone:Month |
| Model 8 | LnCE $\sim$ Year+DepCat+Vessel+Month+DayNight+Zone+Zone:DepCat |

Table 13.112. Offshore Ocean Perch from zones 10 and 20 in depths $200-700 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the adjusted $r^{2}$ and the change in adjusted $r^{2}$. The optimum is Zone:Month (model 7).

|  | Year |  | DepCat | Vessel | Month | DayNight | Zone |  |  | Zone:Mth | Zone:DepC |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: | :---: |
| AIC | 20116 | 8803 | 571 | -1484 | -1722 | -1751 | -3645 | -2100 |  |  |  |
| RSS | 98877 | 84734 | 75671 | 73618 | 73381 | 73351 | 71513 | 72964 |  |  |  |
| MSS | 2097 | 16240 | 25303 | 27356 | 27594 | 27623 | 29462 | 28011 |  |  |  |
| Nobs | 75911 | 75512 | 75512 | 75512 | 75512 | 75512 | 75512 | 75512 |  |  |  |
| Npars | 26 | 51 | 206 | 217 | 220 | 221 | 232 | 246 |  |  |  |
| adj_r2 | 2.045 | 16.028 | 24.855 | 26.883 | 27.116 | 27.144 | 28.960 | 27.505 |  |  |  |
| \%Change | 0.000 | 13.983 | 8.827 | 2.028 | 0.233 | 0.028 | 1.816 | -1.455 |  |  |  |



Figure 13.130. The relative influence of each factor used on the final trend in the optimal standardization for Offshore Ocean Perch from zones 10 and 20. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.


Figure 13.131. Offshore Ocean Perch, depths $>200$ for trawl and AutoLongLine, in zones 10 and 20. Catches through time taken by trawl and by AutoLongLine. Some of the decline in trawl catches in recent years have been made up by the AutoLong Lining. Geometric mean catch rates for Offshore Ocean Perch in depth $200-700$ metres for both trawl and autolongline; scaled to the mean of each series for comparison.


Figure 13.132. Depth distribution of catches of Offshore Ocean Perch, depths 200-700 for trawl, 01000 m for AutoLongLine. Most catches by AutoLongLine are taken in the same depths as trawl catches.

### 13.43 Inshore Ocean Perch, Z1020 (REG - 37287001 - H. percoides) 0-200m

In the November 2009 Slope RAG meeting a separate analysis was required for the Inshore Ocean Perch. These were defined as all those Ocean Perch reported as caught between $0-299 \mathrm{~m}$ to avoid overlap with Offshore Ocean Perch. However, in 2010 this decision was reversed and so the analysis was repeated for depths $0-200 \mathrm{~m}$.

[^3]| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:DepC | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 262.446 | 339 | 15.239 | 50 | 6.8543 | 0.8363 | 0.0000 |
| 1987 | 198.347 | 406 | 11.971 | 58 | 5.9511 | 0.9828 | 0.0919 |
| 1988 | 186.712 | 518 | 16.548 | 59 | 7.2891 | 1.1184 | 0.0884 |
| 1989 | 206.258 | 443 | 15.392 | 52 | 8.0367 | 1.0713 | 0.0924 |
| 1990 | 180.560 | 450 | 15.614 | 45 | 7.7738 | 1.1429 | 0.0936 |
| 1991 | 223.188 | 498 | 20.364 | 43 | 8.1374 | 1.2818 | 0.0927 |
| 1992 | 169.669 | 258 | 13.830 | 28 | 9.5229 | 1.6960 | 0.1042 |
| 1993 | 259.310 | 467 | 25.080 | 38 | 10.1873 | 1.9063 | 0.0956 |
| 1994 | 257.241 | 558 | 23.340 | 35 | 9.4326 | 1.7357 | 0.0925 |
| 1995 | 239.951 | 600 | 21.200 | 35 | 8.7548 | 1.2840 | 0.0902 |
| 1996 | 263.235 | 688 | 21.307 | 39 | 7.0539 | 1.1194 | 0.0895 |
| 1997 | 296.334 | 572 | 16.365 | 40 | 5.9056 | 1.0464 | 0.0923 |
| 1998 | 292.098 | 646 | 15.628 | 41 | 5.7524 | 0.9151 | 0.0910 |
| 1999 | 290.643 | 675 | 15.978 | 40 | 4.9974 | 0.8112 | 0.0901 |
| 2000 | 269.827 | 1326 | 30.551 | 39 | 4.5708 | 0.9859 | 0.0861 |
| 2001 | 281.541 | 1035 | 23.397 | 34 | 4.2075 | 0.9800 | 0.0878 |
| 2002 | 255.307 | 1422 | 25.185 | 36 | 2.6164 | 0.6996 | 0.0866 |
| 2003 | 322.581 | 1085 | 17.438 | 40 | 2.3132 | 0.5408 | 0.0875 |
| 2004 | 315.869 | 962 | 15.461 | 41 | 2.2440 | 0.5522 | 0.0891 |
| 2005 | 316.769 | 898 | 19.849 | 41 | 2.9880 | 0.6250 | 0.0898 |
| 2006 | 237.601 | 602 | 9.339 | 35 | 2.2501 | 0.5206 | 0.0929 |
| 2007 | 180.579 | 395 | 8.745 | 21 | 3.5455 | 0.7329 | 0.0991 |
| 2008 | 184.267 | 330 | 7.969 | 21 | 4.2486 | 0.9001 | 0.1025 |
| 2009 | 173.879 | 289 | 6.671 | 21 | 4.1335 | 0.7656 | 0.1065 |
| 2010 | 195.594 | 307 | 7.136 | 21 | 3.8363 | 0.8049 | 0.1052 |
| 2011 | 186.639 | 275 | 6.431 | 19 | 3.6642 | 0.9448 | 0.1074 |



Figure 13.133. Inshore Ocean Perch from zones 10 and 20 in depths $0-200 \mathrm{~m}$ by trawl. The top left is the depth distribution of all records reporting Ocean Perch, the top right graph depicts the depth distribution of shots containing Inshore Ocean Perch from zones 10 and 20 in depths $0-200 \mathrm{~m}$ by trawl. The middle left diagram depicts the distribution of catch by depth within zones 10 and 20 ( 20 is top red line), the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Ocean Perch catches (top line, black is total catches, middle line, red, are those used in the analysis.


Figure 13.134. Inshore Ocean Perch from zones 10 and 20 in depths $0-200 \mathrm{~m}$ by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.114. Inshore Ocean Perch from zones 10 and 20 in depths $0-200 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE~Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel+DepCat |
| Model 4 | LnCE $\sim$ Year+Vessel+DepCat+Month |
| Model 5 | LnCE $\sim$ Year+Vessel+DepCat+Month+DayNight |
| Model 6 | LnCE $\sim$ Year+Vessel+DepCat+Month+DayNight+Zone |
| Model 7 | LnCE $\sim$ Year+Vessel+DepCat+Month+DayNight+Zone+Zone:Month |
| Model 8 | LnCE $\sim$ Year+Vessel+DepCat+Month+DayNight+Zone+Zone:DepCat |

Table 13.115. Inshore Ocean Perch from zones 10 and 20 in depths $0-200 \mathrm{~m}$ by trawl. Model selection criteria, including the AIC, the adjusted $r^{2}$ and the change in adjusted $r^{2}$. The optimum is Zone:DepCat (model 8).

|  | Year | Vessel | DepCat | Month | DayNight | Zone |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 5669 | 2214 | 1345 | 1274 | 1227 | 1155 | 1157 | 1065 |
| AIC | 5669 | 1650 | Zone:DepC |  |  |  |  |  |
| RSS | 22769 | 18036 | 16666 | 16567 | 16511 | 16433 | 16412 | 16318 |
| MSS | 3779 | 8512 | 9883 | 9981 | 10037 | 10116 | 10137 | 10230 |
| Nobs | 16044 | 16044 | 15645 | 15645 | 15645 | 15645 | 15645 | 15645 |
| Npars | 26 | 168 | 178 | 189 | 192 | 193 | 204 | 203 |
| adj_r2 | 14.101 | 31.347 | 36.507 | 36.836 | 37.039 | 37.333 | 37.369 | 37.731 |
| \%Change | 0.000 | 17.246 | 5.160 | 0.329 | 0.203 | 0.295 | 0.035 | 0.362 |



Figure 13.135. The relative influence of each factor used on the final trend in the optimal standardization for Inshore Ocean Perch from zones 10 and 20. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.44 John Dory (DOJ - 37264004) Zeus faber

Zones 10 and 20 in depths $0-200 \mathrm{~m}$

Table 13.116. John Dory from Zones 10 and 20 in depths 0 to 200 m by trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr)}$ ). Zone:DepCat is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:DepC | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 231.715 | 6418 | 202.235 | 90 | 7.6948 | 1.5483 | 0.0000 |
| 1987 | 206.090 | 4663 | 181.591 | 78 | 8.5155 | 1.7713 | 0.0208 |
| 1988 | 181.984 | 4538 | 161.563 | 73 | 8.3856 | 1.6647 | 0.0211 |
| 1989 | 217.924 | 4813 | 188.443 | 70 | 9.5319 | 1.8250 | 0.0210 |
| 1990 | 167.853 | 3700 | 136.764 | 60 | 8.7451 | 1.6491 | 0.0230 |
| 1991 | 172.291 | 4041 | 126.696 | 53 | 7.1954 | 1.3564 | 0.0226 |
| 1992 | 130.849 | 3809 | 100.026 | 48 | 5.6282 | 1.1209 | 0.0231 |
| 1993 | 240.438 | 5446 | 181.622 | 56 | 7.0963 | 1.4522 | 0.0214 |
| 1994 | 267.868 | 6573 | 209.897 | 55 | 6.7516 | 1.3690 | 0.0204 |
| 1995 | 185.672 | 6070 | 168.531 | 52 | 5.9610 | 1.1594 | 0.0205 |
| 1996 | 160.753 | 6411 | 146.769 | 59 | 4.5279 | 0.8950 | 0.0204 |
| 1997 | 87.766 | 4473 | 79.224 | 60 | 3.3776 | 0.7008 | 0.0224 |
| 1998 | 109.029 | 5091 | 98.479 | 53 | 3.6350 | 0.7274 | 0.0215 |
| 1999 | 132.842 | 5553 | 121.021 | 56 | 3.9411 | 0.8410 | 0.0212 |
| 2000 | 164.053 | 7094 | 147.876 | 58 | 3.5716 | 0.7920 | 0.0203 |
| 2001 | 129.300 | 6789 | 116.224 | 51 | 2.9450 | 0.6650 | 0.0205 |
| 2002 | 150.974 | 6670 | 136.130 | 49 | 3.1506 | 0.6556 | 0.0208 |
| 2003 | 156.740 | 6559 | 137.336 | 51 | 3.1538 | 0.6388 | 0.0207 |
| 2004 | 165.858 | 7093 | 147.526 | 51 | 3.4191 | 0.6745 | 0.0204 |
| 2005 | 107.390 | 4934 | 88.640 | 48 | 2.6772 | 0.5638 | 0.0222 |
| 2006 | 85.401 | 3727 | 71.625 | 43 | 2.8463 | 0.6354 | 0.0237 |
| 2007 | 62.479 | 2844 | 51.685 | 23 | 2.8023 | 0.5790 | 0.0259 |
| 2008 | 16.789 | 3852 | 102.992 | 26 | 4.3014 | 0.8620 | 0.0239 |
| 2009 | 91.707 | 3148 | 79.746 | 23 | 4.1921 | 0.7984 | 0.0252 |
| 2010 | 61.785 | 3074 | 52.258 | 24 | 2.6414 | 0.5177 | 0.0256 |
| 2011 | 72.253 | 3426 | 57.450 | 22 | 2.7474 | 0.5373 | 0.0248 |



Figure 13.136. John Dory from Zones 10 and 20 in depths 0 to 200 m by trawl. The top left is the depth distribution of all records reporting John Dory, the top right graph depicts the depth distribution of shots containing John Dory from Zones 10 and 20 in depths 0 to 200 m by trawl. The middle left diagram depicts the distribution of catch by depth within zones 10 and 20 ( 20 is top red line), the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the John Dory catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.137. John Dory from Zones 10 and 20 in depths 0 to 200 m by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.117. John Dory from Zones 10 and 20 in depths 0 to 200 m by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE~Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel+DepCat |
| Model 4 | LnCE $\sim$ Year+Vessel+DepCat+DayNight |
| Model 5 | LnCE $\sim$ Year+Vessel+DepCat+DayNight + Month |
| Model 6 | LnCE $\sim$ Year+Vessel+DepCat+DayNight + Month+Zone |
| Model 7 | LnCE $\sim$ Year+Vessel+DepCat+DayNight + Month+Zone + Zone:Month |
| Model 8 | LnCE $\sim$ Year+Vessel+DepCat+DayNight + Month + Zone + Zone:DepCat |

Table 13.118. John Dory from Zones 10 and 20 in depths 0 to 200 m by trawl. Model selection criteria, including the AIC, the adjusted $\mathrm{r}^{2}$ and the change in adjusted $\mathrm{r}^{2}$. The optimum is Zone:DepCat (model 8).

|  | Year | Vessel | DepCat | DayNight | Month | Zone | Zone:Mth | Zone:DepC |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 26049 | 10945 | 9295 | 7667 | 6995 | 6953 | 6164 | 5772 |
| RSS | 159570 | 141821 | 138945 | 137206 | 136475 | 136428 | 135578 | 135171 |
| MSS | 22595 | 40344 | 43220 | 44959 | 45690 | 45737 | 46588 | 46995 |
| Nobs | 130809 | 130809 | 129730 | 129730 | 129730 | 129730 | 129730 | 129730 |
| Npars | 26 | 186 | 196 | 199 | 210 | 211 | 222 | 221 |
| adj_r2 | 12.387 | 22.037 | 23.611 | 24.565 | 24.961 | 24.986 | 25.447 | 25.672 |
| \%Change | 0.000 | 9.650 | 1.574 | 0.954 | 0.396 | 0.025 | 0.461 | 0.224 |



Figure 13.138. The relative influence of each factor used on the final trend in the optimal standardization for John Dory from Zones 10 and 20. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.45 Mirror Dory (DOM - 37264003 Zenopsis nebulosus)

Only data from Zones 10 to 50 in depths $0-600 \mathrm{~m}$. All vessels reporting Mirror Dory were included.


Figure 13.139. The catches and geometric mean catch rates from 1986-2010 for Mirror Dory split between east (Zones 10-30) and west (zones 40 and 50. The general trends in catch rates, in periods of significant catches, are similar across zones within the east and west. This implies that the assumption that there are no year x zone interactions is valid.


Figure 13.140. Standardized trends for Month and DepCat factors for Mirror Dory taken by trawl.

Table 13.119. Mirror Dory from Zones 10 to 50 in depths 0 to 600 m by trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr}$ ). Zone:Month is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 402.048 | 3199 | 375.385 | 91 | 18.6423 | 1.2117 | 0.0000 |
| 1987 | 450.766 | 3103 | 429.090 | 92 | 19.7476 | 1.2147 | 0.0308 |
| 1988 | 346.014 | 3189 | 328.220 | 88 | 16.9455 | 1.1875 | 0.0306 |
| 1989 | 591.631 | 3068 | 524.863 | 84 | 23.1957 | 1.4715 | 0.0311 |
| 1990 | 295.764 | 1906 | 264.346 | 73 | 20.6077 | 1.3530 | 0.0357 |
| 1991 | 240.313 | 2230 | 183.737 | 77 | 13.9567 | 1.1550 | 0.0343 |
| 1992 | 166.980 | 2228 | 147.170 | 71 | 11.3487 | 0.9990 | 0.0345 |
| 1993 | 306.220 | 3290 | 285.221 | 72 | 13.7999 | 1.0955 | 0.0314 |
| 1994 | 297.268 | 3828 | 280.195 | 70 | 11.4667 | 0.9802 | 0.0306 |
| 1995 | 244.924 | 4209 | 234.433 | 70 | 10.0782 | 0.9047 | 0.0300 |
| 1996 | 352.722 | 5835 | 327.514 | 84 | 8.9039 | 0.8782 | 0.0287 |
| 1997 | 459.626 | 6681 | 436.446 | 80 | 9.6820 | 0.9362 | 0.0284 |
| 1998 | 355.794 | 5572 | 346.706 | 68 | 9.0983 | 0.8485 | 0.0290 |
| 1999 | 309.481 | 5543 | 298.167 | 74 | 8.0995 | 0.7039 | 0.0292 |
| 2000 | 171.066 | 5613 | 165.229 | 79 | 4.6519 | 0.4847 | 0.0294 |
| 2001 | 243.362 | 7016 | 233.924 | 75 | 5.1157 | 0.5653 | 0.0288 |
| 2002 | 449.555 | 8199 | 435.035 | 69 | 7.1647 | 0.7542 | 0.0283 |
| 2003 | 613.832 | 7796 | 560.887 | 71 | 8.6661 | 0.9179 | 0.0283 |
| 2004 | 507.392 | 6485 | 452.616 | 69 | 8.2044 | 0.8819 | 0.0291 |
| 2005 | 579.886 | 6190 | 523.814 | 66 | 9.3924 | 0.9762 | 0.0292 |
| 2006 | 419.556 | 4293 | 363.075 | 54 | 9.7517 | 0.9631 | 0.0308 |
| 2007 | 289.603 | 3400 | 268.103 | 33 | 9.5152 | 0.9292 | 0.0324 |
| 2008 | 396.242 | 3377 | 376.364 | 34 | 12.2034 | 1.1150 | 0.0325 |
| 2009 | 476.515 | 3567 | 461.781 | 32 | 13.1797 | 1.2256 | 0.0322 |
| 2010 | 579.973 | 3702 | 561.230 | 32 | 12.8612 | 1.1642 | 0.0321 |
| 2011 | 516.330 | 3903 | 505.245 | 33 | 10.8311 | 1.0831 | 0.0318 |



Figure 13.141. Mirror Dory from Zones 10 to 50 in depths 0 to 600 m by trawl. The top left is the depth distribution of all records reporting Mirror Dory, the top right graph depicts the depth distribution of shots containing Mirror Dory from Zones 10 to 50 in depths 0 to 600 m by trawl. The middle left diagram depicts the distribution of catch by depth within zones 10 to 50 , the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Mirror Dory catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.142. Mirror Dory from Zones 10 to 50 in depths 0 to 600 m by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

| Table 13.120. Mirror Dory from Zones 10 to 50 in depths 0 to 600 m by trawl. Statistical model |
| :--- |
| structures used in this analysis. DepCat is a series of 20 metre depth categories. |


| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel+Month |
| Model 4 | LnCE $\sim$ Year+Vessel+Month+DepCat |
| Model 5 | LnCE $\sim$ Year+Vessel+Month+DepCat+DayNight |
| Model 6 | LnCE $\sim$ Year+Vessel+Month+DepCat+DayNight+Zone |
| Model 7 | LnCE $\sim$ Year+Vessel+Month+DepCat+DayNight+Zone+Zone:Month |
| Model 8 | LnCE $\sim$ Year+Vessel+Month + DepCat+DayNight+Zone+Zone:DepCat |

Table 13.121. Mirror Dory from Zones 10 to 50 in depths 0 to 600 m by trawl. Model selection criteria, including the AIC, the adjusted $\mathrm{r}^{2}$ and the change in adjusted $\mathrm{r}^{2}$. The optimum is Zone:Month (model 7).

|  | Year | Vessel | DepCat | Month | DayNight | Zone |  | Zone:Mth Zone:DepC |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 71566 | 50600 | 49043 | 38502 | 37257 | 36610 | 32447 | 35740 |
| RSS | 215899 | 179983 | 177579 | 161686 | 159964 | 159070 | 153386 | 157566 |
| MSS | 16240 | 52155 | 54560 | 70453 | 72175 | 73069 | 78753 | 74573 |
| Nobs | 117422 | 117422 | 117422 | 116819 | 116819 | 116819 | 116819 | 116819 |
| Npars | 26 | 225 | 236 | 266 | 269 | 273 | 317 | 393 |
| adj_r2 | 6.976 | 22.319 | 23.350 | 30.191 | 30.933 | 31.317 | 33.746 | 31.896 |
| \%Change | 0.000 | 15.343 | 1.030 | 6.841 | 0.742 | 0.384 | 2.429 | -1.850 |



Figure 13.143. The relative influence of each factor used on the final trend in the optimal standardization for Mirror Dory from Zones 10 to 50. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.46 Mirror Dory East (DOM - 37264003 Zenopsis nebulosus)

Only data from Zones 10 to 30 in depths $0-600 \mathrm{~m}$. All vessels reporting Mirror Dory were included.

Table 13.122. Mirror Dory from Zones 10 to 30 in depths 0 to 600 m by trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr}$ ). Zone:Month is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 402.048 | 3141 | 367.985 | 80 | 18.7487 | 1.1585 | 0.0000 |
| 1987 | 450.766 | 2961 | 413.571 | 70 | 19.9429 | 1.1556 | 0.0322 |
| 1988 | 346.014 | 3067 | 313.237 | 77 | 16.8882 | 1.1336 | 0.0318 |
| 1989 | 591.631 | 2997 | 513.736 | 70 | 23.1617 | 1.3791 | 0.0323 |
| 1990 | 295.764 | 1811 | 254.380 | 61 | 20.5538 | 1.2896 | 0.0373 |
| 1991 | 240.313 | 2021 | 170.954 | 68 | 14.2052 | 1.1339 | 0.0366 |
| 1992 | 166.980 | 2022 | 138.871 | 56 | 11.7312 | 0.9845 | 0.0366 |
| 1993 | 306.220 | 3013 | 267.091 | 62 | 14.1976 | 1.0792 | 0.0332 |
| 1994 | 297.268 | 3498 | 262.033 | 62 | 11.6924 | 0.9448 | 0.0323 |
| 1995 | 244.924 | 3500 | 196.290 | 59 | 10.2913 | 0.8577 | 0.0322 |
| 1996 | 352.722 | 4397 | 212.369 | 69 | 7.7998 | 0.7617 | 0.0309 |
| 1997 | 459.626 | 4775 | 288.136 | 65 | 8.6425 | 0.8100 | 0.0308 |
| 1998 | 355.794 | 4103 | 230.495 | 55 | 8.0944 | 0.7297 | 0.0314 |
| 1999 | 309.481 | 4225 | 234.873 | 59 | 7.8713 | 0.6626 | 0.0315 |
| 2000 | 171.066 | 4633 | 142.768 | 63 | 4.7885 | 0.4995 | 0.0315 |
| 2001 | 243.362 | 4570 | 128.644 | 55 | 4.0443 | 0.4995 | 0.0318 |
| 2002 | 449.555 | 5038 | 194.433 | 53 | 5.2594 | 0.6208 | 0.0313 |
| 2003 | 613.832 | 5362 | 405.679 | 58 | 7.7688 | 0.9105 | 0.0308 |
| 2004 | 507.392 | 4275 | 292.676 | 57 | 7.2635 | 0.8636 | 0.0320 |
| 2005 | 579.886 | 4417 | 423.631 | 55 | 9.9946 | 1.1053 | 0.0319 |
| 2006 | 419.556 | 3230 | 297.559 | 44 | 10.3893 | 1.1074 | 0.0337 |
| 2007 | 289.603 | 2223 | 203.162 | 22 | 11.4463 | 1.1963 | 0.0369 |
| 2008 | 396.242 | 2495 | 317.705 | 26 | 14.4563 | 1.3389 | 0.0363 |
| 2009 | 476.515 | 2232 | 338.488 | 27 | 15.8458 | 1.4196 | 0.0373 |
| 2010 | 579.973 | 2105 | 383.480 | 25 | 14.3976 | 1.1806 | 0.0376 |
| 2011 | 516.330 | 2254 | 347.067 | 26 | 12.7502 | 1.1774 | 0.0372 |



Figure 13.144. Mirror Dory from Zones 10 to 30 in depths 0 to 600 m by trawl. The top left is the depth distribution of all records reporting Mirror Dory, the top right graph depicts the depth distribution of shots containing Mirror Dory from Zones 10 to 30 in depths 0 to 600 m by trawl. The middle left diagram depicts the distribution of catch by depth within zones 10 to 30 , the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Mirror Dory catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.145. Mirror Dory from Zones 10 to 30 in depths 0 to 600 m by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.123. Mirror Dory from Zones 10 to 30 in depths 0 to 600 m by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel+DepCat |
| Model 4 | LnCE $\sim$ Year+Vessel+DepCat+Month |
| Model 5 | LnCE $\sim$ Year+Vessel+DepCat+Month+DayNight |
| Model 6 | LnCE $\sim$ Year+Vessel+DepCat+Month+DayNight+Zone |
| Model 7 | LnCE $\sim$ Year+Vessel+DepCat+Month+DayNight+Zone+Zone:Month |
| Model 8 | LnCE $\sim$ Year+Vessel+DepCat+Month+DayNight+ZOne+Zone:DepCat |

Table 13.124. Mirror Dory from Zones 10 to 30 in depths 0 to 600 m by trawl. Model selection criteria, including the AIC, the adjusted $\mathrm{r}^{2}$ and the change in adjusted $\mathrm{r}^{2}$. The optimum is Zone:Month (model 7).

|  | Year | Vessel | DepCat | Month | DayNight | Zone Zone:Mth Zone:DepC |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 58701 | 43347 | 33534 | 31899 | 31243 | 30638 | 29161 | 30360 |
| RSS | 171604 | 43670 | 128072 | 125680 | 124738 | 123877 | 121751 | 123317 |
| MSS | 18473 | 46407 | 62005 | 64397 | 65339 | 66200 | 68326 | 66760 |
| Nobs | 88365 | 88365 | 87914 | 87914 | 87914 | 87914 | 87914 | 87914 |
| Npars | 26 | 199 | 229 | 240 | 243 | 245 | 267 | 305 |
| adj_r2 | 9.693 | 24.245 | 32.446 | 33.699 | 34.194 | 34.647 | 35.752 | 34.898 |
| \%Change | 0.000 | 14.552 | 8.201 | 1.253 | 0.495 | 0.453 | 1.105 | -0.855 |



Figure 13.146. The relative influence of each factor used on the final trend in the optimal standardization for Mirror Dory from Zones 10 to 30. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.47 Mirror Dory West (DOM - 37264003 Zenopsis nebulosus)

Only data from Zones 40 to 50 in depths $0-600 \mathrm{~m}$. All vessels reporting Mirror Dory were included.

Table 13.125. Mirror Dory from Zones 40 to 50 in depths 0 to 600 m by trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr)}$ ). Zone:Month is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 402.048 | 57 | 7.374 | 10 | 13.7130 | 2.3666 | 0.0000 |
| 1987 | 450.766 | 142 | 15.519 | 23 | 16.0832 | 1.5740 | 0.1989 |
| 1988 | 346.014 | 122 | 14.983 | 17 | 18.4525 | 1.2966 | 0.2076 |
| 1989 | 591.631 | 71 | 11.127 | 15 | 24.6757 | 1.6475 | 0.2191 |
| 1990 | 295.764 | 95 | 9.966 | 14 | 21.6631 | 1.1110 | 0.2226 |
| 1991 | 240.313 | 209 | 12.783 | 17 | 11.7670 | 0.7909 | 0.1963 |
| 1992 | 166.980 | 205 | 8.289 | 20 | 8.1608 | 0.6541 | 0.1979 |
| 1993 | 306.220 | 276 | 18.010 | 18 | 10.1017 | 0.7732 | 0.1931 |
| 1994 | 297.268 | 330 | 18.162 | 20 | 9.3264 | 0.6758 | 0.1915 |
| 1995 | 244.924 | 709 | 38.143 | 23 | 9.0896 | 0.8594 | 0.1886 |
| 1996 | 352.722 | 1438 | 115.145 | 26 | 13.3473 | 1.2433 | 0.1885 |
| 1997 | 459.626 | 1906 | 148.310 | 24 | 12.8686 | 1.2670 | 0.1880 |
| 1998 | 355.794 | 1469 | 116.211 | 20 | 12.6121 | 1.2404 | 0.1884 |
| 1999 | 309.481 | 1318 | 63.294 | 23 | 8.8763 | 0.8103 | 0.1886 |
| 2000 | 171.066 | 980 | 22.461 | 27 | 4.0569 | 0.4281 | 0.1896 |
| 2001 | 243.362 | 2446 | 105.280 | 29 | 7.9361 | 0.7384 | 0.1879 |
| 2002 | 449.555 | 3156 | 240.252 | 28 | 11.7181 | 1.0837 | 0.1876 |
| 2003 | 613.832 | 2429 | 154.899 | 27 | 11.0165 | 0.9361 | 0.1879 |
| 2004 | 507.392 | 2208 | 159.809 | 25 | 10.3786 | 0.9388 | 0.1881 |
| 2005 | 579.886 | 1769 | 100.006 | 23 | 8.0456 | 0.7379 | 0.1883 |
| 2006 | 419.556 | 1061 | 65.351 | 19 | 8.0395 | 0.6304 | 0.1894 |
| 2007 | 289.603 | 1177 | 64.941 | 16 | 6.7120 | 0.5653 | 0.1892 |
| 2008 | 396.242 | 879 | 58.533 | 17 | 7.5767 | 0.6335 | 0.1898 |
| 2009 | 476.515 | 1333 | 123.246 | 14 | 9.7010 | 0.9617 | 0.1886 |
| 2010 | 579.973 | 1596 | 177.550 | 14 | 11.0745 | 1.1448 | 0.1885 |
| 2011 | 516.330 | 1644 | 156.846 | 16 | 8.6540 | 0.8911 | 0.1886 |



Figure 13.147. Mirror Dory from Zones 40 to 50 in depths 0 to 600 m by trawl. The top left is the depth distribution of all records reporting Mirror Dory, the top right graph depicts the depth distribution of shots containing Mirror Dory from Zones 40 to 50 in depths 0 to 600 m by trawl. The middle left diagram depicts the distribution of catch by depth within zones 40 to 50 , the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Mirror Dory catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.148. Mirror Dory from Zones 40 to 50 in depths 0 to 600 m by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.126. Mirror Dory from Zones 40 to 50 in depths 0 to 600 m by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE~Year |
| :--- | :--- |
| Model 2 | LnCE~Year+Vessel |
| Model 3 | LnCE~Year+Vessel+Month |
| Model 4 | LnCE~Year+Vessel+Month+DepCat |
| Model 5 | LnCE~Year+Vessel+Month+DepCat+DayNight |
| Model 6 | LnCE~Year+Vessel+Month+DepCat+DayNight+Zone |
| Model 7 | LnCE~Year+Vessel+Month+DepCat+DayNight + Zone + Zone:Month |
| Model 8 | LnCE~Year+Vessel+Month + DepCat + DayNight + Zone + Zone:DepCat |

Table 13.127. Mirror Dory from Zones 40 to 50 in depths 0 to 600 m by trawl. Model selection criteria, including the AIC, the adjusted $\mathrm{r}^{2}$ and the change in adjusted $\mathrm{r}^{2}$. The optimum is Zone:Month (model 7).

|  | Year | Vessel | Month | DepCat | DayNight | Zone Zone:Mth Zone:DepC |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 9354 | 2688 | 1318 | 115 | -514 | -849 | -1190 | -881 |
| RSS | 39990 | 31597 | 30117 | 28697 | 28073 | 27747 | 27400 | 27672 |
| MSS | 2012 | 10406 | 11885 | 13306 | 13930 | 14255 | 14602 | 14330 |
| Nobs | 29025 | 29025 | 29025 | 28873 | 28873 | 28873 | 28873 | 28873 |
| Npars | 26 | 112 | 123 | 146 | 149 | 150 | 161 | 173 |
| adj_r2 | 4.709 | 24.485 | 27.993 | 31.334 | 32.820 | 33.597 | 34.402 | 33.723 |
| \%Change | 0.000 | 19.777 | 3.508 | 3.340 | 1.486 | 0.777 | 0.805 | -0.679 |



Figure 13.149. The relative influence of each factor used on the final trend in the optimal standardization for Mirror Dory from Zones $40-50$. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

### 13.48 Ribaldo (RBD - 37224002 - Mora moro)

Only data from Zones 10 to 50 in depths $0-1000 \mathrm{~m}$.

Table 13.128. Ribaldo from Zones 10 to 50 in depths 0 to 1000 m by trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr}$ ). Zone:Mth is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Mth | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 4.104 | 72 | 3.524 | 11 | 14.6630 | 2.2797 | 0.0000 |
| 1987 | 7.941 | 158 | 7.292 | 14 | 10.2593 | 1.2772 | 0.1379 |
| 1988 | 10.898 | 123 | 8.049 | 22 | 16.5570 | 2.0037 | 0.1542 |
| 1989 | 11.342 | 136 | 7.711 | 14 | 18.2556 | 1.8029 | 0.1526 |
| 1990 | 3.668 | 58 | 2.259 | 11 | 8.9113 | 1.4196 | 0.1728 |
| 1991 | 7.808 | 145 | 5.162 | 22 | 7.9930 | 1.3647 | 0.1522 |
| 1992 | 13.333 | 226 | 11.689 | 26 | 9.7616 | 1.3480 | 0.1441 |
| 1993 | 22.777 | 330 | 19.762 | 37 | 11.2449 | 1.1172 | 0.1439 |
| 1994 | 41.938 | 423 | 23.622 | 30 | 11.8156 | 1.2565 | 0.1416 |
| 1995 | 90.323 | 1147 | 86.299 | 26 | 12.3128 | 1.3011 | 0.1382 |
| 1996 | 82.278 | 1492 | 77.012 | 32 | 10.1757 | 1.0009 | 0.1379 |
| 1997 | 103.111 | 1714 | 96.567 | 30 | 9.8023 | 0.8776 | 0.1376 |
| 1998 | 99.924 | 1667 | 92.015 | 33 | 9.6696 | 0.8530 | 0.1377 |
| 1999 | 72.157 | 1133 | 59.668 | 32 | 8.7093 | 0.7871 | 0.1386 |
| 2000 | 66.791 | 1174 | 53.845 | 37 | 7.4217 | 0.7152 | 0.1385 |
| 2001 | 82.479 | 1122 | 52.390 | 37 | 6.7639 | 0.6655 | 0.1384 |
| 2002 | 157.878 | 1142 | 57.271 | 30 | 6.7944 | 0.6229 | 0.1387 |
| 2003 | 181.036 | 1310 | 66.180 | 35 | 6.7153 | 0.6103 | 0.1384 |
| 2004 | 180.961 | 1257 | 66.417 | 33 | 7.2233 | 0.6613 | 0.1386 |
| 2005 | 90.375 | 671 | 30.046 | 32 | 6.3488 | 0.5715 | 0.1404 |
| 2006 | 122.615 | 637 | 32.083 | 34 | 6.3304 | 0.6153 | 0.1405 |
| 2007 | 78.314 | 404 | 15.571 | 24 | 3.2493 | 0.4015 | 0.1433 |
| 2008 | 78.475 | 367 | 17.618 | 24 | 4.7326 | 0.5556 | 0.1438 |
| 2009 | 104.960 | 572 | 33.410 | 20 | 5.6978 | 0.6149 | 0.1409 |
| 2010 | 92.104 | 685 | 37.305 | 22 | 5.5851 | 0.6321 | 0.1401 |
| 2011 | 94.029 | 864 | 44.555 | 20 | 5.8331 | 0.6447 | 0.1392 |



Figure 13.150. Ribaldo from Zones 10 to 50 in depths 0 to 1000 m by trawl. The top left is the depth distribution of all records reporting Ribaldo, the top right graph depicts the depth distribution of shots containing Ribaldo from Zones 10 to 50 in depths 0 to 1000 m by trawl. The middle left diagram depicts the distribution of catch by depth within zones 10 to 50 , the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Ribaldo catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.151. Ribaldo from Zones 10 to 50 in depths 0 to 1000 m by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates.

Table 13.129. Ribaldo from Zones 10 to 50 in depths 0 to 1000 m by trawl. Statistical model structures used in this analysis. DepCat is a series of 50 metre depth categories.

| Model 1 | LnCE $\sim$ Year |
| :--- | :--- |
| Model 2 | LnCE $\sim$ Year+Vessel |
| Model 3 | LnCE $\sim$ Year+Vessel+DepCat |
| Model 4 | LnCE $\sim$ Year+Vessel+DepCat+Zone |
| Model 5 | LnCE $\sim$ Year+Vessel+DepCat+Zone+DayNight |
| Model 6 | LnCE $\sim$ Year+Vessel+DepCat+Zone+DayNight+Month |
| Model 7 | LnCE $\sim$ Year+Vessel+DepCat+Zone+DayNight+Month+Zone:Month |
| Model 8 | LnCE $\sim$ Year+Vessel+DepCat+Zone+DayNight+Month+Zone:DepCat |

Table 13.130. Ribaldo from Zones 10 to 50 in depths 0 to 1000 m by trawl. Model selection criteria, including the AIC, the adjusted $\mathrm{r}^{2}$ and the change in adjusted $\mathrm{r}^{2}$. The optimum is Zone:Month (model 7).

|  | Year | Vessel | DepCat | Zone DayNight | Month |  | Zone:Mth | Zone:DepC |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | -2112 | -4223 | -6347 | -6990 | -7097 | -7127 | -7595 | -7373 |
| RSS | 16983 | 15011 | 13226 | 12779 | 12702 | 12667 | 12300 | 12241 |
| MSS | 1605 | 3577 | 5363 | 5810 | 5887 | 5921 | 6289 | 6347 |
| Nobs | 19029 | 19029 | 18893 | 18893 | 18893 | 18893 | 18893 | 18893 |
| Npars | 26 | 145 | 195 | 199 | 202 | 213 | 257 | 413 |
| adj_r2 | 8.516 | 18.629 | 28.111 | 30.528 | 30.933 | 31.081 | 32.923 | 32.678 |
| \%Change | 0.000 | 10.113 | 9.482 | 2.417 | 0.405 | 0.149 | 1.841 | -0.244 |



Figure 13.152. The relative influence of each factor used on the final trend in the optimal standardization for Ribaldo from Zones 10 to 50 . The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.


Figure 13.153. Ribaldo, all zones 10-50, plus the GAB and north of Barrenjoey. Catches by the two main methods, trawl and AutoLongLine. As with trawling, most catches by AutoLongLine are taken in zones 20-50.

### 13.48.1 The Effect of Closures

An alternative analysis was conducted (Haddon (2010) that included a factor for inside and outside of the current deepwater closures. After the other single factors had been included in the standardization there was no significant effect of being inside or outside of a closure.

By considering the current deepwater closures and identifying each shot with respect to its starting position the catches within and outside the closures can be characterized (Figure 13.154, Figure 13.155, Table 13.131).


Figure 13.154. The annual catches of Ribaldo taken by trawl inside and outside of declared deepwater closures. The low catches taken from 2008 onwards derive from the precision of the available location data to discern all shots that are taken along the outer edge of a closure.


Figure 13.155. The annual catches of Ribaldo taken by AutoLine inside and outside of declared deepwater closures. The low catches taken from 2008 onwards derive from the precision of the available location data to discern all shots that are taken along the outer edge of a closure.

Table 13.131. Catches of Ribaldo by Trawl and AutoLine in open and closed areas.
Trawl
AutoLine

| Year | Open | Closed | Open | Closed |
| :---: | :---: | :---: | :---: | :---: |
| 1986 | 3.054 | 0.470 |  |  |
| 1987 | 6.173 | 1.119 |  |  |
| 1988 | 6.082 | 1.967 |  |  |
| 1989 | 5.266 | 2.445 |  |  |
| 1990 | 2.144 | 0.115 |  |  |
| 1991 | 4.278 | 0.884 |  |  |
| 1992 | 11.354 | 0.335 |  |  |
| 1993 | 18.539 | 1.223 |  |  |
| 1994 | 21.175 | 2.447 |  |  |
| 1995 | 54.787 | 31.512 |  |  |
| 1996 | 57.647 | 19.365 |  |  |
| 1997 | 67.066 | 29.501 | 1.375 | 0.030 |
| 1998 | 59.008 | 33.007 | 1.753 |  |
| 1999 | 33.724 | 25.944 | 1.687 | 0.260 |
| 2000 | 33.332 | 20.513 | 8.486 | 0.553 |
| 2001 | 30.350 | 22.040 | 15.110 | 0.610 |
| 2002 | 35.577 | 21.695 | 88.474 | 7.023 |
| 2003 | 42.289 | 23.891 | 93.960 | 8.922 |
| 2004 | 43.108 | 23.309 | 87.052 | 9.537 |
| 2005 | 21.626 | 8.421 | 33.177 | 4.013 |
| 2006 | 22.564 | 9.520 | 63.046 | 2.329 |
| 2007 | 14.802 | 0.769 | 27.887 | 0.239 |
| 2008 | 16.348 | 1.271 | 56.655 | 0.118 |
| 2009 | 30.393 | 3.017 | 66.447 | 1.827 |
| 2010 | 35.040 | 2.265 | 50.216 | 1.472 |
| 2011 | 41.630 | 2.925 | 46.364 | 0.113 |

### 13.49 Ribaldo (RBD - 37224002 - Mora moro) AutoLine

Table 13.132. Ribaldo taken by Autoline in Zones $20,3040,50,81,82,83,84,85$ in depths 0 to 1000 m . Total Catch is the total reported in the database, Records is the number of records used in the analysis, Catch T is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr}$ ). Zone:Month is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Month | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2001 | 82.479 | 63 | 15.720 | 2 | 157.4316 | 1.0602 | 0.0000 |
| 2002 | 157.878 | 259 | 95.497 | 4 | 135.9460 | 2.3811 | 0.1817 |
| 2003 | 181.036 | 337 | 102.882 | 7 | 75.0323 | 1.6703 | 0.1786 |
| 2004 | 180.961 | 714 | 96.589 | 11 | 51.6307 | 1.3981 | 0.1736 |
| 2005 | 90.375 | 308 | 37.189 | 7 | 44.5029 | 0.7914 | 0.1801 |
| 2006 | 122.615 | 605 | 65.374 | 8 | 39.5786 | 0.8238 | 0.1739 |
| 2007 | 78.314 | 393 | 28.125 | 6 | 25.0254 | 0.4843 | 0.1771 |
| 2008 | 78.475 | 401 | 56.772 | 6 | 39.2440 | 0.5789 | 0.1750 |
| 2009 | 104.960 | 432 | 68.270 | 6 | 49.8911 | 0.5870 | 0.1727 |
| 2010 | 92.104 | 381 | 51.687 | 5 | 47.4986 | 0.5594 | 0.1751 |
| 2011 | 94.029 | 356 | 46.476 | 5 | 45.6603 | 0.6656 | 0.1754 |



Figure 13.156. Standardized catch rates for Ribaldo by Autoline. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates. The vertical black lines represent 1.96 times the standard errors. The same statistical models that were used for the trawl analysis were also used here (Table 13.129).


Figure 13.157. Ribaldo by Autoline. The top left is the depth distribution of all records reporting Ribaldo taken by autoline, the top right graph depicts the depth distribution of shots containing Ribaldo taken by Autoline as used in the standardization. The middle left diagram depicts the distribution of catch by depth within each zone, the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Ribaldo catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).

Table 13.133. Ribaldo taken by Autoline. Model selection criteria, including the AIC, the adjusted $\mathrm{r}^{2}$ and the change in adjusted $r^{2}$. The optimum is Zone:Month (model 7).

|  | Year | Vessel | DepCat | Zone | DayNight | Month | Zone:Mth | Zone:DepC |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 3784 | 2277 | 1991 | 1925 | 1903 | 1841 | 1707 | 2113 |
| RSS | 10300 | 7185 | 6580 | 6455 | 6413 | 6287 | 5874 | 5875 |
| MSS | 648 | 3764 | 4369 | 4493 | 4536 | 4661 | 5074 | 5074 |
| Nobs | 4250 | 4250 | 4236 | 4236 | 4236 | 4236 | 4236 | 4236 |
| Npars | 11 | 23 | 63 | 70 | 73 | 84 | 161 | 364 |
| adj_r2 | 5.700 | 34.036 | 39.011 | 40.063 | 40.413 | 41.428 | 44.240 | 41.314 |
| \%Change | 0.000 | 28.337 | 4.974 | 1.053 | 0.350 | 1.014 | 2.812 | -0.114 |

### 13.50 Ocean Jackets (LTC - 37465006 - Nelusetta ayraudi)

## Alternate: LeatherJackets (LTH - 37465000)

Only data from Zones 10 to 50 in depths $0-300 \mathrm{~m}$. All vessels and records reporting leatherjackets are included. This is the first year this data has been considered.

[^4]| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:DepCat | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 56.429 | 2473 | 44.715 | 75 | 5.0337 | 0.6900 | 0.0000 |
| 1987 | 53.354 | 1445 | 28.151 | 61 | 5.1085 | 0.7302 | 0.0357 |
| 1988 | 66.304 | 1911 | 45.725 | 66 | 6.2067 | 0.8855 | 0.0332 |
| 1989 | 71.666 | 1808 | 32.778 | 65 | 4.8860 | 0.7620 | 0.0338 |
| 1990 | 90.969 | 1548 | 33.157 | 46 | 4.9715 | 0.7474 | 0.0357 |
| 1991 | 170.481 | 1329 | 24.788 | 46 | 4.4265 | 0.6489 | 0.0375 |
| 1992 | 88.884 | 1127 | 22.074 | 40 | 4.7352 | 0.6445 | 0.0392 |
| 1993 | 71.897 | 1342 | 29.245 | 42 | 5.0852 | 0.7226 | 0.0380 |
| 1994 | 74.438 | 1455 | 35.044 | 45 | 5.9717 | 0.8095 | 0.0366 |
| 1995 | 140.179 | 2237 | 59.316 | 42 | 5.9904 | 0.8226 | 0.0331 |
| 1996 | 199.571 | 2576 | 72.307 | 54 | 6.3230 | 0.8599 | 0.0323 |
| 1997 | 177.419 | 2009 | 52.492 | 51 | 5.4540 | 0.7762 | 0.0341 |
| 1998 | 189.899 | 2488 | 68.017 | 44 | 5.2603 | 0.7632 | 0.0327 |
| 1999 | 202.805 | 2691 | 88.415 | 52 | 7.0029 | 0.9046 | 0.0321 |
| 2000 | 198.811 | 2983 | 73.176 | 51 | 5.1836 | 0.7116 | 0.0318 |
| 2001 | 222.570 | 3160 | 63.794 | 55 | 4.2040 | 0.6267 | 0.0317 |
| 2002 | 378.516 | 4863 | 199.088 | 61 | 5.4894 | 0.7472 | 0.0298 |
| 2003 | 482.582 | 5503 | 187.624 | 58 | 5.0890 | 0.7125 | 0.0293 |
| 2004 | 692.874 | 6214 | 313.391 | 60 | 8.3226 | 1.1598 | 0.0289 |
| 2005 | 890.644 | 5162 | 342.889 | 54 | 9.8920 | 1.3488 | 0.0297 |
| 2006 | 741.530 | 4636 | 301.737 | 50 | 10.2758 | 1.4852 | 0.0303 |
| 2007 | 564.833 | 3092 | 285.396 | 27 | 14.0314 | 1.7918 | 0.0326 |
| 2008 | 490.402 | 3554 | 318.317 | 29 | 13.7150 | 1.6963 | 0.0320 |
| 2009 | 609.980 | 3260 | 376.112 | 28 | 16.0145 | 1.8956 | 0.0325 |
| 2010 | 484.039 | 3258 | 300.273 | 29 | 13.2712 | 1.5906 | 0.0325 |
| 2011 | 487.141 | 3220 | 277.118 | 29 | 12.3501 | 1.4670 | 0.0325 |



Figure 13.158. Ocean Jackets from Zones 10 to 50 in depths 0 to 300 m by trawl. The top left is the depth distribution of all records reporting Leatherjackets, the top right graph depicts the depth distribution of shots containing Ocean Jackets from Zones 10 to 50 in depths 0 to 300 m by trawl. The middle left diagram depicts the distribution of catch by depth within zones 10 to 50 , the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Leatherjacket catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.159. Ocean Jackets from Zones 10 to 50 in depths 0 to 300 m by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates. The fine vertical lines are the $95 \%$ confidence intervals. The horizontal blue line is at one, which is the average of the time series. If the standardization is only applied to data from Zones 10 and 20 differences occur only at the third decimal place in the standardization.

Table 13.135. Ocean Jackets from Zones 10 to 50 in depths 0 to 300 m by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

| Model 1 | LnCE~Year |
| :--- | :--- |
| Model 2 | LnCE~Year+Vessel |
| Model 3 | LnCE~Year+Vessel+DepCat |
| Model 4 | LnCE~Year+Vessel+DepCat+Month |
| Model 5 | LnCE $\sim$ Year+Vessel+DepCat+Month+Zone |
| Model 6 | LnCE $\sim$ Year+Vessel+DepCat+Month+Zone+DayNight |
| Model 7 | LnCE $\sim$ Year+Vessel+DepCat+Month+Zone+DayNight + Zone:Month |
| Model 8 | LnCE $\sim$ Year+Vessel+DepCat + Month + Zone + DayNight + Zone:DepCat |

Table 13.136. Ocean Jackets from Zones 10 to 50 in depths 0 to 300 m by trawl. Model selection criteria, including the AIC, the adjusted $r^{2}$ and the change in adjusted $r^{2}$. The optimum is Zone:DepCat (model 8).

|  | Year | Vessel | DepCat | Month | Zone |  |  | DayNight |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | Zone:Mth Zone:DepC (



Figure 13.160. The relative influence of each factor used on the final trend in the optimal standardization for Ocean Jackets from Zones 10 to 50. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.


Figure 13.161. Ocean Jackets from Zones 10 to 50 in depths 0 to 300 m by trawl. The catches taken in each of the four main SESSF zones is depicted with the total catch across these zones. The scales on the $y$-axis changes between graphs.


Figure 13.162. Trends in catches and geometric mean catch rates for zones 10 and 20. The catches in the other zones remains too low to be informative about catch rates.

Table 13.137. Ocean Jackets, catch by zone. Only those zones reporting more than 5 tonnes over the years are included (Zone 40 is included for completeness). Zones 82 and 83 are in the GAB and are not included in the CPUE standardization above. See Figure 13.161

| Zone | $\mathbf{1 0}$ | $\mathbf{2 0}$ | $\mathbf{3 0}$ | $\mathbf{4 0}$ | $\mathbf{5 0}$ | $\mathbf{6 0}$ | $\mathbf{8 2}$ | $\mathbf{8 3}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1977 |  |  |  |  |  |  | 4.465 | 2.487 |
| 1978 |  |  |  |  |  |  | 373.114 | 25.268 |
| 1979 |  |  |  |  |  |  | 40.664 | 0.517 |
| 1985 | 15.299 | 6.450 | 0.205 |  | 0.551 | 2.097 |  |  |
| 1986 | 32.330 | 12.456 | 0.040 |  | 0.366 | 1.505 | 6.900 | 1.620 |
| 1987 | 19.467 | 9.655 | 0.025 |  | 0.075 | 0.329 | 22.794 | 0.168 |
| 1988 | 33.812 | 12.793 | 0.042 |  | 0.175 | 1.186 | 15.259 | 0.361 |
| 1989 | 23.302 | 10.381 |  |  | 0.265 | 0.603 | 24.463 | 11.057 |
| 1990 | 22.457 | 12.412 | 0.075 |  | 1.335 | 2.109 | 4.336 | 47.775 |
| 1991 | 18.174 | 8.984 | 0.065 |  | 0.560 | 0.834 | 49.655 | 91.172 |
| 1992 | 17.790 | 8.742 |  |  | 0.030 | 0.406 | 19.869 | 41.280 |
| 1993 | 19.052 | 11.435 | 0.010 |  | 0.050 | 0.398 | 19.048 | 20.096 |
| 1994 | 21.729 | 13.920 | 0.060 |  | 0.229 | 0.363 | 21.415 | 15.757 |
| 1995 | 36.659 | 22.843 | 0.130 |  | 0.160 | 0.434 | 34.935 | 44.667 |
| 1996 | 39.696 | 31.988 | 0.270 | 0.035 | 1.280 | 0.592 | 60.575 | 64.399 |
| 1997 | 32.263 | 19.767 | 0.101 | 0.005 | 1.393 | 0.735 | 57.039 | 64.664 |
| 1998 | 43.919 | 23.707 | 0.070 |  | 0.901 | 0.577 | 51.033 | 66.430 |
| 1999 | 56.431 | 30.112 | 0.106 |  | 2.784 | 1.528 | 43.265 | 66.750 |
| 2000 | 41.329 | 26.344 | 0.705 | 0.001 | 5.116 | 0.152 | 42.115 | 82.373 |
| 2001 | 38.970 | 23.082 | 0.722 |  | 3.830 | 0.545 | 68.411 | 79.797 |
| 2002 | 134.630 | 61.212 | 1.611 | 0.021 | 4.488 | 0.193 | 79.197 | 69.986 |
| 2003 | 145.927 | 43.126 | 2.257 | 0.005 | 4.174 | 0.650 | 99.595 | 183.790 |
| 2004 | 242.908 | 64.483 | 0.973 | 0.321 | 8.713 | 0.885 | 146.522 | 220.003 |
| 2005 | 262.599 | 80.953 | 0.860 | 0.015 | 6.454 | 2.130 | 222.886 | 304.358 |
| 2006 | 192.177 | 108.936 | 0.358 |  | 5.870 | 2.600 | 143.171 | 270.716 |
| 2007 | 194.781 | 99.582 | 0.045 | 0.065 | 5.530 | 1.410 | 116.410 | 140.539 |
| 2008 | 235.622 | 96.248 | 0.137 |  | 1.947 | 1.439 | 42.070 | 105.828 |
| 2009 | 277.180 | 106.624 | 0.015 |  | 1.355 | 1.647 | 83.024 | 137.576 |
| 2010 | 220.167 | 86.237 | 0.045 |  | 1.400 | 1.637 | 82.178 | 86.630 |
| 2011 | 197.143 | 78.452 | 0.225 |  | 12.764 | 2.262 | 65.644 | 129.319 |
| Total | 2615.811 | 1110.923 | 9.149 | 0.467 | 71.793 | 29.245 | 2040.050 | 2375.382 |
|  |  |  |  |  |  |  |  |  |
| 10 |  |  |  |  |  |  |  |  |

### 13.51 Ocean Jackets - GAB (LTC - 37465006 - N. ayraudi)

## Alternate: LeatherJackets (LTH - 37465000)

Only data from Zones 82 and 83 in the GAB in depths $0-300 \mathrm{~m}$. All vessels and records reporting leatherjackets are included. This is the first year this data has been considered.

Table 13.138. Ocean Jackets from Zones 82 and 83 in depths 80 to 220 m by trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr}$ ). Zone:Month is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

| Year | TotCatch | Records | CatchT | Vessels | GeoMean | Zone:Month | StDev |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 56.429 | 141 | 8.490 | 1 | 11.5206 | 1.2913 | 0.0000 |
| 1987 | 53.354 | 212 | 22.632 | 3 | 13.7002 | 1.0634 | 0.1115 |
| 1988 | 66.304 | 245 | 15.590 | 7 | 14.0350 | 1.2440 | 0.1950 |
| 1989 | 71.666 | 576 | 34.714 | 7 | 11.9652 | 1.2598 | 0.1932 |
| 1990 | 90.969 | 920 | 51.380 | 11 | 11.1086 | 0.8543 | 0.1907 |
| 1991 | 170.481 | 1252 | 139.797 | 8 | 15.0694 | 1.0924 | 0.1901 |
| 1992 | 88.884 | 954 | 59.534 | 7 | 9.0287 | 0.9602 | 0.1900 |
| 1993 | 71.897 | 819 | 38.764 | 4 | 6.3105 | 0.6543 | 0.1899 |
| 1994 | 74.438 | 745 | 36.660 | 5 | 5.7741 | 0.5675 | 0.1907 |
| 1995 | 140.179 | 1316 | 78.832 | 5 | 6.2242 | 0.7460 | 0.1893 |
| 1996 | 199.571 | 1725 | 123.469 | 6 | 7.8262 | 0.8679 | 0.1889 |
| 1997 | 177.419 | 2135 | 121.064 | 9 | 6.4622 | 0.7189 | 0.1889 |
| 1998 | 189.899 | 1799 | 116.437 | 9 | 7.1373 | 0.7804 | 0.1890 |
| 1999 | 202.805 | 1585 | 108.970 | 7 | 7.8084 | 0.9009 | 0.1893 |
| 2000 | 198.811 | 1540 | 121.614 | 5 | 7.8119 | 0.9264 | 0.1895 |
| 2001 | 222.570 | 1877 | 138.429 | 6 | 8.7175 | 0.9556 | 0.1894 |
| 2002 | 378.516 | 1788 | 147.551 | 6 | 9.0818 | 1.0093 | 0.1894 |
| 2003 | 482.582 | 2837 | 279.605 | 9 | 10.8621 | 1.1495 | 0.1891 |
| 2004 | 692.874 | 3433 | 364.440 | 9 | 12.7575 | 1.2370 | 0.1890 |
| 2005 | 890.644 | 4317 | 522.910 | 10 | 13.9012 | 1.3275 | 0.1889 |
| 2006 | 741.530 | 3609 | 408.448 | 11 | 12.0564 | 1.0231 | 0.1890 |
| 2007 | 564.833 | 2647 | 254.851 | 8 | 10.2989 | 0.9211 | 0.1893 |
| 2008 | 490.402 | 2351 | 146.362 | 6 | 7.4758 | 0.8048 | 0.1894 |
| 2009 | 609.980 | 2160 | 219.965 | 4 | 10.4196 | 1.1106 | 0.1894 |
| 2010 | 484.039 | 1792 | 168.203 | 4 | 12.6091 | 1.2685 | 0.1898 |
| 2011 | 487.141 | 1877 | 192.596 | 4 | 13.0498 | 1.2653 | 0.1897 |



Figure 13.163. Ocean Jackets from Zones 82 and 83 in depths 80 to 220 m by trawl. The top left is the depth distribution of all records reporting Leatherjackets, the top right graph depicts the depth distribution of shots containing Ocean Jackets from Zones 82 and 83 in depths 80 to 220 m by trawl. The middle left diagram depicts the distribution of catch by depth within zones 82 and 83 , the middle right hand graph depicts the number of vessels through time. The bottom left reflects the number of records used in analysis, and bottom right is the Leatherjacket catches (top line, black is total catches, middle line, blue, are those used in the analysis, and bottom, red, are catches $<30 \mathrm{Kg}$ ).


Figure 13.164. Ocean Jackets from Zones 82 and 83 in depths 80 to 220 m by trawl. The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates. The graph standardizes catch rates relative to the mean of the standardized catch rates. The fine vertical lines are the $95 \%$ confidence intervals. The horizontal blue line is at one, which is the average of the time series. This is the first time this analysis has been conducted.

Table 13.139. Ocean Jackets from Zones 82 and 83 in depths 80 to 220 m by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.
Model $1 \quad$ LnCE $\sim$ Year
Model 2 LnCE~Year+Vessel
Model 3 LnCE $\sim$ Year+Vessel+DepCat
Model $4 \quad$ LnCE~Year+Vessel+DepCat+Zone
Model 5 LnCE~Year+Vessel+DepCat+Zone+DayNight
Model 6 LnCE~Year+Vessel+DepCat+Zone+DayNight+Month
Model 7 LnCE~Year+Vessel+DepCat+Zone+DayNight+Month+Zone:Month
Model $8 \quad$ LnCE $\sim$ Year+Vessel+DepCat+Zone+DayNight+Month+Zone:DepCat

Table 13.140. Ocean Jackets from Zones 82 and 83 in depths 80 to 220 m by trawl. Model selection criteria, including the AIC, the adjusted $\mathrm{r}^{2}$ and the change in adjusted $\mathrm{r}^{2}$. The optimum is Zone:Month (model 8).

|  | Year | Vessel | DepCat | Zone | DayNight | Month | Sth | :DepC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AIC | 5879 | 1182 | -1071 | -3298 | -4471 | -4498 | -4744 | -4516 |
| RSS | 50877 | 45790 | 43094 | 40913 | 39823 | 39797 | 39557 | 39754 |
| MSS | 3204 | 8291 | 10987 | 13167 | 14257 | 14284 | 14524 | 14327 |
| Nobs | 44652 | 44652 | 44238 | 44238 | 44238 | 44238 | 44238 | 44238 |
| Npars | 26 | 29 | 44 | 79 | 90 | 91 | 102 | 106 |
| adj_r2 | 5.872 | 15.277 | 20.238 | 24.214 | 26.214 | 26.262 | 26.688 | 26.316 |
| \%Change | 0.000 | 9.405 | 4.960 | 3.977 | 2.000 | 0.048 | 0.426 | -0.372 |



Figure 13.165. The relative influence of each factor used on the final trend in the optimal standardization for Ocean Jackets from Zones 82 and 83 . The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.


Figure 13.166. Trends in catches and geometric mean catch rates for zones 82 and 83 in the GAB. The catches in the other zones remains too low to be informative about catch rates.

### 13.52 Bibliography

A collection of publications relating to the analysis of catch rates, only some of which are referred to explicitly here but the rest are included as a resource for anyone interested in pursuing this subject further.

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# 14. Catch Rate Standardization Updates with Data to Oct 2012 

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### 14.1 Summary

In order that the most recent catch rate data might influence the TAC setting procedures the most up-to-data catch and effort data were sent for analysis from AFMA. Using standard extracts and analyses the commercial catch and effort data, to October 2012, were standardized and the ratio of the 2012 and 2011 indices were compared and used as the basis for calculating the TAC Multiplier in each fishery.

A total of 24 standardizations were conducted which related to a total of 16 TACs considered. In 14 of the 16 fisheries the TAC Multiplier was between $0.9-1.1$ meaning they implied less than a $10 \%$ increase or decrease to the allocated TAC. However, there were two fisheries, Royal Red Prawn and Blue grenadier where the final TAC multiplier was only 0.86 and 0.84 respectively. However, an inspection of the standardization with its error bars indicates that the catch rates have simply returned to the long run average. Blue grenadier certainly exhibits a strong decline in its catch rates for the non-spawning fishery.

This year Inshore and Offshore Ocean Perch were treated separately. Inshore Ocean Perch were only just above the $10 \%$ threshold for a reduction.

The fishery for Silver Trevally was a special case because of the influence of the Batemans Bay marine protected area. Previously, catch rate standardizations have removed records taken within State waters within the MPA. However, now that there is sufficient data available to determine empirically where the fishers are no longer reporting catches it becomes apparent that removing all State waters is too stringent and removes too many records. An improved analysis is given here but it has no effect on the update calculations which only relate to the last two years.

The TAC Multiplier is based upon a comparison of two years of catch rate data. The Tier 4 method usually uses an average of the four most recent years while the tier 3 is a fishing mortality based method that ignores catch rates. It should be remembered that the CPUE multiplier places exceptional emphasis on catch rates, which is a potential problem for those species where catch rates are becoming less informative.

The RAG has not considered that applying this catch rate meta-rule to those species where Industry members report avoidance is occurring or which are not directly targeted may produce anomalous results leading to reductions in TAC that are not strictly warranted. Further discussion is required to confirm whether this TAC Multiplier approach should be approved for all fisheries or only selected ones.

### 14.2 Introduction

### 14.2.1 The Catch Rate Multiplier Rule

In recent years, industry members have voiced concerns that the SESSF stock assessments, and therefore the resultant TACs, are conducted using data that are at least 12 months old. For example, to calculate the RBCs and set the TACs for 2013, data from the fisheries up to the end of 2011 are analysed but usually data from 2012 are unavailable. The SESSF industry were therefore concerned that the most recent information regarding stock availability and/or relative abundance was not being used to inform the most recent assessments of stock status and thus influence the TAC setting procedures. To mitigate this perceived problem it was proposed that management needed to include the most recent year's CPUE data when deriving the proposed TACs.

CSIRO, in dialogue with RAG Chairs, developed a rule by which TACs can be adjusted based on recent CPUE trends (described in the Methods). MSE testing of the procedure was applied to TIERs 1, 3 , and 4 so it was assumed that this rule could be applied to the outcomes of assessments from all Tiers. The rule requires generating the natural $\log$ of the ratio of catch rates between the most recent year and the previous year (e.g. 2012 with 2011). If there is more than one significant fishing method within a particular fishery, or different regions in the fishery are considered separately in the assessment, then the RAGs decided that a catch weighted ratio of the analysis on both approaches or multiple regions (or both) was to be used. The use of log-transformed catch rates focuses on proportional changes and ensures that increases and decreases in catch rates give rise to the same adjustment to the RBC, up or down (Thomson et al., 2008). The effects of this rule have been explored using management strategy evaluation (Wayte et al., 2009). This rule has not been applied by the Shark RAG or Deep RAG. The GAB RAG has an alternative meta-rule (Haddon, 2012b).

The simulation tests (Wayte et al, 2009) indicate that the use of the catch rate adjustment rule does not significantly alter the performance of the harvest strategy being applied to a particular fishery. This means there is no increased risk to the stock and overall catch levels are not affected in the long term, although, not surprisingly, applying the rule does significantly increase year-to-year catch variability.

Nevertheless, there are some other issues with the use of this rule. It constitutes an increase in the weight given to the most recent CPUE trends above other factors in stock assessments (although this is only a temporary increase in influence). It is questionable whether it should be applied to those fisheries assessed using a TIER 4 approach. The TIER 4 method estimates the RBC using the average of the catch rates from the last four year's. By focussing attention on the catch rates from the very last year this is changing the intent of the assessment method.

After discussions with RAG chairs, stock assessment scientists and industry, the preference is to apply the rule to all species for the purposes of calculating final TACs for each fishing season. However, while the MSE did not find any negative effects of using catch rates from just the last year (the outcome was effectively adding noise to the time series) this still appears to circumvent the intention of the assessment method and this should be discussed. Similarly with the TIER 3 assessments, the TIER 3 approach is based around estimating fishing mortality based performance measures and it should
also be discussed whether the intent of the method is being subverted by using the one year catch rate rule.

### 14.3 Methods

### 14.3.1 The Balanced Multiplier Rule

The TAC modifying rule is simply the original TAC from the given stock assessment multiplied by a proportional increase or decrease depending on the ratio of the most recent catch rate to the previous catch rate:

$$
\begin{equation*}
T A C_{\text {new }}=T A C_{\text {ass }}(1+\alpha \hat{R}) \tag{4}
\end{equation*}
$$

In order for the proportional change to the TAC $(\alpha \hat{R})$ to be equivalent whether it be increasing or decreasing, the method needs to natural log transform (i.e. using base $e$ ) the catch rate ratio:

$$
\begin{equation*}
\hat{R}=\operatorname{Ln}\left(C E_{y+1} / C E_{y}\right) \tag{5}
\end{equation*}
$$

where $C E_{y}$ is the standardized catch rate in year $y$. Last year this log-ratio was divided by $e^{l}$ but the same effect can be induced by setting the scaling factor to a value of $\alpha=$ 0.367879 (instead of $\alpha=1.0$ ), which was found to provide outcomes acceptable to the RAG; this value was used in the forthcoming analyses. This rule provides a symmetrical relationship between changes in the catch rates and in the TAC. Thus, for any catch rate ratio the percentage change in TAC is the same whether the ratio steps up from a smaller to a larger catch rate, or down from a larger to a smaller catch rate. What this means is that if the catch rate trajectory starts at $X$, then increases to $Y$ but then decreases back to $X$ again, the TAC imposed at $X$ is the same. Without this symmetrical relationship the TAC at $X$ would end up larger after these changes, which would be an undesirable side effect of such a rule.

The same species/fisheries as analysed in Haddon (2012) are presented here except for those species which have only been allocated a bycatch TAC. Only those catch rate standardizations are produced which were used in an assessment.


Figure 14.1. The proportional change in TAC when the proportional change in catch rate is either up (solid line) or down (dotted line). The lines overlap and are therefore hard to distinguish (Copied from Wayte et al, 2009).

### 1.1.1 Catch Weighting

The assessment for five species, Jackass Morwong, Flathead, Pink Ling, Ocean Perch, and now Mirror Dory are based on multiple time series of catch rates (from different zones) and, in some cases, multiple methods. Only one TAC Multiplier is used to modify the allocated TAC and so, in these species the separate TAC Multipliers from the different time series need to be combined in some way. They are combined by weighting each contribution by the proportion of the total catch taken in each region of the fishery or by each method. The areas are combined first (within methods), and then the methods are combined. The relative catches are determined using the same catch and effort data used in the catch rate standardizations.

### 1.1.2 Data Manipulations

Data from 2011 to at least the end of the third week in October for each fishery were sent from AFMA and processed in the usual way (Danish Seine vessels are identified, Zones are added, DayNight is added, depths are standardized across different ways of estimating it through the years, and the vessel IDs are made consistent through time).

The same standard set of database extracts were used as those used to generate the standardized catch rates (Haddon, 2012). These were designed to identify positive shots containing the species of interest in each case. For each species the standardization analyses were restricted to data from particular methods, zones and depth ranges as listed in Haddon (2012). A repeat standardization for Pink Ling Autoline was also conducted to add this method to the weighted ratio analysis for Pink Ling.

The statistical software $R$ was used in all analyses (R Development Core Team, 2009), which, because of the large size of the datasets, required the use of the library "biglm". The extra analyses needed for the application of the rule and the calculation of the TAC Multiplier, Eq. 4, were also programmed into the R scripts.

### 14.4 Results

A summary of the TAC Multiplier for each fishery demonstrates that there were only two fisheries (Royal Red Prawn and Blue Grenadier) for which the CPUE ratio change was greater than $10 \%$, which is the cut-off level in the meta-rule for action to be taken.

Table 14.1. Summary results for each analysis. Only using the final TAC multiplier values for those
species where a weighted TAC multiplier was generated, there were only two reducing multipliers (those
which were either greater than 1.1 or less than 0.9 . The Final TAC multipliers relate to whole fisheries,
while the sub-group multipliers relate to individual component fisheries.

| Species | Zone | Final TAC Multiplier | Sub-Group |
| :---: | :---: | :---: | :---: |
| 1 School Whiting | 60 | 1.0272 |  |
| 2 Jackass Morwong Summary |  | 1.0052 |  |
| Jackass Morwong | 10,20 |  | 1.0276 |
| Jackass Morwong | 30 |  | 1.1016 |
| Jackass Morwong | 40,50 |  | 0.8761 |
| 3 Flathead Summary |  | 1.0083 |  |
| Flathead | 10,20 |  | 1.0364 |
| Flathead | 30 |  | 1.1058 |
| Flathead DS | 20,60 |  | 0.9596 |
| 4 RedFish | 10 | 0.9352 |  |
| 5 Silver Trevally | 10,20 | 0.9617 |  |
| 6 Royal Red Prawn | 10 | 0.8604 |  |
| 7 Blue Eye | 10-50 | 0.9373 |  |
| 8 Blue Grenadier Summary |  | 0.8358 |  |
| Blue Grenadier Spawning | 40 |  | 0.8549 |
| Blue Grenadier Non-Sp | 10-60 |  | 0.7768 |
| 9 Spotted/Silver Warehou | 10-50 | 0.9157 |  |
| 10 Pink Ling Summary |  | 0.9854 |  |
| Pink Ling, | 10,20,30 |  | 0.9755 |
| Pink Ling, | 40,50 |  | 0.9874 |
| Pink Ling AutoLine | 20-50,83- |  | 1.0322 |
| 11 Western Gemfish | 40,50 | 0.9973 |  |
| 12 Offshore Ocean Perch | 10,20 | 1.0317 |  |
| 13 Inshore Ocean Perch | 10,20 | 0.9004 |  |
| 14 John Dory | 10,20 | 0.9931 |  |
| 15 Mirror Dory Summary | 10-50 | 0.9187 |  |
| Mirror Dory East | 10,20,30 |  | 0.9384 |
| Mirror Dory West | 40,50 |  | 0.8691 |
| 16 Ribaldo | 10-50 | 0.9295 |  |

### 14.5 School Whiting (WHS - 37330014) Sillago flindersi

School Whiting are taken primarily by Danish Seine (and within State waters). In Commonwealth waters the catches are primarily within Zone 60, and in depths less than or equal to 100 m . Catch rates were expressed as the natural $\log$ of catch per shot.

Only data from Zone 60 in depths $0-100 \mathrm{~m}$ taken by Danish Seine were used. All vessels reporting School Whiting were included.


Figure 14.2. School Whiting reported from Danish Seine in Zone 60 , in depths 0 to 100 m . The solid black line represents the optimal standardized catch rates (the model including the Month:DepCat interaction term) to October 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization. The right hand graph illustrates $2 \times$ Standard Errors around the curve.

The catch rates in 2012 are slightly higher than in 2011, overall leading to a TAC multiplier of 1.0272 .

Table 14.2. The standardized catch rates for the alternative statistical models for School Whiting are reported from Zone 60 in depths 0 to 100 m . The optimal model was Model 7.

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 0.8351 |
| 2012_CE | 0.8992 |
| Ln(CE_Ratio) | 0.0739 |
| Rprime | 0.0272 |
| TAC_Multiplier | 1.0272 |

### 14.6 Jackass Morwong Summary (MOR-37377003 Nemadactylus macropterus)

Three regions are assessed in the formal assessment and so a weighted TAC Multiplier is required that combines all three regions.

Table 14.3. Catch in tonnes by Zone (Zones $10-20$, 30, and $40-50$ ) of Jackass Morwong across all depths for the years $2000-2011$. The relative percent between assessment regions are labelled with a $\%$.

| Year | $10-20$ | 30 | $40-50$ | $\% 1020$ | $\% 30$ | $\% 4050$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2000 | 605.099 | 126.264 | 121.507 | 70.95 | 14.80 | 14.25 |
| 2001 | 312.265 | 113.019 | 287.745 | 43.79 | 15.85 | 40.36 |
| 2002 | 379.784 | 110.840 | 255.919 | 50.87 | 14.85 | 34.28 |
| 2003 | 290.756 | 196.717 | 177.131 | 43.75 | 29.60 | 26.65 |
| 2004 | 255.636 | 205.965 | 190.392 | 39.21 | 31.59 | 29.20 |
| 2005 | 326.526 | 151.993 | 199.892 | 48.13 | 22.40 | 29.46 |
| 2006 | 324.883 | 166.055 | 192.288 | 47.55 | 24.30 | 28.14 |
| 2007 | 251.842 | 118.953 | 122.517 | 51.05 | 24.11 | 24.84 |
| 2008 | 375.240 | 122.652 | 105.954 | 62.14 | 20.31 | 17.55 |
| 2009 | 264.633 | 55.928 | 65.965 | 68.46 | 14.47 | 17.07 |
| 2010 | 231.117 | 59.890 | 40.983 | 69.62 | 18.04 | 12.34 |
| 2011 | 247.835 | 51.259 | 86.373 | 64.29 | 13.30 | 22.41 |
| 2012 | 199.503 | 77.302 | 27.702 | 65.52 | 25.39 | 9.10 |
| Average percent between 2000 and 2011 |  | 54.99 | 20.30 | 24.71 |  |  |

Table 14.4. Calculation of the weighted TAC Multiplier using the regional multipliers from Zones 10 20,30 , and $40-50$. The Proportion is the proportion of the total catch from the specified zones. The TACm are the regional multipliers, and the Contribution is the first two rows multiplied together, The Weighted TAC Multiplier is then obtained by summing the contributions.

| Region | 1020 | 30 | 4050 |
| :--- | ---: | ---: | ---: |
| Proportion | 0.5499 | 0.2030 | 0.2471 |
| TACm | 1.0276 | 1.1016 | 0.8761 |
| Contribution | 0.5650 | 0.2237 | 0.2165 |
| Weighted TAC Multiplier for Jackass Morwong |  | 1.0052 |  |

14.7 Jackass Morwong Z1020 (MOR-37377003 N. macropterus)


Figure 14.3. Jackass Morwong reported from trawl in Zones 10 - 20, in depths 70 to 300 m . The solid black line represents the optimal standardized catch rates (the model including the Zone:Month interaction term) to October 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization. The right hand graph illustrates $2 \times$ Standard Errors around the curve.

The catch rates in 2012 are somewhat higher than in 2011, overall leading to a TAC multiplier of 1.0276 . However, to obtain an appropriately weighted single TAC multiplier it is necessary to weight the analyses for each of the three assessment regions with respect to their relative catches.

Table 14.5. The standardized catch rates for the alternative statistical models for Jackass Morwong reported from trawl in Zones $10-20$, in depths 70 to 300 m . The optimal model was Model 8. The weighted TAC Multiplier is shown in Table 14.4.

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 0.4641 |
| 2012_CE | 0.5002 |
| Ln(CE_Ratio) | 0.0750 |
| Rprime | 0.0276 |
| TAC_Multiplier | 1.0276 |

14.8 Jackass Morwong Z30 (MOR - 37377003 N. macropterus)

Only data from zone 30 were used with depths between 70 and 300 m .


Figure 14.4. Jackass Morwong reported from trawl in Zone 30, in depths 70 to 300 m . The solid black line represents the optimal standardized catch rates (the model including the Month:DepCat interaction term) to Oct 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization. The right hand graph illustrates $2 \times$ Standard Errors around the curve.

The catch rates in 2012 are rather higher than in 2011, overall leading to a TAC multiplier of 1.1016 . However, to obtain an appropriately weighted single TAC multiplier it is necessary to weight the analyses for each of the three assessment regions with respect to their relative catches.

Table 14.6. The standardized catch rates for the alternative statistical models for Jackass Morwong reported from trawl in Zone 30, in depths 70 to 300 m . The optimal model was Model 6. The weighted TAC Multiplier is shown in Table 14.4.

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 0.2689 |
| 2012_CE | 0.3545 |
| Ln(CE_Ratio) | 0.2763 |
| Rprime | 0.1016 |
| TAC_Multiplier | 1.1016 |

14.9 Jackass Morwong Z4050 (MOR - 37377003 N. macropterus)

The data restrictions used in selecting the data for analysis were trawl caught Jackass Morwong from Zones 40 and 50, and depths between 70 and 360 m .


Figure 14.5. Jackass Morwong reported from trawl in Zones 40 - 50, in depths 70 to 360 m . The solid black line represents the optimal standardized catch rates (the model including the Zone:Month interaction term) to October 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization. The right hand graph illustrates $2 \times$ Standard Errors around the curve.

The catch rates in 2012 are rather lower than in 2011, overall leading to a TAC multiplier of 0.8761 . However, to obtain an appropriately weighted single TAC multiplier it is necessary to weight the analyses for each of the three assessment regions with respect to their relative catches.

Table 14.7. The standardized catch rates for the alternative statistical models for Jackass Morwong reported from trawl in Zones $40-50$, in depths 70 to 360 m . The optimal model was Model 8. The weighted TAC Multiplier is shown in Table 14.4.

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 0.4439 |
| 2012_CE | 0.3169 |
| Ln(CE_Ratio) | -0.3369 |
| Rprime | -0.1239 |
| TAC_Multiplier | 0.8761 |

### 14.10 Flathead Summary (FLT - 37296001 - Neoplatycephalus richardsoni)

The flathead stock assessment uses trawl caught catches from Zones $10-20$ and from Zone 30 separately, it also uses Danish Seine catches (from Zones 20 and 60).

Table 14.8. Annual catches of Flathead by Zone and method. The average percent over the years 2000 2008 are given at the bottom.

| Year | 1020 | 30 | TW | DS | $\% 1020$ | $\% 30$ | $\%$ TW | $\% \mathrm{DS}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2000 | 1699.619 | 83.664 | 1783.283 | 1038.428 | 95.31 | 4.69 | 63.20 | 36.80 |
| 2001 | 1388.282 | 101.308 | 1489.590 | 1005.508 | 93.20 | 6.80 | 59.70 | 40.30 |
| 2002 | 1447.242 | 212.158 | 1659.400 | 1147.232 | 87.21 | 12.79 | 59.12 | 40.88 |
| 2003 | 1599.972 | 240.860 | 1840.832 | 1214.571 | 86.92 | 13.08 | 60.25 | 39.75 |
| 2004 | 1349.783 | 477.426 | 1827.209 | 1254.558 | 73.87 | 26.13 | 59.29 | 40.71 |
| 2005 | 1159.545 | 388.383 | 1547.928 | 1126.864 | 74.91 | 25.09 | 57.87 | 42.13 |
| 2006 | 1154.279 | 288.378 | 1442.657 | 970.529 | 80.01 | 19.99 | 59.78 | 40.22 |
| 2007 | 1077.223 | 173.180 | 1250.404 | 1182.317 | 86.15 | 13.85 | 51.40 | 48.60 |
| 2008 | 1332.344 | 173.739 | 1506.083 | 1283.714 | 88.46 | 11.54 | 53.99 | 46.01 |
| 2009 | 1063.573 | 100.225 | 1163.798 | 1169.299 | 91.39 | 8.61 | 49.88 | 50.12 |
| 2010 | 1127.232 | 104.186 | 1231.418 | 1173.237 | 91.54 | 8.46 | 51.21 | 48.79 |
| 2011 | 1096.784 | 131.274 | 1228.059 | 1125.254 | 89.31 | 10.69 | 52.18 | 47.82 |
| 2012 | 884.820 | 139.493 | 1024.313 | 1104.554 | 86.38 | 13.62 | 48.12 | 51.88 |
| Average percent across the years $2000-2010$ | 86.52 | 13.48 | 56.49 | 43.51 |  |  |  |  |

Table 14.9. Calculation of the weighted TAC Multiplier using the regional multipliers from Zones 10 20 and 30, and from Trawl and Danish Seine. The Proportions are from Table 14.8. The TACm are the regional multipliers, and the Contribution is the first two rows multiplied together. The Trawl TACm is the left hand two contributions added together. The Weighted TAC Multiplier is then obtained by summing the contributions from the Trawl and the Danish Seine.

| Region/Method | TW10202 | TW30 | Trawl | DS |
| :--- | ---: | ---: | ---: | ---: |
| Proportion | 0.8652 | 0.1348 | 0.5649 | 0.4351 |
| TACm | 1.0364 | 1.1058 | 1.0458 | 0.9596 |
| Contribution | 0.8967 | 0.1490 | 0.5907 | 0.4175 |
| Weighted TAC Multiplier |  |  |  | 1.0083 |

### 14.10.1 Flathead Trawl Z1020(FLT - 37296001 - N. Richardsoni)

Only trawl data from zones 10 and 20 were used from depths less than 400 m .


Figure 14.6. Flathead reported from trawl in Zones $10-20$, in depths 0 to 400 m . The solid black line represents the optimal standardized catch rates (the model including the DepCat:Zone interaction term) to October 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization. The right hand graph illustrates $2 \times$ Standard Errors around the curve.

The catch rates in 2012 are somewhat higher than in 2011, overall leading to a TAC multiplier of 1.0364 . However, to obtain an appropriately weighted single TAC multiplier it is necessary to weight the analyses for each of the three assessment regions and two methods with respect to their relative catches.

Table 14.10. The standardized catch rates for the alternative statistical models for Flathead reported from trawl in Zones $10-20$, in depths 0 to 400 m . The optimal model was Model 8. The weighted TAC Multiplier is shown in Table 14.9.

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 1.0511 |
| 2012_CE | 1.1604 |
| Ln(CE_Ratio) | 0.0990 |
| Rprime | 0.0364 |
| TAC_Multiplier | 1.0364 |

14.10.2 Flathead Trawl Z30 (FLT - 37296001 - N. Richardsoni)

Only trawl data from zone 30 were used from depths between $0-400 \mathrm{~m}$.


Figure 14.7. Flathead reported from trawl in Zone 30, in depths 0 to 400 m . The solid black line represents the optimal standardized catch rates (the model including the DepCat:Month interaction term) to October 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization. The right hand graph illustrates $2 \times$ Standard Errors around the curve. Table 14.9).

Table 14.11.The standardized catch rates for the alternative statistical models for Flathead reported from trawl in Zone 30, in depths 0 to 400 m . The optimal model was Model 6. The weighted TAC Multiplier is shown in Table 14.9.

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 0.9760 |
| 2011_CE | 1.3010 |
| Ln(CE_Ratio) | 0.2875 |
| Rprime | 0.1058 |
| TAC_Multiplier | 1.1058 |

### 14.10.3 Flathead Danish Seine (FLT - 37296001 - N. Richardsoni)

Only Danish Seine data from zones 20, and 60 were used (i.e. Otter Trawl vessels were excluded), and from depths less than 200 m .



Figure 14.8. Flathead reported from Danish Seine in Zones 20 and 60 in depths 0 to 200 m . The solid black line represents the optimal standardized catch rates (the model including the DepCat:Month interaction term) to October 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization. The right hand graph illustrates 2 x Standard Errors around the curve.

The catch rates in 2012 are again lower than in 2011, overall leading to a TAC multiplier of 0.9596 . However, to obtain an appropriately weighted single TAC multiplier it is necessary to weight the analyses for each of the three assessment regions and two methods with respect to their relative catches (see Table 14.9).

Table 14.12.The standardized catch rates for the alternative statistical models for Flathead reported from Danish Seine in Zones 20 and 60 in depths 0 to 200 m . The optimal model was Model 6. The weighted TAC Multiplier is shown in Table 14.9

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 0.8972 |
| 2012_CE | 0.8038 |
| Ln(CE_Ratio) | -0.1099 |
| Rprime | -0.0404 |
| TAC_Multiplier | 0.9596 |

### 14.11 RedFish Zone 10 (RED - 37258003 - Centroberyx affinis)

Only data taken by trawl from Zone 10 were used from depths less than 400 m .


Figure 14.9. Redfish reported from trawl in Zone 10, in depths 0 to 400 m . The solid black line represents the optimal standardized catch rates (the model including the Month:DepCat interaction term) to October 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization. The right hand graph illustrates $2 \times$ Standard Errors around the curve.

The catch rate in 2012 was again somewhat lower than in 2011, overall leading to a TAC multiplier of 0.9352 .

Table 14.11. The standardized catch rates for the alternative statistical models for Redfish reported from trawl in Zone 10, in depths 0 to 400 m . The optimal model was Model 6.

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 0.2506 |
| 2012_CE | 0.2101 |
| Ln(CE_Ratio) | -0.1760 |
| Rprime | -0.0648 |
| TAC_Multiplier | 0.9352 |

### 14.12 Silver Trevally (TRE - 37337062 - Pseudocaranx dentex)

Only data from zones 10 and 20 combined were used, depths less than 200 m . Previously, to discount the influence of catches taken within the Batemans Bay MPA, all data in State waters within the MPA have been excluded from the analysis. This usage removed more records than required in practice. There are now sufficient data to permit an empirical determination of what areas to exclude from the analysis (Figure 14.10).


Figure 14.10. Schematic map of individual sites where fishing for Silver Trevally has been reported. The blue dots are all records from 1986 - 2012, whereas the red dots are all data from 2009 - 2012; the few red dots inside the green exclusion area are assumed to be erroneous reporting from inside the MPA. The green line indicates the region whose previous catches were excluded from the analysis.


Figure 14.11. Silver Trevally from Zones 10 and 20, in depths 0 to 200 m , minus the records from those areas within the Bateman's Bay MPA, from which recent catches have been excluded (Figure 14.10). The solid black line represents the optimal standardized catch rates (the model including the Zone:Month interaction term) to October 20122, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization and the vertical lines are 2 x Standard Error.

This analysis of catch rates is slightly different from the analysis given in the original standardization document. This can be seen from the differences between this analysis and that used in the standardization document (Haddon, 2012). However, it would appear to be more valid to remove only those catches from areas where catches are no longer taken than from all apparently closed areas (Figure 14.10). The new analysis more closely follows the geometric mean catch rate except in the latest year where the standardized rates go down while the geometric rates go up. The total catches in this analysis are greater than that in Haddon (2012). This use of an alternative analysis has had no effect on the difference between the last two years and so has not influenced the outcome of the update analysis.

The catch rates in 2012 are somewhat lower than in 2011, overall leading to a TAC multiplier of 0.9617 .

Table 14.12. The standardized catch rates for the alternative statistical models for Silver Trevally from Zones 10 and 20, in depths 0 to 200 m , minus the records from State Waters within the Bateman's Bay MPA. The optimal model was Model 7.

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 0.9399 |
| 2011_CE | 0.8469 |
| Ln(CE_Ratio) | -0.1041 |
| Rprime | -0.0383 |
| TAC_Multiplier | 0.9617 |

### 14.13 Royal Red Prawn (PRR - 28714005 - Haliporoides sibogae)

Only data taken by trawl from Zone 10 were used between depths of $200-700 \mathrm{~m}$.


Figure 14.12. Royal Red Prawn reported from trawl in Zone 10, in depths 200 to 700 m . The solid black line represents the optimal standardized catch rates (the model including the Month:DepCat interaction term) to October 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization. The right hand graph illustrates $2 \times$ Standard Errors around the curve.

The catch rates in 2012 are rather higher than in 2011, overall leading to a TAC multiplier of 0.8604 . This analysis fails to take into account the use of specialized prawn nets so the TAC setting committee should consider whether to use this result.

Table 14.13. The standardized catch rates for the alternative statistical models for Royal Red Prawn reported from trawl in Zone 10, in depths 200 to 700 m . The optimal model was Model 6.

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 1.3482 |
| 2012_CE | 0.9223 |
| Ln(CE_Ratio) | -0.3796 |
| Rprime | -0.1396 |
| TAC_Multiplier | 0.8604 |

### 14.14 Blue Eye, AL \& DL (TBE - 37445001 - H. antarctica)

Depths between $200-600 \mathrm{~m} \mathrm{~m}$. All data from auto-longlining and droplining combined. Zones 20, 30, 40, 50, 83, 84, and 85 included ( $83-85$ are in the GAB).


Figure 14.13. Blue Eye reported from Autolong line and Drop line in Zones $20-50$ and $83-85$, from depths 200 to 600 m . The solid black line represents the optimal standardized catch rates (the model including the Month:Zone interaction term) to October 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization. The right hand graph illustrates $2 \times$ Standard Errors around the curve.

The catch rates in 2012 are somewhat higher than in 2011, overall leading to a TAC multiplier of 0.9373 .

Table 14.14. The standardized catch rates for the alternative statistical models for Blue Eye reported from Auto-long-line and Drop line in Zones $20-50$ and $83-85$, from depths 200 to 600 m . The optimal model was Model 7.

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 0.6704 |
| 2012_CE | 0.5653 |
| Ln(CE_Ratio) | -0.1704 |
| Rprime | -0.0627 |
| TAC_Multiplier | 0.9373 |

### 14.15 Blue Grenadier Summary (GRE - 37227001 Macruronus novaezelandiae)

There is a spawning fishery and a non-spawning fishery, each of which is currently used in the stock assessment. In future the spawning time series is no longer going to be used in the assessment but for the setting of the multi-year TAC the pair was used. There is therefore a need to weight their respective TAC Multipliers by their relative contribution to the total catches. Because of the Multi-year TAC this analysis would appear to be redundant.

Table 14.15. Annual catches of Blue Grenadier by fishery with their relative contribution to the total catch.

| Year | Non-Spawning | Spawning | \%NonSp | $\%$ Sp |
| ---: | ---: | ---: | ---: | ---: |
| 2000 | 1715.640 | 5616.097 | 23.40 | 76.60 |
| 2001 | 1013.169 | 7281.279 | 12.22 | 87.78 |
| 2002 | 1126.150 | 6782.001 | 14.24 | 85.76 |
| 2003 | 670.094 | 7132.933 | 8.59 | 91.41 |
| 2004 | 1204.672 | 4603.280 | 20.74 | 79.26 |
| 2005 | 1166.433 | 2829.270 | 29.19 | 70.81 |
| 2006 | 1293.355 | 2034.510 | 38.86 | 61.14 |
| 2007 | 1196.315 | 1728.897 | 40.90 | 59.10 |
| 2008 | 1256.788 | 2805.832 | 30.94 | 69.06 |
| 2009 | 1112.922 | 2700.396 | 29.19 | 70.81 |
| 2010 | 1131.383 | 3280.615 | 25.64 | 74.36 |
| 2011 | 878.786 | 3525.454 | 19.95 | 80.05 |
| 2012 | 339.396 | 3746.786 | 8.31 | 91.69 |
| Average percent contribution from $2000-2010$ | 24.96 | 75.04 |  |  |

Table 14.16. Calculation of the weighted TAC Multiplier using the fishery multipliers from the NonSpawning and Spawning fisheries. The Proportions are from . The TACm are the fishery multipliers, and the Contribution is the first two rows multiplied together, The Weighted TAC Multiplier is then obtained by summing the contributions from the Non-Spawning and Spawning fisheries

|  | Non-Spawning | Spawning |
| :--- | :---: | :---: |
| Proportion | 0.2449 | 0.7551 |
| TACm | 0.7768 | 0.8549 |
| Contribution | 0.1902 | 0.6456 |
| Weighted TAC Multiplier |  | 0.8358 |

14.15.1 Blue Grenadier Spawning (GRE - 37227001 - M. novaezelandiae)

Data from Zone 40 in months June to August, depths between 100 and 1000 m , all vessels that reporting fishing in the spawning fishery were included in the analysis.


Figure 14.14. Blue Grenadier from the Spawning fishery reported from Trawl in Zone 40 during June to August from depths 200 to 1000 m . The solid black line represents the optimal standardized catch rates (the model included no significant interaction terms) to October 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization. The right hand graph illustrates $2 \times$ Standard Errors around the curve.

The catch rates in 2012 are slighter lower than in 2011, overall leading to a TAC multiplier of 0.8549 . However, this multiplier needs to be weighted with that from the non-spawning fishery.

Table 14.17. The standardized catch rates for the alternative statistical models for Blue Eye reported from Auto-long-line and Drop line in Zones $20-50$ and $83-85$, from depths 200 to 600 m . The optimal model was Model 7.

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 0.9651 |
| 2012_CE | 0.6505 |
| Ln(CE_Ratio) | -0.3945 |
| Rprime | -0.1451 |
| TAC_Multiplier | 0.8549 |

### 14.15.2 Blue Grenadier Non-Spawning (GRE - 37227001 - M. novaezelandiae)

Data from zones 10 to 60 , except Zone 40 in months June to August, depths less than 1000 m and greater than 0 m .


Figure 14.15. Blue Grenadier from the Non-Spawning fishery reported from Trawl in Zones $10-60$ (except Zone 40 in June to August) from depths 0 to 1000 m . The solid black line represents the optimal standardized catch rates (the model including the Zone:Month interaction term) to October 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization. The right hand graph illustrates $2 \times$ Standard Errors around the curve.

The catch rates in 2012 are rather lower than in 2011, overall leading to a TAC multiplier of 0.7768 . However, this multiplier needs to be weighted with that from the spawning fishery.

Table 14.18. The standardized catch rates for the alternative statistical models for Blue Eye reported from Auto-long-line and Drop line in Zones $20-50$ and $83-85$, from depths 200 to 600 m . The optimal model was Model 7.

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 0.6735 |
| 2012_CE | 0.3671 |
| Ln(CE_Ratio) | -0.6068 |
| Rprime | -0.2232 |
| TAC_Multiplier | 0.7768 |

### 14.16 Silver Warehou (TRS - 37445006 - Seriolella punctata)

Trawl data for Silver/Spotted Warehou from zones 10 to 50 , depths greater than 0 m .


Figure 14.16. Silver Warehou data from Zones $10-50$ and depths greater than 0 m The solid black line represents the optimal standardized catch rates (the model including the Zone:Month interaction term) to October 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization. The right hand graph illustrates $2 \times$ Standard Errors around the curve.

The catch rates in 2012 are rather lower than in 2011, overall leading to a TAC multiplier of 0.9157 .

Table 14.19. The standardized catch rates for the alternative statistical models for Silver Warehou data from Zones $10-50$ and depths greater than 0 m The optimal model was Model 7.

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 0.4954 |
| 2012_CE | 0.3939 |
| Ln(CE_Ratio) | -0.2292 |
| Rprime | -0.0843 |
| TAC_Multiplier | 0.9157 |

### 14.17 Pink Ling Summary (LIG - 37228002 - Genypterus blacodes)

Table 14.20. Trawl Catch of Pink Ling by Zonal region, including Zones $10-30$ and Zones $40-50$. The percent contribution by each region is shown in the $\% 102030$ and $\% 4050$ columns. The relative weightings given the two areas are 0.5534 and 0.4466 .

| Year | $10-30$ | $40-50$ | $\% 102030$ | $\% 4050$ |
| ---: | ---: | ---: | ---: | ---: |
| 2000 | 658.895 | 508.949 | 56.42 | 43.58 |
| 2001 | 483.313 | 498.442 | 49.23 | 50.77 |
| 2002 | 360.184 | 429.004 | 45.64 | 54.36 |
| 2003 | 444.627 | 358.987 | 55.33 | 44.67 |
| 2004 | 346.188 | 304.417 | 53.21 | 46.79 |
| 2005 | 324.814 | 195.212 | 62.46 | 37.54 |
| 2006 | 321.107 | 207.895 | 60.70 | 39.30 |
| 2007 | 202.762 | 284.511 | 41.61 | 58.39 |
| 2008 | 325.428 | 211.797 | 60.58 | 39.42 |
| 2009 | 208.330 | 258.294 | 44.65 | 55.35 |
| 2010 | 265.716 | 268.810 | 49.71 | 50.29 |
| 2011 | 287.423 | 345.186 | 45.43 | 54.57 |
| 2012 | 180.264 | 193.406 | 48.24 | 51.76 |
| Average Annual Percentage from 2000 to 2010 | 52.08 | 47.92 |  |  |

Table 14.21. Catch of Pink Ling by Method. The percent contribution by each region is shown in the \%AutoLine and \%Trawl columns. The relative weightings given to the two methods are 0.3805 and 0.6195. 2002 was selected as the start year because the Auto Line method had become established by then.

| Year | AutoLine | Trawl | \%AutoLine | \%Trawl |
| ---: | ---: | ---: | ---: | ---: |
| 2000 | 54.720 | 1167.844 | 4.48 | 95.52 |
| 2001 | 176.418 | 981.755 | 15.23 | 84.77 |
| 2002 | 379.354 | 789.188 | 32.46 | 67.54 |
| 2003 | 382.861 | 803.614 | 32.27 | 67.73 |
| 2004 | 704.479 | 650.605 | 51.99 | 48.01 |
| 2005 | 524.440 | 520.026 | 50.21 | 49.79 |
| 2006 | 419.985 | 529.002 | 44.26 | 55.74 |
| 2007 | 294.705 | 487.273 | 37.69 | 62.31 |
| 2008 | 365.753 | 537.225 | 40.51 | 59.49 |
| 2009 | 253.504 | 466.624 | 35.20 | 64.80 |
| 2010 | 318.338 | 534.526 | 37.33 | 62.67 |
| 2011 | 373.726 | 632.609 | 37.14 | 62.86 |
| 2012 | 198.487 | 373.670 | 34.69 | 65.31 |
| Average Annual Percentage from 2002 to 2010 | 34.90 | 65.10 |  |  |

Table 14.22. Calculation of the weighted TAC Multiplier using the regional multipliers from Zones 10 30 and $40-50$, and from Trawl and Danish Seine. The Proportions are from Table 14.20 and Table 14.21. The TACm are the regional multipliers, and the Contribution is the first two rows multiplied together. The Trawl TACm is the left hand two contributions added together. The Weighted TAC Multiplier is then obtained by summing the contributions from the Trawl and the Danish Seine.

| Region/Method | TW102030 | TW4050 | Trawl | AutoL |
| :--- | ---: | ---: | ---: | ---: |
| Proportion | 0.5208 | 0.4792 | 0.3490 | 0.6510 |
| TACm | 0.9755 | 0.9874 | 0.9812 | 0.9876 |
| Contribution | 0.5080 | 0.4732 | 0.3424 | 0.6429 |
| Weighted TAC Multiplier |  |  |  | 0.9854 |

## 

Data from zones 10, 20 and 30, depths greater than 0 m and less than 600 m .


Figure 14.17. Pink Ling trawl data from Zones $10-30$ and depths between 0 m and 600 m . The solid black line represents the optimal standardized catch rates (the model including the Zone:DepCat interaction term) to October 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization.

The catch rates in 2012 are lower than those in 2011, overall leading to a TAC multiplier of 0.9755 .

Table 14.23. The standardized catch rates for the alternative statistical models for Pink Ling trawl data from Zones $10-30$ and depths between 0 m and 600 m . The optimal model was Model 8.

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 0.8095 |
| 2012_CE | 0.7573 |
| Ln(CE_Ratio) | -0.0667 |
| Rprime | -0.0245 |
| TAC_Multiplier | 0.9755 |

### 14.17.2 Pink Ling, Z4050 (LIG - 37228002 - G. blacodes)

Data from zones 40 and 50, depths greater than 200 m and less or equal to 800 m .


Figure 14.18. Pink Ling trawl data from Zones $40-50$ and depths between 200 m and 800 m . The solid black line represents the optimal standardized catch rates (the model including the Zone:Month interaction term) to October 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization.

The catch rates in 2012 are slightly lower than in 2011, overall leading to a TAC multiplier of 0.9874 .

Table 14.24. The standardized catch rates for the alternative statistical models for Pink Ling trawl data from Zones $40-50$ and depths between 200 m and 800 m . The optimal model was Model 7 .

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 0.8478 |
| 2012_CE | 0.8193 |
| Ln(CE_Ratio) | -0.0343 |
| Rprime | -0.0126 |
| TAC_Multiplier | 0.9874 |

### 14.17.3 Pink Ling, AutoLine (LIG - 37228002 - G. blacodes)

For the TAC Multiplier to be appropriately calculated a standardized catch rate series for Auto-Long-Line caught Pink Ling was required. This was produced before the multiplier calculations. Methods were the same as in Haddon (2012).


Figure 14.19. Pink Ling as reported from Auto-Long-Line in Zones $20-50$ and $83-85$ in depths 200 to 800 m . The dashed black line represents the geometric mean catch rate and the solid black line the standardized catch rates (the model including the Zone:Month interaction term) to October 2012. The blue line on top of the black is last year's analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates.

The catch rates in 2012 are slightly higher than in 2011, overall leading to a TAC multiplier of 1.0322 .

Table 14.25. The standardized catch rates for the alternative statistical models for Pink Ling Auto-LongLine data from Zones $20-50$ and $83-85$ in depths between 200 m and 800 m . The optimal model was Model 7.

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 0.4748 |
| 2012_CE | 0.5181 |
| Ln(CE_Ratio) | 0.0874 |
| Rprime | 0.0322 |
| TAC_Multiplier | 1.0322 |

14.18 Western Gemfish Z4050 (GEM - 37439002 - R. solandri)

Data from zones 40 and 50, depths greater than 200 and less than or equal to 600 m .


Figure 14.20. Western Gemfish from Zones $40-50$ and depths 200 to 600 m . The solid black line represents the optimal standardized catch rates (the model including the Zone:Month interaction term) to October 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization.

The catch rates in 2012 are effectively the same as in 2011, overall leading to a TAC multiplier of 0.9973 .

Table 14.26. The standardized catch rates for the alternative statistical models for Western Gemfish from Zones $40-50$ and depths 200 to 600 m . The optimal model included Zone:Month.

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 0.7417 |
| 2012_CE | 0.7362 |
| Ln(CE_Ratio) | -0.0074 |
| Rprime | -0.0027 |
| TAC_Multiplier | 0.9973 |

### 14.19 Offshore Ocean Perch, Z1020 (REG - 37287001 - H. percoides) 200m

Previous analyses identified Offshore Ocean Perch by selecting records of Ocean Perch from depths between $300-700 \mathrm{~m}$. In the July 2010 Slope RAG again revised this figure down to $200-700 \mathrm{~m}$ to avoid overlap with inshore Ocean Perch. The following analyses are therefore restricted to data from $200-700 \mathrm{~m}$ by trawl from Zones $10-20$. The TAC this year is being set using both the Inshore and the Offshore Ocean Perch catch rate series, so a weighted TAC Multiplier combining Offshore and Inshore Ocean Perch is required.


Figure 14.21. Offshore Ocean Perch by trawl from Zones $10-20$ and depths 300 to 700 m . The solid black line represents the optimal standardized catch rates (the model including the Zone:Month interaction term) to October 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization.

The catch rates in 2012 are somewhat higher than in 2011, overall leading to a TAC multiplier of 1.0317 .

Table 14.27. The standardized catch rates for the alternative statistical models for Offshore Ocean Perch by trawl from Zones $10-20$ and depths 300 to 700 m . The optimal model was Model 7.

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 0.8616 |
| 2012_CE | 0.9391 |
| Ln(CE_Ratio) | 0.0862 |
| Rprime | 0.0317 |
| TAC_Multiplier | 1.0317 |

### 14.20 Inshore Ocean Perch, Z1020 (REG - 37287001 - H. percoides) $0-200 \mathrm{~m}$

The catch rate series for inshore Ocean Perch complements the Offshore Ocean Perch in that only data taken by trawl in Zones 10 and 20 between 0 and 200 m depth are used.


Figure 14.22. Inshore Ocean Perch by trawl from Zones $10-20$ and depths 0 to 299 m . The solid black line represents the optimal standardized catch rates (the model including the Zone:DepCat interaction term) to October 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization.

The catch rates in 2012 are slightly lower than in 2011, overall leading to a TAC multiplier of 0.9004 .

Table 14.28. The standardized catch rates for the alternative statistical models for Inshore Ocean Perch by trawl from Zones $10-20$ and depths 0 to 200 m . The optimal model was Model 8.

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 1.0016 |
| 2012_CE | 0.7640 |
| Ln(CE_Ratio) | -0.2708 |
| Rprime | -0.0996 |
| TAC_Multiplier | 0.9004 |

### 14.21 John Dory (DOJ - 37264004) Zeus faber

Only included are trawl catches from Zones 10 and 20 in depths $0-200 \mathrm{~m}$


Figure 14.23. John Dory from Zones $10-20$ in depths 0 to 200 m . The solid black line represents the optimal standardized catch rates (the model including the Zone:DepCat interaction term) to October 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization.

The catch rates in 2012 are very slightly lower than in 2011, overall leading to a TAC multiplier of 0.9931 .

Table 14.29. The standardized catch rates for the alternative statistical models for John Dory from Zones $10-20$ in depths 0 to 200 m . The optimal model was Model 8.

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 0.5509 |
| 2012_CE | 0.5407 |
| Ln(CE_Ratio) | -0.0187 |
| Rprime | -0.0069 |
| TAC_Multiplier | 0.9931 |

### 14.22 Mirror Dory Summary (DOM - 37264003 Zenopsis nebulosus)

To generate an overall TAC multiplier for Mirror Dory the multipliers from east and west need to be combined.

Table 14.30. Trawl Catch of Mirror Dory by Zonal region, including Zones $10-30$ and Zones $40-50$. The percent contribution by each region is shown in the $\% 102030$ and $\% 4050$ columns. The relative weightings given the two areas are 0.7194 and 0.2806 .

| Year | $10-30$ | $40-50$ | $\% 102030$ | $\% 4050$ |
| :---: | ---: | ---: | ---: | ---: |
| 2000 | 142.938 | 22.361 | 86.47 | 13.53 |
| 2001 | 128.790 | 104.890 | 55.11 | 44.89 |
| 2002 | 194.460 | 240.337 | 44.72 | 55.28 |
| 2003 | 406.214 | 153.916 | 72.52 | 27.48 |
| 2004 | 293.861 | 159.782 | 64.78 | 35.22 |
| 2005 | 424.496 | 99.625 | 80.99 | 19.01 |
| 2006 | 298.028 | 64.647 | 82.17 | 17.83 |
| 2007 | 203.162 | 63.157 | 76.29 | 23.71 |
| 2008 | 317.705 | 57.233 | 84.74 | 15.26 |
| 2009 | 338.488 | 122.938 | 73.36 | 26.64 |
| 2010 | 385.470 | 176.825 | 68.55 | 31.45 |
| 2011 | 347.527 | 155.046 | 69.15 | 30.85 |
| 2012 | 241.126 | 62.666 | 79.37 | 20.63 |
| Average Annual Percentage from 2000 to 2010 | 71.27 | 28.73 |  |  |

Table 14.31. Calculation of the weighted TAC Multiplier using the fishery multipliers from the east and west Mirror Dory fisheries. The Proportions are from Table 14.30. The TACm are the fishery multipliers, and the Contribution is the first two rows multiplied together, The Weighted TAC Multiplier is then obtained by summing the contributions from each fisheries

|  | $\% 10-30$ | $\% 40-50$ |
| :--- | :---: | :---: |
| Proportion | 0.7157 | 0.2843 |
| TACm | 0.9384 | 0.8691 |
| Contribution | 0.6716 | 0.2471 |
| Weighted TAC Multiplier |  | 0.9187 |

### 14.23 Mirror Dory East (DOM - 37264003 Z. nebulosus)

Catches of Mirror Dory are primarily taken by trawl. Other methods are ignored in this analysis. Only data from Zones 10 to 30 in depths $0-600 \mathrm{~m}$. All vessels reporting Mirror Dory were included.


Figure 14.24. Mirror Dory reported from trawling in Zones 10 to 30 , in depths 0 to 600 m . The solid black line represents the optimal standardized catch rates (the model including the Zone:Month interaction term) to October 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization.

The catch rates in 2012 are lower than in 2011, overall leading to a TAC multiplier of 0.9384 .

Table 14.32. The standardized catch rates for the alternative statistical models for Mirror Dory are reported from Zones 10 to 50 in depths 0 to 600 m . The optimal model was Model 7.

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 1.1908 |
| 2012_CE | 1.0072 |
| Ln(CE_Ratio) | -0.1674 |
| Rprime | -0.0616 |
| TAC_Multiplier | 0.9384 |

### 14.24 Mirror Dory West (DOM - 37264003 Z. nebulosus)

As with Eastern Mirror Dory, the majority of catches are taken by trawl and all other methods are ignored.

Only data from Zones 40 to 50 in depths $0-600 \mathrm{~m}$. All vessels reporting Mirror Dory were included.


Figure 14.25. Mirror Dory reported from trawling in Zones 40 to 50 , in depths 0 to 600 m . The solid black line represents the optimal standardized catch rates (the model including the Zone:Month interaction term) to October 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization.

The catch rates in 2012 are somewhat lower than in 2011, overall leading to a TAC multiplier of 0.8691 . The fact that the catch rate trends in the east and west are different indicate that the decision to analyse the east and west separately is well founded.

Table 14.33. The standardized catch rates for the alternative statistical models for Mirror Dory are reported from Zones 40 to 50 in depths 0 to 600 m . The optimal model was Model 7.

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 0.9097 |
| 2012_CE | 0.6373 |
| Ln(CE_Ratio $)$ | -0.3558 |
| Rprime | -0.1309 |
| TAC_Multiplier | 0.8691 |

### 14.25 Ribaldo (RBD - 37224002 - Mora moro)

Catches of Ribaldo are taken by trawl, and more recently also by autoline. Other methods are ignored in this analysis.

Table 14.34. Catch in tonnes by Method of Ribaldo across all zones and depths. Unk is unknown method, AL is autoline, TW is trawl, and Other is all other methods (with minor amounts being reported from 10 other methods).

| Year | Unk | AL | TW | Other |
| ---: | ---: | ---: | ---: | ---: |
| 1986 |  |  | 4.104 |  |
| 1987 |  |  | 7.941 |  |
| 1988 |  |  | 10.898 |  |
| 1989 |  |  | 3.342 |  |
| 1990 | 0.295 |  | 7.568 |  |
| 1991 | 0.495 |  | 12.838 |  |
| 1992 |  |  | 22.761 | 0.016 |
| 1993 | 0.060 |  | 41.938 |  |
| 1994 | 0.070 |  | 90.263 |  |
| 1995 | 0.341 | 1.480 | 100.208 |  |
| 1996 | 0.190 | 1.853 | 96.686 | 0.854 |
| 1997 | 0.225 | 2.197 | 67.976 | 1.195 |
| 1998 | 0.080 | 9.159 | 55.140 | 1.759 |
| 1999 | 0.229 | 15.720 | 59.724 | 2.412 |
| 2000 |  | 95.497 | 60.770 | 6.806 |
| 2001 |  | 103.017 | 77.575 | 1.611 |
| 2002 |  | 103.062 | 77.155 | 0.443 |
| 2003 |  | 37.209 | 53.071 | 0.744 |
| 2004 |  | 66.167 | 56.164 | 0.095 |
| 2005 |  |  | 49.080 | 0.284 |
| 2006 |  | 57.415 | 20.136 | 0.510 |
| 2007 |  | 68.921 | 35.121 | 0.923 |
| 2008 |  | 51.940 | 39.764 | 0.918 |
| 2009 |  | 36.521 | 46.915 | 0.400 |
| 2010 |  |  | 24.150 | 0.593 |
| 2011 |  |  |  | 3.802 |
| 2012 |  |  |  |  |

Only data from Zones 10 to 50 in depths $0-1000 \mathrm{~m}$. A significant amount of Ribaldo is now taken by Autoline and if future assessments include this aspect of the fishery this will need to be included in future assessments and updates of catch rates.


Figure 14.26. Ribaldo reported from trawling in Zones 10 to 50 , in depths 0 to 1000 m . The solid black line represents the optimal standardized catch rates (the model including the Zone:Month interaction term) to October 2012, while the blue overlapping line represents the same optimal standardized catch rates from the 2011 analysis. The graph standardizes catch rates relative to the mean of the standardized catch rates. The dashed line is the geometric mean catch rate, used to illustrate the effect of the standardization.

The catch rates in 2012 are lower than in 2011, leading to a TAC multiplier of 0.9095 . The catch by Autoline is now equal to or greater than the trawl catch and should be included in future assessments (Table Table 14.34).

Table 14.35. The standardized catch rates for the alternative statistical models for Ribaldo are reported from Zones 10 to 50 in depths 0 to 1000 m , taken by trawl. The optimal model was Model 7 .

| Statistic | Value |
| :---: | :---: |
| 2011_CE | 0.6519 |
| 2012_CE | 0.5097 |
| Ln(CE_Ratio) | -0.2460 |
| Rprime | -0.0905 |
| TAC_Multiplier | 0.9095 |

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## 15. Standardization of Bight Redfish in the GAB 2000/2001 Feb 2011/2012. Catch Rate Update

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### 15.1 Summary

The change in catch rates between 2009/2010 and July-Feb 2010/2011 is less than 20\% (+7.71) (Figure 15.3; Table 15.7), therefore the control rule suggests no change should be made to the default TAC.

### 15.2 Methods

Data was provided from July 2000 to February 2012 for catches of bight redfish from the GAB (Table 15.1). Records were only included in the analysis that adhered to the following selection criteria:

Depths were between 50 - 500 metres (Table 15.2; Figure 15.1),
Non-zero catches of bight redfish,
Shot length $>1.0$ and $<10$ hours, Only from Zone 80 (GAB), The DayNight factor only used Day, Night, and Mixed (Unknown was omitted).

The analysis conducted included all vessels which had reported catches of bight redfish as well as adhering to the conditions listed above (Table 15.3).

Seven statistical models (Table 15.4) were examined using six different factors:

Fishing Year (July - June),
Vessel,
Depth Category ( 50 metre categories),
Month,
SubZone ( 5 degree of longitude subdivisions),
DayNight (Day, Night, Mixed - a small number of Unknown were omitted).
All statistical models were plotted after dividing each series by the average of each series. This means that the average of each series becomes one, and this ensures they are all on the same scale and hence directly comparable.

The percent difference of the catch rates between years is calculated as:

$$
\% \mathrm{D}=100 \times\left(\mathrm{CE}_{10 / 11}-\mathrm{CE}_{09 / 10}\right) / \mathrm{CE}_{09 / 10}
$$

### 15.3 Results

Catch rates exhibited a highly skewed distribution which was approximately lognormally distribution but a log transformation approximately normalizes the data prior to analysis; Figure 15.2). There are numerous records grouped around catch rates of 1, $2,5,10,15$, and $30 \mathrm{~kg} / \mathrm{hr}$, which appear as spikes in the observed $\log$ of catch rates; this seems likely to be due to rounding to nearest convenient weight of catch.

The optimum statistical model was the most complex having the most parameters (Table 15.5; Table 15.6; Figure 15.2). Catch rates for bight redfish from the GAB initially increased to a peak in 2003/2004 and then after which catch rates have remained relatively stable varying slightly up and down until 2009/2010 when they started to decline. However, in the latest year, 20111/2012, catch rates increased by 7.71\% (Figure 15.2; Table 15.7).

The standardization analysis with this year's data follows essentially the same trajectory as that produced by last year's analysis Figure 15.3 which indicates that data to the end of February are sufficient to describe each year's trends.

The GABTF Harvest Strategy decision rules, applied to both deepwater flathead and bight redfish are:

The FIS and the collection of age and length frequency data as well as the monitoring of catch and effort information will be ongoing regardless of whether an assessment is to take place in that year. The information obtained from these sources will be analysed and presented to the RAG each year well prior to the date at which a decision on the TAC for the next year is made.

- Any adjustment to the TAC limit through the application of the decision rules would apply to the default TAC
- When the Fishery Independent Survey (FIS) has been conducted in two consecutive years, the catch rates from the first leg of the survey will be the indicator of abundance used to make any adjustment to the default TAC.
- In a year when the Fishery Independent Survey (FIS) is not conducted, the standardised commercial catch rate for the period July-February inclusive is the indicator of abundance used to make any adjustment to the default TAC, comparing the current year to the immediately preceding year.
- If there is a change of $\geq 20 \%$ to the indicator of abundance, a $10 \%$ (increase or decrease) to the default TAC will occur.
- If the RAG is concerned with any indicators over the period between stock assessments (length frequency distributions, standardised commercial catch rates, age distributions etc), then it can decide to undertake a full assessment in that year


### 15.4 Conclusion

The change in catch rates between 2010/2011 and July-Feb 2011/2012 is less than 20\% (+7.71) Figure 15.3; Table 15.7), therefore the control rule suggests no change should be made to the default TAC.

### 15.5 Acknowledgements

John Garvey of AFMA is thanked for providing the original data extract. Dr Neil Klaer of CSIRO is especially thanked for pre-processing the catch and effort data so rapidly.

Table 15.1. The frequency of catch rate observations in each month and fishing year (financial year July/June) for Bight Redfish from the GAB following data selection.

| Mth | $00 / 01$ | $01 / 02$ | $02 / 03$ | $03 / 04$ | $04 / 05$ | $05 / 06$ | $06 / 07$ | $07 / 08$ | $08 / 09$ | $09 / 10$ | $10 / 11$ | $11 / 12$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 7 | 41 | 33 | 77 | 188 | 152 | 178 | 187 | 142 | 159 | 145 | 68 | 121 |
| 8 | 39 | 89 | 63 | 211 | 185 | 222 | 231 | 204 | 108 | 152 | 67 | 213 |
| 9 | 143 | 160 | 147 | 181 | 305 | 253 | 335 | 280 | 196 | 196 | 186 | 208 |
| 10 | 181 | 219 | 136 | 337 | 317 | 294 | 316 | 281 | 268 | 211 | 177 | 182 |
| 11 | 161 | 216 | 201 | 338 | 346 | 287 | 321 | 307 | 229 | 200 | 180 | 200 |
| 12 | 99 | 89 | 103 | 192 | 241 | 224 | 198 | 244 | 164 | 163 | 155 | 182 |
| 1 | 124 | 114 | 235 | 304 | 485 | 437 | 219 | 334 | 206 | 184 | 181 | 228 |
| 2 | 159 | 159 | 259 | 276 | 492 | 371 | 235 | 229 | 192 | 207 | 144 | 211 |
| 3 | 176 | 180 | 225 | 289 | 521 | 363 | 332 | 223 | 197 | 248 | 202 | 0 |
| 4 | 211 | 134 | 218 | 272 | 290 | 313 | 325 | 230 | 189 | 261 | 157 | 0 |
| 5 | 210 | 133 | 204 | 242 | 234 | 248 | 169 | 208 | 227 | 153 | 195 | 0 |
| 6 | 71 | 93 | 169 | 118 | 230 | 343 | 215 | 171 | 181 | 125 | 133 | 0 |
| Total | 1615 | 1619 | 2037 | 2948 | 3798 | 3533 | 3083 | 2853 | 2316 | 2245 | 1845 | 1545 |

Table 15.2. The relative frequency of depths records for Bight Redfish from the GAB see (Figure 15.1).

| Depth M | Count |
| ---: | ---: |
| 0 | 0 |
| 50 | 8 |
| 100 | 6232 |
| 150 | 21359 |
| 200 | 1638 |
| 250 | 181 |
| 300 | 10 |
| 350 | 0 |
| 400 | 2 |
| 450 | 4 |
| 500 | 3 |

Table 15.3. Summary statistics characterizing the data included in the standardization.

|  | Records | Catches | Effort | GeomCE | Vessels |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| $2000 / 2001$ | 1615 | 261.868 | 8422 | 13.448 | 5 |
| $2001 / 2002$ | 1619 | 200.566 | 8466 | 10.410 | 7 |
| $2002 / 2003$ | 2037 | 294.920 | 11034 | 12.630 | 9 |
| $2003 / 2004$ | 2948 | 541.632 | 16041 | 17.453 | 11 |
| $2004 / 2005$ | 3798 | 712.731 | 20591 | 18.617 | 10 |
| $2005 / 2006$ | 3533 | 586.826 | 18928 | 16.236 | 11 |
| $2006 / 2007$ | 3083 | 599.814 | 16194 | 19.014 | 9 |
| $2007 / 2008$ | 2853 | 532.261 | 14876 | 17.290 | 7 |
| $2008 / 2009$ | 2316 | 470.236 | 11975 | 19.573 | 4 |
| $2009 / 2010$ | 2245 | 396.187 | 11644 | 16.682 | 4 |
| $2010 / 2011$ | 1845 | 277.004 | 9463 | 14.269 | 4 |
| $2011 / 2012$ | 1545 | 214.974 | 8253 | 14.026 | 4 |

```
Table 15.4. The seven statistical models examined for Bight Redfish from the GAB.
Model 1 Fyear
Model 2 Fyear + Vessel
Model 3 Fyear + Vessel + DepCat
Model 4 Fyear + Vessel + DepCat + Month
Model 5 Fyear + Vessel + DepCat + Month + SubZone
Model 6 Fyear + Vessel + DepCat + Month + SubZone + DN
Model 7 Fyear + Vessel + DepCat + Month + SubZone + DN + DepCat:Month
```

Table 15.5. The standardized catch rates for the alternative statistical models for Bight Redfish from the GAB in depths 50 to 500 m . Values are relative to the mean of the standardized catch rates so that the average of the series remains 1.0 . Fishing Years were from July/June, DepCat were 50 m categories, Subzones were $5^{\circ}$ of Longitude, and DN relates to DayNight categories.

|  | FYear | DN | Month | Vessel | Subzone | DepCat | DepCat:Mth | StErr |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $00 / 01$ | 0.8502 | 0.8547 | 0.8355 | 0.8062 | 0.8983 | 0.8978 | $\mathbf{0 . 8 8 8 0}$ | 0.0000 |
| $01 / 02$ | 0.6589 | 0.6467 | 0.6481 | 0.6605 | 0.7390 | 0.7447 | $\mathbf{0 . 7 4 6 5}$ | 0.0404 |
| $02 / 03$ | 0.7992 | 0.7943 | 0.7593 | 0.7689 | 0.8233 | 0.8220 | $\mathbf{0 . 8 2 1 5}$ | 0.0392 |
| $03 / 04$ | 1.1044 | 1.1231 | 1.1252 | 1.0975 | 1.1432 | 1.1466 | $\mathbf{1 . 1 4 3 8}$ | 0.0382 |
| $04 / 05$ | 1.1779 | 1.2040 | 1.1202 | 1.1830 | 1.1656 | 1.1671 | $\mathbf{1 . 1 6 4 1}$ | 0.0374 |
| $05 / 06$ | 1.0273 | 1.0154 | 1.0014 | 1.0680 | 1.0737 | 1.0678 | $\mathbf{1 . 0 6 4 7}$ | 0.0376 |
| $06 / 07$ | 1.2031 | 1.2364 | 1.2582 | 1.1930 | 1.0865 | 1.0820 | $\mathbf{1 . 0 8 4 1}$ | 0.0380 |
| $07 / 08$ | 1.0940 | 1.1294 | 1.1598 | 1.1432 | 1.0914 | 1.0980 | $\mathbf{1 . 1 2 8 3}$ | 0.0394 |
| $08 / 09$ | 1.2386 | 1.2784 | 1.3393 | 1.2119 | 1.1782 | 1.1894 | $\mathbf{1 . 1 8 3 1}$ | 0.0404 |
| $09 / 10$ | 1.0557 | 1.0202 | 1.0045 | 1.0362 | 1.0151 | 1.0223 | $\mathbf{1 . 0 1 8 6}$ | 0.0408 |
| $10 / 11$ | 0.9030 | 0.8510 | 0.8527 | 0.8918 | 0.8725 | 0.8592 | $\mathbf{0 . 8 4 6 1}$ | 0.0433 |
| $11 / 12$ | 0.8877 | 0.8464 | 0.8958 | 0.9398 | 0.9133 | 0.9031 | $\mathbf{0 . 9 1 1 3}$ | 0.0458 |

Table 15.6. Model selection criteria, including the AIC, the adjusted $\mathrm{r}^{2}$, and the proportional change in adj $R^{2}$. Optimal model was model 7: FYear + Vessel + DepCat + Month + SubZone + DayNight + DepCat:Month. The Daynight factor is clearly the most influential with Bight Redfish.

|  | Year | DN | Month | Vessel | Subzone | DepCat | DepCat:Month |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 17921 | 13286 | 10246 | 8800 | 7875 | 7802 | $\mathbf{7 6 2 9}$ |
| RSS | 54066 | 46185 | 41622 | 39623 | 38357 | 38240 | $\mathbf{3 7 8 9 8}$ |
| MSS | 826 | 8708 | 13271 | 15269 | 16536 | 16653 | $\mathbf{1 6 9 9 4}$ |
| Nobs | 29437 | 29437 | 29437 | 29437 | 29437 | 29437 | $\mathbf{2 9 4 3 7}$ |
| Npars | 12 | 14 | 25 | 26 | 42 | 50 | $\mathbf{9 6}$ |
| Adj_r2 | 1.469 | 15.827 | 24.115 | 27.755 | 30.027 | 30.220 | $\mathbf{3 0 . 7 3 6}$ |
| \%Change |  | 14.358 | 8.288 | 3.641 | 2.271 | 0.194 | $\mathbf{0 . 5 1 5}$ |

Table 15.7. The optimum standardized catch rate model relative to the unstandardized geometric mean catch rates (Fyear) with the percent difference between years for each. The value of interest is at the bottom right showing the difference between $10 / 11$ and $11 / 12$.

|  | Fyear | Diff | Optimum | Diff |
| :--- | ---: | ---: | ---: | ---: |
| $00 / 01$ | 0.8502 |  | 0.8880 |  |
| $01 / 02$ | 0.6589 | -22.51 | 0.7465 | -15.93 |
| $02 / 03$ | 0.7992 | 21.31 | 0.8215 | 10.05 |
| $03 / 04$ | 1.1044 | 38.18 | 1.1438 | 39.23 |
| $04 / 05$ | 1.1779 | 6.66 | 1.1641 | 1.77 |
| $05 / 06$ | 1.0273 | -12.79 | 1.0647 | -8.54 |
| $06 / 07$ | 1.2031 | 17.11 | 1.0841 | 1.82 |
| $07 / 08$ | 1.0940 | -9.06 | 1.1283 | 4.08 |
| $08 / 09$ | 1.2386 | 13.21 | 1.1831 | 4.86 |
| $09 / 10$ | 1.0557 | -14.77 | 1.0186 | -13.91 |
| $10 / 11$ | 0.9030 | -14.46 | 0.8461 | -16.94 |
| $11 / 12$ | 0.8877 | -1.69 | 0.9113 | 7.71 |



Figure 15.1. The relative frequency of depth records from Bight Redfish from the GAB. The lower graph is a repeat of the upper graph except with more detail. Data is from 2000/2001 - Feb 2011/2012.


Figure 15.2. The catch rates for Bight Redfish are normalized by a natural log transformation. Data is from 2000/2001 - Feb 2011/2012. The spikes in the distribution, which distort the distribution away from a strict log-normal, relate to catch rates of $1,2,5,10,15$, and $30 \mathrm{~kg} / \mathrm{hr}$. There are a very few very large catch rates, but they are so few they do not influence the standardized catch rate trend.


Figure 15.3. The standardized catch rates for Bight Redfish from the GAB. The dashed line is the unstandardized geometric mean catch rates see Table 15.7. The lower graph depicts the percent difference between consecutive fishing years (see Table 15.7).


Figure 15.4. Comparison of this year's analysis (black line) with last year's (red line - scaling this year's analysis to the mean of $00 / 01-10 / 11$ to make it comparable with last year's analysis).

## 16. Standardization of Deepwater Flathead in the GAB 2000/2001 - Feb 2011/2012. Catch rate Update.

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### 16.1 Summary

As the change in catch rates between 2010/2011 and July-Feb 2011/2012 is greater than $-20 \%$ (Figure 16.3; Table 16.7) the control rule suggests a $10 \%$ decrease should be made to the default TAC.

### 16.2 Methods

Data was provided from July 2000 to February 2012 for catches of deepwater flathead from the GAB (Table 16.1). Records were only included in the analysis that adhered to the following selection criteria (Table 16.3):

Depths were between 50 - 500 metres (Table 16.2; Figure 16.1),
Non-zero catches of deepwater flathead,
Shot length $>0.5$ and $<10$ hours,
Only from Zone 80 (GAB),
The DayNight factor only used Day, Night, and Mixed (Unknown was omitted).
Only Vessels in the fishery for more than 2 years were included.
Seven statistical models (Table 16.4) were examined using six different factors:

Fishing Year (July - June),
Vessel,
Depth Category ( 50 metre categories),
Month,
SubZone ( 5 degree of longitude subdivisions),
DayNight (Day, Night, Mixed - a small number of Unknown were omitted).
Catch rates are log-normally distributed but a log transformation successfully normalizes the data prior to analysis (Figure 16.2).

The percent difference between years is calculated as:

$$
\% \mathrm{D}=100 \times\left(\mathrm{CE}_{11 / 12}-\mathrm{CE}_{10 / 11}\right) / \mathrm{CE}_{10 / 11}
$$

### 16.3 Results

The optimum statistical model was the most complex having the most parameters (Table 16.6; Figure 16.3).

Catch rates for Deepwater Flathead from the GAB initially increased to a peak in 2002/2003 and 2003/2004 and then declined to half the maximum levels in 2005/2006 after which catch rates have exhibited a slow increase although almost all the gains over the last four years appear to have been lost in this most recent year. In the latest year catch rates decreased by $25.59 \%$ (Figure 16.3; Table 16.7).

The standardization analysis with this year's data follows essentially the same trajectory as that produced by last year's analysis (Figure 16.4) with only a very slight deviation in the $10 / 11$ points, which again indicates that data to the end of February are sufficient to describe each year's trends.

The GABTF Harvest Strategy decision rules, applied to both deepwater flathead and bight redfish are:

The FIS and the collection of age and length frequency data as well as the monitoring of catch and effort information will be ongoing regardless of whether an assessment is to take place in that year. The information obtained from these sources will be analysed and presented to the RAG each year well prior to the date at which a decision on the TAC for the next year is made.

- Any adjustment to the TAC limit through the application of the decision rules would apply to the default TAC
- When the Fishery Independent Survey (FIS) has been conducted in two consecutive years, the catch rates from the first leg of the survey will be the indicator of abundance used to make any adjustment to the default TAC.
- In a year when the Fishery Independent Survey (FIS) is not conducted, the standardised commercial catch rate for the period July-February inclusive is the indicator of abundance used to make any adjustment to the default TAC, comparing the current year to the immediately preceding year.
- If there is a change of $\geq 20 \%$ to the indicator of abundance, a $10 \%$ (increase or decrease) to the default TAC will occur.
- If the RAG is concerned with any indicators over the period between stock assessments (length frequency distributions, standardised commercial catch rates, age distributions etc), then it can decide to undertake a full assessment in that year


### 16.4 Conclusion

As the change in catch rates between 2010/2011 and July-Feb 2011/2012 is greater than $-20 \%$ (Figure 16.3; Table 16.7) the control rule suggests a $10 \%$ decrease should be made to the default TAC.

### 16.5 Acknowledgements

John Garvey of AFMA is thanked for providing the original data extract. Dr Neil Klaer of CSIRO is especially thanked for pre-processing the catch and effort data so rapidly.

Table 16.1. The frequency of catch rate observations in each month and fishing year (financial year July/June) for deepwater flathead from the GAB following data selection.

| Mth | $00 / 01$ | $01 / 02$ | $02 / 03$ | $03 / 04$ | $04 / 05$ | $05 / 06$ | $06 / 07$ | $07 / 08$ | $08 / 09$ | $09 / 10$ | $10 / 11$ | $11 / 12$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 215 | 184 | 320 | 430 | 648 | 624 | 304 | 455 | 287 | 278 | 236 | 329 |
| 2 | 189 | 246 | 380 | 412 | 595 | 550 | 317 | 294 | 263 | 304 | 177 | 278 |
| 3 | 255 | 259 | 354 | 491 | 627 | 480 | 401 | 301 | 280 | 315 | 267 | 55 |
| 4 | 270 | 184 | 353 | 492 | 414 | 463 | 432 | 368 | 285 | 332 | 310 | 0 |
| 5 | 308 | 195 | 323 | 394 | 467 | 436 | 311 | 345 | 312 | 221 | 287 | 0 |
| 6 | 121 | 147 | 250 | 242 | 418 | 497 | 317 | 222 | 234 | 207 | 221 | 0 |
| 7 | 60 | 58 | 111 | 237 | 287 | 282 | 273 | 187 | 186 | 205 | 131 | 189 |
| 8 | 69 | 139 | 106 | 288 | 303 | 337 | 318 | 274 | 129 | 198 | 106 | 316 |
| 9 | 206 | 211 | 208 | 295 | 425 | 380 | 421 | 339 | 226 | 248 | 293 | 288 |
| 10 | 250 | 306 | 241 | 450 | 473 | 402 | 403 | 376 | 337 | 298 | 269 | 291 |
| 11 | 268 | 331 | 316 | 470 | 488 | 472 | 419 | 447 | 315 | 320 | 275 | 327 |
| 12 | 187 | 167 | 154 | 276 | 367 | 358 | 275 | 326 | 224 | 249 | 251 | 281 |
| Total | 2398 | 2427 | 3116 | 4477 | 5512 | 5281 | 4191 | 3934 | 3078 | 3175 | 2823 | 2354 |

Table 16.2. The relative frequency of depths records for Deepwater Flathead from the GAB see Figure 15.1). Data from 2000/2001 to Feb 2011/2012

| Depth M | Count |
| ---: | ---: |
| 0 | 531 |
| 50 | 19 |
| 100 | 8705 |
| 150 | 30314 |
| 200 | 2916 |
| 250 | 884 |
| 300 | 353 |
| 350 | 37 |
| 400 | 39 |
| 450 | 10 |
| 500 | 6 |
| 550 | 3 |
| 600 | 9 |
| 650 | 1 |
| 750 | 1 |
| 800 | 3 |
| 850 | 2 |
| 900 | 1 |
| 950 | 3 |
| 1000 | 1 |
| 1100 | 1 |
| 1250 | 1 |
| 1350 | 1 |
| 1500 | 1 |

Table 16.3. Summary statistics characterizing the data included in the standardization

|  | Records | Catches | Effort | GeomCE | Vessels |
| ---: | ---: | ---: | ---: | ---: | ---: |
| $2000 / 2001$ | 2398 | 771.247 | 12248 | 44.134 | 6 |
| $2001 / 2002$ | 2427 | 906.838 | 12486 | 52.982 | 6 |
| $2002 / 2003$ | 3116 | 1613.140 | 16763 | 73.451 | 9 |
| $2003 / 2004$ | 4477 | 2157.200 | 24119 | 68.429 | 11 |
| $2004 / 2005$ | 5512 | 2087.262 | 29731 | 55.122 | 10 |
| $2005 / 2006$ | 5281 | 1341.148 | 28037 | 37.579 | 11 |
| $2006 / 2007$ | 4191 | 952.492 | 21895 | 32.912 | 10 |
| $2007 / 2008$ | 3934 | 957.801 | 20243 | 36.186 | 7 |
| $2008 / 2009$ | 3078 | 775.565 | 15766 | 41.015 | 5 |
| $2009 / 2010$ | 3175 | 805.629 | 16333 | 38.679 | 4 |
| $2010 / 2011$ | 2823 | 932.788 | 14351 | 50.702 | 4 |
| $2011 / 2012$ | 2299 | 570.140 | 12197 | 40.247 | 4 |

Table 16.4. The seven statistical models examined for Deepwater Flathead from the GAB.
Model 1 Fyear
Model 2 Fyear + Vessel
Model 3 Fyear + Vessel + DepCat
Model 4 Fyear + Vessel + DepCat + Month
Model 5 Fyear + Vessel + DepCat + Month + SubZone
Model 6 Fyear + Vessel + DepCat + Month + SubZone + DN
Model 7 Fyear + Vessel + DepCat + Month + SubZone + DN + DepCat:Month

Table 16.5. The standardized catch rates for the alternative statistical models for Deepwater Flathead from the GAB in depths 50 to 500 m . Values are relative to the mean of the standardized catch rates so that the average of the series remains 1.0. Fishing Years were from July/June, DepCat were 50 m categories, Subzones were $5^{\circ}$ of Longitude, and DN relates to DayNight categories.

|  | FYear | Vessel | DepCat | Month | Subzone | DN | DepCat:Mth | StErr |
| :--- | :--- | :--- | ---: | :--- | ---: | :--- | ---: | :--- |
| $00 / 01$ | 0.9244 | 0.9253 | 0.9400 | 0.9267 | 0.9353 | 0.9350 | 0.9655 | 0.0000 |
| $01 / 02$ | 1.1100 | 1.1289 | 1.1348 | 1.1296 | 1.1031 | 1.1046 | 1.1306 | 0.0196 |
| $02 / 03$ | 1.5388 | 1.5810 | 1.6039 | 1.5818 | 1.5497 | 1.5486 | 1.5588 | 0.0190 |
| $03 / 04$ | 1.4336 | 1.5073 | 1.5307 | 1.5196 | 1.5325 | 1.5333 | 1.5226 | 0.0187 |
| $04 / 05$ | 1.1548 | 1.2029 | 1.2257 | 1.2340 | 1.2102 | 1.2117 | 1.1756 | 0.0183 |
| $05 / 06$ | 0.7873 | 0.7946 | 0.7900 | 0.7936 | 0.7770 | 0.7779 | 0.7677 | 0.0184 |
| $06 / 07$ | 0.6895 | 0.6571 | 0.6578 | 0.6687 | 0.6968 | 0.6967 | 0.6867 | 0.0190 |
| $07 / 08$ | 0.7867 | 0.7161 | 0.7273 | 0.7274 | 0.7519 | 0.7521 | 0.7506 | 0.0195 |
| $08 / 09$ | 0.8593 | 0.8293 | 0.8491 | 0.8519 | 0.9028 | 0.9029 | 0.9060 | 0.0202 |
| $09 / 10$ | 0.8103 | 0.7999 | 0.8102 | 0.8217 | 0.8050 | 0.8035 | 0.8109 | 0.0202 |
| $10 / 11$ | 1.0622 | 1.0303 | 0.9635 | 0.9703 | 0.9635 | 0.9652 | 0.9891 | 0.0211 |
| $00 / 01$ | 0.8432 | 0.8271 | 0.7670 | 0.7748 | 0.7723 | 0.7684 | 0.7360 | 0.0225 |

Table 16.6. Model selection criteria, including the AIC, the adjusted $\mathrm{r}^{2}$, and the proportional change in adj R ${ }^{2}$. Optimal model was model 7: FYear + Vessel + DepCat + Month + SubZone + DayNight + DepCat:Month.

|  | FYear | Vessel | DepCat | Month | subzone | DN | DepCat:Month |
| :--- | :---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | -21129 | -25011 | -26709 | -28313 | -29770 | -30215 | -33478 |
| RSS | 25781 | 23512 | 22580 | 21732 | 20995 | 20774 | 19172 |
| MSS | 2659 | 4928 | 5860 | 6708 | 7445 | 7666 | 9268 |
| Nobs | 42439 | 42439 | 42439 | 42439 | 42439 | 42439 | 42439 |
| Npars | 12 | 26 | 35 | 46 | 49 | 51 | 123 |
| Adj_r2 | 9.327 | 17.279 | 20.541 | 23.507 | 26.095 | 26.869 | 32.395 |
| \%Change |  | 7.952 | 3.262 | 2.966 | 2.588 | 0.774 | 5.526 |

Table 16.7. The optimum standardized catch rate model relative to the unstandardized geometric mean catch rates (Fyear) with the percent difference between years for each. The value of interest is at the bottom right showing the difference between 10/11 and 11/12.

|  | Fyear | Diff | Optimum | Diff |
| :--- | ---: | ---: | ---: | ---: |
| $00 / 01$ | 0.9244 |  | 0.9655 |  |
| $01 / 02$ | 1.1100 | 20.08 | 1.1306 | 17.09 |
| $02 / 03$ | 1.5388 | 38.63 | 1.5588 | 37.88 |
| $03 / 04$ | 1.4336 | -6.84 | 1.5226 | -2.32 |
| $04 / 05$ | 1.1548 | -19.45 | 1.1756 | -22.79 |
| $05 / 06$ | 0.7873 | -31.83 | 0.7677 | -34.70 |
| $06 / 07$ | 0.6895 | -12.42 | 0.6867 | -10.55 |
| $07 / 08$ | 0.7867 | 14.10 | 0.7506 | 9.31 |
| $08 / 09$ | 0.8593 | 9.22 | 0.9060 | 20.70 |
| $09 / 10$ | 0.8103 | -5.70 | 0.8109 | -10.49 |
| $10 / 11$ | 1.0622 | 31.09 | 0.9891 | 21.96 |
| $11 / 12$ | 0.8432 | -20.62 | 0.7360 | $-\mathbf{2 5 . 5 9}$ |



Figure 16.1. The relative frequency of depth records from Deepwater Flathead from the GAB. The lower graph is a repeat of the upper graph except with more detail. Data is from 2000/2001 - Feb 2011/2012.



Figure 16.2. The catch rates for Deepwater Flathead are normalized by a natural $\log$ transformation. Data is from 2000/2001 - Feb 2011/2012.


Figure 16.3. The standardized catch rates for Deepwater Flathead from the GAB. The dashed line is the unstandardized geometric mean catch rates see Table 15.7. The lower graph depicts the percent difference between consecutive fishing years (see Table 15.7).


Figure 16.4. Comparison of this year's analysis (black line) with last year's (red line - scaling this year's analysis to the mean of 00/01-10/11 to make it comparable with last year's analysis).

# 17. Standardized Catch Rates for the SESSF Gummy Shark Fishery: Data from 1976-2011 

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### 17.1 Summary

Reported catches of gummy sharks has declined from a high in 2008, although interpreting this is made more complex because of the 16 month TAC put in place for the 2007/2008 season. Nevertheless, the recent decline is real and is related to parallel declines in catches from South Australia and Bass Strait. Catches from South Australia decreased further in 2011 but recovered slightly in Bass Strait. These changes appear related to the introduction of gillnet fishery closures to protect Australian Sea Lions and dolphins in South Australian waters. At the same time the proportion of catches taken by gillnets declined over the period 2001-2011.

Standardized catch rates in South Australia have also exhibited a decline since 2008, however, the general trend since 1984 remains flat but noisy. The most recent mean estimate is slightly below the long term average, which again is thought to be related to the influence of the marine closures in South Australia.

In Bass Strait, standardized catch rates have also declined since 2008 but they are now still above or at the long term average depending on how the standardization for positive shots is combined with the standardization of the probability of obtaining a positive shot. Catches in the gummy shark fishery continue to be greatest in Bass Strait.

Standardized catch rates in Tasmania also remain noisy but flat. There is some indication of a very slow decline since about 2000 but given the variation surrounding the mean estimates the apparent decline is not yet statistically significant.

### 17.2 Introduction

The shark fishery off southern Australia has a long history starting with a long-line fishery which began in the 1920s which switched to gillnets in the 1960s and 1970s when the primary target also switched to gummy sharks (Mustelus antarcticus; Punt et al., 2000; Punt \& Gason, 2006; Thomson \& Punt, 2010). This gillnet fishery now mainly targets gummy sharks although used to target relatively large quantities of School sharks (Galeorhinus galeus) but this is now a bycatch only species. In this shark fishery there are significant amounts of the common saw shark (Pristiophorus cirratus) and southern saw shark ( $P$. nudipinnis; not distinguished from each other in the catch effort records) as well as elephant fish (Callorhinchus milii) taken as bycatch.

In 1990 - 1995 some major management changes were introduced. These included the amalgamation of endorsements and a reduction in the net unit from 6000 m to $4,200 \mathrm{~m}$ (by 1993 in Bass Strait and by 1995 in South Australia and Tasmania). With respect to gummy sharks the next big change came in 2001 when Individual Transferable Quotas (based on catch histories from 1994 - 1997) were introduced for both gummy and
school sharks. The structural adjustment package across 2006/2007 led to 26 gillnet vessel SFRs and 17 shark hook vessel SFRs leaving the fishery.

Previous attempts to standardize commercial catch rates for sharks in Australia began with Punt et al. (2000) who used the Delta method, which analyses any trend in the probability of obtaining a positive shot and separately any trends in catch rates in the positive shots and then combining these two trends to obtain a single standardized catch rate for the fishery. Punt et al. (2000) focused on school sharks but their method was revised and extended when it was later applied to Gummy sharks (Punt \& Gason, 2006; Thomson \& Punt, 2010).

As Kimura (1981, p211) says: "Since the 1950s it has been recognized that fishing power generally differs among vessels, and if c.p.u.e. is to be proportional to abundance, effort measurements must be standardized." The most commonly used method of standardization is to include the various factors thought to effect catch rates into a generalized linear model and to include Year as a factor, in this way the parameters derived for each year become the indices of relative abundance (Venables \& Dichmont, 2004).

After standardization we are left with a set of yearly coefficients that represent the catch rate relative to some reference year (usually with reference to the mean of the time series, which simplifies visual comparisons with other times series). Unfortunately, even if the standardization accounts for a large proportion of the variability in the data there are no guarantees that catch effort, even standardized catch effort, can act as a good proxy for stock size. Instead of the statistical success of the standardization, one should be able to argue from the nature of the fishery and the species concerned whether or not there is likely to be even an approximate relationship between catch rates and the exploitable biomass.

In this present work we focus on the catch rates for gummy shark, treating South Australia, Bass Strait, and Tasmania separately, because this reflects the assumed stock structure.

### 17.3 Methods

### 17.3.1 Catch Rate Standardization

The original data was provided in a text file named CANDE10.txt. This contained 421,977 records each with 23 fields (described in Haddon, 2012). The data provided received some pre-treatment in order to add the catch rate variables of interest and identify those records for inclusion in the analyses. Catch rates were calculated where there were positive catches of gummy sharks associated with positive effort levels. Where catch rates could be calculated they were also log transformed in preparation for the log-linear modelling of positive catches. Depth information, where present, was subdivided into 10 metre depth categories for inclusion in the standardization. Finally, a field was added that identified which records contained positive catches of gummy sharks. This latter was necessary as a separate analysis is conducted to characterize the occurrence of zero shots (the complement of positive shots) and whether their incidence has altered through time (see below).

In previous standardizations (Punt \& Gason, 2006; Thomson \& Punt, 2010) a wide array of criteria were used to select records for analysis. An important aspect of any standardization where the trend in the probability of zero shots is included is how to identify zero shots, which relate to targeted effort that fails to catch the species of interest. In the SESSF trawl fishery identifying those shots that might have captured a species but didn't is extremely difficult because targeting is so difficult to establish. Fortunately, in the shark gillnet fishery this is less of a problem because gillnet shark fishers are targeting sharks, especially gummy sharks. The problem thus becomes one of focusing attention on those vessels and areas where the fishery is a main focus of effort. The primary data selection criteria are to select the years where the fishery was operating normally (as defined by the SharkRAG), to use records only where gillnets with mesh sizes of $6 ", 6.5 "$ and $7 "$ were used, to select only those vessels catching a defined minimum total catch per year and a defined number of years in the fishery, to include only those areas which were the main focus of the fishery, and to exclude those records with effort less than 1000 m . In addition, records used were limited to particular gears and finally, those vessels that only caught small amounts of gummy shark across the years of the study were also excluded. The sensitivity of the analyses to the specific values selected as being a minimum reported catch for each area and vessel was tested by comparing an array of different combinations. In addition, the minimum number of years for a vessel to be active in the fishery was also considered.

### 17.3.2 The Delta Distribution

Including zero shots has two parts: 1) First, determine the relative probability of obtaining a positive catch. 2) Secondly, conduct a log-linear standardization on those records containing positive catches. These two analyses are then combined to provide the overall estimate of the yearly changes in catch rates required for inclusion in stock assessments.

### 17.3.3 Zero Catches

To estimate the probability of a positive observation (i.e. the species of interest is present in a shot) a binomial GLM (using a logit link function) is used to determine the effect of an array of factors on the probability $p_{i}$, which is the probability that the species of interest is present in the $i^{\text {th }}$ shot:

$$
\begin{equation*}
\ln \left(\frac{p_{i}}{1-p_{i}}\right)=\beta_{0}+\beta_{1} x_{i, 1}+\beta_{2} x_{i, 2}+\sum_{j=3}^{N} \beta_{j} x_{i j} \tag{1}
\end{equation*}
$$

where $p_{i}$ is the probability that the species of interest was present in the $i^{\text {th }}$ shot, and $x_{i j}$ are the values of the explanatory variables, $j$, for the $i^{\text {th }}$ shot and the $\beta_{j}$ are the coefficients for the $N$ factors $j$, to be estimated ( $\beta_{0}$ is the intercept, $\beta_{1}$ the coefficient for the first factor, etc.).

The catch rate standardizations all used individual records from the database, which in a number of cases appeared to be aggregated data, potentially aggregated within months, although there were also many individual shots recorded. This is apparent because the reported effort as net length is sometimes in the 100's of thousands of metres for a single record. The catch rate data for positive catches were normalized by using a natural-log transformation. General Linear Models were used with this transformed data rather than using Generalized Linear Models on the untransformed data with a log-link;
the approach used has advantages in terms of normalizing the data while stabilizing the variance, which the Generalized Linear Model approach does not always achieve appropriately (Venables \& Dichmont, 2004).

Up to eight different log-linear models were fitted and compared in an effort to account for the effects of year, area, month of fishing, vessel, which depth category was used, which gear was used, and any interactions between area and month, and area and gear (see Haddon, 2011). All variables were treated as categorical variables (alternatively termed factors). The optimum statistical model was selected on the basis of the Akaike's Information Criterion (Burnham \& Anderson, 1998), and the adjusted $\mathrm{r}^{2}$ (Neter et al.,1996). The resulting optimal model was plotted in comparison with the geometric mean catch rate, both being scaled to the mean of each series for ease of visual comparison. The standardized catch rates for the year factor can be used in assessment models as the index of relative abundance through time.

Standard analyses were conducted in each case and all were coded in the statistical software R (R development Core Team, 2009). In each case, catch rates, as kilograms per metre of gillnet fished, were natural log-transformed to normalize the data and stabilize the variance. The General Linear Models all had the same form:

$$
\begin{equation*}
\operatorname{Ln}\left(C P U E_{i}\right)=\alpha_{0}+\alpha_{1} x_{i, 1}+\alpha_{2} x_{i, 2}+\sum_{j=3}^{N} \alpha_{j} x_{i j}+\varepsilon_{i j} \tag{2}
\end{equation*}
$$

where $\operatorname{Ln}\left(C P U E_{i}\right)$ is the natural logarithm of the catch rate $(\mathrm{kg} / \mathrm{m})$ for the $i$-th record, $x_{i j}$ are the values of the explanatory variables $j$ for the $i$-th shot (i.e. Year, Disting, Month, etc), and the $\alpha_{j}$ are the coefficients for the $N$ factors $j$ to be estimated ( $\alpha_{0}$ is the intercept, $\alpha_{1}$ is the coefficient for the first factor, etc.), and $\varepsilon_{i j}$ are the normal random residual errors.

### 17.3.4 The Year Effect

The standardised overall year effect for the fishery is calculated as the product of the Year coefficients from the binomial and log-linear GLMs (Eqs (1) and (2)) transformed back onto their original scales. For back-transformation all other predictor variables were set to zero, indicating the reference level of each categorical factor. The expected probability (back-transformed from logit) of a non-zero catch in year $t$ is therefore

$$
\begin{equation*}
\hat{p}_{t}=\frac{\exp \left(\beta_{0}+\lambda_{t}\right)}{1+\exp \left(\beta_{0}+\lambda_{t}\right)} \tag{3}
\end{equation*}
$$

where $p_{t}$ is the probability of a non-zero catch in year $t, \beta_{0}$ is the intercept and the $\lambda_{t}$ is the Year coefficient for year $t$. As a test of the procedure the back transformation of the simple $\mathrm{PA}=$ Year model should deliver the annual proportion of positive shots.

For the log-normal model the expected back-transformed year effect involves a biascorrection for log-normality; the back transformation without the correction estimates the median of the distribution rather than the mean, adding $\sigma^{2} / 2$ before backtransformation improves the approximation to the mean of the distribution:

$$
\begin{equation*}
C P U E_{t}=e^{\left(\gamma_{t}+\sigma_{t}^{2} / 2\right)} \tag{4}
\end{equation*}
$$

where $\gamma_{\mathrm{t}}$ is the Year coefficient for year $t$ and $\sigma_{t}$ is its standard error.
Total standardised catch rates for year $t$ are calculated as the product of Eqs (3) and (4), stated relative to the average of all values:

$$
\begin{equation*}
\bar{Y}=\frac{\sum_{t=1}^{n} p_{t} C P U E_{t}}{n} \tag{5}
\end{equation*}
$$

where $n$ is the number of years of data. So the standardized catch rates are given relative to the mean of the series. This implies that the average of the time series of standardized catch rates will always be one, and hence each series is directly comparable with all the others:

$$
\begin{equation*}
Y_{t}=\frac{p_{t} C P U E_{t}}{\bar{Y}} \tag{6}
\end{equation*}
$$

The factors considered in the analyses were all taken as categorical variables and were:
Year the standard calendar year,
Disting each vessel is uniquely and confidentially identified,
Month standard calendar months,
Area Standard shark statistical reporting blocks (Figure 17.1).
Gear $6.0^{\prime \prime}, 6.5$ ", or $7.0^{\prime \prime}$ mesh nets.
DepCat $\quad 10 \mathrm{~m}$ categories (novel this year)
Area:Month An interaction term used to include any seasonal changes across areas.

### 17.3.5 Data Selection Gummy Sharks

Data selection occurred with the years of data used by zone, the gear used, the depths, used, and with areas only included if total catches exceeded a given limit, vessels only included if their average annual catches exceeded a given limit, and they were reporting catches for more than a given number of years in the fishery (Table 17.7).

There were also some records where no effort data were included (effort $=-1$ ) and these could not be included in the standardization. In addition, if the reported effort was $<$ 1000 m these records were also excluded.

Depth data was not provided from South Australia until after 1997 so depth cannot be included in the South Australia standardization.

There are a large number of vessels contributing to the final analysis, even with the restricted number of years and areas used. To remove noise generated by those vessels reporting very small amounts of gummy sharks those vessels reporting less than an average of 2 tonne per year (for the years in which they reported saw sharks) were removed from the analysis. In addition, if they reported for less than 3 years they were excluded.

### 17.3.6 Disjunction in the 1990s

Major changes appear to have occurred in the data from the fishery during the early 1990s. To illustrate this disjunction the catch per vessel per year (as identified by their distinguishing marks) can be tabulated. From this table it is possible to sum the catches per vessel from 1976-1993 and, separately, the catches by vessel from 1994 - 2010. These data can then be used to estimate the proportional representation of the catches by vessel across these two periods.

### 17.4 Results

### 17.4.1 The Shark Fishery



Figure 17.1. Map of shark statistical reporting areas along with the statistical regions. WA is Western Australia, WSA is Western South Australia, CSA is Central South Australia, ESA is Eastern South Australia (sometimes known as SAV - South Australia Victoria), WBS is Western Bass Strait, EBS is Eastern Bass Strait, NSW is New South Wales, ETS is Eastern Tasmania and WTS is Western Tasmania.

### 17.4.2 The Gummy Shark Fishery

Following the decline in the school shark fishery, the non-trawl shark fishery is now dominated by the gummy shark fishery (Figure 17.2, Figure 17.3,Table 17.4).


Figure 17.2. The total annual catch and number of records for the three main regions in the Gummy shark fishery for all gears. The thick lines represent the range of years chosen by the SharkRAG to represent the fishery, while the fine lines represent the available data in the log book data base. The grey vertical lines relate to 1995 and 2009.


Figure 17.3. Total reported catches of gummy sharks, 1976-2011 from the log-books. The grey lines relate to the individual regions. These data relate to all gillnet catches by all mesh sizes. The vertical grey lines relate to 1995 and 2009

There is a clear disjunction between the available data prior to 1995 and that after, and this is especially apparent in South Australia and Tasmania (Figure 17.2, Figure 17.3 and Figure 17.4). This also becomes apparent in the standardized catch rates. Total catches have been relatively stable since 1995 although have been declining since 2009, primarily in South Australia, where the Australian Sea Lion closures began to impact the gummy shark fishery.


Figure 17.4 The relationship between the number of records and the resulting catch each year in each of the three regions. This data relate only to catches taken with 6 ", 6.5 " and 7 " nets are include data from all years 1970 to 2011.

Reported catches of gummy sharks has declined from a high in 2008, although interpreting this is made more complex because of the 16 month TAC put in place for the 2007/2008 season (Table 17.8; Figure 17.13). Nevertheless, the recent decline is real and is related to parallel declines of catches in South Australia and Bass Strait. Catches from South Australia decreased further in 2011 but recovered slightly in Bass Strait. These changes appear to be related to the introduction of closures to protect Australian Sea Lions and dolphins in South Australian waters.

At the same time the proportion of catches taken by gillnets declined over the period 2001-2011 (Table 17.8).

### 17.5 South Australia

The standardization of the South Australian gummy shark catch-rates reduces the variation exhibited by the trend through time, with the geometric mean catch rates having a CV of $22.8 \%$ while the optimum model has a CV of $16.8 \%$ (Figure 17.5; Table 17.1). Nevertheless, each mean estimate is relatively uncertain, as indicated by the $95 \%$ confidence intervals on the graphs) and only a few years could be considered statistically significant. It should be noted that the width of these confidence intervals are likely to be under-estimates owing to the various influential factors that have not been able to be included in the analysis.


Figure 17.5. Standardized catch rates for South Australia gummy sharks using data relating to $6.0^{\prime \prime}, 6.5$ ", and 7.0 " mesh gear, from areas that reported more than 10 tonnes across the 28 years considered (19842011), and from vessels with average catches greater than 2 tonnes per annum which had been present in the fishery for at least 3 years. The top panel represents the probability of obtaining a positive catch, the dashed line being the proportion of positive catches in the raw data and the solid line being the statistical optimum model (with the $95 \%$ error bars surrounding the trend). The central panel represents the loglinear modelling of positive catches. The dashed line is the geometric mean while the solid line is the optimal model, and again the bars are the $95 \%$ confidence limits on the mean estimates. The bottom panel represents the final standardized catch rates, combining the results from the log-linear modelling and the binomial modelling of the probability of a positive catch. All trends have been scaled to the mean of each series to ease visual comparison.

When the analysis of positive catches is combined with the analysis of the relative incidence of positive shots then there does not appear to have been the overall trend through time. Perhaps catch rates were lower pre-1995 and generally higher after 1995. The decline from the high in 2008 is associated with a reduction in the catch landed in South Australia and with a reduction of greater than $40 \%$ in the number of records in

2011 brought about by the Australian sea lion closures. These closures appear likely to lead to a decline in observed catch rates.

The overall conclusion is that the catch rates for gummy shark in South Australia remain flat but noisy about the long term average.


Figure 17.6. South Australian Gummy shark: The relative influence of each factor used on the final trend in the optimal standardization. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

Table 17.1. The different standardization models fitted to the South Australian gummy shark data. The models are cumulative across the table with the optimum being the Area:Month model.

| Year | GeoMean | Vessel | Area | Month | Gear | Area:Month | Area:Gear |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1984 | 1.4462 | 1.2708 | 1.3151 | 1.2845 | 1.1569 | 1.1954 | 1.1687 |
| 1985 | 1.3532 | 1.3147 | 1.2931 | 1.2586 | 1.0986 | 1.1439 | 1.1168 |
| 1986 | 1.0231 | 1.1240 | 1.1092 | 1.1093 | 0.9751 | 1.0258 | 0.9794 |
| 1987 | 0.7198 | 0.7100 | 0.7179 | 0.7164 | 0.6328 | 0.6471 | 0.6242 |
| 1988 | 0.9903 | 1.0242 | 1.0434 | 1.0316 | 0.9262 | 0.9465 | 0.9081 |
| 1989 | 0.9935 | 1.0399 | 1.0136 | 1.0058 | 0.9069 | 0.9332 | 0.9465 |
| 1990 | 1.0002 | 1.1116 | 1.0201 | 1.0335 | 0.9832 | 0.9749 | 1.0305 |
| 1991 | 0.9340 | 1.0252 | 0.9866 | 0.9967 | 0.9551 | 0.9613 | 0.9997 |
| 1992 | 0.8779 | 0.9482 | 0.9450 | 0.9580 | 0.9128 | 0.9117 | 0.9372 |
| 1993 | 0.8826 | 1.0155 | 1.0942 | 1.1027 | 0.9711 | 0.9543 | 0.9478 |
| 1994 | 1.0748 | 1.1738 | 1.1809 | 1.2001 | 1.0888 | 1.1133 | 1.0933 |
| 1995 | 1.2560 | 1.0781 | 0.9922 | 0.9678 | 0.9414 | 0.9582 | 0.9528 |
| 1996 | 1.6133 | 1.4242 | 1.3536 | 1.3343 | 1.3151 | 1.3253 | 1.2860 |
| 1997 | 0.8717 | 0.8339 | 0.8751 | 0.8433 | 0.9042 | 0.9082 | 0.8990 |
| 1998 | 0.5469 | 0.5709 | 0.5832 | 0.5827 | 0.6275 | 0.6297 | 0.6101 |
| 1999 | 0.6535 | 0.6953 | 0.6877 | 0.6943 | 0.7485 | 0.7554 | 0.7262 |
| 2000 | 0.8939 | 0.8998 | 0.8969 | 0.9036 | 0.9737 | 0.9704 | 0.9625 |
| 2001 | 0.9554 | 0.9147 | 0.9102 | 0.9082 | 0.9822 | 0.9588 | 0.9716 |
| 2002 | 1.0623 | 1.0151 | 0.9914 | 0.9942 | 1.0770 | 1.0589 | 1.0626 |
| 2003 | 1.0556 | 1.0176 | 1.0604 | 1.0637 | 1.1516 | 1.1316 | 1.1392 |
| 2004 | 1.0550 | 1.0352 | 1.0850 | 1.0923 | 1.1821 | 1.1633 | 1.1726 |
| 2005 | 0.9800 | 1.0080 | 1.0272 | 1.0402 | 1.1271 | 1.1006 | 1.1224 |
| 2006 | 0.9856 | 1.0010 | 1.0112 | 1.0211 | 1.1035 | 1.0818 | 1.0932 |
| 2007 | 1.0328 | 1.0234 | 1.0427 | 1.0501 | 1.1347 | 1.1103 | 1.1248 |
| 2008 | 1.2250 | 1.1964 | 1.2200 | 1.2342 | 1.3351 | 1.2936 | 1.3302 |
| 2009 | 0.9737 | 0.9785 | 0.9850 | 0.9943 | 1.0768 | 1.0774 | 1.0732 |
| 2010 | 0.8082 | 0.8177 | 0.8454 | 0.8543 | 0.9259 | 0.8990 | 0.9251 |
| 2011 | 0.7354 | 0.7322 | 0.7137 | 0.7242 | 0.7863 | 0.7702 | 0.7967 |
| CV | 22.84 | 19.00 | 18.49 | 18.12 | 17.15 | 16.81 | 17.09 |

Table 17.2. The statistical diagnostics for the South Australian gummy shark standardization. The smallest AIC and largest adjusted $r^{2}$ indicates the optimum statistical model.

|  | GeoMean | Vessel | Area | Month | Gear | Area:Month | Area:Gear |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 28566 | 23815 | 20658 | 19942 | 19883 | 17901 | 19474 |
| RSS | 117772 | 111032 | 106857 | 105920 | 105840 | 102609 | 105184 |
| MSS | 4160 | 10899 | 15075 | 16012 | 16092 | 19322 | 16748 |
| Nobs | 83813 | 83813 | 83813 | 83813 | 83813 | 83813 | 83813 |
| Npars | 28 | 122 | 150 | 161 | 163 | 471 | 219 |
| adj_r2 | 3.380 | 8.807 | 12.207 | 12.966 | 13.029 | 15.372 | 13.510 |
| \%Change | 0.000 | 5.427 | 3.400 | 0.758 | 0.064 | 2.343 | 0.481 |

### 17.6 Bass Strait

The transition in the character of the gillnet commercial catch and effort data before and after 1995 is clearly apparent in the catch rate standardization, although in the case of Bass Strait this is only apparent in the standardization of the probability of obtaining a positive shot. Zero shots for gummy sharks became far less likely following 1995 (Figure 17.7) which corresponds to changes in allowable net length and other related management changes. This transition is very apparent in the plot of the influence of each factor on the trend in the standardized catch rates (Figure 17.8).


Figure 17.7. Standardized catch rates for Bass Strait gummy sharks using data relating to $6.0^{\prime \prime}, 6.5^{\prime \prime}$, and $7.0^{\prime \prime}$ mesh gear, from areas that reported more than 10 tonnes across the 36 years considered ( $1976-$ 2011), and from vessels with average catches greater than 2 tonnes per annum which had been present in the fishery for at least 3 years. The top panel represents the probability of obtaining a positive catch, the dashed line being the proportion of positive catches in the raw data and the solid line being the statistical optimum model (with the $95 \%$ error bars surrounding the trend). The central panel represents the loglinear modelling of positive catches. The dashed line is the geometric mean while the solid line is the optimal model, and again the bars are the $95 \%$ confidence limits on the mean estimates. The bottom panel represents the final standardized catch rates, combining the results from the log-linear modelling and the binomial modelling of the probability of a positive catch. All trends have been scaled to the mean of each series to ease visual comparison.
As with the positive catches in South Australia, the gummy shark catch rates in Bass Strait are noisy and flat relative to the long term catch rate. The probability of a positive catch, however, undergoes a significant change between 1993 - 1997 so when these two series are combined the net result is stable catch rates from 1976 - about 1990 followed by a gradual increase up until 2008, followed by a decline to the present day. Despite the decline the catch rates since 2008 the catch rates are still above the long term average. The catch rates (for the positive catches only) for both South Australia and

Bass Strait (Figure 17.7) follow approximately the same trajectory through time (Figure 17.9).


Figure 17.8.The relative influence of each factor used on the final trend in the optimal standardization for Bass Strait Gummy shark. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor 3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

In case the large transition in the probability of a positive shot in the 1990s had a large influence on the outcome the trend was re-plotted using the same trend for the positive catches, but the probability of a positive shot was rescaled to the mean of the estimates between 1996-2011. When this is done there is still an increase in catch rates from 1987 through to 2008 but the current catch rates are now at the long term average rather than being above (Figure 17.10).


Figure 17.9. A comparison of the optimum standardized catch rates for positive catches for South Australia (SA) and Bass Strait (BS), both scaled to a mean of 1.0 over the years 1984 - 2011.

Table 17.3. The statistical diagnostics for the South Australian gummy shark standardization. The smallest AIC and largest adjusted $\mathrm{r}^{2}$ indicates the optimum statistical model.

|  | GeoMean | Vessel | Area | Month | Gear | Depth | Area:Month |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 42045 | 36699 | 31733 | 30962 | 30951 | 29680 | 27166 |
| RSS | 188934 | 181559 | 175186 | 174198 | 174179 | 172545 | 168995 |
| MSS | 5351 | 12727 | 19099 | 20088 | 20107 | 21740 | 25291 |
| Nobs | 139990 | 139990 | 139990 | 139990 | 139990 | 139990 | 139990 |
| Npars | 36 | 150 | 168 | 179 | 181 | 205 | 403 |
| adj_r2 | 2.730 | 6.451 | 9.723 | 10.225 | 10.234 | 11.060 | 12.767 |
| \%Change |  | 3.721 | 3.272 | 0.502 | 0.009 | 0.826 | 1.706 |



Figure 17.10. A comparison of the overall combined standardized catch rates for Bass Strait (Combined) with the same time series for the positive catches, but the probability of a positive shot being scaled to the mean of the time series over the years 1984 - 2011. This was to illustrate the influence of the large change in the probability of a positive shot during the 1990s.

Table 17.4. The different standardization models fitted to the Bass Strait gummy shark data. The models are cumulative across the table with the optimum being the Area:Month model. The CVs reflect the relative variability of each time series.

| Year | GeoMean | Vessel | Area | Month | Gear | Depth | Area:Month |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1976 | 1.3817 | 1.2373 | 1.2609 | 1.2555 | 1.2249 | 1.2134 | 1.1860 |
| 1977 | 1.4091 | 1.2128 | 1.2138 | 1.2264 | 1.2347 | 1.2437 | 1.2189 |
| 1978 | 1.0946 | 1.0350 | 1.0671 | 1.0894 | 1.0931 | 1.1210 | 1.1217 |
| 1979 | 0.8321 | 0.7942 | 0.8212 | 0.8256 | 0.8230 | 0.8389 | 0.8631 |
| 1980 | 0.9121 | 0.8892 | 0.8956 | 0.8841 | 0.8819 | 0.8993 | 0.8985 |
| 1981 | 0.9972 | 0.9996 | 1.0161 | 1.0108 | 1.0058 | 1.0201 | 1.0393 |
| 1982 | 1.0260 | 1.0459 | 1.0697 | 1.0613 | 1.0570 | 1.0740 | 1.0806 |
| 1983 | 0.8532 | 0.8788 | 0.9132 | 0.9076 | 0.9051 | 0.9319 | 0.9422 |
| 1984 | 0.7172 | 0.7666 | 0.7856 | 0.7864 | 0.7854 | 0.8089 | 0.8150 |
| 1985 | 0.7268 | 0.7474 | 0.7727 | 0.7692 | 0.7727 | 0.7787 | 0.7949 |
| 1986 | 0.7430 | 0.7263 | 0.7580 | 0.7580 | 0.7613 | 0.7760 | 0.7728 |
| 1987 | 0.6979 | 0.6761 | 0.7069 | 0.7081 | 0.7111 | 0.7259 | 0.7285 |
| 1988 | 0.8135 | 0.8072 | 0.8364 | 0.8295 | 0.8325 | 0.8509 | 0.8561 |
| 1989 | 0.9686 | 0.9520 | 0.9334 | 0.9268 | 0.9302 | 0.9447 | 0.9588 |
| 1990 | 0.9548 | 0.9669 | 0.9497 | 0.9533 | 0.9553 | 0.9711 | 0.9935 |
| 1991 | 0.8763 | 0.8772 | 0.8954 | 0.8998 | 0.9016 | 0.9168 | 0.9347 |
| 1992 | 1.2095 | 1.2311 | 1.2705 | 1.2764 | 1.2767 | 1.3047 | 1.3101 |
| 1993 | 1.2479 | 1.2693 | 1.2951 | 1.3077 | 1.3090 | 1.3269 | 1.3393 |
| 1994 | 0.9295 | 0.9592 | 0.9935 | 0.9983 | 0.9993 | 1.0076 | 1.0234 |
| 1995 | 1.1091 | 1.1215 | 1.1769 | 1.1760 | 1.1771 | 1.1959 | 1.1871 |
| 1996 | 0.8604 | 0.8565 | 0.8799 | 0.8729 | 0.8737 | 0.8867 | 0.8997 |
| 1997 | 0.6595 | 0.6392 | 0.6431 | 0.6414 | 0.6420 | 0.6497 | 0.6556 |
| 1998 | 0.7687 | 0.7612 | 0.7586 | 0.7535 | 0.7541 | 0.7567 | 0.7674 |
| 1999 | 0.9177 | 0.9005 | 0.8823 | 0.8805 | 0.8812 | 0.8792 | 0.8889 |
| 2000 | 0.9195 | 0.9117 | 0.8965 | 0.8968 | 0.8975 | 0.8790 | 0.8827 |
| 2001 | 1.2038 | 1.1918 | 1.1415 | 1.1479 | 1.1488 | 1.1225 | 1.1153 |
| 2002 | 0.9579 | 0.9588 | 0.9212 | 0.9212 | 0.9220 | 0.8967 | 0.8919 |
| 2003 | 0.9779 | 0.9818 | 0.9432 | 0.9370 | 0.9377 | 0.9077 | 0.9014 |
| 2004 | 0.9581 | 0.9708 | 0.9478 | 0.9554 | 0.9558 | 0.9288 | 0.9089 |
| 2005 | 1.0871 | 1.1325 | 1.0708 | 1.0693 | 1.0700 | 1.0402 | 1.0277 |
| 2006 | 1.1554 | 1.1881 | 1.1174 | 1.1120 | 1.1128 | 1.0820 | 1.0584 |
| 2007 | 1.4187 | 1.4534 | 1.3847 | 1.3812 | 1.3822 | 1.3407 | 1.3187 |
| 2008 | 1.5009 | 1.5749 | 1.5268 | 1.5278 | 1.5289 | 1.4902 | 1.4646 |
| 2009 | 1.2362 | 1.3051 | 1.2755 | 1.2831 | 1.2841 | 1.2582 | 1.2376 |
| 2010 | 0.9949 | 1.0279 | 1.0164 | 1.0126 | 1.0134 | 0.9937 | 0.9785 |
| 2011 | 0.8833 | 0.9522 | 0.9626 | 0.9573 | 0.9582 | 0.9376 | 0.9383 |
| CV | 21.66 | 21.52 | 20.22 | 20.42 | 20.34 | 19.62 | 18.88 |

### 17.7 Tasmania

Even though the RAG decided to use the years 1990 onwards there are major changes prior to 1995 . The catches are all $<20$ from $1979-2004$, and the number of records jumps from <200 to >800 between 1994 and 1995 (Figure 17.7). Nevertheless, the trend in the probability of a positive catch is effectively flat throughout the time series and so is the standardized catch rates for positive shots, at least since 1996 (Figure 17.11)


Figure 17.11. Standardized catch rates for Tasmanian gummy sharks using data relating to $6.0^{\prime \prime}, 6.5^{\prime \prime}$, and 7.0 " mesh gear, from areas that reported more than 10 tonnes across the 36 years considered ( $1976-$ 2011), and from vessels with average catches greater than 2 tonnes per annum which had been present in the fishery for at least 3 years. The top panel represents the probability of obtaining a positive catch, the dashed line being the proportion of positive catches in the raw data and the solid line being the statistical optimum model (with the $95 \%$ error bars surrounding the trend). The central panel represents the loglinear modelling of positive catches. The dashed line is the geometric mean while the solid line is the optimal model, and again the bars are the $95 \%$ confidence limits on the mean estimates. The bottom panel represents the final standardized catch rates, combining the results from the log-linear modelling and the binomial modelling of the probability of a positive catch. All trends have been scaled to the mean of each series to ease visual comparison.


Figure 17.12. The relative influence of each factor used on the final trend in the optimal standardization for Tasmanian Gummy shark. The top graph depicts the geometric mean (the black line) and the optimum model (the red line). The difference between them is illustrated by the vertical bars with blue bars indicating the optimum model is higher than the geometric mean and red bars indicating it is lower. The top graph bars are the sum of all the bars in the graphs below. The graphs for individual factors are cumulative. Thus the second graph has the geometric mean (grey line) and the effect of adding Year + factor2 (model 2). In the third graph, the grey line represents model 2 and the black line the effect of adding factor3 to the model. The remaining graphs continue in the same cumulative manner except for the interaction terms which are added singularly to the final single factor model.

Some large changes occurred in the Tasmanian fishery prior to 1997 with respect to both the vessels doing the fishing and the gear that was used ( 6 ", 6.5 ", or 7 "). Otherwise there were few differences between the geometric mean catch rates and the optimum model, so other factors only contributed very little to changes in the observed trend.

Table 17.5. The different standardization models fitted to the Tasmanian gummy shark data. The models are cumulative across the table with the optimum being the Area:Month model. The CVs reflect the relative variability of each time series.

| Year | GeoMean | Vessel | Area | Month | Gear | Depth | Area:Month |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1990 | 0.5216 | 0.4727 | 0.4540 | 0.4933 | 0.5386 | 0.5214 | 0.5102 |
| 1991 | 0.7267 | 0.7851 | 0.7480 | 0.7156 | 0.7380 | 0.7182 | 0.7268 |
| 1992 | 1.1218 | 0.8897 | 1.1116 | 1.1414 | 1.0937 | 0.9816 | 1.0389 |
| 1993 | 0.7605 | 1.3214 | 1.6681 | 1.8050 | 1.8253 | 1.7128 | 1.7786 |
| 1994 | 0.8685 | 1.7159 | 1.9366 | 2.1388 | 2.1129 | 2.0013 | 2.0572 |
| 1995 | 2.3132 | 1.5672 | 1.6532 | 1.6929 | 1.6759 | 1.7498 | 1.6829 |
| 1996 | 1.6851 | 1.1283 | 1.1798 | 1.1847 | 1.1859 | 1.2516 | 1.2222 |
| 1997 | 1.2163 | 0.9804 | 1.0461 | 1.0337 | 1.0412 | 1.0303 | 1.0617 |
| 1998 | 0.8585 | 0.9230 | 0.8434 | 0.8551 | 0.8661 | 0.8915 | 0.9149 |
| 1999 | 0.8518 | 0.9883 | 0.9039 | 0.9076 | 0.9162 | 0.9418 | 0.9568 |
| 2000 | 0.6864 | 0.7419 | 0.7197 | 0.7085 | 0.7170 | 0.7456 | 0.7703 |
| 2001 | 1.4667 | 1.2770 | 1.1798 | 1.1534 | 1.1760 | 1.1858 | 1.2189 |
| 2002 | 0.9812 | 0.9400 | 0.8490 | 0.8118 | 0.8264 | 0.8448 | 0.8279 |
| 2003 | 0.8824 | 1.0858 | 0.9873 | 0.9439 | 0.9473 | 0.9739 | 0.9715 |
| 2004 | 1.0043 | 1.0923 | 0.9720 | 0.9236 | 0.9308 | 0.9523 | 0.9516 |
| 2005 | 0.9734 | 0.8934 | 0.7988 | 0.7546 | 0.7582 | 0.7750 | 0.7630 |
| 2006 | 1.2969 | 1.0886 | 1.0522 | 0.9897 | 0.9545 | 0.9667 | 0.9282 |
| 2007 | 0.7882 | 0.8542 | 0.8203 | 0.8047 | 0.7970 | 0.8096 | 0.8001 |
| 2008 | 0.7282 | 0.7414 | 0.7071 | 0.6841 | 0.6736 | 0.6855 | 0.6589 |
| 2009 | 0.8839 | 0.8717 | 0.8223 | 0.7905 | 0.7860 | 0.7991 | 0.7885 |
| 2010 | 0.8267 | 0.8408 | 0.7930 | 0.7700 | 0.7584 | 0.7710 | 0.7404 |
| 2011 | 0.5577 | 0.8008 | 0.7538 | 0.6971 | 0.6811 | 0.6902 | 0.6304 |
| CV | 40.55 | 28.02 | 35.34 | 40.16 | 39.59 | 37.56 | 38.84 |

Table 17.6. The statistical diagnostics for the Tasmanian gummy shark standardization. The smallest AIC and largest adjusted $r^{2}$ indicates the optimum statistical model.

|  | GeoMean | Vessel | Area | Month | Gear | Depth | Area:Month |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 3470 | 588 | 550 | 380 | 328 | 302 | 117 |
| RSS | 13718 | 10101 | 10058 | 9859 | 9793 | 9720 | 9449 |
| MSS | 864 | 4480 | 4523 | 4722 | 4788 | 4861 | 5132 |
| Nobs | 9601 | 9601 | 9601 | 9601 | 9601 | 9601 | 9601 |
| Npars | 22 | 50 | 52 | 63 | 69 | 92 | 135 |
| adj_r2 | 5.716 | 30.370 | 30.652 | 31.947 | 32.361 | 32.700 | 34.278 |
| \%Change | 0.000 | 24.654 | 0.282 | 1.295 | 0.414 | 0.339 | 1.578 |

### 17.8 Extra Tables

Table 17.7. The annual catches and reported number of records for each of the three main regions. The greyed cells illustrate the years used in the analyses for each region.

|  | Bass Strait |  | South Australia |  | Tasmania |  | Unknown |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | Catch | Records | Catch | Records | Catch | Records | Catch | Records |
| 1976 | 471.093 | 2185 | 52.926 | 490 | 27.485 | 71 |  |  |
| 1977 | 578.470 | 2351 | 47.503 | 479 | 28.538 | 93 |  |  |
| 1978 | 544.502 | 2921 | 44.285 | 304 | 27.808 | 63 |  |  |
| 1979 | 444.641 | 2925 | 23.885 | 187 | 2.354 | 23 |  |  |
| 1980 | 508.746 | 3463 | 22.351 | 178 | 10.666 | 71 |  |  |
| 1981 | 492.028 | 3433 | 23.761 | 118 | 7.573 | 47 |  |  |
| 1982 | 678.532 | 4152 | 9.916 | 176 | 6.171 | 17 |  |  |
| 1983 | 609.786 | 4600 | 91.031 | 324 | 2.053 | 17 |  |  |
| 1984 | 532.883 | 4242 | 299.894 | 513 | 1.829 | 11 |  |  |
| 1985 | 458.243 | 4287 | 306.673 | 575 | 4.718 | 67 |  |  |
| 1986 | 526.546 | 4379 | 409.420 | 972 | 2.582 | 25 |  |  |
| 1987 | 449.632 | 4515 | 482.599 | 1416 | 2.978 | 12 |  |  |
| 1988 | 480.744 | 4550 | 540.592 | 1255 | 9.171 | 150 |  |  |
| 1989 | 450.105 | 3829 | 575.254 | 1204 | 8.363 | 153 |  |  |
| 1990 | 525.466 | 3539 | 465.773 | 1597 | 18.741 | 278 |  |  |
| 1991 | 562.129 | 4408 | 375.650 | 1621 | 11.637 | 131 |  |  |
| 1992 | 732.125 | 4706 | 317.716 | 1435 | 23.068 | 218 |  |  |
| 1993 | 809.244 | 5019 | 331.363 | 1253 | 17.376 | 162 |  |  |
| 1994 | 605.967 | 5163 | 375.433 | 1219 | 7.655 | 126 |  |  |
| 1995 | 950.690 | 6387 | 415.897 | 1248 | 105.810 | 818 |  |  |
| 1996 | 744.193 | 6545 | 507.163 | 919 | 122.991 | 836 |  |  |
| 1997 | 586.505 | 6614 | 537.713 | 4846 | 89.276 | 825 |  |  |
| 1998 | 730.726 | 6656 | 473.153 | 9020 | 84.490 | 883 |  |  |
| 1999 | 948.874 | 7285 | 520.038 | 7494 | 98.170 | 1203 |  |  |
| 2000 | 922.686 | 6714 | 431.015 | 5631 | 73.967 | 839 |  |  |
| 2001 | 1118.283 | 6133 | 381.404 | 5507 | 67.015 | 574 |  |  |
| 2002 | 887.361 | 6035 | 420.466 | 5439 | 104.135 | 716 |  |  |
| 2003 | 916.533 | 6412 | 498.628 | 6725 | 99.713 | 838 |  |  |
| 2004 | 873.027 | 5963 | 475.288 | 5855 | 120.701 | 898 | 0.303 | 10 |
| 2005 | 815.843 | 5145 | 484.716 | 5293 | 87.476 | 690 | 2.011 | 7 |
| 2006 | 735.996 | 4139 | 554.176 | 6227 | 115.736 | 682 |  |  |
| 2007 | 875.038 | 3511 | 438.039 | 4665 | 93.865 | 839 | 1.837 | 4 |
| 2008 | 954.048 | 3691 | 540.465 | 5011 | 62.183 | 648 | 0.21 | 5 |
| 2009 | 833.293 | 4125 | 410.399 | 5283 | 68.633 | 546 | 0.33 | 2 |
| 2010 | 744.537 | 4468 | 382.473 | 5430 | 76.467 | 553 |  |  |
| 2011 | 797.564 | 5270 | 229.194 | 3467 | 102.800 | 699 |  |  |

Table 17.8. A comparison of reported weights with landed weights from the CDR database. Quotas were only introduced in 2001, which was when this data began to be reported in the CDRs. LogBook relate to all methods, GillNets relates to GillNet catches reported in the logbooks.

| Year | Commonwealth | Log-Books | GillNets | \%LogBook | \%GillNet | TAC |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2001 | 1702.654 | 1660.869 | 1566.702 | 97.55 | 92.02 |  |
| 2002 | 1605.165 | 1494.666 | 1411.962 | 93.12 | 87.96 |  |
| 2003 | 1678.243 | 1618.277 | 1514.874 | 96.43 | 90.27 |  |
| 2004 | 1735.455 | 1656.367 | 1469.319 | 95.44 | 84.66 | 1717 |
| 2005 | 1644.881 | 1570.52 | 1390.046 | 95.48 | 84.51 | 1717 |
| 2006 | 1645.733 | 1577.138 | 1405.908 | 95.83 | 85.43 | 1717 |
| 2007 | 1665.106 | 1574.951 | 1408.779 | 94.59 | 84.61 | 2467 |
| 2008 | 1865.681 | 1727.945 | 1556.906 | 92.62 | 83.45 | 1717 |
| 2009 | 1646.200 | 1500.789 | 1312.655 | 91.17 | 79.74 | 1717 |
| 2010 | 1540.178 | 1404.716 | 1203.477 | 91.20 | 78.14 | 1717 |
| 2011 | 1516.728 | 1348.002 | 1129.558 | 88.88 | 74.47 | 1717 |



Figure 17.13. A comparison of the landings reported against quota in the CDRs and catches reported in the log-books, both across all methods and for Gill Nets only.

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# 18. Standardized Catch Rates for the SESSF Saw Shark and Elephant Fish Fisheries. Data from 1980-2011 

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### 18.1 Summary

As recommended by the RAG, catch rates for sawshark were standardized for the years 1980 - 1991 and 1998 - 2011, while those for elephant fish were standardized for the years 1980 - 2011. Both were treated as fisheries across their full geographical ranges but, in addition, in an attempt to focus on the approximate details of the geographical range of the two species of sawsharks, these were also briefly considered as two populations split across eastern and western Bass Strait but because this made only very minor differences to the analyses it was not pursued further. To account for the occurrence of zero catches, the standardizations used a Delta method whereby the probability of obtaining a positive catch is estimated using a Generalized Linear Model with a binomial error structure (to describe the presence or absence of catches). This probability is combined with the yearly indices from a log-linear statistical model that standardizes those catch rates coming from positive catches. Data selection for saw sharks was restricted to the years used (1980-1991 and 1998-2011), those statistical areas from which, cumulatively across the 25 years, more than 10 tonne of sawshark were reported, those vessels that had an average annual catch greater than 0.25 tonnes, and from depths $<160 \mathrm{~m}$. For elephant fish (Callorhinchus milii), data selection was a minimum cumulative catch by statistical area of 4 tonnes, a minimum annual catch per vessel of 0.25 tonnes, and depths $<200 \mathrm{~m}$. For both species only the records pertaining to 6" mesh gear were used. The depth threshold for elephant fish (family Callorhinchidae) is designed to exclude catches of ghost sharks (family Chimaeridae), which are included in the quota allocation for elephant fish; when trunked these can be difficult to separate.

For sawsharks, taking into account the approximate $95 \%$ confidence intervals around the mean estimates for each year, the combined standardized catch rates were noisy but approximately flat from 1981-2011, although 1980 differed significantly from this and 1988 - 1990 appeared to be below the average while 1998 - 2000 appear to be above the long term average. The 2010 and 2011 values appear to be below the scaled average of 1.0. A declining trend to 2010 appears to have begun in 2008 but catch rates in 2011 were the same as in 2010. The combined standardization was robust to different data selection criteria and to splitting the data into eastern and western fisheries. The relatively flat combined catch rate arose because a declining catch rate for positive catches was counter-acted by an increase in the probability of obtaining a positive catch. The drop in 2010 resulted from a recent decline in the relative probability of a positive catch combined with a continuation of the decline in the catch rate of positive catches. Vessels accounted for most variation in the catch rates followed by year, area, and depth category. The Area $x$ Month interaction term accounted for more than twice the variation accounted by Month indicating that seasonal patterns are expressed more by where fishing occurs than by when fishing occurs.

Trawl caught sawsharks exhibited a similar pattern of standardized catch rates to those seen in the GHT for the positive catches. The seasonality of sawshark availability is clearly apparent in the monthly catch rates.

For elephant fish, the standardized catch rates were more variable than those for saw sharks and there was a significant decline between 1984 and 1991. However, catch rates could not be distinguished from the average across the time series from 1992 - 2006. A significant rise from 2007 - 2009 has been reversed and the values for 2010 and 2011 have declined and are not significantly different from the mean of the complete time series. This recent decline is a result of a small decrease in the standardized catch rates for positive catches combined with a decrease in the relative probability of a positive shot. Most of the variation accounted for in the log-linear modelling was driven by Vessel followed by year. Area, month, and depth category were all minor contributors, although, like saw-sharks, the Area x Month interaction was important, suggesting that location of fishing changes with the season which emphasizes that spatial details in this fishery are as important as in the other shark fisheries.

### 18.2 Introduction

The shark fishery off southern Australia has a long history starting with a long-line fishery which began in the 1920s which switched to gillnets in the 1960s and 1970s when the primary target also switched from school sharks (Galeorhinus galeus) to gummy sharks (Mustelus antarcticus; Punt et al., 2000; Punt \& Gason, 2006; Thomson \& Punt, 2010). This gillnet fishery now mainly targets gummy sharks but also used to target relatively large quantities of School sharks although since this became depleted to low levels of the unfished spawning biomass this is now a bycatch only species.

An attempt was made at age-structured stock assessment modelling for both Saw Sharks and Elephant Fish (Punt, et al., 2004). This suggested that pup production in 2003 for both Saw Sharks and Elephant Fish was below $40 \%$ of the 1950 pup production (the assumed virgin stock). However, the catch rate series used was that from the Gummy shark fishery, the analysis was restricted to Bass Strait only, owing to a lack of data, and the effect of combining both species of Saw Sharks was unknown. As the authors stated "The analyses of this paper are clearly preliminary".

Attempts at stock assessment of Saw Sharks and Elephant Fish since Punt et al. (2004) have so far been limited to the application of the SESSF Tier 4 empirical assessment rule in 2009 (Rodriguez \& McLoughlin, 2009a, b) and in 2010 and 2011 (Haddon, 2010a, 2012). These Tier 4 assessments are based upon time series of catches and of standardized catch rates (Little et al., 2011; Haddon, 2012). The intent of the present document is to conduct a standardization of the catch per unit effort data available for both saw sharks and elephant fish in preparation for conducting a third Tier 4 analysis for each group.

As Kimura (1981, p211) says: "Since the 1950s it has been recognized that fishing power generally differs among vessels, and if c.p.u.e. is to be proportional to abundance, effort measurements must be standardized." The most commonly used method of standardization is to include the various factors thought to effect catch rates into a generalized linear model and to include Year as a factor, in this way the
parameters derived for each year become the indices of relative abundance (Venables \& Dichmont, 2004).

After standardization we are left with a set of yearly coefficients that represent the catch rate relative to some reference year (usually scaled to the mean of the time series; thus the average of the series equals one). Unfortunately, even if the standardization accounts for a large proportion of the variability in the data there are no guarantees that catch effort, even standardized catch effort, can act as a good proxy for stock size. Instead of the statistical success of the standardization, one should be able to argue from the nature of the fishery and the species concerned whether or not there is likely to be even an approximate relationship between catch rates and the exploitable biomass.

### 18.3 Methods

### 18.3.1 Catch Rate Standardization

The original data was provided in a text file named CANDE11.txt. This contained 421,977 records each with 23 fields (Table 18.1). There are numerous fields that contain codes in this data set with the codes used for the different regions (Table 18.2) and gears (Table 18.3) being necessary for appropriate record selection (Table 18.6). The data provided received some pre-treatment in order to add the catch rate variables of interest and identify those records for inclusion in the analyses. Catch rates were calculated where there were positive catches of saw sharks or, separately, elephant fish associated with positive effort levels. Where catch rates could be calculated they were also log transformed in preparation for the log-linear modelling. Finally, two fields were added that identified which records contained positive catches of saw sharks and of elephant fish. This latter was necessary as a separate analysis is conducted to characterize the occurrence of zero shots and whether their incidence has altered through time (see below).

In previous standardizations (Rodriguez \& McLoughlin, 2009a, b; Haddon, 2010b) various criteria were used to select records for analysis. An important aspect of any standardization is the number of zero shots, which relate to targeted effort that fails to catch the species of interest. In the SESSF trawl fishery identifying those shots that might have captured a species but didn't is extremely difficult because targeting is so difficult to establish. Fortunately, in the shark gillnet fishery this is less of a problem; gillnet shark fishers are targeting sharks. However, for bycatch species, such as saw sharks and elephant fish, especially those not captured in all areas, there is still an issue in deciding what records to include in the analysis. Both saw sharks and elephant fish are bycatch species and so record selection in each case focused on excluding those areas where few saw sharks or elephant fish are taken (see below; Table 18.9, Table 18.10, Table 18.11, Table 18.12). In addition, records used were limited to particular gears and finally, those vessels that caught very few of either of the two species groups were omitted from consideration.

### 18.3.2 The Delta Distribution

Catch rates are known, generally, to be highly variable, ranging from very high catch rates to shots that contain none of the species of interest. The inclusion of these zero shots is important if there is a trend in the likelihood of failing to catch a species (Stefánsson, 1996). Including zero shots has two parts: 1) First, determine the relative probability of obtaining a positive catch. 2) Secondly, conduct a log-linear standardization on those records containing positive catches. These two analyses are
then combined to provide the overall estimate of the yearly changes in catch rates required for inclusion in stock assessments.

### 18.3.3 Zero Catches

To estimate the probability of a positive observation (i.e. the species of interest is present in a shot) a binomial GLM (using a logit link function) is used to determine the effect of an array of factors on the probability $p_{i}$, which is the probability that the species of interest is present in the $i^{\text {th }}$ shot:

$$
\begin{equation*}
\ln \left(\frac{p_{i}}{1-p_{i}}\right)=\beta_{0}+\beta_{1} x_{i, 1}+\beta_{2} x_{i, 2}+\sum_{j=3}^{N} \beta_{j} x_{i j} \tag{1}
\end{equation*}
$$

where $p_{i}$ is the probability that the species of interest was present in the $i^{\text {th }}$ shot, and $x_{i j}$ are the values of the explanatory variables, $j$, for the $i^{\text {th }}$ shot and the $\beta_{j}$ are the coefficients for the $N$ factors $j$, to be estimated ( $\beta_{0}$ is the intercept, $\beta_{I}$ the coefficient for the first factor, etc.).

The catch rate standardizations all used individual records from the database, which in a number of cases appeared to be aggregated data, potentially aggregated within months, although there were also many individual shots recorded. The catch rate data for positive catches were normalized by using a natural-log transformation. General Linear Models were used with this transformed data rather than using Generalized Linear Models on the untransformed data with a log-link; the approach used has advantages in terms of normalizing the data while stabilizing the variance, which the Generalized Linear Model approach does not always achieve appropriately (Venables \& Dichmont, 2004).

Up to six different log-linear models were fitted and compared in an effort to account for the effects of year, area, month of fishing, vessel, which depth category was used, and any interactions between area and month (Table 18.4). All variables were treated as categorical variables (alternatively termed factors). The optimum statistical model was selected on the basis of the Akaike's Information Criterion (Burnham \& Anderson, 1998), and the adjusted $\mathrm{r}^{2}$ (Neter et al.,1996). The resulting optimal model was plotted in comparison with the geometric mean catch rate, both being scaled to the mean of each series for ease of visual comparison. The standardized catch rates for the year factor are used in the Tier 4 assessment as the index of relative abundance through time.

Standard analyses were conducted in each case and all were coded in the statistical software $R$ ( R development Core Team, 2009). In each case, catch rates, as kilograms per metre of gillnet fished, were natural log-transformed to normalize the data and stabilize the variance. The General Linear Models all had the same form:

$$
\begin{equation*}
\operatorname{Ln}\left(C P U E_{i}\right)=\alpha_{0}+\alpha_{1} x_{i, 1}+\alpha_{2} x_{i, 2}+\sum_{j=3}^{N} \alpha_{j} x_{i j}+\varepsilon_{i j} \tag{2}
\end{equation*}
$$

where $\operatorname{Ln}\left(C P U E_{i}\right)$ is the natural logarithm of the catch rate $(\mathrm{kg} / \mathrm{m})$ for the $i$-th record, $x_{i j}$ are the values of the explanatory variables $j$ for the $i$-th shot (i.e. Year, Disting, Month, etc), and the $\alpha_{j}$ are the coefficients for the $N$ factors $j$ to be estimated ( $\alpha_{0}$ is the intercept, $\alpha_{1}$ is the coefficient for the first factor, etc.), and $\varepsilon_{i j}$ are the normal random residual errors.

### 18.3.4 The Year Effect

The standardised overall year effect for the fishery is calculated as the product of the Year coefficients from the binomial and log-linear GLMs (Eqs (1) and (2)) transformed back onto their original scales. For back-transformation all other predictor variables were set to zero, indicating the reference level of each categorical factor. The expected probability (back-transformed from logit) of a non-zero catch in year $t$ is therefore

$$
\begin{equation*}
\hat{p}_{t}=\frac{\exp \left(\beta_{0}+\lambda_{t}\right)}{1+\exp \left(\beta_{0}+\lambda_{t}\right)} \tag{3}
\end{equation*}
$$

where $p_{t}$ is the probability of a non-zero catch in year $t, \beta_{0}$ is the intercept and the $\lambda_{t}$ is the Year coefficient for year $t$. As a test of the procedure the back transformation of the simple PA = Year model should deliver the annual proportion of positive shots.

For the log-normal model the expected back-transformed year effect involves a biascorrection for log-normality; the back transformation without the correction estimates the median of the distribution rather than the mean, adding $\sigma^{2} / 2$ before backtransformation improves the approximation to the mean of the distribution:

$$
\begin{equation*}
C P U E_{t}=e^{\left(\gamma_{t}+\sigma_{t}^{2} / 2\right)} \tag{4}
\end{equation*}
$$

where $\gamma_{\mathrm{t}}$ is the Year coefficient for year $t$ and $\sigma_{t}$ is its standard error.
Total standardised catch rates for year $t$ are calculated as the product of Eqs (3) and (4), stated relative to the average of all values:

$$
\begin{equation*}
\bar{Y}=\frac{\sum_{t=1}^{n} p_{t} C P U E_{t}}{n} \tag{5}
\end{equation*}
$$

where $n$ is the number of years of data. So the standardized catch rates are given relative to the mean of the series. This implies that the average of the time series of standardized catch rates will always be one, and hence each series is directly comparable with all the others:

$$
\begin{equation*}
Y_{t}=\frac{p_{t} C P U E_{t}}{\bar{Y}} \tag{6}
\end{equation*}
$$

The factors considered in the analyses were all taken as categorical variables and were:
Year the standard calendar year,
Disting each vessel is uniquely and confidentially identified,
Month standard calendar months,
Area Standard shark statistical reporting blocks.
DepCat 20 m categories (novel this year)
Area:Month An interaction term used to include any seasonal changes across areas.

### 18.3.5 Saw Sharks

Shark RAG decided (May 2009) that operator behaviour before 1980 was sufficiently different that pre-1980 data should be excluded. It was also decided that conditions (targeting increased so that catches and catch rates increased markedly) changed between 1992 and 1996 before quotas were introduced for saw sharks (Rodriguez \& McLoughlin, 2009a). Therefore Shark RAG's recommendation of standardizing the CPUE for the years 1980 - 1991 and 1997 - 2009 was adopted in the base case against which all other standardizations were compared.

The fishing gear used has an important influence of catches. Predominantly saw sharks have been taken by 6 inch mesh gill nets (with some taken by unknown mesh sizes; (Figure 18.4; Table 18.7). Thus, only those records where 6 inch gill mesh was reported were used in the base case standardization.

There were also some records where no effort data were included (effort $=-1$ ) and these could not be included in the standardization.

The zero shots considered in the binomial standardization are very influential on the final combined standardization so the selection of which areas to include is very important. There are nine shark regions and multiple shark areas identified within the SESSF (Figure 18.1) and catches are distributed very heterogeneously across these regions and areas in a manner that reflects the geographical distribution of sawsharks. If the total catches taken in each area using 6 inch gear in the years $1980-1991$ and 1997 - 2010 are considered there are 25 areas catching more than $10 \mathrm{t}, 36$ areas with less than 10 tonnes reported, within which 27 reported less than two tonnes Table 18.10). Thus, inspection of available data suggested omitting those areas that reported less than 10 tonnes over the years 1980-1991 and 1997-2009; Table 18.10). This area selection excluded 34 areas and two categories of unknown areas but this only removed about $1.18 \%$ of the reported catch taken by 6 " mesh gear in the years under consideration. The main regions reporting saw shark catches are Eastern Bass Strait and Western Bass Strait, with smaller amounts coming from Eastern South Australia and Eastern Tasmania (Table 18.9, Table 18.12; Figure 18.11). The amount of effort expended in Central South Australia is quite high relative to the saw shark catches (Table 18.5; Table 18.7). If the bulk of the Central South Australian catches were to be included the number of zero shots seems likely to be increased in an inappropriate manner. The main geographical area where saw sharks are found is in Bass Strait with peripheral areas surrounding (Figure 18.11, Table 18.10, Table 18.9).

There are a large number of vessels contributing to the final analysis, even with the restricted number of years and areas used. To remove noise generated by those vessels reporting very small amounts of saw sharks those vessels reporting less than an average of 0.25 tonne per year (for the years in which they reported saw sharks) were removed from the analysis. This removed a further 25 t of catches ( $0.6 \%$ of the catches) from consideration and left a total of 134021 records (Table 18.11).

Finally, to provide depth information the reported minimum and maximum depths were averaged for each record. Previously these were then categorized into depth categories of $0-19 \mathrm{~m}, 20-80 \mathrm{~m}$, and $>80 \mathrm{~m}$ (for comparability with Rodriguez \& McLoughlin, 2009a). However, by plotting the average depths) it was clear that most catches were taken between 0 and 160 m . Removing those records that had no depth information
excluded a further 101 t of catch so that the exclusions left out about $1.2 \%$ of the catches that could have been analysed. Removing catches reported in depths greater than 160 m eliminated about $2 \%$ of catches leaving 129,263 records for the analysis of the probability of a positive catch and 92,323 records for the standardization of positive catch rates (Table 18.11).

Despite these data selection criteria appearing to be a reasonable choice, sensitivity tests were made by conducting the analyses with somewhat different choices. Thus, the results from the 10 t by area, 0.25 t by vessel and gear as 6 inch mesh were contrasted with a limit of 5 t per area, 1.0 t per vessel, and 6 inch, 6.5 inch and 7 inch gear considered together.

In addition, the base case trends across the fishery were compared to analyses conducted separately split across Bass Strait in an effort to isolate at least the southern sawshark. Gomon, Bray, and Kuiter (2008) provide approximate distribution maps which indicate significant overlap although the southern saw shark is not indicated in north east Bass Strait, which is a significant catching area for saw sharks.

### 18.3.6 Elephant Fish

As with the saw sharks, Shark RAG decided (May 2009) that operator behaviour before 1980 was sufficiently different that pre-1980 data should be excluded. No other years were considered necessary for exclusion, therefore Shark RAG's recommendation of standardizing the CPUE for the years including and following 1980 was adopted.

Rodrigues \& McLoughlin (2009b) excluded areas 99 to 108,112 to 115 , 126 to 140 , 148, 149, and 201. These were also eliminated here but in addition areas 122, 144, 155, and 158 were excluded on the grounds of minimal catches. Excluding those areas that had reports of less than 4 tonnes of elephant fish over the period 1980-2010 led to 30 areas being excluded along with two categories of unknown areas. Out of 1,177 tonnes in total this selection excluded about 27 tonnes. Finally, there were hundreds of vessels in the database but a large proportion never reported catching elephant fish. With the much lower catches of elephant fish, those vessels reporting less than an annual average catch of 250 kg were omitted from consideration. Increasing this value to 500 kg made very little difference to the overall standardized trend so the lower value was used to maintain a larger number of observations. Finally, by excluding shots with depths $>$ $200 \mathrm{~m}, 60.5 \%$ of all catches taken between $1980-2010$ were included in the analysis; the biggest reduction in catches used was from excluding gear other than 6 " mesh nets (Table 18.21).

### 18.4 Results

### 18.4.1 The General Southern Shark Fishery

The southern shark fishery extends across from New South Wales, around Tasmania, and across to Western Australia (Figure 18.1).


Figure 18.1. Map of shark statistical reporting areas along with the statistical regions. WA is Western Australia, WSA is Western South Australia, CSA is Central South Australia, ESA is Eastern South Australia (sometimes known as SAV - South Australia Victoria), WBS is Western Bass Strait, EBS is Eastern Bass Strait, NSW is New South Wales, ETS is Eastern Tasmania and WTS is Western Tasmania.


Figure 18.2. The amount of mesh effort across the whole shark fishery applied over the period 1970 to 2011. 8.0' mesh net effort is not shown but only achieved an average of $\sim 850$ across the years it was used. Unknown increased dramatically from 2003 onwards.


Figure 18.3. The seasonality of effort across the whole shark fishery applied over the period 2000 to 2011. The horizontal grey line is the overall average monthly effort across the 12 years.

### 18.4.2 Saw Sharks

### 18.4.2.1 Saw Shark Catches

Saw shark have always been taken mostly by 6.0 " mesh nets and only minor amounts by other gears (Table 18.7). Total catches were approximately 250 tonnes $\pm 100 \mathrm{t}$ from 1980 - 2000 but since then have slowly declined to about 100 tonnes (Figure 18.4).

Saw shark are caught predominantly in depths of $30-80 \mathrm{~m}$ (Figure 18.5) with slightly over $99 \%$ of all catches taken in depths of 130 m and less (Table 18.8).

There is some evidence of a seasonal trend in the catches with more being taken in the November - January and in May - July periods than other months (Figure 18.7). The pattern of seasonality is rather different to that exhibited by the effort through the year (Figure 18.3) so the seasonal fluctuations in catch may reflect changes in availability.


Figure 18.4. Catches of saw sharks by fishing gear method. The hatched areas relate to the periods of exclusion decided upon by SharkRAG (1970-1980, 1992-1996).


Figure 18.5. Number of records and catches of saw sharks by 10 m depth category using data from 1980 1991 and 1997-2011.


Figure 18.6. The percentage of catches and of records taken between 0 and 100 metres depth by five year group. The first ten years data were taken from deeper water on average than in the last 15 years.


Figure 18.7. Catch by Month for saw sharks since 1998 along with an average across years to illustrate the approximate seasonality of the fishery and its variation through time. In each graph the average catch per month across years is illustrated by the grey line.


Figure 18.8. Relative frequency of statistical reporting areas reporting different levels of total catch of saw shark. There are 30 reporting less than 10 tonnes (left panel) with most of those areas reporting less than 4 tonnes (right panel). The data considered related to gear $=6$ " mesh. Effort $>0$, years 1980-91, 1997-2011.

Only a few statistical areas have relatively high catches while 30 areas out of a total of 61 had catches $<10 \mathrm{t}$ (Figure 18.8). Saw shark are mostly taken in Eastern and Western Bass Strait with the next most abundant catches being in Eastern South Australia (Figure 18.9, Figure 18.10, Figure 18.11; Table 18.9, Table 18.12); these three regions dominate the fishery. Relatively minor catches are also taken in eastern and western Tasmania and Central South Australia.


Figure 18.9. Catches by region through time for saw sharks. The hatched areas relate to periods of exclusion decided upon by SharkRAG (1970-1980, 1992-1996).


Figure 18.10. Total 6 inch mesh effort in thousands of kilometres across the years. The amount of effort in Central South Australia is far greater than the catches and would contribute, inappropriately, to zero catches if included.


Figure 18.11. Sawshark catches by area for years 1980-1991 and 1997-2011. Only areas with catches $>$ 10 t are used in the CPUE standardizations. Data used included areas with catches $>10 \mathrm{t}$, and Effort $>0$.


Figure 18.12. The annual catch in tonnes by different methods recorded in the log book data from the SESSF. The grey vertical line indicates post-structural adjustment.

### 18.4.2.2 Saw Shark Catch Rate Standardization

The base case standardization removed those areas that reported less than 10 tonnes over the total period examined, in addition, vessels with an average annual catch less than 0.25 tonnes were also removed (this left 176 vessels). The log-linear modelling of the positive catches tends to reduce the variation exhibited by the geometric mean catch rates, although they generally followed the same trends as the unstandardized CPUE. Catch rates appear relatively flat in the early period (1980-1991) and exhibit an almost $40 \%$ decline in the second period from 1997 to 2011 (Figure 18.13; Table 18.13). Overall, from 1980 to 2011 catch rates of positive catches decline by about $60 \%$, but there are many reasons to consider ignoring the earlier time series as there appear to be many changes in fishing practices between the two periods modelled. The binomial modelling of the probability of obtaining a positive catch exhibits a slight decline in the early period with a higher probability of a positive catch in the second period with an almost $35 \%$ increase from 1997 to 2009, followed by a drop of about $25 \%$ in 2011.

When the two analyses are combined, Equ (6), the optimum model exhibits a downward trend during the early period followed by a relatively flat series in the second period (Figure 18.13; Table 18.13), with a final downturn in 2010 that continued in to 2011. All factors in the log-linear standardization had important impacts on the trends in catch rates, although the interaction term between area and month only had a relatively minor influence (Figure 18.14; Table 18.14). In the binomial modelling the same order of factors were influential except that the depth category factor was more important than the month factor. When the contribution to changes in the trends are graphed (Figure 18.14) then it becomes very clear that there are major differences between the two periods in the data with respect to the vessels operating, the areas operated in, the depths in which operations occur, and to a lesser extent the seasonality of fishing. This is a strong indication that the two data time series are not strictly comparable.

The log-linear modelling is relatively robust to different assumptions about which data to include. Reducing the total catch per area to 5.0 tonnes across the years and increasing the average catch per annum to 1 tonne had very little effect on the outcome. Even when the reported catch per area was 1.0 tonnes, average catch per vessel was
0.25 tonne, and gears used included $6 ", 6.5 "$, and $7 "$ mesh, only a slight difference was observed in 2006 but otherwise the curves were effectively coincident.


Figure 18.13. Standardized catch rates for saw sharks using data relating to 6 " mesh gear, from areas that reported more than 10 tonnes across the 24 years considered, and from vessels with average catches greater than 0.25 tonnes per annum. The top panel represents the log-linear modelling of positive catches. The dashed line is the geometric mean while the solid line is the optimal model. The central panel represents the probability of obtaining a positive catch, the dashed line being the proportion of positive catches in the raw data and the solid line being the statistical optimum model. The bottom panel represents the final standardized catch rates, combining the results from the log-linear modelling and the binomial modelling of the probability of a positive catch (the dashed line represents the optimum loglinear model while the solid line is the optimum combined model, the dotted line is the mean of the optimum combined model and the short vertical red lines are the approximate $95 \%$ confidence intervals).

When the sawshark fishery is divided through the middle of Bass Strait, most of the records and catches are in the east (Table 18.9; Figure 18.23).

In conclusion, for saw sharks, when conducted without attention being paid to the probability of catching saw sharks the log-linear modelling exhibits no real trend from 1980 to 1991 being noisy but flat. However, between 1997 to 2011 there was a significant decline in both nominal and standardized catch rates. With the binomial modelling of the probability of a positive catch, the 1980 to 1991 period was noisy but was lower than the overall average, however, the 1997 to 2010 period had a higher than average probability of a positive catch, which mostly, had the effect of cancelling out
the declining nominal catch rates. Thus, with the two time series combined, the inclusion of 1980 leads to the appearance of a large initial decline by 1981, followed by a relatively flat period during which a reduction occurs but only between 1988 to 1990. The periods from 1981 to 2010 only differs significantly from a flat line between 1988 1990 (with inter-annual noise) and most importantly in 2010 and 2011 where the standardized catch rates are below the average, possibly continuing a trend that has occurred since 2008.


Figure 18.14. The relative contribution of each factor to changes in the trends in the log-linear modelling. Blue bars indicate the line moves above the previous combination of factors while a red line indicates the opposite. The number in each graph is the sum of squared differences between the trends on each graph. The grey line is the black line from the previous graph (top to bottom). The black line in the top graph is the geometric mean and the red line the final trend.

Significant catches of saw sharks are taken as a bycatch in the trawl fishery (Table 18.15). When the catch rates for positive shots in this fishery are standardized (Table 18.16 and Table 18.17) the trends observed are remarkably similar to those seen in the positive shots within the gillnet fishery (Figure 18.25; Figure 18.26). In addition, the
relative catch rates in each month also follow a strong cycle (Figure 18.24), which is also seen in the gill net fishery.

There conditions and operation of the fishery were clearly different in the first time series relative to the second. Because the two time series do not appear to be comparable it is not valid to compare the two separate time series.

### 18.4.3 Elephant fish

### 18.4.3.1 Elephant Fish Catches

The majority of elephant fish catches are taken in Eastern Bass Strait, with smaller amounts in Eastern Tasmania and Western Bass Strait (Table 18.18). Only minor catches are taken in other regions.


Figure 18.15. Catches in the shark fishery by region from 1970 - 2011 for elephant fish; these data include all methods, all vessels, and all areas. The hatched area relates to the periods of exclusion decided upon by SharkRAG (the years prior to 1980).

Like the saw sharks, catches of elephant fish are primarily taken now by 6.0 inch mesh gillnets although similar quantities were taken in the 1980s by unknown mesh nets (Figure 18.16; Table 18.19). Approximately the same seasonal pattern of catches are seen in elephant fish as is seen in saw sharks with more being taken in the November January and in May - July periods than other months. As with saw sharks, the pattern of seasonality is rather different to that exhibited by the effort through the year (Figure 18.17) so the seasonal fluctuations in catch may reflect changes in availability. Total
catches have declined from about 80 t in 1980 to about 60 t in 2000, with present catches at about 50 t (Figure 18.16).


Figure 18.16 Catches of elephant fish by fishing gear method 1970 - 2011; these data include all vessels and areas. The hatched area relates to the period of exclusion decided upon by SharkRAG (1980).

As with sawsharks, elephant fish are mostly taken in small amounts with most statistical areas only reporting less than seven tonnes over the 30 year period from 1980-2011 (Figure 18.20). The distribution of catches is mostly focussed across Bass Strait with the largest amounts from eastern Bass Strait (areas 9, 10, and 11). Eastern Tasmania has reported catches whereas western Tasmania only reports minor catches (Figure 18.20; Table 18.18).


Figure 18.17. Catch by Month for elephant fish since 1998 along with a total across years to illustrate the approximate seasonality of the fishery and its variation through time.




Figure 18.18. Catch by depth. Left panel is the relative frequency of records of elephant fish catches at different depths and the right panel is the catch in tonnes across the years $1980-2011$ as taken by 6 " mesh nets.


Figure 18.19. Catch by depth. Left panel is the relative frequency of records of elephant fish catches at different depths and the right panel is the catch in tonnes across the years $1980-2011$ as taken by 6 " mesh nets.


Figure 18.20. Elephant fish catches by area taken by 6 " gear. The relative catch levels across the years 1980 - 2011. Only those areas with catches greater than 4 t (not the empty areas) were used. If 6.5 " gear is included then areas 150 and 151 show green and areas 5 and 54 shows blue. All areas remain as they were in the analysis using data up until the end of 2010.

### 18.4.3.2 Elephant Fish Standardized Catch Rates

The base case standardization removed those areas reporting less than 4 tonnes across the 30 years from 1980 - 2009. In addition, vessels with an annual average catch less than 0.25 tonnes were also removed, which reduces the number of vessels from 296 down to 86 but only reduces the catch accounted for by about $4.5 \%$ (Table 18.21). The
selection by gear had a much greater impact on the available records and the catch accounted for in the analysis, which stemmed for the large amount of catch taken with unknown mesh size in the early years of the fishery.


Figure 18.21. Standardized catch rates for elephant fish using data relating to $6 "$ mesh gear, from areas that reported more than 4 tonnes across the 32 years considered, and from vessels with average catches greater than 0.25 tonnes per annum. The top panel represents the log-linear modelling of positive catches. The dashed line is the geometric mean while the solid line is the optimal model. The central panel represents the probability of obtaining a positive catch, the dashed line being the proportion of positive catches in the raw data and the solid line being the statistical optimum model. The bottom panel represents the final standardized catch rates, combining the results from the log-linear modelling and the binomial modelling of the probability of a positive catch (the dashed line represents the optimum loglinear model while the solid line is the optimum combined model, the dotted line is the mean of the optimum combined model and the short vertical red lines are the approximate $95 \%$ confidence intervals).

The log-linear modelling of positive catch rates was relatively flat with some oscillations up and down from 1985 to 1995. The standardization is somewhat less variable than the geometric mean catch rates. Over the last 12 years there has been a slight upward trend but essentially these catch rates were flat (Table 18.23; Figure 18.22).


Figure 18.22. The relative contribution of each factor to changes in the trends in the log-linear modelling. Blue bars indicate the line moves above the previous combination of factors while a red line indicates the opposite. The number in each graph is the sum of squared differences between the trends on each graph. The grey line is the black line from the previous graph (top to bottom). The black line in the top graph is the geometric mean and the red line the final trend.

The probability of obtaining elephant fish in a set is much more variable than the loglinear modelling. The probability of catching elephant fish has more than doubled since 1980. There was a decline in the probability of catching elephant fish from 1980 to about 1987, following which there was a slow increase to a peak in 2008 and 2009. The binomial modelling follows the same general trend as the simple probability of a positive catch but is somewhat less variable. The combination of the increasing trend in the probability of a positive catch with the slow increase in catch rates from the loglinear modelling has led to the combined model being very different from the log-linear model (which ignores the effects of zero shots). By considering the $95 \%$ confidence intervals about the yearly mean estimates, for long periods of time the catch rate trend does not differ from the overall average. However, there appears to have been a
significant decline from 1984 - 1990 and the last three years (2007-2009) have exhibited above average catch rates followed by three years of decline back to the long term average (Figure 18.21). As with sawsharks the most influential factors were Vessel followed by Area, though the Area x Month interaction term contributed $\sim 5.7 \%$; the year factor only accounted for $4.2 \%$ of variability (Table 18.22).

The standardization was insensitive or robust to altering the data filtering so that with a minimum catch per area of 10 t and a minimum annual catch of 0.5 t , there were only minor and insignificant differences between both the log-linear modelling of positive catch rates and the probability of obtaining a positive set.

### 18.5 Discussion

### 18.5.1 Saw Sharks

The sawshark standardization is relatively robust to an array of alternative data selection criteria and to the data being split between east and west Bass Strait. The amount of catch in the early period taken with unknown mesh size adds a degree of uncertainty to the analysis. The reporting in the later period appears to be more complete.

The inclusion of the zero shots had a significant effect on the overall catch rate trend. Because the probability of obtaining a positive shot increased in recent years this had the effect of converting a slightly decreasing catch rate trend into an essentially flat pattern. For 1981 to 2010 (ignoring events in 1992-1997) the catch rates for sawsharks appears to show no overall trends. There is a significant drop in 1980 to 1981 (and this is also an exceptional year in the elephant fish standardization). It is possible that starting the time series in 1980 instead of 1981 is affecting the outcome. Fishing behaviour was clearly very different in the 1970s with catch rates of sawsharks very much higher than later (mainly due to a higher probability of capturing a saw shark. The reason for the Shark RAG choosing the years that were selected need to be made clear, although the influence of 1980 on the TIER 4 analysis will be minor.

### 18.5.2 Elephant Fish

The elephant fish standardization is also robust to alternative data selection scenarios. Increasing the total catch per area for inclusion to 10 tonnes, and increasing the average annual catch per vessel to 0.5 tonnes had almost no effect on either the log-linear modelling of the positive catches or the binomial modelling of the probability of catching elephant fish. The elephant fish standardization is based on records of much smaller catches (usually less than half the catches of sawsharks) and the geographical distribution of elephant fish is more restricted than that of saws sharks. Again there was a significant amount of elephant fish taken with unknown mesh sizes in the 1980s and this may have influenced the results.

The standardization of positive shots exhibited some variation from 1985 to 1995 but otherwise was relatively flat. However, the probability of catching elephant fish has increased markedly over the study period with a recent decline in 2010. Since about 1990 the probability of capturing elephant fish appears to have trebled. When the two time series are combined the trend is approximately flat for many years, although it declined significantly between 1984 and 1991 but has increased significantly above average between 2007 - 2009 followed by a large decrease in 2010. The increase in
$2007-2009$ is due to a small rise in the log-linear modelling of positive catches but is especially due to a recent increase in the probability of catching elephant fish. The 2010 decrease was a combination of both a decline in the log-linear modelling and the probability of a positive catch decreasing.

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### 18.7 Additional Graphs

### 18.7.1 Saw Shark



Figure 18.23. Top panel: The relative catches of sawshark when the fishery is split across Bass Strait. Eastern Bass Strait dominates in the second part of the time series. Middle panel: the observed geometric mean catch rates for the east and western Bass Strait regions, and the bottom panel depicts the geometric mean catch rates scaled to a mean of one to simplify the visual comparison of the trends. Data only until 2010.


Figure 18.24. The relative catch rates of saw sharks taken by trawl in the SET as a function of season (by month), illustrating a clear seasonal trend in catch rates.


Figure 18.25. The standardized catch rates for trawl caught saw sharks. The dotted line is the unstandardized geometric mean catch rate. The solid line is the optimum model with approximate $95 \%$ confidence bounds. Following the introduction of quotas to the trawl fishery a clear change occurred. When the gillnet fishery analysis includes the 1992-1996 period it exhibits a huge increase in catch rates, completely inconsistent with the trawl fishery. As with the gill net fishery, the early time series of trawl data for saw sharks, prior to 1995, does not appear comparable to the later series.


Figure 18.26. A comparison between the trawl caught catch rates and the standardization of positive shots in the GHT, and the GHT standardization where the probability of a positive shot is included. The trends in the positive shots are surprisingly similar.


Figure 18.27. A comparison between the trawl caught catch rates and the standardization of positive shots in the GHT, and the GHT standardization where the probability of a positive shot is included for the years 1997-2011 only.

### 18.7.2 Elephant Fish



### 18.8 Tables

### 18.8.1 General Methods

Table 18.1. Data fields contained in the original file, CANDE11.csv, used in the analyses. The fields from CE down (24-28) were added prior to analysis.

| FieldNo | Column | Contents |
| :--- | :--- | :--- |
| 1 | Year | Calendar Year |
| 2 | Month | Calendar Month |
| 3 | Vessel | Vessel Name - only available consistently in two years |
| 4 | Disting | Vessel Distinguishing mark - across all years |
| 5 | orig | Presumably region of original port |
| 6 | op | Operation within the month |
| 7 | Gear | Type of fishing gear mesh size, hooks, or unknown |
| 8 | Region | Fishery Region |
| 9 | Zone | Fishery zone name : BS, SA, TS, or UN |
| 10 | Gummy | Gummy shark catches |
| 11 | School | School shark catches |
| 12 | Comb | Combined School and Gummy shark catches |
| 13 | Saw | Saw shark catches |
| 14 | Eleph | Elephant fish/shark catches |
| 15 | other | Other sharks - seven gill, etc. |
| 16 | Scale | Scalefish catches |
| 17 | Effort | Fishing effort: $-1=$ no data |
| 18 | sh |  |
| 19 | Area | Statistical reporting area |
| 20 | dmin | minimum depth |
| 21 | dmax | maximum depth |
| 22 | gear2 | Second type of gear when used |
| 23 | effort2 | Effort in second type of gear where used. |
| 24 | CE | Catch rate where catches $>0$ and effort $>0$ |
| 25 | LnCE | Log of CE, where CE is valid |
| 26 | Dav | Average of dmin and dmax. |
| 27 | DepCat | Four depth categories: $>90,<=90 \&>75,<=75 \&>30,<=30$ |
| 28 | PA | Positive School shark catches vs zero School shark catch |


| Table 18.2. The regions codes (column originally headed r in <br> CANDE11.csv). <br> Code | Region |
| :--- | :--- |
| -1 | Unknown |
| 1 | Western South Australia |
| 2 | Central South Australia |
| 3 | Eastern Southern Australia |
| 4 | Western Bass Strait |
| 5 | Eastern Bass Strait |
| 6 | Western Tasmania |
| 7 | Eastern Tasmania |
| 8 | New South Wales |
| 10 | Western Australia |


| Table 18.3. <br> The gear codes (column headed $g$ in CANDE11.csv). <br> Code | Meaning |
| :---: | :--- |
| -1 | Unknown? |
| 0 | Unknown? |
| 1 | Unknown |
| 2 | Line |
| 3 | Unknown Mesh Gillnet |
| 5 | 6.5" Gillnet |
| 6 | 6.0" Gillnet |
| 7 | $7.0^{\prime \prime}$ Gillnet |
| 8 | $8.0^{\prime \prime}$ Gillnet |

Table 18.4. The statistical models used in the log-linear modelling of positive catches. The same models were used in the binomial modelling except the LnCE term was replaced by PA (present/absent). DepCat relates to three classes ( $0-19 \mathrm{~m}, 20-80 \mathrm{~m}$, and $>80 \mathrm{~m}$ ).

| Year | LnCE $\sim$ Year |
| :--- | :--- |
| Area | LnCE $\sim$ Year + Disting |
| Disting | LnCE $\sim$ Year + Disting + Area |
| Month | LnCE $\sim$ Year + Disting + Area + Month |
| DepCat | LnCE $\sim$ Year + Disting + Area + Month + DepCat |
| Month:Gear | LnCE $\sim$ Year + Disting + Area + Month + DepCat + Month:Area |

Table 18.5. Effort in the southern shark fishery; in all cases divided by 1000 .

| Year | Unknown | Line | UnknownMesh | 6.5 mesh | 6.0mesh | 7.0mesh | 8.0mesh |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 |  | 204.060 | 21.948 |  |  |  |  |
| 1971 |  | 2690.294 | 7888.331 |  |  |  |  |
| 1972 |  | 936.330 | 12623.219 |  | 21.945 |  |  |
| 1973 |  | 930.871 | 2411.539 | 24.690 | 2220.639 | 12915.206 | 2717.387 |
| 1974 |  | 1084.127 | 2186.466 |  | 8483.933 | 5997.578 | 2217.489 |
| 1975 |  | 956.823 | 1747.700 |  | 11236.080 | 2836.997 | 2343.468 |
| 1976 |  | 631.105 | 3648.162 |  | 11522.868 | 2490.020 | 1114.297 |
| 1977 |  | 659.227 | 5701.002 |  | 14882.550 | 432.528 | 843.523 |
| 1978 |  | 855.017 | 7748.690 | 79.527 | 16119.190 | 221.336 | 155.920 |
| 1979 |  | 401.625 | 7915.678 | 158.060 | 14988.435 | 492.960 | 35.700 |
| 1980 |  | 473.647 | 10342.961 | 463.562 | 16682.365 | 535.434 | 3.000 |
| 1981 |  | 1074.182 | 11119.481 | 327.838 | 16141.491 | 553.200 | 33.000 |
| 1982 |  | 608.045 | 12542.753 | 23.200 | 20592.177 | 423.493 | 110.250 |
| 1983 |  | 874.825 | 5388.237 | 27.000 | 23286.048 | 6467.716 | 137.102 |
| 1984 |  | 1614.408 | 198.780 | 153.130 | 23163.088 | 14329.863 | 4128.200 |
| 1985 |  | 2224.698 | 569.083 | 1761.740 | 23000.926 | 14624.391 | 1481.500 |
| 1986 |  | 3525.794 | 1663.080 | 826.300 | 25866.748 | 25082.980 | 1707.000 |
| 1987 | 16.000 | 3477.219 | 669.831 | 570.360 | 27407.196 | 37276.910 | 2285.851 |
| 1988 |  | 3471.042 | 2874.292 | 1656.346 | 25597.776 | 32898.900 | 1964.710 |
| 1989 |  | 4654.487 | 4982.911 | 2140.300 | 23429.352 | 30523.901 | 753.045 |
| 1990 |  | 2765.441 | 996.928 | 3567.600 | 24783.737 | 28014.864 | 326.700 |
| 1991 | 2.000 | 4219.227 | 7881.462 | 4586.325 | 25594.586 | 24664.335 | 138.200 |
| 1992 |  | 4600.058 | 2736.286 | 2675.800 | 23987.503 | 19639.390 | 87.350 |
| 1993 |  | 3944.153 | 2506.150 | 2578.600 | 23528.226 | 17727.660 | 1004.400 |
| 1994 |  | 2713.568 | 1242.540 | 3986.100 | 23534.358 | 18854.380 | 542.100 |
| 1995 |  | 2850.390 |  | 8138.380 | 28739.786 | 19098.260 | 161.400 |
| 1996 |  | 2427.501 |  | 9117.120 | 30656.291 | 13262.300 | 297.600 |
| 1997 |  | 3692.668 | 2472.850 | 20970.240 | 30205.536 | 3948.900 | 442.300 |
| 1998 |  | 2939.503 | 102.600 | 29436.280 | 27698.295 | 2594.100 | 76.300 |
| 1999 |  | 1475.458 | 292.600 | 23289.672 | 30104.841 | 2263.200 | 19.000 |
| 2000 |  | 570.050 | 304.300 | 16245.780 | 28452.620 | 501.600 |  |
| 2001 |  | 1454.511 | 413.150 | 14702.870 | 27051.340 | 128.600 | 79.400 |
| 2002 |  | 283.900 | 602.900 | 13339.380 | 29097.982 |  |  |
| 2003 | 9702.759 | 440.605 |  | 15438.285 | 32353.550 | 430.400 |  |
| 2004 | 9953.572 | 252.808 |  | 16057.994 | 29540.560 | 34.000 |  |
| 2005 | 10412.111 | 230.600 |  | 14571.380 | 26884.860 |  |  |
| 2006 | 11841.305 | 1.979 |  | 19351.326 | 21998.770 |  |  |
| 2007 | 9056.540 | 2.686 |  | 14895.725 | 19624.580 |  |  |
| 2008 | 9035.687 | 2.247 |  | 14250.907 | 20941.455 | 4.200 | 28.800 |
| 2009 | 7926.485 | 2.584 |  | 13827.510 | 22881.940 |  |  |
| 2010 | 7575.732 | 2.493 |  | 13626.920 | 25668.159 |  |  |
| 2011 | 6920.496 | 3.715 |  | 8443.700 | 27852.090 |  |  |

### 18.8.2 Saw Sharks

Table 18.6. Selection criteria for which records to include in the standardization of saw shark.

Criteria
Years Included
Gear Types
Depth
Areas
Vessels
No minimum effort

## Values

1980-1991 \& 1997-2009
6 inch mesh gillnet
10 m depth classes $0-160 \mathrm{~m}$
Reporting > 10 t over years.
Average annual catch $>0.25 \mathrm{t}$
Remove effort $=-1$ records

Table 18.7. Catch by gear for sawsharks across the years of interest.

| Year Unknown | Line UnkMesh |  | $6.0^{\prime \prime}$ | $6.5^{\prime \prime}$ | $7.0^{\prime \prime}$ | $8.0^{\prime \prime}$ |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1980 |  | 0.114 | 20.040 | 199.781 | 2.302 | 5.732 |  |
| 1981 |  | 0.104 | 39.775 | 146.912 | 0.984 | 5.817 |  |
| 1982 |  | 0.119 | 42.140 | 199.480 | 0.238 | 2.070 |  |
| 1983 | 0.888 | 0.348 | 50.154 | 182.923 | 0.220 | 0.140 |  |
| 1984 |  | 0.259 | 67.008 | 162.274 | 0.400 | 0.524 |  |
| 1985 | 0.012 | 2.658 | 89.538 | 166.407 | 2.302 | 1.996 |  |
| 1986 | 0.970 | 4.583 | 94.038 | 176.238 | 0.828 | 3.841 | 0.031 |
| 1987 | 1.895 | 3.839 | 99.034 | 209.040 | 0.790 | 11.556 | 1.211 |
| 1988 | 0.125 | 2.942 | 61.455 | 177.952 | 1.761 | 4.401 | 0.072 |
| 1989 | 2.561 | 4.660 | 67.549 | 132.396 | 1.231 | 4.088 | 0.105 |
| 1990 | 0.935 | 6.351 | 44.211 | 123.735 | 1.279 | 3.612 |  |
| 1991 | 0.301 | 6.313 | 52.778 | 147.796 | 1.942 | 2.476 |  |
|  |  |  |  |  |  |  |  |
| 1997 | 1.819 | 0.613 | 3.849 | 187.891 | 33.976 | 5.084 | 0.097 |
| 1998 | 1.131 | 0.358 | 0.002 | 226.693 | 15.731 | 0.232 |  |
| 1999 | 1.760 | 0.399 | 0.000 | 199.836 | 10.179 | 0.599 |  |
| 2000 | 0.717 | 0.371 | 0.386 | 186.984 | 12.415 | 0.137 |  |
| 2001 | 0.161 | 4.307 | 0.136 | 162.149 | 9.059 |  |  |
| 2002 | 0.124 | 0.066 | 1.458 | 156.478 | 8.665 |  |  |
| 2003 | 0.162 | 0.174 |  | 197.944 | 6.134 | 0.119 |  |
| 2004 | 3.886 | 0.146 |  | 179.798 | 9.483 |  |  |
| 2005 | 1.023 | 0.209 |  | 159.531 | 11.885 |  |  |
| 2006 | 0.212 | 0.061 |  | 148.534 | 9.857 |  |  |
| 2007 | 1.092 | 0.062 |  | 99.325 | 7.398 |  |  |
| 2008 | 0.407 | 0.097 |  | 105.845 | 8.856 | 0.015 | 0.090 |
| 2009 | 0.542 | 0.129 |  | 83.271 | 5.497 |  |  |
| 2010 | 0.492 | 0.368 |  | 81.260 | 10.597 |  |  |
| 2011 | 0.223 | 0.187 |  | 97.747 | 4.807 |  |  |
| Total | 21.438 | 39.836 | 733.551 | 4298.219 | 178.815 | 52.439 | 1.606 |


| Table 18.8. Catches of Saw <br> Depth | Shark by depth category from $1980-1991$ <br> Catch t | \& $1997-2011$. <br> Cumulative $\%$ | Cumulative C |
| ---: | ---: | ---: | ---: |
| 0 | 7.337 | 0.00186 | 7.337 |
| 10 | 15.494 | 0.00580 | 22.831 |
| 20 | 73.622 | 0.02451 | 96.453 |
| 30 | 212.803 | 0.07860 | 309.256 |
| 40 | 580.605 | 0.22616 | 889.862 |
| 50 | 1390.688 | 0.57960 | 2280.550 |
| 60 | 819.352 | 0.78783 | 3099.902 |
| 70 | 481.329 | 0.91016 | 3581.230 |
| 80 | 133.327 | 0.94405 | 3714.557 |
| 90 | 83.634 | 0.96530 | 3798.191 |
| 100 | 37.720 | 0.97489 | 3835.910 |
| 110 | 29.456 | 0.98238 | 3865.366 |
| 120 | 17.748 | 0.98689 | 3883.114 |
| 130 | 25.843 | 0.99345 | 3908.957 |
| 140 | 18.125 | 0.99806 | 3927.082 |
| 150 | 7.541 | 0.99998 | 3934.623 |
| 160 | 0.088 | 1.00000 | 3934.711 |

Table 18.9. Catch of saw sharks included in the analysis and those excluded, sub-divided by region. Totals relate to years $1980-1991 \& 1997-2011$. The included data also constituted those areas reporting more than 10 t over the years and those vessels reporting an average catch greater than 0.25 t per annum, Effort $>-1$, and depths $<160 \mathrm{~m}$. The total data was only restricted to the same years.

| Region | Included | Total | Excluded t | \% Excluded |
| ---: | ---: | ---: | ---: | ---: |
| Unknown |  | 0.369 | 0.369 | 100 |
| Western SA |  | 31.624 | 31.328 | 100 |
| Central SA |  | 129.219 | 126.402 | 100 |
| Eastern SA | 233.067 | 325.627 | 92.560 | 58.28 |
| Western BS | 1559.598 | 1941.783 | 382.186 | 55.46 |
| Eastern BS | 2237.704 | 2717.051 | 479.347 | 54.84 |
| Western Tas | 17.183 | 66.121 | 48.938 | 79.37 |
| Eastern Tas | 44.956 | 113.387 | 68.432 | 71.61 |
| NSW | 0.106 | 0.702 | 0.596 | 86.88 |
| West Australia |  | 0.023 | 0.023 | 100 |
| Total | 4092.613 | 5325.905 | 1246.944 | 56.55 |

Table 18.10. Catches of saw shark in the shark fishery in tonnes by statistical reporting area. 80-10 refers to all catches taken between 1980-1991 plus 1997 and 2011 by 6 " mesh. $70-11$ relates to all catches in Cande11.csv from 1970 to 2011 using all gears. Those areas reporting less than 10 tonnes during the analysis period (the right hand three columns) were omitted from the analysis. Areas 0 and -1 are unknown. Percent is the proportion of the total catch in the year group included in the catch rate analysis.

| Area | $\mathbf{8 0 - 1 1}$ | $\mathbf{7 0 - 1 1}$ | Area | $\mathbf{8 0 - 1 1}$ | $\mathbf{7 0 - 1 1}$ |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 10 | 708.244 | 1081.944 | 151 | 8.984 | 37.428 |
| 19 | 548.839 | 1071.368 | 150 | 6.783 | 40.213 |
| 22 | 466.817 | 943.440 | 158 | 4.567 | 22.859 |
| 11 | 386.200 | 814.643 | 155 | 3.733 | 16.767 |
| 32 | 290.191 | 558.667 | 54 | 3.407 | 13.063 |
| 5 | 219.384 | 391.833 | 144 | 2.872 | 8.251 |
| 21 | 196.351 | 350.929 | 55 | 2.524 | 27.703 |
| 31 | 172.273 | 329.012 | 114 | 2.273 | 9.918 |
| 23 | 165.803 | 379.417 | 48 | 2.237 | 22.726 |
| 20 | 146.466 | 274.280 | 149 | 2.031 | 15.442 |
| 34 | 138.323 | 256.891 | 105 | 1.225 | 3.238 |
| 6 | 122.955 | 266.578 | 112 | 1.215 | 2.624 |
| 7 | 106.643 | 242.960 | 115 | 1.087 | 7.232 |
| 18 | 106.121 | 273.715 | 148 | 0.889 | 10.819 |
| 35 | 102.855 | 216.939 | 104 | 0.869 | 2.555 |
| 9 | 92.495 | 143.869 | 113 | 0.826 | 4.587 |
| 8 | 62.686 | 123.739 | 126 | 0.741 | 9.268 |
| 33 | 59.609 | 114.893 | 107 | 0.709 | 2.480 |
| 30 | 40.488 | 98.770 | 108 | 0.660 | 6.408 |
| 4 | 23.971 | 60.872 | 138 | 0.567 | 7.436 |
| 42 | 23.481 | 57.643 | 0 | 0.551 | 0.601 |
| 12 | 19.876 | 48.154 | 139 | 0.541 | 3.530 |
| 41 | 17.768 | 80.609 | -1 | 0.441 | 3.673 |
| 49 | 15.092 | 47.334 | 106 | 0.331 | 1.057 |
| 56 | 14.568 | 56.589 | 102 | 0.279 | 0.604 |
| Total | 4247.494 | 8285.084 | 140 | 0.100 | 5.188 |
|  |  |  | 128 | 0.079 | 1.671 |
|  |  |  | 132 | 0.074 | 0.339 |
| Percent | 98.82 | 96.61 | 101 | 0.071 | 1.166 |
|  |  |  | 129 | 0.036 | 1.096 |
|  |  |  | 99 | 0.020 | 0.020 |
|  |  |  | 103 | 0.008 | 0.498 |
|  |  |  | 122 |  | 0.007 |
|  |  |  | 136 |  | 0.050 |
|  |  |  | Total | 50.726 | 290.525 |
|  |  |  |  | 0.010 |  |

Table 18.11. Reduction in records and associated catch of saw sharks made by the data selection decisions. Years used refers to $1980-1991 \& 1997-2011$ (with data available from 1970 - 2012), Gear Used refers to 6 inch gill nets, Effort $<0$ refers to those records for which no effort data was available, areas used refers to those remaining after the removal of areas reporting $<10 \mathrm{t}$ over the time period, Vessels Used refers to those vessels whose average annual catch was $>0.25 \mathrm{t}$. The estimation of the probability of a positive catch had 135740 records and the standardized catch rates for positive catches had 96932 records. The \% loss is relative to each line, whereas the \%records and \%catch relates to the totals following selection by YearsUsed.

|  | Records | Catch | CatchReduct | \%Catch | \%Records | \%Loss |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Records | 431156 | 8575.609 | 0 |  | 0 | 0 |
| YearsUsed | 323687 | 5325.905 | 3249.704 | 100.00 | 100.00 | 37.89 |
| GearUsed | 165323 | 4298.219 | 1027.686 | 80.70 | 51.07 | 11.98 |
| Effort<0 | 163575 | 4274.312 | 23.907 | 80.26 | 50.53 | 0.28 |
| Areasused | 148563 | 4223.667 | 50.646 | 79.30 | 45.90 | 0.59 |
| VesselsUsed | 140370 | 4202.008 | 21.659 | 78.90 | 43.37 | 0.25 |
| Depths | 135740 | 4092.613 | 109.396 | 76.84 | 41.94 | 1.28 |
| '+veCatch | 96932 | 4092.613 | 0.000 | 76.84 | 29.95 | 0.00 |

Table 18.12. Total catches t , of saw shark, by assessment region, in CANDE11.csv

| Year | WestSA | CentSA | EastSA | WestBS | EastBS | WestTas | EastTas | NSW | WA |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1970 | 0 | 0 | 1.728 | 7.247 | 15.839 | 0 | 1.503 | 0 | 0 |
| 1971 | 0 | 0 | 1.197 | 26.772 | 16.309 | 0 | 0.034 | 0 | 0 |
| 1972 | 0 | 0 | 4.591 | 35.638 | 25.717 | 0 | 3.976 | 0 | 0 |
| 1973 | 0 | 0.224 | 13.384 | 46.323 | 87.869 | 0.209 | 0.377 | 0 | 0 |
| 1974 | 0.028 | 0.197 | 3.701 | 45.441 | 162.459 | 2.393 | 13.059 | 0 | 0 |
| 1975 | 0 | 0.020 | 4.570 | 49.025 | 166.018 | 0.008 | 1.198 | 0 | 0 |
| 1976 | 0 | 0.431 | 35.518 | 85.531 | 122.478 | 0.452 | 4.240 | 0 | 0 |
| 1977 | 0 | 0.320 | 13.665 | 98.608 | 114.037 | 2.028 | 1.719 | 0 | 0 |
| 1978 | 0 | 0 | 11.417 | 133.508 | 116.783 | 0.171 | 7.321 | 0 | 0 |
| 1979 | 0 | 0 | 5.391 | 118.550 | 108.596 | 0 | 4.223 | 0 | 0 |
| 1980 | 0 | 0.120 | 4.023 | 113.508 | 108.781 | 0.370 | 1.167 | 0 | 0 |
| 1981 | 0.018 | 0 | 4.462 | 84.540 | 102.892 | 0.441 | 1.239 | 0 | 0 |
| 1982 | 0 | 0 | 8.292 | 130.106 | 101.910 | 0.960 | 2.779 | 0 | 0 |
| 1983 | 0 | 0 | 6.425 | 95.327 | 130.641 | 0.575 | 1.705 | 0 | 0 |
| 1984 | 0 | 0.194 | 4.206 | 92.811 | 124.495 | 3.621 | 5.138 | 0 | 0 |
| 1985 | 0.025 | 1.452 | 5.239 | 109.432 | 134.874 | 5.219 | 6.672 | 0 | 0 |
| 1986 | 1.603 | 2.538 | 4.595 | 114.500 | 143.532 | 8.793 | 4.968 | 0 | 0 |
| 1987 | 5.190 | 8.396 | 13.374 | 159.258 | 131.112 | 7.230 | 2.805 | 0 | 0 |
| 1988 | 1.167 | 7.483 | 13.210 | 108.335 | 107.515 | 4.399 | 6.599 | 0 | 0 |
| 1989 | 0.751 | 5.810 | 7.314 | 80.008 | 109.515 | 1.969 | 7.223 | 0 | 0 |
| 1990 | 2.669 | 3.753 | 4.472 | 70.764 | 90.263 | 5.200 | 3.002 | 0 | 0 |
| 1991 | 2.824 | 4.383 | 17.320 | 89.838 | 89.511 | 2.709 | 5.021 | 0 | 0 |
| 1992 | 1.179 | 3.356 | 6.951 | 85.717 | 90.742 | 14.755 | 6.542 | 0 | 0 |
| 1993 | 0.211 | 2.137 | 8.629 | 113.184 | 142.494 | 17.241 | 5.309 | 0 | 0 |
| 1994 | 0.996 | 2.538 | 7.480 | 129.354 | 176.588 | 5.348 | 5.102 | 0 | 0 |
| 1995 | 0.473 | 6.951 | 31.963 | 185.733 | 152.381 | 2.518 | 10.964 | 0 | 0 |
| 1996 | 0.623 | 3.431 | 16.519 | 136.674 | 136.033 | 5.591 | 11.956 | 0 | 0 |
| 1997 | 2.528 | 7.980 | 28.404 | 86.981 | 95.393 | 2.536 | 9.507 | 0 | 0 |


|  |  |  |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1998 | 2.866 | 8.600 | 16.997 | 81.192 | 126.555 | 2.019 | 5.918 | 0 | 0 |
| 1999 | 1.969 | 7.007 | 18.561 | 60.195 | 117.536 | 1.924 | 5.581 | 0 | 0 |
| 2000 | 0.826 | 5.802 | 12.247 | 39.917 | 134.874 | 1.424 | 5.920 | 0 | 0 |
| 2001 | 1.059 | 10.114 | 8.187 | 22.578 | 129.058 | 1.406 | 3.410 | 0 | 0 |
| 2002 | 0.722 | 4.247 | 15.288 | 44.423 | 95.083 | 1.511 | 5.517 | 0 | 0 |
| 2003 | 1.170 | 5.079 | 18.542 | 66.343 | 106.741 | 2.510 | 4.149 | 0 | 0 |
| 2004 | 0.492 | 5.050 | 15.069 | 74.666 | 92.126 | 0.808 | 5.061 | 0.041 | 0 |
| 2005 | 1.948 | 6.989 | 23.809 | 65.962 | 68.609 | 0.845 | 3.928 | 0.558 | 0 |
| 2006 | 0.917 | 5.895 | 33.697 | 34.519 | 73.476 | 6.505 | 3.631 | 0 | 0 |
| 2007 | 0.556 | 5.770 | 12.760 | 28.365 | 55.522 | 1.156 | 3.745 | 0.004 | 0 |
| 2008 | 0.679 | 7.105 | 9.805 | 22.834 | 71.450 | 0.189 | 3.202 | 0.028 | 0.020 |
| 2009 | 0.592 | 5.027 | 6.873 | 21.464 | 52.866 | 0.367 | 1.903 | 0 | 0.003 |
| 2010 | 0.706 | 6.477 | 8.206 | 19.951 | 55.383 | 0.278 | 1.645 | 0.071 | 0 |
| 2011 | 0.348 | 3.949 | 4.251 | 23.967 | 67.339 | 1.157 | 1.953 | 0 | 0 |
| Total | 35.134 | 148.824 | 492.331 | 3239.088 | 4351.394 | 116.835 | 190.910 | 0.702 | 0.023 |

Table 18.13. The optimum models and geometric mean catch rates and the proportion of positive catches through time for saw sharks. YearS is the geometric mean catch rate relative to the mean of the series, AreaMonth is the optimum statistical model for the log-linear modelling, YearP is the proportion of positive shots and GearP is the proportion of positive shots after standaridzation. Finally, Combined is the combination of both optimum series. This is the base case ( $>10 \mathrm{t} /$ area, $>0.25 \mathrm{t} / \mathrm{vessel}, 6$ " mesh).

| Year | YearS | AreaMonth | YearP | Area:MthP | Combined | StErr |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1980 | 1.6280 | 1.4620 | 1.0488 | 1.0433 | 1.6610 | 0.0000 |
| 1981 | 1.3428 | 1.3343 | 0.9666 | 0.7461 | 1.0841 | 0.1080 |
| 1982 | 1.5023 | 1.3047 | 0.9904 | 0.6220 | 0.8837 | 0.1037 |
| 1983 | 1.3690 | 1.3075 | 0.9871 | 0.7083 | 1.0085 | 0.1034 |
| 1984 | 1.2182 | 1.2778 | 0.9924 | 0.7237 | 1.0070 | 0.1058 |
| 1985 | 1.2617 | 1.3389 | 0.9649 | 0.6928 | 1.0101 | 0.1051 |
| 1986 | 1.2418 | 1.1863 | 0.9297 | 0.6023 | 0.7781 | 0.1025 |
| 1987 | 1.2948 | 1.1603 | 0.9321 | 0.5881 | 0.7430 | 0.1036 |
| 1988 | 1.3964 | 1.2824 | 0.8732 | 0.5156 | 0.7200 | 0.1031 |
| 1989 | 1.3763 | 1.2902 | 0.8606 | 0.4573 | 0.6425 | 0.1070 |
| 1990 | 1.2052 | 1.1968 | 0.8594 | 0.4819 | 0.6281 | 0.1085 |
| 1991 | 1.5343 | 1.4175 | 0.9524 | 0.6384 | 0.9854 | 0.1057 |
|  |  |  |  |  |  |  |
| 1997 | 0.9482 | 0.9382 | 1.0657 | 0.9902 | 1.0117 | 0.1007 |
| 1998 | 0.9100 | 0.9468 | 1.0628 | 1.3012 | 1.3416 | 0.0995 |
| 1999 | 0.8388 | 0.9284 | 1.0669 | 1.3577 | 1.3726 | 0.0980 |
| 2000 | 0.8139 | 0.9104 | 1.0623 | 1.2391 | 1.2284 | 0.0994 |
| 2001 | 0.7969 | 0.8559 | 1.0328 | 1.2011 | 1.1195 | 0.1001 |
| 2002 | 0.6897 | 0.7582 | 1.0410 | 1.2484 | 1.0308 | 0.0998 |
| 2003 | 0.7907 | 0.8621 | 1.0234 | 1.1354 | 1.0659 | 0.0989 |
| 2004 | 0.7044 | 0.7981 | 1.0541 | 1.3145 | 1.1425 | 0.1005 |
| 2005 | 0.6420 | 0.7187 | 1.0488 | 1.3597 | 1.0642 | 0.1025 |
| 2006 | 0.7348 | 0.7508 | 1.0639 | 1.4312 | 1.1701 | 0.1057 |
| 2007 | 0.5889 | 0.6419 | 1.0319 | 1.3690 | 0.9570 | 0.1081 |
| 2008 | 0.6433 | 0.6725 | 1.0614 | 1.4235 | 1.0424 | 0.1082 |
|  | Stock Assessment for SESSF Species: | AFMA Project 2010/0818 |  |  |  |  |


| 2009 | 0.5201 | 0.5848 | 1.0372 | 1.4172 | 0.9026 | 0.1070 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2010 | 0.5063 | 0.5290 | 0.9739 | 1.2080 | 0.6958 | 0.1068 |
| 2011 | 0.5011 | 0.5455 | 1.0164 | 1.1840 | 0.7034 | 0.1067 |

Table 18.14. Statistical model diagnostics for the log-linear modelling for saw-sharks. The optimum model (smallest AIC plus greatest adjusted $\mathrm{r}^{2}$ ) included all factors.

|  | Year | Vessel | Area | DepCat | Montht | Area:Month |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 45873 | 37788 | 30522 | 25652 | 23275 | 18307 |
| RSS | 150122 | 136923 | 126289 | 119643 | 116512 | 109645 |
| MSS | 10538 | 23738 | 34372 | 41018 | 44149 | 51016 |
| Nobs | 90465 | 90465 | 90465 | 90465 | 90465 | 90465 |
| Npars | 27 | 147 | 171 | 181 | 192 | 456 |
| adj_r2 | 6.533 | 14.637 | 21.246 | 25.382 | 27.326 | 31.409 |
| \%Change | 0.000 | 8.105 | 6.609 | 4.136 | 1.944 | 4.083 |

Table 18.15. SawShark catches in the Table 18.17. Statistical model diagnostics for the log-linear modelling of trawl caught saw-sharks. The optimum model (smallest AIC plus greatest adjusted r2) included all factors up to Zone:Month. Commonwealth Logbooks. 37023000 and 37023900 are both generic sawshark codes, while 37023001 is the Southern sawshark and 37023002 is the Common sawshark. DS is Danish Seine, GN is gill net, and TW is trawl.

| Year | 37023000 | 37023001 | 37023002 | 37023900 | DS | GN | TW |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 19.478 |  |  |  | 3.578 |  | 15.900 |
| 1987 | 16.431 |  |  | 0.015 | 2.402 |  | 14.044 |
| 1988 | 30.514 |  |  | 0.505 | 4.832 |  | 26.187 |
| 1989 | 18.608 |  |  | 3.983 | 2.479 |  | 20.112 |
| 1990 | 17.598 |  |  | 9.601 | 3.441 |  | 23.578 |
| 1991 | 23.931 |  |  | 14.442 | 2.541 |  | 34.649 |
| 1992 | 25.541 |  |  | 25.265 | 4.597 |  | 45.031 |
| 1993 | 31.782 |  |  | 20.506 | 3.962 |  | 48.316 |
| 1994 | 43.078 |  |  | 17.149 | 7.446 |  | 52.781 |
| 1995 | 32.762 |  |  | 24.375 | 4.822 |  | 52.196 |
| 1996 | 37.963 |  |  | 29.537 | 6.964 |  | 60.536 |
| 1997 | 194.616 |  |  | 27.611 | 4.018 | 157.406 | 59.769 |
| 1998 | 278.915 |  |  | 25.726 | 6.750 | 249.079 | 48.323 |
| 1999 | 177.741 | 33.985 | 65.618 | 23.123 | 6.464 | 241.592 | 51.660 |
| 2000 | 69.471 | 138.485 | 136.310 | 8.108 | 7.165 | 274.250 | 69.851 |
| 2001 | 75.549 | 107.596 | 155.001 |  | 7.029 | 262.152 | 65.856 |
| 2002 | 97.507 | 52.710 | 105.540 |  | 24.454 | 158.055 | 72.094 |
| 2003 | 126.951 | 59.937 | 131.019 |  | 22.429 | 190.646 | 104.482 |
| 2004 | 121.206 | 68.145 | 125.265 |  | 24.336 | 192.374 | 96.756 |
| 2005 | 124.542 | 66.292 | 105.837 |  | 17.418 | 171.412 | 106.638 |
| 2006 | 159.035 | 51.031 | 107.632 |  | 18.028 | 158.508 | 140.957 |
| 2007 | 106.644 | 38.155 | 69.736 |  | 21.624 | 107.724 | 85.020 |
| 2008 | 96.580 | 41.003 | 74.307 |  | 22.596 | 114.904 | 73.873 |
| 2009 | 102.026 | 29.769 | 59.670 |  | 21.127 | 88.997 | 80.897 |
| 2010 | 99.820 | 29.005 | 63.712 |  | 17.043 | 92.256 | 82.762 |


| 2011 | 91.004 | 28.817 | 74.147 | 25.997 | 102.554 | 64.997 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Table 18.16. Standardization of trawl catch rates for Saw Sharks in the Commonwealth SET fishery. The optimum model included all factors up to Zone:Month. Zone: DepCat is excluded as being less optimal.

|  | Year | Vessel | DepCat | Month | DayNight | Zone | Zone:Month |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | ---: |
| 1986 | 1.5798 | 1.4120 | 1.4517 | 1.4240 | 1.4173 | 1.4742 | $\mathbf{1 . 4 9 8 9}$ |
| 1987 | 1.3047 | 1.4302 | 1.4359 | 1.4631 | 1.4535 | 1.5010 | $\mathbf{1 . 4 9 1 9}$ |
| 1988 | 1.4740 | 1.4419 | 1.3992 | 1.3951 | 1.3921 | 1.4235 | $\mathbf{1 . 4 6 5 5}$ |
| 1989 | 1.4040 | 1.2808 | 1.2897 | 1.3359 | 1.3328 | 1.3265 | $\mathbf{1 . 3 4 9 8}$ |
| 1990 | 1.4467 | 1.2639 | 1.3049 | 1.3369 | 1.3329 | 1.3133 | $\mathbf{1 . 3 1 9 0}$ |
| 1991 | 1.4639 | 1.3249 | 1.3568 | 1.3615 | 1.3582 | 1.3373 | $\mathbf{1 . 3 5 4 2}$ |
| 1992 | 1.5595 | 1.4458 | 1.4541 | 1.4672 | 1.4679 | 1.4655 | $\mathbf{1 . 4 7 5 7}$ |
| 1993 | 1.6356 | 1.6490 | 1.5862 | 1.5945 | 1.5978 | 1.5923 | $\mathbf{1 . 5 5 6 1}$ |
| 1994 | 1.2002 | 1.3569 | 1.3192 | 1.3196 | 1.3228 | 1.3076 | $\mathbf{1 . 3 1 8 4}$ |
| 1995 | 0.8748 | 0.9927 | 1.0114 | 1.0240 | 1.0260 | 1.0024 | $\mathbf{1 . 0 0 7 2}$ |
| 1996 | 0.9264 | 1.0956 | 1.1248 | 1.1145 | 1.1144 | 1.0731 | $\mathbf{1 . 0 7 0 7}$ |
| 1997 | 0.8274 | 0.9511 | 0.9835 | 0.9747 | 0.9751 | 0.9582 | $\mathbf{0 . 9 7 9 3}$ |
| 1998 | 0.7548 | 0.8432 | 0.8639 | 0.8358 | 0.8363 | 0.8446 | $\mathbf{0 . 8 3 3 4}$ |
| 1999 | 0.9081 | 0.9220 | 0.9225 | 0.9109 | 0.9110 | 0.9219 | $\mathbf{0 . 9 0 4 9}$ |
| 2000 | 0.8978 | 0.8649 | 0.8479 | 0.8418 | 0.8412 | 0.8376 | $\mathbf{0 . 8 3 9 4}$ |
| 2001 | 0.7915 | 0.8351 | 0.8277 | 0.8283 | 0.8296 | 0.8265 | $\mathbf{0 . 8 1 4 7}$ |
| 2002 | 0.6808 | 0.7015 | 0.7052 | 0.7182 | 0.7196 | 0.7282 | $\mathbf{0 . 7 1 7 1}$ |
| 2003 | 0.6159 | 0.6508 | 0.6423 | 0.6487 | 0.6504 | 0.6568 | $\mathbf{0 . 6 5 5 8}$ |
| 2004 | 0.6745 | 0.6585 | 0.6500 | 0.6464 | 0.6483 | 0.6484 | $\mathbf{0 . 6 3 3 6}$ |
| 2005 | 0.6807 | 0.6610 | 0.6408 | 0.6357 | 0.6382 | 0.6387 | $\mathbf{0 . 6 3 1 9}$ |
| 2006 | 0.7984 | 0.7585 | 0.7315 | 0.7165 | 0.7181 | 0.7094 | $\mathbf{0 . 7 0 8 7}$ |
| 2007 | 0.7028 | 0.6231 | 0.6058 | 0.6008 | 0.6021 | 0.6032 | $\mathbf{0 . 5 9 4 8}$ |
| 2008 | 0.6393 | 0.6356 | 0.6316 | 0.6231 | 0.6246 | 0.6341 | $\mathbf{0 . 6 2 0 9}$ |
| 2009 | 0.8225 | 0.8119 | 0.8138 | 0.8002 | 0.8036 | 0.8102 | $\mathbf{0 . 8 0 6 5}$ |
| 2010 | 0.6900 | 0.7176 | 0.7324 | 0.7324 | 0.7347 | 0.7225 | $\mathbf{0 . 7 2 0 6}$ |
| 2011 | 0.6460 | 0.6716 | 0.6670 | 0.6503 | 0.6515 | 0.6431 | $\mathbf{0 . 6 3 1 3}$ |

Table 18.17. Statistical model diagnostics for the log-linear modelling of trawl caught saw-sharks. The optimum model (smallest AIC plus greatest adjusted $\mathrm{r}^{2}$ ) included all factors up to Zone:Month.

|  | Year | Vessel | DepCat | Month | DayNight | Zone | Zone:Month | Zone:DepCat |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 21681 | 8606 | 7332 | 5948 | 5910 | 5028 | 3376 | 4411 |
| RSS | 79526 | 61662 | 59653 | 58071 | 58022 | 57044 | 55177 | 56160 |
| MSS | 3698 | 21562 | 23571 | 25152 | 25202 | 26180 | 28047 | 27064 |
| Nobs | 52793 | 52793 | 52307 | 52307 | 52307 | 52307 | 52307 | 52307 |
| Npars | 26 | 204 | 229 | 240 | 243 | 247 | 291 | 347 |
| adj_r2 | 4.398 | 25.622 | 28.008 | 29.902 | 29.958 | 31.133 | 33.331 | 32.070 |
| \%Change | 0.000 | 21.224 | 2.386 | 1.894 | 0.056 | 1.176 | 2.197 | 0.937 |

### 18.8.3 Elephant Fish

Table 18.18. Total catches of elephant fish, in tonnes, by assessment region, in the data table provided, these data are from all methods.

| Year | WestSA | CentSA | EastSA | WestBS | EastBS | WestTas | EastTas | NSW |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 |  |  | 0 | 0.884 | 5.166 | 0.476 | 0.007 |  |
| 1971 |  | 0 | 0 | 1.155 | 2.852 | 0 | 0.034 |  |
| 1972 |  | 0 | 0 | 2.333 | 8.686 | 0 | 0.041 |  |
| 1973 | 0 | 0 | 0 | 7.715 | 22.253 | 0.024 | 1.324 |  |
| 1974 | 0.029 | 0.170 | 0.931 | 4.651 | 50.927 | 0.135 | 8.551 |  |
| 1975 | 0 | 0 | 0 | 5.870 | 56.790 | 0 | 6.894 |  |
| 1976 | 0 | 0 | 0.030 | 10.903 | 28.981 | 0.452 | 1.822 |  |
| 1977 | 0 | 0 | 2.087 | 10.219 | 54.200 | 0 | 1.828 |  |
| 1978 | 0 | 0 | 0.029 | 20.587 | 36.249 | 0 | 8.710 |  |
| 1979 | 0 | 0 | 0.339 | 12.968 | 80.013 | 0 | 7.261 |  |
| 1980 | 0 | 0 | 0.215 | 20.337 | 54.688 | 0.217 | 6.826 |  |
| 1981 | 0 | 0.003 | 0.357 | 24.093 | 52.571 | 0.155 | 4.886 |  |
| 1982 | 0 | 0 | 0.254 | 24.494 | 28.959 | 0.615 | 4.341 |  |
| 1983 | 0 | 0 | 0.356 | 27.246 | 47.466 | 0.244 | 5.166 |  |
| 1984 | 0 | 0.003 | 0.073 | 16.601 | 45.888 | 3.627 | 12.003 |  |
| 1985 | 0.148 | 0.182 | 0.053 | 19.479 | 42.485 | 3.642 | 42.998 |  |
| 1986 | 0 | 0 | 0.052 | 11.962 | 33.639 | 2.365 | 17.350 |  |
| 1987 | 0 | 0.325 | 0.263 | 17.128 | 24.393 | 0.742 | 20.512 |  |
| 1988 | 0 | 0.442 | 0.623 | 16.882 | 21.230 | 1.046 | 26.877 |  |
| 1989 | 0 | 0.065 | 0.080 | 11.178 | 22.757 | 0.478 | 27.551 |  |
| 1990 | 0 | 0.300 | 0.058 | 13.618 | 13.599 | 1.194 | 27.023 |  |
| 1991 | 0.022 | 0.025 | 0.027 | 13.689 | 39.226 | 0.093 | 16.118 |  |
| 1992 | 0 | 0.116 | 0.371 | 14.543 | 26.426 | 5.894 | 23.721 |  |
| 1993 | 0 | 0.007 | 0.025 | 15.537 | 12.642 | 4.370 | 21.754 |  |
| 1994 | 0 | 0.057 | 0.031 | 10.551 | 12.739 | 1.859 | 34.265 |  |
| 1995 | 0 | 1.867 | 0.906 | 21.388 | 12.001 | 1.589 | 14.085 |  |
| 1996 | 0 | 1.267 | 0.718 | 24.274 | 21.291 | 1.794 | 27.767 |  |
| 1997 | 0 | 2.306 | 3.072 | 15.511 | 17.999 | 0.797 | 20.172 |  |
| 1998 | 0.012 | 2.264 | 0.409 | 15.443 | 20.580 | 1.761 | 12.363 |  |
| 1999 | 0.008 | 4.501 | 1.267 | 10.825 | 30.737 | 0.480 | 11.381 |  |
| 2000 | 0.148 | 3.133 | 0.509 | 8.417 | 25.942 | 0.655 | 15.084 |  |
| 2001 | 0.047 | 6.597 | 0.833 | 3.289 | 29.320 | 1.242 | 6.002 |  |
| 2002 | 0 | 2.086 | 0.519 | 6.654 | 24.328 | 0.084 | 6.899 |  |
| 2003 | 0.115 | 3.905 | 0.627 | 6.168 | 30.463 | 1.465 | 6.090 |  |
| 2004 | 0.152 | 1.689 | 0.830 | 4.588 | 20.913 | 0.661 | 6.732 | 0.020 |
| 2005 | 0.173 | 2.041 | 0.149 | 6.998 | 20.896 | 0.463 | 5.568 | 0.013 |
| 2006 | 0.858 | 1.498 | 0.086 | 3.227 | 21.524 | 1.275 | 4.827 | 0 |
| 2007 | 0.332 | 2.492 | 0.121 | 2.559 | 20.270 | 0.368 | 8.587 | 0.040 |
| 2008 | 0.184 | 2.604 | 0.399 | 3.493 | 27.290 | 0.210 | 6.272 | 0.020 |
| 2009 | 0.035 | 2.932 | 0.234 | 6.088 | 29.718 | 0.105 | 4.992 | 0 |
| 2010 | 0.058 | 3.170 | 0.248 | 5.103 | 22.501 | 0.055 | 3.582 | 0.038 |
| 2011 | 0.014 | 4.329 | 0.506 | 4.688 | 20.805 | 0.334 | 3.230 | 0 |
| Total | 2.334 | 50.376 | 17.687 | 483.335 | 1221.400 | 40.966 | 491.495 | 0.131 |


| Year | Unknown | Line | Unknown Mesh | 6.5 " | $6 "$ | $7{ }^{\prime \prime}$ | 8" |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 | 0.483 | 2.942 | 3.108 |  |  |  |  |
| 1971 | 0.132 | 0.148 | 3.761 |  |  |  |  |
| 1972 | 3.324 | 1.143 | 6.593 |  | 0 |  |  |
| 1973 | 0.486 | 0.425 | 0.399 | 0 | 2.160 | 26.108 | 1.738 |
| 1974 | 29.045 | 0.932 | 4.368 |  | 20.045 | 10.834 | 0.170 |
| 1975 | 18.786 | 2.361 | 7.206 |  | 38.340 | 2.861 | 0 |
| 1976 | 4.752 | 1.163 | 4.623 |  | 28.692 | 2.928 | 0.030 |
| 1977 | 7.753 | 0 | 8.599 |  | 51.979 | 0.003 | 0 |
| 1978 | 1.762 | 0 | 16.712 | 0.130 | 46.966 | 0.005 | 0 |
| 1979 | 0 | 0 | 23.797 | 0.285 | 75.706 | 0.793 | 0 |
| 1980 | 0 | 0.002 | 24.897 | 3.998 | 52.320 | 1.066 | 0 |
| 1981 | 0 | 0 | 20.179 | 0.341 | 60.991 | 0.554 | 0 |
| 1982 | 0 | 0.064 | 15.461 | 0.100 | 43.014 | 0.024 | 0 |
| 1983 | 0.408 | 0.020 | 19.950 | 0.020 | 60.016 | 0.064 | 0 |
| 1984 | 0 | 0.236 | 39.972 | 0.486 | 37.498 | 0.003 | 0 |
| 1985 | 0 | 0.565 | 72.741 | 0.026 | 35.440 | 0.215 | 0 |
| 1986 | 0.530 | 2.589 | 33.395 | 0.059 | 28.795 | 0 | 0 |
| 1987 | 0 | 0.069 | 33.299 | 0 | 29.449 | 0.277 | 0.269 |
| 1988 | 0.015 | 7.735 | 30.750 | 0 | 26.890 | 1.710 | 0 |
| 1989 | 0.784 | 1.022 | 36.314 | 0 | 23.857 | 0.107 | 0.025 |
| 1990 | 0.245 | 2.295 | 31.905 | 0.181 | 20.856 | 0.310 | 0 |
| 1991 | 0.184 | 1.380 | 21.913 | 0 | 45.698 | 0.025 | 0 |
| 1992 | 0.022 | 6.899 | 25.674 | 0.070 | 37.512 | 0.893 | 0.001 |
| 1993 | 0.943 | 3.762 | 26.413 | 0 | 19.896 | 0.911 | 2.410 |
| 1994 | 0.798 | 0.504 | 38.013 | 0.021 | 19.550 | 0.156 | 0.460 |
| 1995 | 0.092 | 0.291 |  | 2.395 | 40.482 | 8.576 | 0 |
| 1996 | 2.209 | 0.437 |  | 1.046 | 58.894 | 14.525 | 0 |
| 1997 | 2.908 | 0.069 | 0.500 | 4.437 | 42.999 | 8.944 | 0 |
| 1998 | 2.373 | 1.347 | 0 | 2.303 | 46.809 | 0 | 0 |
| 1999 | 5.390 | 0.435 | 0.498 | 4.557 | 48.278 | 0.041 | 0 |
| 2000 | 3.058 | 0.033 | 0.120 | 4.204 | 46.473 | 0 | 0 |
| 2001 | 0.231 | 0.053 | 0.122 | 5.917 | 41.007 | 0 | 0 |
| 2002 | 0.010 | 0.122 | 0.089 | 2.069 | 38.280 |  |  |
| 2003 | 0.863 | 0.096 |  | 2.932 | 44.740 | 0.202 |  |
| 2004 | 3.220 | 0.525 |  | 2.265 | 29.575 | 0 |  |
| 2005 | 2.065 | 0 |  | 1.674 | 32.562 |  |  |
| 2006 | 1.081 | 0.003 |  | 3.347 | 28.863 |  |  |
| 2007 | 0.273 | 0.037 |  | 4.893 | 29.565 |  |  |
| 2008 | 0.003 | 0.007 |  | 3.057 | 37.389 | 0.005 | 0.010 |
| 2009 | 0.019 | 0.002 |  | 2.703 | 41.412 |  |  |
| 2010 | 0.014 | 0.004 |  | 2.929 | 31.808 |  |  |
| 2011 | 0.004 | 0.025 |  | 3.032 | 30.845 |  |  |

Table 18.20. Catch of elephant fish by 20 m depth category from 1980-2011. All catches below 200m were ignored which removed 4.761 t from 1116.357 t .

| $\mathbf{Y e a r}$ | $\mathbf{0}$ | $\mathbf{2 0}$ | $\mathbf{4 0}$ | $\mathbf{6 0}$ | $\mathbf{8 0}$ | $\mathbf{1 0 0}$ | $\mathbf{1 2 0}$ | $\mathbf{1 4 0}$ | $\mathbf{1 6 0}$ | $\mathbf{1 8 0}$ | $\mathbf{2 0 0}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{1 9 8 0}$ | 5.748 | 9.424 | 23.601 | 8.140 | 0.739 | 1.062 | 3.009 | 0.110 | 0 | 0 |  |
| $\mathbf{1 9 8 1}$ | 7.880 | 18.109 | 20.086 | 8.914 | 0.757 | 0.221 | 0.212 | 0.330 |  |  |  |
| $\mathbf{1 9 8 2}$ | 4.067 | 2.581 | 17.990 | 10.967 | 6.033 | 0.200 | 0 | 0.060 | 0 | 0.145 |  |
| $\mathbf{1 9 8 3}$ | 13.094 | 7.133 | 23.48 | 10.797 | 2.249 | 0.574 | 0.248 | 0.070 | 0 | 0 |  |
| $\mathbf{1 9 8 4}$ | 5.906 | 7.164 | 16.942 | 4.080 | 2.028 | 0.386 | 0.302 | 0.262 | 0.070 | 0 |  |
| $\mathbf{1 9 8 5}$ | 0.914 | 6.482 | 19.662 | 4.279 | 1.591 | 0.623 | 0.462 | 0.210 | 0 | 0 |  |
| $\mathbf{1 9 8 6}$ | 0.216 | 5.424 | 15.129 | 5.415 | 1.166 | 0.434 | 0.122 | 0.100 | 0 | 0 |  |
| $\mathbf{1 9 8 7}$ | 1.063 | 2.871 | 17.202 | 3.830 | 0.825 | 1.641 | 0.388 | 0.103 |  |  |  |
| $\mathbf{1 9 8 8}$ | 3.550 | 3.448 | 10.218 | 5.401 | 1.180 | 1.012 | 0.090 | 0.475 | 0 | 0 |  |
| $\mathbf{1 9 8 9}$ | 1.947 | 5.022 | 10.290 | 4.244 | 0.288 | 0.215 | 0.666 | 0.072 | 0 | 0 |  |
| $\mathbf{1 9 9 0}$ | 0.886 | 2.633 | 10.646 | 5.108 | 0.658 | 0.181 | 0.055 | 0.015 | 0 | 0 |  |
| $\mathbf{1 9 9 1}$ | 10.12 | 4.204 | 10.020 | 5.410 | 1.679 | 5.454 | 3.312 | 3.122 | 0.121 | 0.899 |  |
| $\mathbf{1 9 9 2}$ | 5.597 | 2.306 | 6.821 | 11.393 | 0.662 | 1.499 | 3.821 | 3.932 | 0.025 | 0 |  |
| $\mathbf{1 9 9 3}$ | 0.919 | 2.232 | 8.716 | 4.955 | 0.629 | 0.048 | 0.606 | 1.146 | 0.015 | 0 |  |
| $\mathbf{1 9 9 4}$ | 1.291 | 1.639 | 10.758 | 4.538 | 0.231 | 0.006 | 0.104 | 0.030 | 0.018 | 0 |  |
| $\mathbf{1 9 9 5}$ | 2.865 | 6.301 | 13.696 | 10.162 | 1.046 | 0.666 | 0.936 | 0.244 | 0.016 | 0.325 |  |
| $\mathbf{1 9 9 6}$ | 2.685 | 12.035 | 19.940 | 11.959 | 3.136 | 0.911 | 0.320 | 0.181 | 0.043 | 0.088 |  |
| $\mathbf{1 9 9 7}$ | 2.005 | 6.182 | 13.023 | 9.775 | 1.161 | 0.108 | 0.025 | 0.038 | 0 | 0.020 | 0 |
| $\mathbf{1 9 9 8}$ | 1.100 | 8.747 | 13.52 | 11.445 | 2.160 | 1.328 | 0.302 | 0.003 | 0.005 | 0.043 | 0.002 |
| $\mathbf{1 9 9 9}$ | 1.576 | 7.726 | 22.027 | 9.365 | 0.760 | 0.332 | 0.025 | 0 | 0 | 0 | 0 |
| $\mathbf{2 0 0 0}$ | 2.119 | 6.441 | 20.671 | 5.791 | 1.662 | 0.304 | 0.070 | 0.016 | 0 | 0 |  |
| $\mathbf{2 0 0 1}$ | 1.638 | 4.607 | 19.975 | 7.756 | 0.814 | 0.805 | 0.936 | 0 | 0 | 0 |  |
| $\mathbf{2 0 0 2}$ | 0.932 | 8.112 | 18.451 | 5.568 | 1.758 | 0.248 | 0.020 | 0 |  | 0 |  |
| $\mathbf{2 0 0 3}$ | 0.737 | 11.775 | 21.407 | 4.966 | 0.523 | 0.008 | 0.001 | 0.004 | 0 |  |  |
| $\mathbf{2 0 0 4}$ | 0.557 | 6.902 | 14.648 | 3.377 | 0.188 | 0.006 | 0.012 |  |  |  |  |
| $\mathbf{2 0 0 5}$ | 1.377 | 6.202 | 17.187 | 4.807 | 0.759 | 0.170 | 0 | 0 | 0.002 | 0.001 |  |
| $\mathbf{2 0 0 6}$ | 0.498 | 8.240 | 13.858 | 3.588 | 0.435 | 0.110 | 0 | 0 | 0 | 0 |  |
| $\mathbf{2 0 0 7}$ | 1.120 | 5.991 | 16.146 | 1.976 | 0.776 | 0.215 | 0 | 0 |  |  |  |
| $\mathbf{2 0 0 8}$ | 1.0305 | 6.402 | 22.144 | 4.500 | 0.878 | 0.117 | 0 | 0.080 | 0 |  |  |
| $\mathbf{2 0 0 9}$ | 1.1725 | 5.476 | 25.697 | 4.899 | 1.374 | 0.581 | 0 | 0.002 |  |  |  |
| $\mathbf{2 0 1 0}$ | 1.260 | 4.843 | 16.186 | 6.937 | 0.636 | 0.243 | 0.045 | 0.110 | 0.006 | 0.004 |  |
| $\mathbf{2 0 1 1}$ | 0.667 | 4.143 | 14.768 | 5.963 | 0.940 | 0.596 | 0.219 | 0.156 | 0 | 0.004 |  |

Table 18.21. Reduction in records and associated catch of elephant fish made by the data selection decisions. Years used refers to 1980-2011 (with data available from 1970-2011), Gear Used refers to 6 inch gill nets, Effort < 0 refers to those records for which no effort data was available, areas used refers to the removal of those areas reporting $<4 \mathrm{t}$ over the time period, Vessels Used refers to eliminating those vessels whose average annual catch was less than 0.25 tonnes. The estimation of the probability of a positive catch had 157551 records and the standardized catch rates for positive catches had 39,085 records. The \% loss and used are relative to the original totals, whereas the \%records and \%catch relates to the totals following selection by year.

|  | Records | Catch | CatchReduct | \%Loss | \%Catch | \%Records |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Records | 431156 | 2307.756 | 0 | 0 |  |  |
| YearsUsed $80-11$ | 381721 | 1843.180 | 464.576 | 20.13 | 100.00 | 100 |
| GearUsed $=6$ " | 196812 | 1211.762 | 631.419 | 34.26 | 65.74 | 51.56 |
| Effort $<0$ | 194929 | 1208.228 | 3.534 | 0.29 | 65.55 | 51.07 |
| Areasused $>4$ t | 168115 | 1179.399 | 28.829 | 2.39 | 63.99 | 44.04 |
| vesselsUsed $>250 \mathrm{~kg}$ | 159589 | 1116.357 | 63.042 | 5.35 | 60.57 | 41.81 |
| Depths $0-200 \mathrm{~m}$ | 157551 | 1111.596 | 4.761 | 0.43 | 60.31 | 41.27 |
| '+veCatch | 39085 | 1111.596 | 0 | 0.00 | 60.31 | 10.24 |

Table 18.22. The statistical outcome of the log-linear modelling for elephant fish.

|  | Year | Vessel | Area | Month | DepCat | Area:Month |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 27121 | 22553 | 21383 | 21155 | 21081 | $\mathbf{1 8 5 1 7}$ |
| RSS | 78101 | 68905 | 66802 | 66376 | 66216 | $\mathbf{6 1 2 8 3}$ |
| MSS | 4464 | 13661 | 15764 | 16190 | 16350 | $\mathbf{2 1 2 8 3}$ |
| Nobs | 39085 | 39085 | 39085 | 39085 | 39085 | $\mathbf{3 9 0 8 5}$ |
| Npars | 32 | 196 | 217 | 228 | 238 | $\mathbf{4 6 9}$ |
| adj_r2 | 5.332 | 16.127 | 18.643 | 19.139 | 19.313 | $\mathbf{2 4 . 8 7 7}$ |
| \%Change | 0.000 | 10.795 | 2.516 | 0.496 | 0.174 | $\mathbf{5 . 5 6 4}$ |

Table 18.23. The optimum models and geometric mean catch rates and the proportion of positive catches through time for elephant fish. YearS is the geometric mean catch rate relative to the mean of the series, AreaMonth is the optimum statistical model for the log-linear modelling, YearP is the proportion of positive shots and Opt_P is the proportion of positive shots after standardization (which was also included the Area:Month interaction term). Finally, Combined is the combination of both optimum series. This is the base case ( $>4 \mathrm{t} /$ area, $>0.25 \mathrm{t} /$ vessel, 6 " mesh). The combined have been scaled to the mean of the time series; making the mean 1.0.

| Year | YearS | AreaMonth | YearP | Opt_P | Combined | $95 \%$ CI |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1980 | 1.2626 | 1.0871 | 1.0225 | 1.0335 | 1.2954 | 0.0000 |
| 1981 | 1.3885 | 1.2275 | 1.1302 | 1.0011 | 1.4169 | 0.2599 |
| 1982 | 0.9419 | 1.0753 | 1.0318 | 0.8914 | 1.1052 | 0.2550 |
| 1983 | 1.1861 | 1.1276 | 1.0020 | 0.8482 | 1.1028 | 0.2526 |
| 1984 | 1.0884 | 1.0503 | 0.8595 | 0.6341 | 0.7679 | 0.2692 |
| 1985 | 0.9441 | 1.0494 | 0.7704 | 0.6016 | 0.7279 | 0.2793 |
| 1986 | 1.2830 | 1.2561 | 0.5386 | 0.4116 | 0.5961 | 0.2900 |
| 1987 | 1.5381 | 1.3630 | 0.4231 | 0.3295 | 0.5178 | 0.3085 |
| 1988 | 1.3081 | 1.3348 | 0.5333 | 0.4489 | 0.6909 | 0.2995 |
| 1989 | 1.2762 | 1.3218 | 0.6614 | 0.5520 | 0.8412 | 0.2965 |
| 1990 | 1.4753 | 1.5785 | 0.6042 | 0.4971 | 0.9046 | 0.3067 |
| 1991 | 1.8630 | 1.8712 | 0.6258 | 0.6211 | 1.3399 | 0.2856 |
| 1992 | 1.2081 | 1.5936 | 0.6352 | 0.7252 | 1.3325 | 0.2816 |
| 1993 | 1.0183 | 1.1842 | 0.4968 | 0.4825 | 0.6588 | 0.2903 |
| 1994 | 1.5961 | 1.5081 | 0.4657 | 0.4444 | 0.7727 | 0.2906 |
| 1995 | 1.2695 | 1.2987 | 0.7906 | 0.7498 | 1.1228 | 0.2558 |
| 1996 | 1.3551 | 1.1502 | 0.9171 | 0.8363 | 1.1090 | 0.2512 |
| 1997 | 0.7479 | 0.7088 | 1.0926 | 1.1338 | 0.9266 | 0.2439 |
| 1998 | 0.6788 | 0.6122 | 1.2089 | 1.2792 | 0.9030 | 0.2428 |
| 1999 | 0.6965 | 0.6422 | 1.2604 | 1.3767 | 1.0194 | 0.2384 |
| 2000 | 0.5863 | 0.6109 | 1.4198 | 1.5370 | 1.0827 | 0.2407 |
| 2001 | 0.6019 | 0.6343 | 1.5081 | 1.7678 | 1.2929 | 0.2419 |
| 2002 | 0.6900 | 0.6787 | 1.1640 | 1.2838 | 1.0046 | 0.2487 |
| 2003 | 0.6869 | 0.6700 | 0.9887 | 1.0050 | 0.7764 | 0.2515 |
| 2004 | 0.5929 | 0.6070 | 0.8907 | 0.8970 | 0.6278 | 0.2570 |
| 2005 | 0.6318 | 0.5787 | 1.2255 | 1.1056 | 0.7377 | 0.2544 |
| 2006 | 0.6399 | 0.6556 | 1.2288 | 1.2060 | 0.9117 | 0.2617 |
| 2007 | 0.6864 | 0.6888 | 1.3665 | 1.6278 | 1.2929 | 0.2643 |
| 2008 | 0.7591 | 0.7917 | 1.5743 | 1.8535 | 1.6919 | 0.2597 |
| 2009 | 0.8618 | 0.8596 | 1.5625 | 1.7884 | 1.7726 | 0.2582 |
| 2010 | 0.6445 | 0.6509 | 1.4737 | 1.5496 | 1.1630 | 0.2610 |
| 2011 | 0.4931 | 0.5332 | 1.5272 | 1.4807 | 0.9103 | 0.2615 |
|  |  |  |  |  |  |  |

## 19. Yield, total mortality values and Tier 3 estimates for selected shelf and slope species in the SESSF 2012

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### 19.1 Summary

This chapter updates yield analyses presented in Klaer (2011a) for major commercial species caught in the Southern and Eastern Scalefish and Shark Fishery (SESSF) on the shelf and slope. Much of the data processing and analysis has been automated, following procedures documented particularly in Thomson (2002a) and Klaer et al. (2008).

Yield and total mortality estimates are provided for major commercial fish species from the shelf and slope in the South East Fishery. Yield estimates were made using a yield-per-recruit model with the following input: selectivity-at-age, length-at-age, weight-atage, age-at-maturity, and natural mortality. Total mortality values corresponding to various reference equilibrium biomass depletions were calculated for each species.

Recent average total mortality was estimated from catch curves constructed from length frequency information. Length frequency data were from ISMP port and/or onboard measurements. The method used to estimate total mortality also estimates average fishery selectivity.

Tier 3 calculations use the estimates of total mortality, natural mortality and average recent catches to decide the Recommended Biological Catch (RBC) for next year. The method used to calculate the Tier 3 RBC is described in Klaer et al. 2008 and Wayte and Klaer (2010). An average length procedure was developed and tested (Klaer et al., 2012) for species with only length data and no age samples are available. The average length method has been applied here for discussion and evaluation, as in Klaer (2011a).

Tier 3 calculations were applied to all SESSF quota species with sufficient available information, regardless of the actual Tier that applies to the species because (a) the Tier that will apply to each species in the current year is decided by the Research Assessment Groups and (b) it is useful to compare Tier 3 results with those from other Tiers to check performance of the methods.
There were no current Tier 3 species without age samples to 2011. While average length results are comparable to age-based catch curves, the performance of Tier 3 using age based catch curves was shown to result in less catch variability (Klaer et al., 2012). As age data are available, there is currently no need to use the average length procedure for Tier 3 species in the SESSF. Consequently, RBC calculations are only shown here that used the age-based catch curve procedure.

At the SESSFRAG meeting in early 2012 it was agreed to allow the investigation of an $M$-based threshold to limit the size of the RBC multiplier produced by Tier 3 analyses. In the results here, $F_{\text {cur }}$ has been limited to a lowest possible value of $M / 10$. Alfonsino, John dory and mirror dory all reached this threshold, so have had the RBC limited by
this rule. RBC values for alfonsino, John dory mirror dory and redfish were all greater than reference average catches.

### 19.2 Methods

### 19.2.1 Zoning

The fishery region and zones referred to here are as shown in Figure 19.1.

Figure 19.1. Map of the SESSF showing 8 statistical zones used in analyses here.


### 19.2.2 Yield analysis

The information required for this calculation was: selectivity-at-age, length-at-age, weight-at-age; age-at-maturity; and natural mortality. The parameters used are shown in Table 19.1. A mix of shelf and slope quota species has been considered and results are presented where the automated process appears to have produced sensible results, and where sufficient data were available.

Table 19.1. Population parameters used for yield analysis.

| Species | $\boldsymbol{M}$ | $\boldsymbol{h}$ | $\boldsymbol{L}_{\boldsymbol{\infty}}$ | $\boldsymbol{k}$ | $\boldsymbol{t}_{\mathbf{0}}$ | $\boldsymbol{a}$ | $\boldsymbol{b}$ | $\boldsymbol{I}_{\mathbf{2 5}}$ | $\boldsymbol{I}_{\mathbf{5 0}}$ | $\boldsymbol{I}_{\text {mat }}$ | $\boldsymbol{a}_{\text {max }}$ | $\mathbf{C C}_{\text {amax }}$ | $\boldsymbol{S}_{\mathbf{2 5}}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Alfonsino | 0.22 | 0.75 | 54.3 | 0.099 | -3.83 | 0.019 | 3.061 | 20 | 25 | 19 | 20 | 10 | 0.8102 |
| John Dory | 0.36 | 0.45 | 53.2 | 0.15 | -1 | 0.0458 | 2.9 | 15.54 | 30 | 31.5 | 20 | 19 | 1.303 |
| Mirror Dory | 0.3 | 0.75 | 57.44 | 0.2345 | 0 | 0.0164 | 3 | 15.54 | 40 | 35 | 20 | 19 | 1.345 |
| Tiger Flathead | 0.22 | 0.62 | 50.87 | 0.168 | -3.053 | 0.0059 | 3.31 | 27.93 | 31.02 | 30 | 20 | 19 | 1.688 |
| Gemfish E | 0.47 | 0.75 | 109.4 | 0.18 | -0.61 | 0.0014 | 3.39 | 26.95 | 31.7 | 70 | 20 | 19 | 0.9612 |
| Gemfish W | 0.47 | 0.75 | 109.4 | 0.18 | -0.61 | 0.0014 | 3.39 | 26.95 | 31.7 | 70 | 20 | 12 | 0.9612 |
| Blue Grenadier | 0.189 | 0.9 | 101 | 0.18 | 0.58 | 0.0038 | 3.013 | 37.8 | 50.73 | 70 | 20 | 19 | 3.185 |
| Pink Ling | 0.268 | 0.8 | 103.4 | 0.166 | 3.139 | 0.0029 | 3.139 | 39.9 | 43 | 67 | 20 | 19 | 6.078 |
| Jackass Morwong E | 0.15 | 0.7 | 36.39 | 0.34 | -0.45 | 0.0429 | 3 | 21.94 | 21.95 | 22 | 25 | 20 | 2.266 |
| Jackass Morwong W | 0.15 | 0.7 | 36.39 | 0.34 | -0.45 | 0.0429 | 3 | 21.94 | 21.95 | 22 | 25 | 20 | 2.266 |
| Ribaldo |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Redfish | 0.1 | 0.75 | 25.28 | 0.224 | -0.719 | 0.0577 | 2.77 | 15.94 | 17.25 | 19 | 40 | 20 | 3.727 |
| Ocean Perch | 0.1 | 0.75 | 43.72 | 0.114 | 0 | 0.0118 | 2.997 | 15.93 | 18.23 | 31 | 40 | 20 | 3.975 |
| Blue-eye Trevalla | 0.1 | 0.75 | 96 | 0.08 | -5.25 | 0.018 | 3.016 | 48 | 50 | 62 | 20 | 19 | 3.414 |
| Silver Trevally | 0.1 | 0.8 | 63.16 | 0.051 | -6.47 | 0.0443 | 2.786 | 22.31 | 22.32 | 28 | 20 | 19 | 2.074 |
| Silver Warehou | 0.3 | 0.75 | 51.25 | 0.464 | -0.65 | 0.0153 | 3 | 31.05 | 40 | 37 | 23 | 20 | 1.357 |
| Blue Warehou | 0.45 | 0.75 | 54.65 | 0.37 | -0.67 | 0.03 | 2.9 | 17.61 | 35 | 33.4 | 25 | 20 | 0.3812 |
| School Whiting | 0.6 | 0.75 | 26 | 0.25 | -1.15 | 0.0132 | 2.93 | 14 | 15 | 16 | 6 | 5 | 1.943 |

For species for which a recent stock assessment has been performed, the population parameters used in the assessment were used here. Otherwise, the primary source of information on population parameters was Smith and Wayte (2002) or, failing that, the Fishbase website (http://www.fishbase.com). A meta-analysis performed by Koopman et al. (2001) was used to provide values for steepness.

### 19.2.2.1 Length- and weight-at-age

Length-at-age was calculated using the von Bertalanffy growth equation (parameters are $l_{\infty}, k$ and $t_{0}$ ) and the weight-at-age using the allometric length-weight relationship (parameters are $a$ and $b$ ). The von Bertalanffy was calculated using length and age data supplied by the Fish Ageing Services (FAS, Kyne Krusic-Golub pers com). The type of length measurement (e.g. standard length or total length) used was specified in the data. It is assumed the parameters of the length-weight relationship (Smith and Wayte, 2002) use the same measures. The units for these parameters are not specified and do not all appear to use the same units. These were manipulated until the results appeared to be in kg per cm . Parameters that were not available from Smith and Wayte (2002) were obtained from the Fishbase website (http://www.fishbase.org), using values that had been calculated from Australian fish or, if necessary, New Zealand fish.

### 19.2.2.2 Female length-at-maturity

Length-at-maturity for females ( $l_{\text {mat }}$ ) (which is converted into a knife-edged function of age using the calculated lengths-at-age) was obtained, where possible, from Wayte and Smith (2002). If separate figures were not available for males and females, that for both sexes combined was used. In some cases several different figures were available and an arbitrary selection was made - when there were three of more figures the median figure was chosen.

### 19.2.2.3 Natural mortality

Natural mortality ( $M$ ) figures were obtained from Smith and Wayte (2002) or by calculating the median of the figures presented by Bax and Knuckey (2001). The value of $M$ for John dory was updated by the Shelf Research Assessment Group in 2005 based on an additional meta-analysis performed by Matt Koopman. The value of $M$ for tiger flathead was updated for the 2010 stock assessment (Klaer, 2011b).

### 19.2.2.4 Selectivity

A logistic selectivity curve is assumed for all species. Selectivity parameters ( $l_{25}, l_{50}$ ) were drawn from Bax and Knuckey's calculated selectivity factors. All parameters used in the present investigation apply to a 90 mm trawl mesh (except for school whiting where 42 mm has been assumed) and non-trawl gear types are not considered. Figures were not available, from Bax and Knuckey, for John dory or silver trevally. Those for mirror dory were applied to John dory because, of all the quota species, mirror dory are most like John dory in shape.

The selectivity parameters used in this study have been estimated from an empirical relationship between fish size and mesh size derived from covered cod end (or trouser haul) experiments on a subset of the species. These pertain purely to gear selectivity, which is not the function often referred to in stock assessments as "selectivity". Fishers are able to target fish of a particular size by fishing in particular areas and in particular different depths -- all SEF quota shelf-associated species show a pattern of larger fish being caught at greater depths. No account is taken in this study of how trawl selectivity changes as a function of gear design or gear deployment (e.g. changing door separation with depth) that have been shown to exert large influences on overall selectivity in other studies.

It has been suggested that practices such as double bagging might reduce the selectivity of commercial trawls below that expected for a 90 mm mesh cod end, however there was no evidence for this, with the possible exception of school whiting and redfish off Eastern Victoria.

The "selectivity" estimated in stock assessment models is a function of both gear selectivity, targeting by the fishery and availability of fish to being caught.

### 19.2.2.5 Maximum age

Maximum observed age $\left(a_{\max }\right)$ values were selected after examining available aged otolith samples. As the maximum age is treated as a plus group, a maximum age for catch curve analysis $\left(c c a_{\max }\right)$ is also required that is normally at least one age less than the maximum. This was chosen after examination of age samples from the last 5 years.

### 19.2.2.6 Stock-recruit relationship

A Beverton-Holt stock-recruit relationship is assumed using the single-parameter formulation suggested by Francis (1992a). The value of this parameter (steepness - $h$ ) was investigated by Koopman et al. (2001) using meta-population analysis. The histograms presented by Koopman et al. were examined and likely figures for steepness chosen. The default figure of 0.75 suggested by Francis (1992b) is used when the results of Koopman et al. do not suggest a clear pattern.

### 19.2.2.7 Management reference points

Using virgin biomass estimates provided by stock reduction analysis in combination with yield-per-recruit analysis, a number of common $F$-based management reference point values were calculated. While $F_{0.1}$ (Gulland and Boerema 1973) and $F_{\text {spr30 }}$ (or $F_{30 \% \mathrm{SPR}}$, Gabriel et al. 1989) are reasonably widely known, the method used to calculate $F_{\text {msy }}$ is given below (from Klaer 2006).

Fisheries management decisions are often based on abundance relative to target and limit reference points. The most common reference point is the population size where maximum sustainable yield (MSY) is achieved. The fully-selected fishing mortality corresponding to MSY, $F_{\mathrm{msy}}$, is defined as the instantaneous rate of fishing mortality at which yield is maximized, i.e.:

$$
\left.\frac{d Y(F)}{d F}\right|_{F_{\text {USY }}}=0
$$

where $Y(F)$ is yield as a function of fully-selected fishing mortality, i.e.:
$Y(F)=\tilde{Y}(F) R(F)$
$\tilde{Y}(F)$ is yield-per-recruit as a function of $F$, and
$R(F)$ is recruitment as a function of $F$.
Yield-per-recruit is defined according to the formula:

$$
\tilde{Y}(F)=\sum_{s} \sum_{a=0}^{x} w_{a}^{s} \frac{S_{a}^{s} F}{Z_{a}^{s}(F)} N_{a}^{s}(F)\left(1-e^{-Z_{a}^{s}(F)}\right)
$$

where $w_{a}^{s}$ is the weight of an animal of sex $s$ and age $a$,
$S_{a}^{s} \quad$ is the selectivity for animals of $\operatorname{sex} s$ and age $a$,
$Z_{a}^{s}(F)$ is the total mortality on fish of sex $s$ and age $a$,
$Z_{a}^{s}(F)=M+S_{a}^{s} F$
$N_{a}^{s}(F)$ is the number of fish of sex $s$ and age $a$ relative to the number of animals of age 0 (both sexes combined):
$N_{a}^{s}(F)= \begin{cases}0.5 & \text { if } a=0 \\ N_{a-1}^{s}(F) e^{-Z_{a-1}^{s}(F)} & \text { if } 0<a<x \\ N_{x-1}^{s}(F) e^{-Z_{x-1}^{s}(F)} /\left(1-e^{-Z_{x}^{s}(F)}\right) & \text { if } a=x\end{cases}$
$x \quad$ is the maximum age-class.
The recruitment as a function of $F$ depends on the assumed form of the stockrecruitment relationship, e.g.:
$R(F)=\frac{S(F)}{\alpha+\beta S(F)}$
where $S(F)$ is spawner biomass as a function of $F$ :
$S(F)=\tilde{S}(F) R(F)$
$\tilde{S}(F)$ is spawner biomass-per-recruit as a function of $F$ :
$\tilde{S}(F)=\sum_{a=1}^{x} f_{a} N_{a}^{\mathrm{fem}}(F)$
$f_{a} \quad$ is fecundity as a function of age.

### 19.2.3 Catch curves

### 19.2.3.1 Data

This investigation used length frequency data from ISMP port measurements (eg Knuckey et al, 2001). For a given year, fleet and population (see below for further detail) length frequencies are catch-weighted and summed to give annual length frequencies.

Age and length data were obtained from the Central Ageing Facility. Age-length keys (ALKs) were constructed from these data.

Two methods were used to convert length frequencies data into age frequencies: ALKs and chopping. The ALK method was used, where possible, to generate age frequencies data by multiplying the length frequency for a given year by the ALK for that same year. No allowances were made for inadequate sampling of an ALK so that, if no age samples were taken from a particular length class then all samples from this length class in the length frequency were ignored. This occurs because the ALK has a zero for all ages for that length class so that the length frequency is always multiplied by zero. 'Chopping' involves using the von Bertalanffy to chop the length frequency into age classes. Catch curve analysis was applied to all resulting age frequencies. In the future it may be desirable to use a chopping method that allows variance in length-at-age about the von Bertalanffy curve.

Age samples from the 2010 and 2011 calendar years became available for both mirror dory and John dory during October 2011, and were used to provide age-based Tier 3 results here for both species. In both cases, all samples from 2010 and 2011 were used to provide an average age-length key that was applied to length data from the most recent 5 years.

### 19.2.3.2 Fleets and Populations

The difference between a fleet and a population is that although the length frequency data are separated for both, the ALK data are separated into populations but are combined across fleets.

For species except tiger flathead, redfish, spotted warehou and blue grenadier, the length frequency data were separated into trawl and non-trawl (including Danish seine) fleets. Tiger flathead was separated into trawl and Danish seine. Non-trawl data for redfish was ignored so that there was only one fleet - a trawl fleet. Spotted warehou was divided into trawl and non-trawl fleets but any Danish seine records were ignored. For blue grenadier the fleets were separated into the summer non-spawning trawl fishery and the winter spawning trawl fishery.

Redfish was divided into two populations - north and south of $36^{\circ} \mathrm{S}$. Population 1 is north and Population 2 south of this latitude.

As there was no recent age data for redfish, all available age data was combined into a single average ALK for that species.

### 19.2.3.3 Automated catch curve analysis

An improved catch curve method for estimating $F_{\text {cur }}$ has been developed. This method uses all selected ages, rather than just the fully-selected ages. $F_{c u r}$ and two selectivity parameters are estimated by fitting an age-structured production model to the observed catches at age over the last five years.

### 19.2.4 Average length method

Catch curve analysis relies on measurement of the decline in numbers at age of a population in equilibrium under constant levels of fishing pressure. If equilibrium conditions apply, the slope of the right hand limb of an age frequency distribution can be used to estimate fishing mortality. For some SESSF fish populations, otoliths have not been collected or aged, sometimes because of the physical difficulty in doing so. Some species, for example, have very tiny otoliths that are both difficult to collect and age. Normally, however, all quota species are measured by onboard observers, or in the port data collection program, so we have reasonably large length frequency samples for most quota species in most years.

The current Tier 3 method for dealing with species with length samples but no age samples is to slice the length-frequency distribution into assumed ages based on the age transitions calculated from the von Bertalanffy parameters, and then apply the standard catch curve analysis to the derived age distribution. This method is not optimal compared to an analysis based on age samples at least because it does not account for the distribution of lengths at age - that the lengths of fish at any age follow a distribution that overlaps with lengths at age for adjacent aged fish.

A procedure has been developed as part of the Reducing Uncertainty in Stock Status (RUSS) project that uses length frequency samples alone to estimate fishing mortality, and is described in detail in Klaer et al. (2012). Management Strategy Evaluation (MSE) testing of the procedure indicated that it works in theory, and provides comparable results to the age-based catch curve method. The greatest disadvantage of the procedure determined by testing was that it produced more variable RBC values than standard catch curve analysis.

The key assumption of the average length method is that the relative number of large fish in the population will reduce as fishing pressure increases. This is intuitively true, and the determination of stock status indicators from average length measurements has a long history (e.g. see Pauly 1984).

The procedure implemented here first requires the selection of a reference length ( $L_{\text {ref }}$ ) where the stock can be assumed to be fully selected. By default, $L_{\text {ref }}$ is assumed to be 2 cm greater than the length at $50 \%$ selection $\left(S_{50}\right)$, as most species are assumed to have relatively knife-edged selection for Tier 3 analyses. The intention was to select a reference length greater than where selectivity effects occur, but as low as possible to allow the largest sample sizes from existing fishery length-frequencies.

Using yield-per-recruit calculations, it is possible to calculate what the average length of the catch above $L_{\text {ref }}$ would be for any level of $F$ (Figure 19.2). To determine current $F$ ( $F_{\text {cur }}$ ) that corresponds to $F_{\text {cur }}$ using catch curves, calculate the average length of the catch above $L_{\text {ref }}$, then use the relationship in Figure 19.2 to determine $F_{\text {cur }}$. The average length of the catch at the limit $F_{20}$ and target $F_{48}$ are shown as dotted lines in Figure 19.2.


Figure 19.2. Average length reference point calculations.

### 19.2.5 Harvest control rule

The method used to calculate the Tier 3 RBC has been improved and is described in Klaer et al. 2008 and Wayte and Klaer (2010). The new Tier 3 control rule that has limit and target fishing levels was implemented and applied for the first time for the 2008 stock assessments.


Figure 19.3. Method for selecting $F_{\text {RBC }}$ based on estimated $F_{\text {cur }}$.

Yield per recruit calculations were used to calculate $F$ values that will reduce the spawning biomass to $20 \%\left(F_{20}\right), 40 \%\left(F_{40}\right)$ and $48 \%\left(F_{48}\right)$ of the unexploited level. The relationship given in Fig. 1 is then used to assign the value of $F_{\text {RBC }}$ using $F_{\text {cur }}$. This relationship has properties similar to the Tier 1 harvest control rule, with $F_{20}$ as the limit and $F_{48}$ as the target fishing mortality rate.

The following formula that adjusts current catch according to the ratio of the intended and current exploitation rates is then used to calculate $C_{\mathrm{RBC}}$ :

$$
C_{R B C}=\frac{\left(1-e^{-F_{R B C}}\right)}{\left(1-e^{-F_{c u r}}\right)} C_{c u r}
$$

where $F_{\text {cur }}$ is the estimated current fishing mortality, $C_{\text {cur }}$ is current catch, $F_{\mathrm{RBC}}$ is the selected $F$ for the recommended biological catch from the control rule, and $C_{\mathrm{RBC}}$ is the recommended biological catch from the control rule.

It can be seen from the above formula that as the $F_{\text {cur }}$ estimate approaches zero, that the multiplier on $C_{\text {cur }}$ exponentially increases to infinity at $M$. Clearly, it is possible for the control rule to generate very large RBC values that are not realistic, and would not result in good behaviour of the HCR. One method for avoiding such behaviour would be to apply direct limits on possible values for the $C_{\text {cur }}$ multiplier. The upper limit of the multiplier on recent average catch was 1.2 in the previous and first implementation of Tier 3 in the SESSF (Klaer and Thomson 2007). To date there has been no agreement via the RAG process on what direct limits may be applied to the new implementation.

The current SESSF application of harvest control rules includes a TAC change limitation rule that was designed to dampen RBC changes from year to year. This applies to all TACs generated from RBCs. In testing the Tier 3 HCR (Wayte and Klaer 2010), the current SESSF catch change limitation rule was also included, which effectively limits the extreme values that may be generated by the Tier 3 HCR. Testing of the Tier 3 rule showed that it was effective in meeting expected management performance measures in the case where the TAC change limitation rule was applied. If such a change limitation rules was not applied, then it is likely that the Tier 3 behaviour would be considerably degraded.

Good performance of the Tier 3 HCR depends on the application of the catch change limitation rule to avoid extreme behaviour. In practice, when the Tier 3 HCR produces unrealistically high or low RBC values due to (1) noise in population age structure data (2) incorrect fixed value for M (3) incorrect biological assumptions in yield-per-recruit calculations (4) incorrect assumptions about fishery selectivity, the behaviour is limited by the TAC change control rule.

In the past, the actual RBC value generated by the Tier 3 HCR has been criticised if it was well above any of the known historical catch levels. The reason why such values are possible using the current HCR have been described here, and how they are correctly dealt with in the overall TAC setting framework. Unexpectedly large RBC values can be generated using the current HCR simply due to the imprecision in the method used to estimate $F_{\text {cur }}$, and it is probably not possible in a short time-frame to determine whether this is the main cause. To avoid misinterpretation of Tier 3 RBCs, both the RBC as generated by the harvest control rule and also the effectively limited values based on the most recent TAC are reported in this document.

A Tier 3 analysis that consistently produces inflated RBC values suggests either that the fishery is having a low impact on the stock, or that some assumptions of the method (e.g. M value) need to be re-examined.

### 19.3 Results

### 19.3.1 Yield per recruit analyses

Figure 19.4. Alfonsino yield per recruit reference point calculations.


Figure 19.5. John dory yield per recruit reference point calculations.
$\underset{\mathrm{V} 2.40}{\text { SESSF }}$ Automated Data Processing - Per Recruit Calculations: John Dory





FO.1 = 0.319
FO.1 = 0.319
F2Olen L = 37.309
F2Olen L = 37.309
Fspr48 = 0.126
Fspr48 = 0.126
F48len L = 38.249
F48len L = 38.249
Fmsy = 0.222
Fmsy = 0.222
BmsY = 0.283
BmsY = 0.283

Figure 19.6. Mirror dory yield per recruit reference point calculations.
$\underset{\mathrm{V}}{\mathrm{SESSF}} \mathbf{2} \boldsymbol{4 0}$ Automated Data Processing - Per Recruit Calculations: Mirror Dory





FO.1 $=0.222$
Fspr20 $=0.355$
F20len $L=0.47 .544$
Fspr48 $=0.147$
F48len L $=48.695$
Fmsy $=0.255$
Bmsy $=0.301$
AvLenRef $=42.000$

Figure 19.7. Tiger flathead yield per recruit reference point calculations.
SESSF Automated Data Processing - Per Recruit Calculations: Tiger Flathead







Figure 19.8. Gemfish east yield per recruit reference point calculations.

```
V SESSF Automated Data Processing - Per Recruit Calculations: Gemfish East
```







```
FO.1 \(=0.266\)
Fspr2o \(=0.252\)
F201en L \(=51.476\)
Fspr48 \(=0.114\)
F48len \(\mathrm{L}=54.143\)
\(\begin{array}{ll}\text { Fmsy } & =0.238 \\ \text { Bmsy } & =0.220\end{array}\)
\(\begin{array}{lll}\text { Bmsy } & 0.220 \\ \text { AvLenRef } & =33.700\end{array}\)
```

Figure 19.9. Gemfish west yield per recruit reference point calculations.


Figure 19.10. Blue grenadier yield per recruit reference point calculations.
$\underset{\mathrm{V}}{\mathrm{SESH} .40} \mathrm{~A}$ Automated Data Processing - Per Recruit Calculations: Blue Grenadier





FO.1 $=0.163$
Fsprzo $=0.244$
201en $\mathrm{L}=70.362$
Fspr48 $=0.097$
Fmsy $=0.226$
Bmsy $=0.221$
AvLenRef $=52.730$

Figure 19.11. Pink ling yield per recruit reference point calculations.


[^5]Figure 19.12. Jackass morwong yield per recruit reference point calculations.

```
SESSF Automated Data Processing - Per Recruit Calculations: Jackass Morwong
```








```
Fspr20 \(=0.294\) F2olen \(L=29.758\) Fspr48 = 0.102
F48len \(L=31.421\)
\(\mathrm{Fmsy}=0.205\)
\(\begin{array}{ll}\text { Fmsy } & =0.205 \\ \text { Bmsy } & =0.285\end{array}\)
AvLenRef \(=23.950\)
```

Figure 19.13. Ribaldo yield per recruit reference point calculations.
$\underset{\mathrm{V}}{\mathrm{SE} 2.40} \mathrm{SF}$ Automated Data Processing - Per Recruit Calculations: Ribaldo






[^6]Figure 19.14. Redfish yield per recruit reference point calculations.


Figure 19.15. Ocean perch yield per recruit reference point calculations.

SESSF Automated Data Processing - Per Recruit Calculations: Ocean Perch V 2.40 12.07.11 CSIRO



| FO.1 | $=0.080$ |
| :--- | :--- |
| Fspr20 | $=0.096$ |
| F20len $L=29.279$ |  |
| Fspr48 | $=0.041$ |
| F48len L | $=30.899$ |
| Fmsy | $=0.082$ |
| Bmsy | $=0.248$ |
| AvLenRef | $=20.230$ |

Figure 19.16. Blue-eye trevalla yield per recruit reference point calculations.

```
SESSF Automated Data Processing - Per Recruit Calculations: Blue-eye Trevalla
    V 2.40 12.07.11 CSIRO
```








```
Fspr20 \(=0.118\)
F2olen \(L=64.572\)
Fspr48 = 0.049
F48len \(L=67.768\)
Fmsy \(=0.106\)
\(\mathrm{Bmsy}=0.232\)
\(\begin{array}{lll}\text { Bmsy } & 0.232 \\ \text { AvLenRef } & =52.000\end{array}\)
```

Figure 19.17. Silver trevally yield per recruit reference point calculations.

> FO.1 $=0.083$
> $\begin{aligned} & \text { Fspr20 }=0.121 \\ & \text { F2olen L }=31.827\end{aligned}$
> $\begin{aligned} & \text { F2olen } L=31.827 \\ & \text { Fspr48 }=0.048\end{aligned}$
> $\begin{aligned} & \text { Fspr48 }=0.048 \\ & \text { F48len } L=34.191\end{aligned}$
> $\begin{aligned} & \text { F48len L }=34 . \\ & \text { Fmsy }=0.100\end{aligned}$
> $\mathrm{Bmsy}=0.253$
> AvLenRef $=24.320$

Figure 19.18. Silver warehou yield per recruit reference point calculations.


Figure 19.19. Blue warehou yield per recruit reference point calculations.
$\underset{\mathrm{V} 2.40}{\text { SESSF Automated Data Processing - Per Recruit Calculations: Blue Warehou }}$






[^7]Figure 19.20. School whiting yield per recruit reference point calculations.


### 19.3.2 Catch curves

The resulting estimates of $Z$ are shown in Figure 19.21 to Figure 19.37. Average catch curve fits to annual age compositions are shown, as well as plots of the estimated $Z$ value versus year per population and fleet.

The results of catch curve analysis are shown together with the total mortality figures ( $Z$ ) that resulted in spawning biomasses of $20 \%$ and $48 \%$ of pristine (dotted horizontal lines).

### 19.3.3 Average length

The resulting estimates of $Z$ using the average length method are shown in Figure 19.38 to Figure 19.54. These results are only presented for information and discussion. Of the current Tier 3 species, only John dory has $F_{\text {cur }}$ estimation based on length data alone in 2010 (and now age-based in 2011).

MSE testing has shown that the average length performs reasonably well with unbiased sampling, but if age data are available, it would be preferable to use age-based catch curves.

Figure 19.21. Alfonsino catch curve results.


Figure 19.22. John dory catch curve results.


Figure 19.23. Mirror dory catch curve results.


Figure 19.24. Flathead catch curve results.


Figure 19.25. Gemfish east catch curve results.


Pap A11 Flt ETraval Length


Figure 19.26. Gemfish west catch curve results.


Figure 19.27. Blue grenadier catch curve results - need to add in port spawn lf.


Figure 19.28. Pink ling catch curve results.


Figure 19.29. Jackass morwong catch curve results.


Figure 19.30. Ribaldo catch curve results place holder - analysis not completed in 2012

Figure 19.31. Redfish catch curve results


Figure 19.32. Ocean perch catch curve results - where are the SEF1 records?


Figure 19.33. Blue-eye trevalla catch curve results.


Figure 19.34. Silver trevally catch curve results.


Figure 19.35. Silver warehou catch curve results.


Figure 19.36 Blue warehou catch curve results.


Figure 19.37. School whiting catch curve results.


Figure 19.38. Alfonsino average length results.


Figure 19.39. John dory average length results.


Figure 19.40. Mirror dory average length results.


Figure 19.41. Tiger flathead average length results.


Figure 19.42. Gemfish east average length results.


Figure 19.43. Gemfish west average length results.


Figure 19.44. Blue grenadier average length results.


Figure 19.45. Pink ling average length results.


Figure 19.46. Jackass morwong average length results.


Figure 19.47. Ribaldo average length results.


Figure 19.48. Redfish average length results.


Figure 19.49. Ocean perch average length results. (Port lengths 2007 require revision)


Figure 19.50. Blue eye trevalla average length results.


Figure 19.51. Silver trevally average length results.


Figure 19.52. Silver warehou average length results.


Figure 19.53. Blue warehou average length results.


Figure 19.54. School whiting average length results.


### 19.4 RBC Calculations

A summary of $Z$ and current $F$ estimates from catch curve analysis is given in Table 2, and the $F$ values resulting in $20 \%$ and $48 \%$ depletion from the previous yield analysis are also shown. Recent $Z$ estimates are taken from the values in Figures 19 to 35 from age-based estimates from fleets that take the majority of catches. The actual values chosen for averaging are highlighted in Appendix 1.

At the SESSFRAG meeting in early 2012 it was agreed to allow the investigation of an $M$-based threshold to limit the size of the RBC multiplier produced by Tier 3 analyses. In the results here, $F_{\text {cur }}$ has been limited to a lowest possible value of $M / 10$. Alfonsino, John dory and mirror dory all reached this threshold, so have had the RBC limited by this rule. Without the limitation, the RBC values were $2,616 \mathrm{t}, 2,131 \mathrm{t}$ and $8,104 \mathrm{t}$ respectively.

At Shelf and Slope RAG October 2012 it was agreed to follow the advice from SESSFRAG in 2011 that non-target species MEY target values may be set to $F_{\text {spr40 }}$ rather than $F_{\text {spr48. }}$. In Table 19.2 the $F_{\text {spr }}$ target used for RBC calculations is highlighted in bold, and the targets for John dory, redfish, silver trevally and blue warehou are now $F_{\text {spr40 }}$. Other species also agreed, but not included in the Tier 3 calculations below were ribaldo and elephant fish.

Table 19.2. $F$ reference points, $Z_{\mathrm{cur}}, C_{\mathrm{cur}}$ and RBC estimates (ribaldo to be included, blue grenadier to be updated).

| Species | Fspr20 | Fspr40 | Fspr48 | Zcur | Fcur | p | ymin | ymax | Ccur | Frbc | Limit? | RBC |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Alfonsino | 0.479 | 0.201 | 0.149 | 0.230 | 0.022 | 6.362 | 2002 | 2011 | 188 | 0.149 | Yes | 1,196 |
| John Dory | 0.287 | 0.159 | 0.126 | 0.370 | 0.036 | 4.145 | 1993 | 2010 | 184 | 0.159 | Yes | 763 |
| Mirror Dory | 0.355 | 0.188 | 0.147 | 0.310 | 0.030 | 4.626 | 1993 | 2010 | 604 | 0.147 | Yes | 2,794 |
| Tiger Fathead | 0.585 | 0.251 | 0.187 | 0.425 | 0.155 | 1.189 | 1993 | 2010 |  |  |  |  |
| Gemfish E | 0.252 | 0.143 | 0.114 | 0.483 | 0.047 | 2.337 | 1993 | 2011 |  |  |  |  |
| Gemfish W | 0.252 | 0.143 | 0.114 | 0.480 | 0.047 | 2.337 | 2000 | 2011 |  |  |  |  |
| Blue Grenadier | 0.244 | 0.125 | 0.097 | 0.444 | 0.255 | 0.000 | 1993 | 2008 |  |  |  |  |
| Pink Ling | 0.250 | 0.134 | 0.105 | 0.282 | 0.027 | 3.769 | 1993 | 2005 |  |  |  |  |
| Monwong East | 0.294 | 0.135 | 0.102 | 0.487 | 0.337 | 0.000 | 1992 | 2009 |  |  |  |  |
| Morwong West | 0.294 | 0.135 | 0.102 | 0.160 | 0.015 | 6.513 | 1992 | 2009 |  |  |  |  |
| Ribaldo |  |  |  |  |  |  |  |  |  |  |  |  |
| Redfish | 0.213 | 0.098 | 0.074 | 0.334 | 0.055 | 1.740 | 1992 | 2008 | 2,209 | 0.098 | No | 3,843 |
| Ocean Perch | 0.096 | 0.052 | 0.041 | 0.261 | 0.161 | 0.000 | 1992 | 2008 |  |  |  |  |
| Blue-eye Trevalla | 0.118 | 0.062 | 0.049 | 0.272 | 0.172 | 0.000 | 1993 | 2008 |  |  |  |  |
| Silver Trevally | 0.121 | 0.062 | 0.048 | 0.798 | 0.698 | 0.000 | 1993 | 2009 |  |  |  |  |
| Silver Warehou | 0.766 | 0.347 | 0.260 | 0.316 | 0.030 | 7.747 | 1992 | 2010 |  |  |  |  |
| Blue Warehou | 0.680 | 0.348 | 0.269 | 0.697 | 0.247 | 1.341 | 1992 | 2011 |  |  |  |  |
| School Whiting | 0.922 | 0.461 | 0.355 | 1.198 | 0.598 | 0.491 | 2007 | 2010 |  |  |  |  |

Notes: Species that were Tier 3 in 2011 are highlighted in bold.
RBC values for alfonsino, John dory, mirror dory and redfish were greater than reference average catches ( $\mathrm{p}>1$ ). Western gemfish, blue grenadier, pink ling, blue-eye trevalla and silver trevally were unable to be assessed using catch curves due to probable dome-shaped selectivity or high recruitment variability.

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Appendix 1 - details of values that were used as estimates of total Z (shown highlighted)

| Species | Pop | Flt | Year | Catch | CCType | lage | Ilen | Zage | Zlen | SSa | SSI |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ALFCCRes | All | AllMethods | 2007 | 175.43 | 1 | -99 | -99 | 0.23 | -99 | 224 | 0 |
| ALFCCRes | All | AllMethods | 2008 | 0 | 1 | -99 | -99 | 0.23 | -99 | 23 | 0 |
| ALFCCRes | All | AllMethods | 2009 | 14.197 | 1 | -99 | -99 | 0.23 | 0.53264 | 148 | 473 |
| ALFCCRes | All | AllMethods | 2010 | 0.0135 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| ALFCCRes | All | AllMethods | 2011 | 210.98 | 1 | -99 | -99 | 0.23 | 0.53264 | 640 | 4229 |
| DOJCCRes | All | NonTrawl | 2007 | 5.8498 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| DOJCCRes | All | NonTrawl | 2008 | 6.9068 | 1 | -99 | -99 | 0.49004 | 0.54451 | 611 | 804 |
| DOJCCRes | All | NonTrawl | 2009 | 6.0869 | 1 | -99 | -99 | 0.49004 | 0.54451 | 611 | 64 |
| DOJCCRes | All | NonTrawl | 2010 | 4.819 | 1 | -99 | -99 | 0.49004 | 0.54451 | 611 | 450 |
| DOJCCRes | All | NonTrawl | 2011 | 11.045 | 1 | -99 | -99 | 0.49004 | 0.54451 | 611 | 525 |
| DOJCCRes | All | Trawl | 2007 | 53.742 | 1 | -99 | -99 | 0.37024 | 0.42608 | 611 | 1062 |
| DOJCCRes | All | Trawl | 2008 | 106.21 | 1 | -99 | -99 | 0.37024 | 0.42608 | 611 | 1573 |
| DOJCCRes | All | Trawl | 2009 | 84.209 | 1 | -99 | -99 | 0.37024 | 0.42608 | 611 | 3363 |
| DOJCCRes | All | Trawl | 2010 | 55.143 | 1 | -99 | -99 | 0.37024 | 0.42608 | 611 | 2603 |
| DOJCCRes | All | Trawl | 2011 | 60.308 | 1 | -99 | -99 | 0.37024 | 0.42608 | 611 | 1890 |
| DOMCCRes | All | ETrawl | 2007 | 204.78 | 1 | -99 | -99 | 0.31024 | 0.44692 | 1634 | 954 |
| DOMCCRes | All | ETrawl | 2008 | 326.32 | 1 | -99 | -99 | 0.31024 | 0.44692 | 1634 | 1600 |
| DOMCCRes | All | ETrawl | 2009 | 343.91 | 1 | -99 | -99 | 0.31024 | 0.44692 | 1634 | 2318 |
| DOMCCRes | All | ETrawl | 2010 | 389.14 | 1 | -99 | -99 | 0.31024 | 0.44692 | 1634 | 2657 |
| DOMCCRes | All | ETrawl | 2011 | 354.12 | 1 | -99 | -99 | 0.31024 | 0.44692 | 1634 | 1757 |
| DOMCCRes | All | WTrawl | 2007 | 66.706 | 1 | -99 | -99 | 0.31024 | 0.47958 | 1634 | 386 |
| DOMCCRes | All | WTrawl | 2008 | 66.053 | 1 | -99 | -99 | 0.31024 | 0.47958 | 1634 | 75 |
| DOMCCRes | All | WTrawl | 2009 | 131.07 | 1 | -99 | -99 | 0.31024 | 0.47958 | 1634 | 299 |
| DOMCCRes | All | WTrawl | 2010 | 187.75 | 1 | -99 | -99 | 0.31024 | 0.47958 | 1634 | 628 |
| DOMCCRes | All | WTrawl | 2011 | 161.89 | 1 | -99 | -99 | 0.31024 | 0.47958 | 1634 | 658 |
| FLDCCRes | All | Trawl | 2007 | 980.85 | 1 | -99 | -99 | 0.54902 | 0.53618 | 650 | 299 |
| FLDCCRes | All | Trawl | 2008 | 783.44 | 1 | -99 | -99 | 0.54902 | 0.53618 | 554 | 467 |
| FLDCCRes | All | Trawl | 2009 | 834.01 | 1 | -99 | -99 | 0.54902 | 0.53618 | 465 | 13911 |
| FLDCCRes | All | Trawl | 2010 | 916.38 | 1 | -99 | -99 | 0.54902 | 0.53618 | 290 | 2502 |
| FLDCCRes | All | Trawl | 2011 | 248.71 | 1 | -99 | -99 | -99 | 0.53618 | 0 | 1006 |
| FLTCCRes | All | DSeine | 2007 | 1310.7 | 1 | -99 | -99 | 0.44778 | 0.40306 | 474 | 2098 |
| FLTCCRes | All | DSeine | 2008 | 1321.4 | 1 | -99 | -99 | 0.44778 | 0.40306 | 714 | 466 |
| FLTCCRes | All | DSeine | 2009 | 1221.4 | 1 | -99 | -99 | 0.44778 | 0.40306 | 1093 | 1100 |
| FLTCCRes | All | DSeine | 2010 | 1231.4 | 1 | -99 | -99 | 0.44778 | 0.40306 | 1134 | 1429 |
| FLTCCRes | All | DSeine | 2011 | 1170 | 1 | -99 | -99 | 0.44778 | 0.40306 | 1130 | 2369 |
| FLTCCRes | All | ETrawl | 2007 | 1149.6 | 1 | -99 | -99 | 0.40197 | 0.3163 | 474 | 4267 |
| FLTCCRes | All | ETrawl | 2008 | 1390.3 | 1 | -99 | -99 | 0.40197 | 0.3163 | 714 | 1614 |
| FLTCCRes | All | ETrawl | 2009 | 1126.4 | 1 | -99 | -99 | 0.40197 | 0.3163 | 1093 | 2109 |
| FLTCCRes | All | ETrawl | 2010 | 1157.1 | 1 | -99 | -99 | 0.40197 | 0.3163 | 1134 | 4016 |
| FLTCCRes | All | ETrawl | 2011 | 1157.9 | 1 | -99 | -99 | 0.40197 | 0.3163 | 1130 | 2942 |
| FLTCCRes | All | TasTrawl | 2007 | 177.23 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |


| FLTCCRes | All | TasTrawl | 2008 | 175.73 | 1 | -99 | -99 | 0.28 | 0.28 | 714 | 101 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FLTCCRes | All | TasTrawl | 2009 | 102.03 | 1 | -99 | -99 | 0.28 | 0.28 | 1093 | 176 |
| FLTCCRes | All | TasTrawl | 2010 | 105.25 | 1 | -99 | -99 | 0.28 | 0.28 | 1134 | 303 |
| FLTCCRes | All | TasTrawl | 2011 | 132.18 | 1 | -99 | -99 | 0.28 | 0.28 | 1130 | 538 |
| GEECCRes | East | NonTrawl | 2007 | 7.5442 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| GEECCRes | East | NonTrawl | 2008 | 15.908 | 1 | -99 | -99 | 0.48 | 0.64845 | 625 | 37 |
| GEECCRes | East | NonTrawl | 2009 | 11.966 | 1 | -99 | -99 | 0.48 | -99 | 396 | 0 |
| GEECCRes | East | NonTrawl | 2010 | 12.26 | 1 | -99 | -99 | 0.48 | 0.64845 | 580 | 122 |
| GEECCRes | East | NonTrawl | 2011 | 7.258 | 1 | -99 | -99 | 0.48 | 0.64845 | 626 | 551 |
| GEECCRes | East | SumTrawl | 2007 | 18.601 | 1 | -99 | -99 | 0.48296 | 0.48405 | 636 | 299 |
| GEECCRes | East | SumTrawl | 2008 | 27.524 | 1 | -99 | -99 | 0.48296 | 0.48405 | 625 | 1109 |
| GEECCRes | East | SumTrawl | 2009 | 18.679 | 1 | -99 | -99 | 0.48296 | 0.48405 | 396 | 218 |
| GEECCRes | East | SumTrawl | 2010 | 15.054 | 1 | -99 | -99 | 0.48296 | 0.48405 | 580 | 835 |
| GEECCRes | East | SumTrawl | 2011 | 13.643 | 1 | -99 | -99 | 0.48296 | 0.48405 | 626 | 1281 |
| GEECCRes | East | WinTrawl | 2007 | 49.447 | 1 | -99 | -99 | 0.48688 | 0.48006 | 636 | 46 |
| GEECCRes | East | WinTrawl | 2008 | 79.051 | 1 | -99 | -99 | 0.48688 | 0.48006 | 625 | 478 |
| GEECCRes | East | WinTrawl | 2009 | 43.522 | 1 | -99 | -99 | 0.48688 | 0.48006 | 396 | 366 |
| GEECCRes | East | WinTrawl | 2010 | 48.673 | 1 | -99 | -99 | 0.48688 | 0.48006 | 580 | 471 |
| GEECCRes | East | WinTrawl | 2011 | 31.131 | 1 | -99 | -99 | 0.48688 | 0.48006 | 626 | 1201 |
| GEWCCRes | All | NonTrawl | 2007 | 7.0588 | 1 | -99 | -99 | 0.63032 | 0.48031 | 722 | 17 |
| GEWCCRes | All | NonTrawl | 2008 | 9.5857 | 1 | -99 | -99 | 0.63032 | 0.48031 | 625 | 76 |
| GEWCCRes | All | NonTrawl | 2009 | 7.2278 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| GEWCCRes | All | NonTrawl | 2010 | 9.7676 | 1 | -99 | -99 | 0.63032 | 0.48031 | 1167 | 34 |
| GEWCCRes | All | NonTrawl | 2011 | 12.738 | 1 | -99 | -99 | 0.63032 | 0.48031 | 925 | 321 |
| GEWCCRes | All | Trawl40 | 2007 | 4.9823 | 1 | -99 | -99 | 0.48297 | 0.94 | 722 | 27 |
| GEWCCRes | All | Trawl40 | 2008 | 4.105 | 1 | -99 | -99 | 0.48297 | 0.94 | 625 | 105 |
| GEWCCRes | All | Trawl40 | 2009 | 5.2583 | 1 | -99 | -99 | 0.48297 | 0.94 | 1002 | 129 |
| GEWCCRes | All | Trawl40 | 2010 | 11.585 | 1 | -99 | -99 | 0.48297 | 0.94 | 1167 | 137 |
| GEWCCRes | All | Trawl40 | 2011 | 14.856 | 1 | -99 | -99 | 0.48297 | 0.94 | 925 | 334 |
| GEWCCRes | All | Trawl50 | 2007 | 58.041 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| GEWCCRes | All | Trawl50 | 2008 | 53 | 1 | -99 | -99 | 0.48008 | 0.48154 | 625 | 112 |
| GEWCCRes | All | Trawl50 | 2009 | 54.167 | 1 | -99 | -99 | 0.48008 | 0.48154 | 1002 | 420 |
| GEWCCRes | All | Trawl50 | 2010 | 78.374 | 1 | -99 | -99 | 0.48008 | 0.48154 | 1167 | 729 |
| GEWCCRes | All | Trawl50 | 2011 | 44.479 | 1 | -99 | -99 | 0.48008 | 0.48154 | 925 | 118 |
| GEWCCRes | All | GABTrawl | 2007 | 324.63 | 1 | -99 | -99 | 0.48 | 0.48 | 722 | 29 |
| GEWCCRes | All | GABTrawl | 2008 | 99.371 | 1 | -99 | -99 | 0.48 | 0.48 | 625 | 117 |
| GEWCCRes | All | GABTrawl | 2009 | 48.961 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| GEWCCRes | All | GABTrawl | 2010 | 42.731 | 1 | -99 | -99 | 0.48 | 0.48 | 1167 | 140 |
| GEWCCRes | All | GABTrawl | 2011 | 17.589 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| GRECCRes | All | TrawlSpawn | 2007 | 1815.4 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| GRECCRes | All | TrawlSpawn | 2008 | 2837.6 | 1 | -99 | -99 | 0.19912 | 0.19905 | 1848 | 116 |
| GRECCRes | All | TrawlSpawn | 2009 | 2712.2 | 1 | -99 | -99 | 0.19912 | 0.19905 | 2086 | 7999 |
| GRECCRes | All | TrawlSpawn | 2010 | 3384.3 | 1 | -99 | -99 | 0.19912 | 0.19905 | 1642 | 5265 |
| GRECCRes | All | TrawlSpawn | 2011 | 3553.8 | 1 | -99 | -99 | 0.19912 | 0.19905 | 2007 | 7484 |
| GRECCRes | All | TrawlSummer | 2007 | 1229.3 | 1 | -99 | -99 | 0.36706 | 0.48737 | 1574 | 1028 |

Yield, Total Mortality Values and Tier 3 Estimates

| GRECCRes | All | TrawlSummer | 2008 | 1305.8 | 1 | -99 | -99 | 0.36706 | 0.48737 | 1848 | 1895 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| GRECCRes | All | TrawlSummer | 2009 | 1144.7 | 1 | -99 | -99 | 0.36706 | 0.48737 | 2086 | 2979 |
| GRECCRes | All | TrawlSummer | 2010 | 1158.1 | 1 | -99 | -99 | 0.36706 | 0.48737 | 1642 | 2499 |
| GRECCRes | All | TrawlSummer | 2011 | 913.14 | 1 | -99 | -99 | 0.36706 | 0.48737 | 2007 | 3321 |
| GRWCCRes | All | TrawlSummer | 2007 | 58.266 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| GRWCCRes | All | TrawlSummer | 2008 | 3.321 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| GRWCCRes | All | TrawlSummer | 2009 | 0.5625 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| GRWCCRes | All | TrawlSummer | 2010 | 5.145 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| GRWCCRes | All | TrawlSummer | 2011 | 4.8425 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| GRWCCRes | All | TrawlSpawn | 2007 | 19.055 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| GRWCCRes | All | TrawlSpawn | 2008 | 0.27 | 1 | -99 | -99 | 0.53872 | 0.26326 | 1848 | 48 |
| GRWCCRes | All | TrawlSpawn | 2009 | 0 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| GRWCCRes | All | TrawlSpawn | 2010 | 0.5925 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| GRWCCRes | All | TrawlSpawn | 2011 | 0.5025 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| LIGCCRes | East | NonTrawl | 2007 | 163.13 | 1 | -99 | -99 | 0.33366 | 0.278 | 558 | 275 |
| LIGCCRes | East | NonTrawl | 2008 | 231.94 | 1 | -99 | -99 | 0.33366 | 0.278 | 910 | 256 |
| LIGCCRes | East | NonTrawl | 2009 | 159.38 | 1 | -99 | -99 | 0.33366 | 0.278 | 1122 | 1027 |
| LIGCCRes | East | NonTrawl | 2010 | 140.36 | 1 | -99 | -99 | 0.33366 | 0.278 | 1036 | 2177 |
| LIGCCRes | East | NonTrawl | 2011 | 159.38 | 1 | -99 | -99 | 0.33366 | 0.278 | 1296 | 483 |
| LIGCCRes | East | Trawl | 2007 | 262.31 | 1 | -99 | -99 | 0.52156 | 0.51698 | 558 | 1218 |
| LIGCCRes | East | Trawl | 2008 | 379.83 | 1 | -99 | -99 | 0.52156 | 0.51698 | 910 | 2064 |
| LIGCCRes | East | Trawl | 2009 | 245.24 | 1 | -99 | -99 | 0.52156 | 0.51698 | 1122 | 2321 |
| LIGCCRes | East | Trawl | 2010 | 298.67 | 1 | -99 | -99 | 0.52156 | 0.51698 | 1036 | 2523 |
| LIGCCRes | East | Trawl | 2011 | 331.68 | 1 | -99 | -99 | 0.52156 | 0.51698 | 1296 | 1155 |
| LIGCCRes | West | NonTrawl | 2007 | 63.409 | 1 | -99 | -99 | 0.61971 | 0.27802 | 225 | 727 |
| LIGCCRes | West | NonTrawl | 2008 | 38.447 | 1 | -99 | -99 | 0.61971 | -99 | 45 | 0 |
| LIGCCRes | West | NonTrawl | 2009 | 52.321 | 1 | -99 | -99 | 0.61971 | 0.27802 | 88 | 21 |
| LIGCCRes | West | NonTrawl | 2010 | 93.7 | 1 | -99 | -99 | 0.61971 | 0.27802 | 177 | 212 |
| LIGCCRes | West | NonTrawl | 2011 | 145.46 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| LIGCCRes | West | Trawl | 2007 | 295.7 | 1 | -99 | -99 | 0.36683 | 0.38321 | 225 | 828 |
| LIGCCRes | West | Trawl | 2008 | 226.06 | 1 | -99 | -99 | 0.36683 | 0.38321 | 45 | 133 |
| LIGCCRes | West | Trawl | 2009 | 271.59 | 1 | -99 | -99 | 0.36683 | 0.38321 | 88 | 231 |
| LIGCCRes | West | Trawl | 2010 | 281.67 | 1 | -99 | -99 | 0.36683 | 0.38321 | 177 | 491 |
| LIGCCRes | West | Trawl | 2011 | 368.64 | 1 | -99 | -99 | 0.36683 | 0.38321 | 95 | 182 |
| LIGCCRes | GAB | NonTrawl | 2007 | 75.354 | 1 | -99 | -99 | -99 | 0.278 | 0 | 107 |
| LIGCCRes | GAB | NonTrawl | 2008 | 102.21 | 1 | -99 | -99 | 0.278 | 0.278 | 69 | 100 |
| LIGCCRes | GAB | NonTrawl | 2009 | 45.784 | 1 | -99 | -99 | 0.278 | 0.278 | 93 | 98 |
| LIGCCRes | GAB | NonTrawl | 2010 | 86.363 | 1 | -99 | -99 | 0.278 | 0.278 | 99 | 1127 |
| LIGCCRes | GAB | NonTrawl | 2011 | 72.47 | 1 | -99 | -99 | 0.278 | 0.278 | 113 | 374 |
| LIGCCRes | GAB | Trawl | 2007 | 15.859 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| LIGCCRes | GAB | Trawl | 2008 | 1.7864 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| LIGCCRes | GAB | Trawl | 2009 | 0.132 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| LIGCCRes | GAB | Trawl | 2010 | 4.699 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| LIGCCRes | GAB | Trawl | 2011 | 3.137 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| MOWCCRes | All | ETrawl | 2007 | 234.92 | 1 | -99 | -99 | 0.28158 | 0.37072 | 193 | 2302 |


| MOWCCRes | All | ETrawl | 2008 | 335.84 | 1 | -99 | -99 | 0.28158 | 0.37072 | 751 | 2475 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| MOWCCRes | All | ETrawl | 2009 | 243.4 | 1 | -99 | -99 | 0.28158 | 0.37072 | 620 | 2425 |
| MOWCCRes | All | ETrawl | 2010 | 199.27 | 1 | -99 | -99 | 0.28158 | 0.37072 | 892 | 1973 |
| MOWCCRes | All | ETrawl | 2011 | 185 | 1 | -99 | -99 | 0.28158 | 0.37072 | 855 | 1362 |
| MOWCCRes | All | DSeine | 2007 | 17.436 | 1 | -99 | -99 | 0.31493 | 0.25666 | 193 | 753 |
| MOWCCRes | All | DSeine | 2008 | 36.779 | 1 | -99 | -99 | 0.31493 | 0.25666 | 751 | 635 |
| MOWCCRes | All | DSeine | 2009 | 18.538 | 1 | -99 | -99 | 0.31493 | 0.25666 | 620 | 50 |
| MOWCCRes | All | DSeine | 2010 | 17.324 | 1 | -99 | -99 | 0.31493 | 0.25666 | 892 | 492 |
| MOWCCRes | All | DSeine | 2011 | 29.4 | 1 | -99 | -99 | 0.31493 | 0.25666 | 855 | 665 |
| MOWCCRes | All | TasTrawl | 2007 | 116.85 | 1 | -99 | -99 | 0.32535 | 0.86766 | 193 | 137 |
| MOWCCRes | All | TasTrawl | 2008 | 121.07 | 1 | -99 | -99 | 0.32535 | 0.86766 | 751 | 43 |
| MOWCCRes | All | TasTrawl | 2009 | 55.817 | 1 | -99 | -99 | 0.32535 | 0.86766 | 620 | 80 |
| MOWCCRes | All | TasTrawl | 2010 | 59.871 | 1 | -99 | -99 | 0.32535 | 0.86766 | 892 | 341 |
| MOWCCRes | All | TasTrawl | 2011 | 50.633 | 1 | -99 | -99 | 0.32535 | 0.86766 | 855 | 555 |
| MOWCCRes | All | WTrawl | 2007 | 121.82 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| MOWCCRes | All | WTrawl | 2008 | 104.28 | 1 | -99 | -99 | 0.16 | 0.29543 | 751 | 156 |
| MOWCCRes | All | WTrawl | 2009 | 64.952 | 1 | -99 | -99 | 0.16 | 0.29543 | 620 | 140 |
| MOWCCRes | All | WTrawl | 2010 | 40.549 | 1 | -99 | -99 | 0.16 | 0.29543 | 892 | 72 |
| MOWCCRes | All | WTrawl | 2011 | 85.874 | 1 | -99 | -99 | 0.16 | 0.29543 | 855 | 208 |
| MOWCCRes | All | GABTrawl | 2007 | 108.01 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| MOWCCRes | All | GABTrawl | 2008 | 89.765 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| MOWCCRes | All | GABTrawl | 2009 | 64.352 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| MOWCCRes | All | GABTrawl | 2010 | 39.148 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| MOWCCRes | All | GABTrawl | 2011 | 24.545 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| REBCCRes | All | Trawl | 2007 | 758.28 | 1 | -99 | -99 | 0.11 | 0.11 | 443 | 141 |
| REBCCRes | All | Trawl | 2008 | 664.9 | 1 | -99 | -99 | 0.11 | 0.11 | 561 | 716 |
| REBCCRes | All | Trawl | 2009 | 463.44 | 1 | -99 | -99 | 0.11 | 0.11 | 668 | 9093 |
| REBCCRes | All | Trawl | 2010 | 275.41 | 1 | -99 | -99 | 0.11 | 0.11 | 148 | 861 |
| REBCCRes | All | Trawl | 2011 | 65.017 | 1 | -99 | -99 | -99 | 0.11 | 0 | 714 |
| REDCCRes | North | Trawl | 2007 | 171.2 | 1 | -99 | -99 | 0.15487 | 0.16336 | 7310 | 646 |
| REDCCRes | North | Trawl | 2008 | 165.35 | 1 | -99 | -99 | 0.15487 | 0.16336 | 7310 | 583 |
| REDCCRes | North | Trawl | 2009 | 145.52 | 1 | -99 | -99 | 0.15487 | 0.16336 | 7310 | 791 |
| REDCCRes | North | Trawl | 2010 | 136.37 | 1 | -99 | -99 | 0.15487 | 0.16336 | 7310 | 1066 |
| REDCCRes | North | Trawl | 2011 | 76.48 | 1 | -99 | -99 | 0.15487 | 0.16336 | 7310 | 479 |
| REDCCRes | South | Trawl | 2007 | 40.809 | 1 | -99 | -99 | 0.11001 | 0.11 | 7677 | 28 |
| REDCCRes | South | Trawl | 2008 | 16.665 | 1 | -99 | -99 | 0.11001 | -99 | 7677 | 0 |
| REDCCRes | South | Trawl | 2009 | 12.444 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| REDCCRes | South | Trawl | 2010 | 13.213 | 1 | -99 | -99 | 0.11001 | 0.11 | 7677 | 160 |
| REDCCRes | South | Trawl | 2011 | 8.8647 | 1 | -99 | -99 | 0.11001 | 0.11 | 7677 | 3 |
| REGCCRes | All | NonTrawl | 2007 | 0.06 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| REGCCRes | All | NonTrawl | 2008 | 0 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| REGCCRes | All | NonTrawl | 2009 | 0.035 | 1 | -99 | -99 | -99 | 0.1112 | 0 | 77 |
| REGCCRes | All | NonTrawl | 2010 | 0 | 1 | -99 | -99 | -99 | 0.1112 | 0 | 342 |
| REGCCRes | All | NonTrawl | 2011 | 0.005 | 1 | -99 | -99 | -99 | 0.1112 | 0 | 14 |
| REGCCRes | All | Trawl | 2007 | 0.761 | 1 | -99 | -99 | -99 | 0.26111 | 0 | 6746 |
| REGCCRes | All | Trawl | 2008 | 0.844 | 1 | -99 | -99 | -99 | 0.26111 | 0 | 1414 |

Yield, Total Mortality Values and Tier 3 Estimates

| REGCCRes | All | Trawl | 2009 | 1.272 | 1 | -99 | -99 | -99 | 0.26111 | 0 | 2590 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| REGCCRes | All | Trawl | 2010 | 1.32 | 1 | -99 | -99 | -99 | 0.26111 | 0 | 2746 |
| REGCCRes | All | Trawl | 2011 | 1.39 | 1 | -99 | -99 | -99 | 0.26111 | 0 | 2400 |
| TBECCRes | All | NonTrawl | 2007 | 362.52 | 1 | -99 | -99 | 0.27235 | 0.375 | 340 | 151 |
| TBECCRes | All | NonTrawl | 2008 | 226.22 | 1 | -99 | -99 | 0.27235 | 0.375 | 557 | 624 |
| TBECCRes | All | NonTrawl | 2009 | 311.52 | 1 | -99 | -99 | 0.27235 | 0.375 | 960 | 1343 |
| TBECCRes | All | NonTrawl | 2010 | 238.74 | 1 | -99 | -99 | 0.27235 | 0.375 | 743 | 2016 |
| TBECCRes | All | NonTrawl | 2011 | 187.12 | 1 | -99 | -99 | -99 | 0.375 | 0 | 2420 |
| TBECCRes | All | Trawl | 2007 | 37.321 | 1 | -99 | -99 | 0.22274 | 0.26357 | 340 | 21 |
| TBECCRes | All | Trawl | 2008 | 35.899 | 1 | -99 | -99 | 0.22274 | 0.26357 | 557 | 37 |
| TBECCRes | All | Trawl | 2009 | 39.343 | 1 | -99 | -99 | 0.22274 | 0.26357 | 960 | 295 |
| TBECCRes | All | Trawl | 2010 | 44.302 | 1 | -99 | -99 | 0.22274 | 0.26357 | 743 | 95 |
| TBECCRes | All | Trawl | 2011 | 23.327 | 1 | -99 | -99 | -99 | 0.26357 | 0 | 162 |
| TRECCRes | All | NonTrawl | 2007 | 2.5192 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| TRECCRes | All | NonTrawl | 2008 | 1.9973 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| TRECCRes | All | NonTrawl | 2009 | 1.0371 | 1 | -99 | -99 | -99 | 0.61157 | 0 | 247 |
| TRECCRes | All | NonTrawl | 2010 | 25.114 | 1 | -99 | -99 | -99 | 0.61157 | 0 | 263 |
| TRECCRes | All | NonTrawl | 2011 | 0.2565 | 1 | -99 | -99 | -99 | 0.61157 | 0 | 212 |
| TRECCRes | All | Trawl | 2007 | 129.81 | 1 | -99 | -99 | -99 | 0.7984 | 0 | 1932 |
| TRECCRes | All | Trawl | 2008 | 101.86 | 1 | -99 | -99 | -99 | 0.7984 | 0 | 1135 |
| TRECCRes | All | Trawl | 2009 | 142.53 | 1 | -99 | -99 | -99 | 0.7984 | 0 | 2750 |
| TRECCRes | All | Trawl | 2010 | 203.27 | 1 | -99 | -99 | -99 | 0.7984 | 0 | 3085 |
| TRECCRes | All | Trawl | 2011 | 186.93 | 1 | -99 | -99 | -99 | 0.7984 | 0 | 1681 |
| TRSCCRes | All | AllMethods | 2007 | 1797.6 | 1 | -99 | -99 | 0.3157 | 1.29926 | 316 | 1392 |
| TRSCCRes | All | AllMethods | 2008 | 1378.1 | 1 | -99 | -99 | 0.3157 | 1.29926 | 547 | 1609 |
| TRSCCRes | All | AllMethods | 2009 | 1285.1 | 1 | -99 | -99 | 0.3157 | 1.29926 | 821 | 3521 |
| TRSCCRes | All | AllMethods | 2010 | 1188.8 | 1 | -99 | -99 | 0.3157 | 1.29926 | 822 | 4001 |
| TRSCCRes | All | AllMethods | 2011 | 1106.5 | 1 | -99 | -99 | 0.3157 | 1.29926 | 852 | 2861 |
| TRTCCRes | All | NonTrawl | 2007 | 1.6461 | 1 | -99 | -99 | 0.46011 | 1.44933 | 123 | 301 |
| TRTCCRes | All | NonTrawl | 2008 | 6.6546 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| TRTCCRes | All | NonTrawl | 2009 | 3.8073 | 1 | -99 | -99 | 0.46011 | 1.44933 | 274 | 357 |
| TRTCCRes | All | NonTrawl | 2010 | 11.384 | 1 | -99 | -99 | 0.46011 | 1.44933 | 428 | 900 |
| TRTCCRes | All | NonTrawl | 2011 | 4.2449 | 1 | -99 | -99 | -99 | 1.44933 | 0 | 985 |
| TRTCCRes | All | Trawl | 2007 | 170.06 | 1 | -99 | -99 | 0.69743 | 0.86593 | 123 | 428 |
| TRTCCRes | All | Trawl | 2008 | 159.26 | 1 | -99 | -99 | 0.69743 | 0.86593 | 597 | 1282 |
| TRTCCRes | All | Trawl | 2009 | 117.14 | 1 | -99 | -99 | 0.69743 | 0.86593 | 274 | 2412 |
| TRTCCRes | All | Trawl | 2010 | 123.18 | 1 | -99 | -99 | 0.69743 | 0.86593 | 428 | 2123 |
| TRTCCRes | All | Trawl | 2011 | 83.546 | 1 | -99 | -99 | -99 | 0.86593 | 0 | 1785 |
| WHSCCRes | All | NonTrawl | 2007 | 444 | 1 | -99 | -99 | 1.19757 | 1.5321 | 415 | 2558 |
| WHSCCRes | All | NonTrawl | 2008 | 393.34 | 1 | -99 | -99 | 1.19757 | 1.5321 | 479 | 894 |
| WHSCCRes | All | NonTrawl | 2009 | 425.43 | 1 | -99 | -99 | 1.19757 | 1.5321 | 421 | 880 |
| WHSCCRes | All | NonTrawl | 2010 | 360.1 | 1 | -99 | -99 | 1.19757 | 1.5321 | 620 | 1179 |
| WHSCCRes | All | NonTrawl | 2011 | 308.17 | 1 | -99 | -99 | 1.19757 | 1.5321 | 581 | 1222 |
| WHSCCRes | All | Trawl | 2007 | 85.053 | 1 | -99 | -99 | 1.36906 | 1.5969 | 415 | 1056 |
| WHSCCRes | All | Trawl | 2008 | 69.16 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |


|  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | ---: | ---: | ---: | ---: |
| WHSCCRes | All | Trawl | 2009 | 29.744 | 1 | -99 | -99 | 1.36906 | 1.5969 | 421 | 288 |
| WHSCCRes | All | Trawl | 2010 | 38.438 | 1 | -99 | -99 | -99 | -99 | 0 | 0 |
| WHSCCRes | All | Trawl | 2011 | 50.689 | 1 | -99 | -99 | 1.36906 | 1.5969 | 581 | 435 |

# 20. Tier 4 Analyses in the SESSF, including Deep Water Species. Data from 1986-2011 

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### 20.1 Summary

Thirty four TIER 4 analyses are documented here which included a number of species where spatial information was available (Blue Warehou and Mirror Dory) leading to analyses for the east and west presumed stock regions. There are also Tier 4 analyses for some species where discard estimates were included in the analysis of catch rates. In addition, some non-key commercial species were assessed at the RAG's request, at a target assuming a proxy of $40 \% B_{0}$ as well as a proxy target assuming $48 \% B_{0}$.

Six fisheries are assessed using Tier 4 methodology: BlueEye Trevalla, Blue Warehou (split east and west), Inshore Ocean Perch and Offshore Ocean Perch, Redfish, Royal Red Prawns, and Silver Trevally. Three fisheries generated zero RBCs and these were Blue Warehou, Jackass Morwong and Redfish.

Alternative analyses were provided for Redfish and Inshore Ocean Perch in which discards were included in the estimation of the catch rate trends. The inclusion of discards in estimating catch rates adds a great deal of noise to the CPUE trends so the uncertainty in these analyses expands. At the same time it is not clear whether to remove the discards from the RBC to generate a TAC or not. The use of this approach for setting RBCs needs further discussion and examination.

The TIER 4 harvest control rule is applied to species for which there is no reliable information on either current biomass levels or current exploitation rates. Ideally, in line with the notion of being more precautionary in the absence of information, the outcome from these analyses should be more conservative than those available from higher TIER analyses; this is now explicitly implemented by imposing a $15 \%$ discount factor on the RBC as a precautionary measure, unless there are good reasons for not imposing such an discount on particular species. The default procedure will now be to apply the discount factor unless RAGs generate advice that alternative and equivalent precautionary measures are in place (such as spatial or temporal closures) or that there is evidence of historical stability of the stock at current catch levels.

TIER 4 analyses require, as a minimum, knowledge of the time series of total catches and of catch rates, either standardized or simple geometric mean catch rates. This year, only standardized catch rates were used except where discards were explicitly included in the analyses.

The TIER 4 analyses conducting this year used the analytical method developed and tested in 2008 and 2009. This has the capacity to provide advice that will manage a fishery in such a manner that it should achieve the target catch rate derived from the chosen reference period. However, the TIER 4 control rule can only succeed if catch
rates do in fact reflect stock size. Many factors could contribute to make this assumption fail so care needs to be taken when applying this control rule.

To ensure consistency and provide for efficient operation once data becomes available, standard analyses were set up in the statistical software, R , which provided the results as the tables and graphs required for the TIER4 analyses. Both the data and results for each analysis are presented for transparency. The TIER 4 harvest control rule formulation essentially uses a ratio of current catch rates with respect to selected limit and target reference points to calculate a scaling factor. This scaling factor is applied to the target catch to generate an RBC. In all cases where individual attention was required by a particular analysis it was more difficult to automate analyses and these therefore took a disproportionate amount of time.

### 20.2 Summary of RBCs and Discards

The Recommended Biological Catch from this year's analyses are compared (Table 20.1) with those from the previous three years (Haddon, 2010, 2011b). Blue Warehou and Mirror Dory are sub-divided spatially as east and west. Those species where the Tier 4 rule is not used to set a TAC have the RBC , given in the specific sections throughout the document, replaced with NA.

The upper group of species are those whose TAC is determined using the Tier 4 and the lower group the remainder.

In addition, this year, a number of species were assessed using a proxy target of $48 \%$ and of $40 \%$ B0. In all such cases the RBC, if it starts above zero, will increase simply because the ratio of the current average catch rate to the new proxy target of $40 \%$ will be greater than the ratio at a proxy target of $48 \%$ (Table 20.2).

Table 20.1. TIER 4 outcomes by species. The RBC in tonnes, while the weighted discards are a percentage. RBC09 are the 2009 estimates and RBC10 are this year's estimates. For those species where the total catches have been sub-divided (Blue Warehou, Silver Trevally, Ocean Perch, and Mirror Dory) the sub-division of catches and discards was done using the ratio of catches, by the respective areas, observed in the catch effort database. Discards $t$ is the weighted estimate of the discards in 2013.

| Species | RBC09 | RBC10 | RBC11 | RBC12 | Discard t |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Blue Eye Trevalla AL | 536 | 521 | 415 | 288 | 4.022 |
| Blue Warehou | 0 | 0 | 0 | 0 | 44.954 |
| Blue Warehou East | 0 | 0 | 0 | 0 | 6.763 |
| Blue Warehou West | 0 | 0 | 0 | 0 | 38.191 |
| Ocean Perch Inshore | 25 | 26 | 35 | 43 | 193.067 |
| Ocean Perch Inshore D | NA | NA | 95 | 126 | 193.067 |
| Ocean Perch Offshore | 219 | 193 | 215 | 196 | 36.906 |
| Redfish | 62 | 0 | 0 | 0 | 52.107 |
| Redfish Discards | NA | NA | 0 | 0 | 52.107 |
| Royal Red Prawn | 336 | 351 | 276 | 352 | 6.672 |
| Silver Trevally | 649 | 754 | 863 | 980 | 6.582 |
| Deep Water Taxa |  |  |  |  |  |
| Cascade Smooth Oreo |  |  | Catch $<10 t$ | 12.3 |  |
| Non-Cascade Smooth Oreo |  |  | Catch $<10 t$ | 12.3 |  |
| Mixed Oreos |  |  |  | 132.213 | 16.2 |
| Eastern Deepwater Sharks |  |  | Catch $<10 t$ | 2.8 |  |
| Western Deepwater Sharks |  |  | Catch $<10 t$ | 2.8 |  |
| llfonsino |  |  |  | NA | 0.0 |
| Non-Tier4 Species |  |  |  |  |  |
| Blue Grenadier | 639 | 729 | 645 | NA | 381.989 |
| Flathead | 2684 | 3071 | 3129 | NA | 353.129 |
| Gemfish Eastern | 324 | 150 | 225 | NA | 141.554 |
| Gemfish Western | 102 | 93 | 109 | NA | 80.108 |
| Jackass Morwong | 0 | 0 | 0 | NA | 43.680 |
| John Dory | 19 | 35 | 25 | NA | 22.597 |
| Mirror Dory | 381 | 422 | 423 | NA | 0.350 |
| Mirror Dory East | NA | 569 | 544 | NA | 182.545 |
| Mirror Dory West | NA | NA | 161 | NA | 78.441 |
| Pink Ling | 347 | 337 | 320 | NA | 30.850 |
| Ribaldo | 160 | 202 | 197 | NA | 4.954 |
| School Whiting | 1213 | 1236 | 1212 | NA | 36.575 |
| Silver Warehou | 1690 | 1507 | 1348 | NA | 240.987 |


| Table 20.2. Comparison of the calculated RBCs for those species/stock |
| :--- | ---: | :--- | ---: |
| combinations that were assessed using catch rates that included the effects of |
| discards and that used alternative proxy targets of $48 \%$ and $40 \% B_{0}$. |
| Species/Stock Proxy Target $\%$ CPUE RBC <br> Inshore Ocean Perch 48 Discard 125.661 <br> Inshore Ocean Perch 40 Discard 173.993 <br> Offshore Ocean Perch 48 Discard 204.026 <br> Offshore Ocean Perch 40 Discard 282.500 <br> Offshore Ocean Perch 48  196.498 <br> Offshore Ocean Perch 40  272.077 <br> John Dory 48 NA  <br> John Dory 40  NA <br> Redfish 48  NA <br> Redfish 48 Discard NA <br> Ribaldo 48  232.054 <br> Ribaldo 40  321.309 |

### 20.3 Introduction

### 20.3.1 Tier 4 Harvest Control Rule

The TIER 4 harvest control rules are the default procedure applied to species for which only limited information is available; specifically no reliable information on either current biomass levels or current exploitation rates.

Ideally, in line with the notion of being more precautionary in the absence of information, the outcome from these analyses should be more conservative than those available from higher TIER analyses; this is now explicitly implemented by imposing a $15 \%$ discount factor on the RBC as a precautionary measure unless there are good reasons for not imposing such a discount on particular species. The application of the discount factor will occur unless RAGs generate explicit advice that alternative equivalent precautionary measures are in place (such as spatial or temporal closures) or that there is evidence of historical stability of the stock at current catch levels (AFMA, 2009).

In essence TIER 4 analyses require, as a minimum, a time series of total catches and of standardized catch rates.

The current TIER 4 analysis and control rule underwent Management Strategy Evaluation (Wayte, 2009), which demonstrated its advantages over an earlier implementation used in 2007 and 2008. Further work has since demonstrated that as long as there is a limit on increases and decreases to the RBC of no more than $50 \%$ then the notion of including a maximum RBC (at 1.25 times the target) is redundant (Little et $a l, 2011)$.

### 20.4 Methods

### 20.4.1 TIER 4 Harvest Control Rule

The data required are time series of catches and catch rates. The analyses have been conducted on total catches across the entire SESSF (including State catches, SEF2 landing records, and any discards). For some species, where there is only a single stock and a single primary fishing method, analyses are presented using standardized CPUE data (Haddon, 2012). For other species, there may be multiple stocks or areas or multiple methods and selecting which time series of catch rates to use in the analyses is not always straightforward. In those cases, the standardized time series for the method now accounting for the majority of current catch was used.

All 2010 data relating to catches and discards, from both State waters and SEF2 data sets, were provided by AFMA, with initial processing by Dr Neil Klaer and Dr Judy Upston of CSIRO. All catch rate data were derived from the standard commercial catch and effort database processed from the AFMA data by Mike Fuller of CSIRO Hobart.

Standard analyses were set up in the statistical software, R (2009), which provided the tables and graphs required for the TIER4 analyses. The data and results for each analysis are presented for transparency. The TIER 4 harvest control rule formulation essentially uses a ratio of current catch rates with respect to the selected limit and target reference points to calculate a scaling factor for the current year $\left(S F_{t}\right)$. This scaling
factor is applied to the target catch to generate an RBC. To generate a TAC, known discards and State catches are first removed and then, if applicable, the $15 \%$ discount is applied. The TAC calculations are conducted by AFMA. This report focusses on providing the estimates of the Recommended Biological Catches.

$$
\begin{gather*}
\text { Scaling Factor }=S F_{t}=\max \left(0, \frac{\overline{C P U E}-C P U E_{\lim }}{C P U E_{\mathrm{targ}}-C P U E_{\mathrm{lim}}}\right)  \tag{1}\\
R B C=C_{\operatorname{targ}} \times S F_{t} \tag{2}
\end{gather*}
$$

If new data becomes available, for example, more State data has become available this year, or other large changes occur in the catch rates then the RBC could undergo large changes. Such changes are constrained by the following limits:

$$
\begin{array}{l|l}
R B C_{y}=1.5 R B C_{y-1} & R B C_{y}>1.5 R B C_{y-1}  \tag{3}\\
R B C_{y}=0.5 R B C_{y-1} & R B C_{y}<0.5 R B C_{y-1}
\end{array}
$$

where
$R B C_{y}$ is the RBC in year $y$
${ }^{C P U E} E_{\text {targ }}$ is the target CPUE for the species; Eq. (5)
$\underline{C P U E}$ lim is the limit CPUE for the species $=0.4 * \mathrm{CPUE}_{\text {targ }}$
$\overline{C P U E}$ the average CPUE over the past $m$ years; $m$ tends to be the most recent four years.
$C_{\text {targ }} \quad$ is a catch target derived from a period of historical catch that has been identified as a desirable target in terms of CPUE, catches and status of the fishery, e.g. 1986 - 1995 (Table 20.3). This is an average of the total removals for the selected reference period, including any discards; Eq. (4).

$$
\begin{equation*}
C_{\mathrm{targ}}=\frac{\sum_{y=y r 1}^{y r 2} L_{y}}{(y r 2-y r 1+1)} \tag{4}
\end{equation*}
$$

where $L_{y}$ represents the landings in year $y$.

$$
\begin{equation*}
C P U E_{\text {targ }}=\frac{\sum_{y=y r 1}^{y r 2} C P U E_{y}}{(y r 2-y r 1+1)} \tag{5}
\end{equation*}
$$

where $C P U E_{y}$ is the catch rate in year $y, y r 2$ and $y r 1$ represent the last and the first years in the reference period respectively.

For each species a table of landings and of standardized catch rates was assembled. These included all catches (Commonwealth landings, Non-trawl catches, combined State catches, and discards). The State catches are available back to 1994 and non-trawl catches are from 1998. Catches prior to 1994 are either taken from an historical catch database or, if no data is available for the species, then they are taken from the AFMA

GenLog Catch and Effort database. The catch rates are standardized, usually from 1986, using methods described in Haddon (2012).

Percent discards are estimated from ISMP observations from 1998 to the current year. Discards for earlier years, prior to ISMP sampling, are estimated by taking the overall average percent discard from 1998 to the 2006 and applying that discard rate to the reported landings for the earlier years. The year 2006 was selected as the final year as discarding practices altered at about that time following the structural adjustment and the introduction of the Harvest Strategy Policy. For Eastern Gemfish the average discard rate was determined for 1998-2002 to allow for the non-target nature of the fishery following 2002. The calculation of the earlier discards is done so that the total catches can be estimated even though only the landed catches are available. To calculate the discards for a given year we used

$$
\begin{equation*}
D_{y}=\frac{C_{y} \bar{D}_{98-06}}{\left(1-\bar{D}_{98-06}\right)} \tag{6}
\end{equation*}
$$

Discard proportions for the projected year for which the RBC is being calculated are taken as a weighted mean of the previous four years:

$$
\begin{equation*}
\mathrm{D}_{\mathrm{CUR}}=\left(1.0 \mathrm{D}_{y-1}+0.5 \mathrm{D}_{y-2}+0.25 \mathrm{D}_{y-3}+0.125 \mathrm{D}_{y-4}\right) / 1.875 \tag{7}
\end{equation*}
$$

Where $D_{C U R}$ is the estimated discard rate for the coming year $y, D_{y-1}$ is the discards rate in year $y-1$. The discard rate in year $y$ is the ratio of discards to the sum of landed catches plus those discards (this can vary between $0-100 \%$ ):

$$
\begin{equation*}
D_{y}=\frac{\text { Discard }_{y}}{\left(\text { Catches }_{y}+\text { Discard }_{y}\right)} \tag{8}
\end{equation*}
$$

For each species, reference years were selected by the RAGs to generate estimates of target catches and target catch rates. In addition, a decision was required as to whether the fishery could be considered as fully developed or otherwise (Table 20.3). Where a fishery was not considered to be fully developed the target catch rate, $C P U E_{\text {targ }}$, was divided by two as a proxy for expected changes to catch rates as the fishery develops and the resource stock size declines towards the target of $48 \%$ unfished biomass.

Plots are given of the total removals illustrating the target catch level. In addition, the standardized catch rates are illustrated with the target catch rate and the limit catch rate. Finally, where the data is available, plots are given of the Total removals contrasted with State removals, and of discards and non-trawl catches.

### 20.4.2 Data Manipulations

The default reference years were 1986-1995, but various species required different reference years to account for the specific development of each fishery; these are noted in each analysis. In addition, Silver Warehou and Ribaldo were two fisheries where the state of development was such that the exhibited catch rates were unlikely to be representative of a developed fishery and so the target catch rates were halved; these details are provided in Table 20.3.

### 20.4.3 The Inclusion of Discards

Some species, especially redfish (Centroberyx affinis) and inshore Ocean Perch (Helicolenus percoides), have experienced high levels of discarding but the reported catch rates relate only to the estimated landed weights. In those species where discarding makes up a significant proportion of the catch (in some years more redfish were discarded than landed and more inshore ocean perch tend to be discarded than landed) it is reasonable to ask how the discards would have affected catch rates. This is an important question because standardized commercial catch rates are used in Australian stock assessments as an index of relative abundance (Haddon, 2010a, b); if ignoring discards leads to a consistent bias this could affect the outcome of the assessments and thus, the assessments should become aware of the effects of discards.

Catch rates are used in assessments as an index of relative abundance through time and it is the trends exhibited by the catch rates that are important rather than their absolute values. If the discard levels are relatively constant through time and evenly distributed amongst the fleet, then their inclusion would not be expected to influence the trends in catch rates except to add noise. In all cases the discard rates are estimates based on subsampling the fleet of vessels. That the estimates are uncertain can be seen simply by considering the summary data tables in this document; where discards rates are not low they are very variable between years. Redfish provide an extreme where in 1998 the estimate was 2324 t , which was nearly $56 \%$ of the total catch, while in 1999 discards estimated at only 69 t , making up on about $5 \%$ of the total catch. So in those cases where discard levels are low, adding ion discards to the estimation of catch rates is not expected to alter outcomes.

For those species, such as redfish and ocean perch, where discard rates are much higher it was decided to include those estimated catches to determine their effect on the outcome of the Tier 4 analyses. In 2010 it was concluded that while the inclusion of discards contributed a great deal of noise to the analyses, for those species where discarding made up significant proportions of the overall catch the discard augmented catch rates should be examined each year as a sensitivity analysis to contrast with the outcome from the un-augmented catch rates (Haddon, 2010).

### 20.4.4 The Analyses Including Discards

Discard rates cannot simply be added to known catches on the way to calculating catch rates. The standardized catch rates are estimated from individual catch and effort records but the estimates of discards are summary estimates for each fishery. While a method for incrementing the standardized catch rates has been developed it should be noted that this ignores all complications relating to unknown aspects of discarding behaviour (is the discard rate constant across all catch sizes, across all vessels, across all areas? etc). This means that including discard catches into the annual catch rate estimates introduces an unknown amount of uncertainty into the analysis. It should also be noted that the discard estimates are highly variable from year to year and derive from relatively small samples of all trips contributing to catches.

The method developed was to find the multiplier needed to adjust ratio mean catch rates and apply that to the standardized catch rates (Haddon, 2010). The ratio mean catch rates require the annual sum of catches for the fishery along with the sum of effort and ratio means calculated for each year. The discard estimates from the fishery can be added to the catch totals and new ratio means calculated and compared. The multiplier
needed to make the same changes to the ratio mean catch rates can then be developed and applied to the standardized catch rates.

The ratio mean is simply the sum of all catches divided by the sum of effort

$$
\begin{equation*}
\hat{I}_{R, t}=\frac{\sum C_{t}}{\sum E_{t}} \tag{9}
\end{equation*}
$$

where $\hat{I}_{R, t}$ is the ratio mean catch rate for year $t, \Sigma C_{t}$ is the sum of landed catches in year $t$, and $\Sigma E_{t}$ is the sum of effort (as hours trawled) in year $t$. If $\Sigma D_{t}$ is the sum of discards in year $t$ then the discard incremented ratio mean catch rate would be

$$
\begin{equation*}
\hat{I}_{D, t}=\frac{\sum C_{t}+\sum D_{t}}{\sum E_{t}} \tag{10}
\end{equation*}
$$

The same values of $\hat{I}_{D, t}$ can also be obtained using the following multiplier

$$
\begin{equation*}
\hat{I}_{D, t}=\left[\left(\sum D_{t} / \sum C_{t}\right)+1\right] \times I_{t} \tag{11}
\end{equation*}
$$

where $I_{t}$ is the catch rate estimate to be modified by the inclusion of discards. If this is the ratio mean from Equ (9) then the augmented catch rates would be identical to those produced by Equ (10). In practice, the catch rates used with the multiplier are the standardized catch rates from Haddon (2010a).

In the case of redfish and inshore ocean perch the discard augmented standardized mean catch rates were calculated, and compared visually with the geometric mean and original standardized catch rates. After the re-analysis of the catch rates these can be introduced into the TIER 4 analysis for Inshore Ocean Perch using the standard methods as described in Haddon (2010b).

Table 20.3. Characteristics used in the TIER 4 method. If a species is not considered to be fully fished during the reference period then the target catch rate is to be divided by two.

| Species | Reference <br> Years | Fully Fished by <br> Reference Period | First year with <br> catches > 100t. |
| :--- | :---: | :---: | :---: |
| Blue Eye Trevalla ALDL | $1997-2006$ | 1 | 1997 |
| Blue Warehou | $1986-1995$ | 1 | 1986 |
| Blue Warehou East | $1986-1996$ | 1 | 1986 |
| Blue Warehou West | $1986-1997$ | 1 | 1986 |
| Ocean Perch Inshore | $1986-1995$ | 1 | 1986 |
| Ocean Perch Inshore Discards | $1986-1996$ | 1 | 1986 |
| Ocean Perch Offshore | $1986-1997$ | 1 | 1986 |
| Royal Red Prawn | $1986-1995$ | 1 | 1986 |
| Silver Trevally | $1992-2001$ | 1 | 1986 |
| Blue Grenadier | $1986-1995$ | 1 |  |
| Flathead | $1986-1995$ | 1 | 1986 |
| Eastern Gemfish | $1993-2002$ | 1 | 1986 |
| Western Gemfish | $1992-2001$ | 1 | 1986 |
| Jackass Morwong | $1986-1995$ | 1 | 1992 |
| John Dory | $1986-1995$ | 1 | 1986 |
| Mirror Dory | $1986-1995$ | 1 | 1986 |
| Mirror Dory East | $1986-1995$ | 1 | 1986 |
| Mirror Dory West | $1996-2005$ | 1 | 1986 |
| Pink Ling | $1986-1995$ | 1 | 1996 |
| Redfish | $86-90 ; 99-03$ | 1 | 1986 |
| Redfish Discards | $86-90 ; 99-04$ | 1 | 1986 |
| Ribaldo | $1995-2004$ | 0.5 | 1986 |
| School Whiting | $1986-1995$ | 1 | 1995 |
| Spotted/Silver Warehou | $1986-1995$ | 0.5 | 1986 |
|  |  |  | 1986 |

### 20.4.5 Selection of Reference Periods

The Tier 4 requires a reference period to be selected in order to establish target and limit levels of catch rates and associated target levels of catch that are deemed by the RAG to act as a proxy for the desired state for the fishery. These act as a proxy for the Harvest Strategy Policy reference points of $48 \%$ and $20 \%$ unfished spawning biomass. The original Tier 4 rule that used a linear regression of the last four year's catch rates to determine whether catches increase or decrease was not able to rebuild a resource towards a desired target level and the current approach was developed so as to be able to manage a fishery towards a target and away from a limit.

The essence of the Tier 4 control rule is that it sets a RAG agreed target catch rate, which has an associated target catch. An estimate of current catch rates (usually the average of the last four years) is compared with the target and a multiplier is estimated which is to be applied to the target catch to generate the recommended biological catch.

To select a reference period requires a time series of comparable catch rates. For this reason the use of standardized catch rates should be an improvement over using, for example, the observed arithmetic or geometric mean catch rates. Catch rate data is available in the SESSF for all targeted species from 1986-2011, although it needs to be noted that the character of the fishery has changed markedly during that period. Little et al. (2009) provide a discussion on how reference periods might be selected. They proposed a default ten year period of 1986 - 1995, stating: "We have assumed that the average CPUE from 1986 to 1995 corresponds to that which would be attained if the stock were at the level that provides the maximum economic yield, $B_{M E Y}$. The limit CPUE is $40 \%$ of this CPUE." (Little et al., 2009, p 234).

For each species, reference years were selected by the RAGs to generate estimates of target catches and target catch rates. In addition, a decision was required as to whether the fishery could be considered as fully developed or otherwise during the reference period or not. Where a fishery was not considered to be fully developed the target catch rate, $C P U E_{\text {targ, }}$, was divided by two as a proxy for expected changes to catch rates as the fishery develops and the resource stock size declines towards the assumed proxy target for $48 \%$ unfished biomass.

Little et al. (2009) proposed three rules used to estimate the CPUE target:

1. The CPUE target for stocks fully exploited at or prior to 1986 is based on the average CPUE from 1986-1995.
2. Where fishing exploitation up to 1986 is thought to be minimal, the CPUE determined in step 1 is halved (to provide a catch rate proxy for $B_{M E Y}$ ).
3. Where fishing exploitation after 1986 is low, the first year in which catches are above $100 t$ signifies the start of the 10 year period for which CPUE targeted is calculated.

Once the average CPUE for the reference period has been selected as the target CPUE then the limit CPUE is defined as $40 \%$ of the target. All of these rules make the assumption that the target catch rates have achieved and equilibrium with the target catches. In other words, if the target catch was maintained long enough the target catch rate would be the result.

### 20.4.6 Treatment of Non-Target Species

In 2012, the SESSF RAG determined that the assessments of those species which do not constitute the economic drivers for a fishery might use the proxy for $\mathrm{B}_{\text {MSY }}$ as the target instead of $\mathrm{B}_{\text {MEY }}$. In practice this means that the target is assumed to be a proxy for $\mathrm{B}_{40}$ rather than $\mathrm{B}_{48}$. For the Tier 4, this means modifying the control rule used to estimate the RBC by multiplying the target catch rate by $5 / 6$. If the original target was a proxy for $48 \% B_{0}$, then $5 / 6^{\text {th }}$ or 0.83333 of this target would be a proxy for $B_{40 \%}$. The graphs illustrate this by a line below the original target.

### 20.4.7 The Assumption underlying the Tier 4

For the Tier 4 analyses to be valid a number of assumptions need to be met:

- There is a linear relationship between catch rates and exploitable biomass; if there is hyper-stability (catch rates remain stable while stock size changes) or hyperdepletion (catch rates decline much faster than stock size changes) then the standard Tier 4 analysis would provide biased results.
- The character of the estimated catch rates has not changed in significant ways through the period from the start of the reference period to the end of the most recent year; If there has been significant effort creep altering the catchability, or there have been changes to the fleet that have altered the relative efficiency of the vessels fishing then the comparability of the catch rates now relative to the target period may be compromised, which would obviously reduce the responsiveness of the Tier 4 method to change.
- The reference period provides a good estimate of the stock when at a depletion level of $48 \%$ unfished spawning biomass; the Tier 4 method is based on catch rates and thus relates to exploitable biomass and not spawning biomass. As a minimum the reference period will refer to a period when the stock was in an acceptable, productive and sustainable state. But there can be no guarantees that the target aimed for is really $\mathrm{B}_{48 \%}$.


### 20.5 Results

### 20.5.1 Blue Eye (TBE - 37445001 - Hyperoglyphe antarctica)

Table 20.4 Blue eye Trevalla data for the TIER 4 calculations. Total is the sum of Discards, State, Non Trawl, SEF2, and ECDW catches. All values in Tonnes. CE is the standardized catch rate for all Zones 10 to 50 in depths $0-1000 \mathrm{~m}$ (Haddon, 2012). GeoMean is the geometric mean catch rates. Discards are estimates from 1998 to present. The ratio of discards to catch over the 1998 - 2006 period was used to estimate the discards between 1986 and 1997, the proportion of which is the PDiscard.

| Year | Catch Discards | Total | State | Non-T | PDiscard | CE |  | GeoMean |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1997 | 732.786 |  | 732.786 | 620.157 | 0.000 | 1.21 | 1.8553 | 258.2795 |
| 1998 | 599.413 | 0.000 | 599.413 | 123.012 | 380.439 | 0.00 | 1.2439 | 226.1524 |
| 1999 | 706.643 | 0.000 | 706.643 | 132.608 | 464.658 | 0.00 | 1.1126 | 189.1263 |
| 2000 | 743.525 | 37.000 | 780.525 | 89.462 | 565.410 | 4.74 | 1.0566 | 177.6127 |
| 2001 | 665.345 | 33.000 | 698.345 | 77.613 | 478.397 | 4.73 | 1.1035 | 202.9873 |
| 2002 | 615.379 | 0.100 | 615.479 | 102.362 | 427.969 | 0.02 | 0.8559 | 163.8436 |
| 2003 | 650.952 | 0.160 | 651.112 | 51.623 | 556.565 | 0.02 | 0.9295 | 148.5823 |
| 2004 | 715.134 | 1.400 | 716.534 | 64.457 | 566.917 | 0.20 | 1.0215 | 91.4807 |
| 2005 | 549.140 | 0.000 | 549.140 | 55.557 | 450.678 | 0.00 | 0.8194 | 88.2645 |
| 2006 | 607.945 | 0.060 | 608.005 | 44.095 | 496.743 | 0.01 | 0.9396 | 121.2856 |
| 2007 | 638.412 | 2.808 | 641.220 | 53.102 | 536.267 | 0.45 | 1.1791 | 333.7817 |
| 2008 | 408.027 | 0.993 | 409.020 | 34.980 | 338.852 | 0.24 | 0.7867 | 214.3734 |
| 2009 | 478.452 | 0.000 | 478.452 | 35.090 | 404.049 | 0.00 | 0.8701 | 259.8521 |
| 2010 | 442.893 | 0.142 | 443.035 | 42.997 | 358.784 | 0.03 | 0.5634 | 142.9654 |
| 2011 | 492.825 | 7.347 | 500.172 | 33.744 | 430.038 | 1.48 | 0.6629 | 177.7306 |

Discards make up approximately $1.2 \%$ of the catch over the 1998-2006 period.
The catch rate time series used came from the combined autolongline and drop line fishery.

Table 20.5 RBC calculations for Blue Eye. $\mathrm{C}_{\text {targ }}$ and $\mathrm{CPUE}_{\text {targ }}$ relate to the period 1997-2006, CPUE $_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7).

| Ref_Year | $1997-2006$ |
| ---: | ---: |
| CE_Targ | 1.0938 |
| CE_Lim | 0.4375 |
| CE_Recent | 0.7208 |
| Wt_Discard | 4.022 |
| Scaling | 0.4316 |
| Last Year’s TAC |  |
| C $_{\text {targ }}$ | 665.798 |
| RBC | $\mathbf{2 8 7 . 3 7 6}$ |

BlueEyeALDL


Figure 20.1 Blue Eye Trevalla. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 20.5.2 Blue Warehou (TRT - 37445005 - Seriolella brama) Zones 10-50

Table 20.6 Blue Warehou data for the TIER 4 calculations. Total is the sum of Discards, State, Non Trawl and SEF2 catches. All values in Tonnes. CE is the standardized catch rate for Zones 10 to 50 in depths $0-400 \mathrm{~m}$ (Haddon, 2012). GeoMean is the geometric mean catch rates. Discards are estimates from 1998 to present. The ratio of discards to catch over the 1998 - 2006 period was used to estimate the discards between 1986 and 1997, the proportion of which is the PDiscard.

| Year | Catch | Discards | Total | State | Non-T | PDiscar <br> d | CE | GeoMean |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 277.200 | 53.638 | 330.838 |  |  | 16.21 | 2.0083 | 24.6419 |
| 1987 | 1010.400 | 195.512 | 1205.912 |  |  | 16.21 | 2.3004 | 38.9818 |
| 1988 | 999.600 | 193.422 | 1193.022 |  |  | 16.21 | 2.5658 | 42.2791 |
| 1989 | 1598.400 | 309.290 | 1907.690 |  |  | 16.21 | 3.5440 | 53.5132 |
| 1990 | 2272.800 | 439.786 | 2712.586 |  |  | 16.21 | 2.5233 | 49.3618 |
| 1991 | 2478.000 | 479.492 | 2957.492 |  |  | 16.21 | 1.9961 | 38.9026 |
| 1992 | 1869.600 | 361.767 | 2231.367 |  |  | 16.21 | 1.4626 | 34.9011 |
| 1993 | 1440.000 | 278.639 | 1718.639 |  |  | 16.21 | 1.1328 | 27.0143 |
| 1994 | 1308.081 | 253.13 | 1561.194 | 458.856 | 0.000 | 16.21 | 1.0921 | 24.5388 |
| 1995 | 1086.315 | 210.201 | 1296.516 | 328.851 | 0.000 | 16.21 | 0.9299 | 19.7435 |
| 1996 | 1223.451 | 236.737 | 1460.189 | 376.605 | 0.000 | 16.21 | 0.9250 | 16.0446 |
| 1997 | 981.513 | 189.922 | 1171.436 | 193.002 | 0.000 | 16.21 | 0.9172 | 13.9027 |
| 1998 | 1271.881 | 86.000 | 1357.881 | 270.399 | 80.448 | 6.33 | 0.9174 | 18.0335 |
| 1999 | 925.892 | 16.000 | 941.892 | 283.422 | 287.791 | 1.70 | 0.4883 | 9.5323 |
| 2000 | 628.918 | 16.000 | 644.918 | 113.511 | 82.121 | 2.48 | 0.4352 | 7.2891 |
| 2001 | 354.866 | 39.000 | 393.866 | 26.249 | 30.742 | 9.90 | 0.2942 | 5.6327 |
| 2002 | 389.328 | 7.370 | 396.698 | 71.962 | 3.720 | 1.86 | 0.2499 | 4.0433 |
| 2003 | 296.069 | 19.490 | 315.559 | 42.301 | 2.077 | 6.18 | 0.2056 | 3.2843 |
| 2004 | 293.191 | 381.440 | 674.631 | 31.188 | 1.719 | 56.54 | 0.2815 | 4.9660 |
| 2005 | 329.935 | 273.920 | 603.855 | 17.249 | 1.318 | 45.36 | 0.2614 | 6.0446 |
| 2006 | 412.776 | 109.480 | 522.256 | 26.282 | 0.732 | 20.96 | 0.2625 | 7.8259 |
| 2007 | 224.990 | 24.929 | 249.919 | 29.306 | 0.780 | 9.97 | 0.2449 | 5.6784 |
| 2008 | 194.125 | 265.391 | 459.516 | 36.859 | 0.976 | 57.75 | 0.2749 | 5.0903 |
| 2009 | 171.807 | 16.561 | 188.368 | 33.663 | 1.704 | 8.79 | 0.2721 | 6.9116 |
| 2010 | 154.353 | 14.878 | 169.231 | 22.624 | 4.584 | 8.79 | 0.2177 | 6.3388 |
| 2011 | 117.773 | 39.535 | 157.308 | 7.316 | 11.805 | 25.13 | 0.1969 | 5.5194 |

Discards make up approximately $16.2 \%$ of the catch over the 1998-2006 period.

| Table 20.7 RBC calculations for Blue Warehou1050. $\mathrm{C}_{\text {targ }}$ and CPUE $_{\text {targ }}$ relate to the period $1986-1995$, CPUE $_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7). |  |
| :---: | :---: |
| Ref_Year | 1986-1995 |
| CE_Targ | 1.9555 |
| CE_Lim | 0.7822 |
| CE_Recent | 0.2404 |
| Wt_Discard | 44.954 |
| Scaling | 0 |
| Last Year's TAC |  |
| $\mathrm{C}_{\text {targ }}$ | 1711.526 |
| RBC | 0 |

## BlueWarehou 1050



Figure 20.2 Blue Warehou. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 20.5.3 Blue Warehou (TRT - 37445005 - Seriolella brama) Zones 10, 20 \& 30

To provide an analysis more relevant to the two stocks of Blue Warehou (east and west) the landed catches, which are reported in total across zones $10-50$, were subdivided in the same ratio as the reported catches from the catch effort log books, the discards were treated in the same fashion. Thus the catches and discards in Table 20.8 and Table 20.10 should sum in each year to the catches and discards in Table 20.6. The separate columns for the State and NonTrawl catches were not adjusted and so, for these analyses are not meaningful.

Table 20.8 Blue Warehou data for the TIER 4 calculations. Total is the sum of Discards, State, Non Trawl and SEF2 catches. All values in Tonnes. CE is the standardized catch rate for Zones 10 to 30 in depths $0-$ 400m (Haddon, 2012). GeoMean is the geometric mean catch rates. Discards are estimates from 1998 to present. The ratio of discards to catch over the $1998-2006$ period was used to estimate the discards between 1986 and 1997. Prop is the proportion of the Commonwealth catch taken in zones $10-30$.

| Year | Catch | Discards | Total | State | Non-T | PDiscard | CE | GeoMean | Prop |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 183.061 | 35.422 | 218.483 |  |  | 16.21 | 1.8787 | 22.9216 | 0.660 |
| 1987 | 442.684 | 85.659 | 528.343 |  |  | 16.21 | 2.2848 | 23.2716 | 0.438 |
| 1988 | 627.614 | 121.443 | 749.057 |  |  | 16.21 | 2.8020 | 34.8726 | 0.628 |
| 1989 | 1424.124 | 275.567 | 1699.692 |  |  | 16.21 | 3.5366 | 52.6588 | 0.891 |
| 1990 | 1432.240 | 277.138 | 1709.378 |  |  | 16.21 | 3.2341 | 46.5510 | 0.630 |
| 1991 | 1035.919 | 200.450 | 1236.368 |  |  | 16.21 | 1.7420 | 23.0208 | 0.418 |
| 1992 | 908.680 | 175.829 | 1084.509 |  |  | 16.21 | 1.4238 | 24.3304 | 0.486 |
| 1993 | 782.415 | 151.397 | 933.812 |  |  | 16.21 | 1.1069 | 20.7054 | 0.543 |
| 1994 | 671.031 | 129.844 | 800.875 | 235.388 |  | 16.21 | 1.0674 | 17.5997 | 0.513 |
| 1995 | 640.884 | 124.011 | 764.895 | 194.010 |  | 16.21 | 0.9784 | 15.3567 | 0.590 |
| 1996 | 909.275 | 175.944 | 1085.220 | 279.895 |  | 16.21 | 0.9835 | 14.6415 | 0.743 |
| 1997 | 612.519 | 118.522 | 731.041 | 120.444 |  | 16.21 | 0.9519 | 11.8760 | 0.624 |
| 1998 | 716.450 | 48.444 | 764.894 | 152.316 | 45.316 | 6.33 | 0.8996 | 13.8592 | 0.563 |
| 1999 | 398.296 | 6.883 | 405.179 | 121.921 | 123.801 | 1.70 | 0.4842 | 5.7097 | 0.430 |
| 2000 | 299.640 | 7.623 | 307.263 | 54.081 | 39.126 | 2.48 | 0.4275 | 5.0072 | 0.476 |
| 2001 | 80.801 | 8.880 | 89.681 | 5.977 | 7.000 | 9.90 | 0.2561 | 2.7867 | 0.228 |
| 2002 | 87.139 | 1.650 | 88.788 | 16.106 | 0.833 | 1.86 | 0.1966 | 2.2036 | 0.224 |
| 2003 | 57.263 | 3.770 | 61.033 | 8.182 | 0.402 | 6.18 | 0.1558 | 1.8331 | 0.193 |
| 2004 | 72.222 | 93.960 | 166.182 | 7.683 | 0.423 | 56.54 | 0.2122 | 2.7248 | 0.246 |
| 2005 | 25.164 | 20.892 | 46.056 | 1.316 | 0.101 | 45.36 | 0.1409 | 1.8011 | 0.076 |
| 2006 | 29.231 | 7.753 | 36.984 | 1.861 | 0.052 | 20.96 | 0.1689 | 2.2327 | 0.071 |
| 2007 | 22.795 | 2.526 | 25.320 | 2.969 | 0.079 | 9.97 | 0.1826 | 1.8677 | 0.101 |
| 2008 | 36.683 | 50.150 | 86.834 | 6.965 | 0.184 | 57.75 | 0.2522 | 2.6539 | 0.189 |
| 2009 | 50.338 | 4.852 | 55.190 | 9.863 | 0.499 | 8.79 | 0.2922 | 3.5956 | 0.293 |
| 2010 | 16.105 | 1.552 | 17.658 | 2.361 | 0.478 | 8.79 | 0.1894 | 2.1227 | 0.104 |
| 2011 | 13.174 | 4.422 | 17.596 | 0.818 | 1.320 | 25.13 | 0.1516 | 1.7081 | 0.112 |

Discards make up approximately $16.2 \%$ for the period 1998 - 2006 .


## BlueWarehouE



Figure 20.3 Blue Warehou zones 10-30. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 20.5.4 Blue Warehou (TRT - 37445005 - Seriolella brama) Zones 40 \& 50

To provide an analysis more relevant to the two stocks of Blue Warehou (east and west) the landed catches, which are reported in total across zones $10-50$, were subdivided in the same ratio as the reported catches from the catch effort log books, the discards were treated in the same fashion. Thus the catches and discards in Table 20.8 and Table 20.10 should sum in each year to the catches and discards in Table 20.6. The separate columns for the State and NonTrawl catches were not adjusted and so, for these analyses are not meaningful.

Table 20.10 Blue Warehou data for the TIER 4 calculations. Total is the sum of Discards, State, Non Trawl and SEF2 catches. All values in Tonnes. CE is the standardized catch rate for Zones 40 to 50 in depths $0-400 \mathrm{~m}$ (Haddon, 2012). GeoMean is the geometric mean catch rates. Discards are estimates from 1998 to present. The ratio of discards to catch over the 1998 - 2006 period was used to estimate the discards between 1986 and 1997. Prop is the proportion of the Commonwealth catch taken in zones 40 50.

| Year | Catch | Discards | Total | State | Non-T | PDiscard | CE | GeoMean | Prop |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 94.139 | 18.216 | 112.355 |  |  | 16.21 | 3.3403 | 34.3927 | 0.340 |
| 1987 | 567.716 | 109.853 | 677.569 |  |  | 16.21 | 3.1136 | 153.6342 | 0.562 |
| 1988 | 371.986 | 71.979 | 443.965 |  |  | 16.21 | 1.2972 | 104.5294 | 0.372 |
| 1989 | 174.276 | 33.722 | 207.998 |  |  | 16.21 | 3.3111 | 91.5270 | 0.109 |
| 1990 | 840.560 | 162.648 | 1003.208 |  |  | 16.21 | 1.4434 | 55.8069 | 0.370 |
| 1991 | 1442.081 | 279.042 | 1721.123 |  |  | 16.21 | 2.2301 | 159.6429 | 0.582 |
| 1992 | 960.920 | 185.938 | 1146.858 |  |  | 16.21 | 1.2798 | 88.9759 | 0.514 |
| 1993 | 657.585 | 127.242 | 784.827 |  |  | 16.21 | 0.9374 | 92.3447 | 0.457 |
| 1994 | 637.050 | 123.269 | 760.319 | 223.468 |  | 16.21 | 1.0262 | 67.3117 | 0.487 |
| 1995 | 445.431 | 86.191 | 531.621 | 134.841 |  | 16.21 | 0.7000 | 45.1964 | 0.410 |
| 1996 | 314.176 | 60.793 | 374.969 | 96.710 |  | 16.21 | 0.4568 | 26.4215 | 0.257 |
| 1997 | 368.994 | 71.400 | 440.394 | 72.558 |  | 16.21 | 0.4907 | 35.6095 | 0.376 |
| 1998 | 555.431 | 37.556 | 592.987 | 118.083 | 35.132 | 6.33 | 0.7327 | 58.9967 | 0.437 |
| 1999 | 527.596 | 9.117 | 536.713 | 161.501 | 163.990 | 1.70 | 0.4211 | 32.5226 | 0.570 |
| 2000 | 329.278 | 8.377 | 337.655 | 59.430 | 42.995 | 2.48 | 0.3439 | 28.0473 | 0.524 |
| 2001 | 274.065 | 30.120 | 304.185 | 20.272 | 23.742 | 9.90 | 0.3634 | 27.5825 | 0.772 |
| 2002 | 302.189 | 5.720 | 307.910 | 55.855 | 2.887 | 1.86 | 0.4896 | 35.4216 | 0.776 |
| 2003 | 238.806 | 15.720 | 254.526 | 34.120 | 1.675 | 6.18 | 0.4422 | 28.1023 | 0.807 |
| 2004 | 220.969 | 287.480 | 508.449 | 23.506 | 1.296 | 56.54 | 0.5050 | 28.4995 | 0.754 |
| 2005 | 304.771 | 253.028 | 557.799 | 15.934 | 1.217 | 45.36 | 0.7955 | 53.5991 | 0.924 |
| 2006 | 383.544 | 101.727 | 485.271 | 24.420 | 0.680 | 20.96 | 0.5634 | 31.8482 | 0.929 |
| 2007 | 202.195 | 22.404 | 224.599 | 26.337 | 0.701 | 9.97 | 0.4887 | 22.9820 | 0.899 |
| 2008 | 157.441 | 215.241 | 372.682 | 29.894 | 0.791 | 57.75 | 0.3739 | 20.3955 | 0.811 |
| 2009 | 121.469 | 11.709 | 133.177 | 23.800 | 1.205 | 8.79 | 0.2769 | 18.4388 | 0.707 |
| 2010 | 138.248 | 13.326 | 151.574 | 20.263 | 4.106 | 8.79 | 0.3152 | 17.5511 | 0.896 |
| 2011 | 104.599 | 35.113 | 139.712 | 6.498 | 10.484 | 25.13 | 0.2618 | 14.3658 | 0.888 |

Discards make up approximately 16.2 \% of the catch over the 1998-2006 period.

Table 20.11 RBC calculations for Blue Warehou West (Zones 40-50). $\mathrm{C}_{\text {targ }}$ and $\mathrm{CPUE}_{\text {targ }}$ relate to the period 1986-1995, $\mathrm{CPUE}_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7).

| Ref_Year | $1986-1995$ |
| ---: | ---: |
| CE_Targ $^{\text {CE_Lim }}$ | 1.8679 |
| CE_Recent | 0.7472 |
| Wt_Discard | 0.307 |
| Scaling | 38.191 |
| Last Year's TAC | 0 |
| C $_{\text {targ }}$ | 738.984 |
| RBC | $\mathbf{0}$ |

## BlueWarehouW



Figure 20.4 Blue Warehou zones $40-50$. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 20.5.5 Inshore Ocean Perch Including Discards (REG - 37287001 - Helicolenus percoides)

Inshore Ocean Perch are subject to relatively high levels of discarding, which was likely to have large effects on the perceived catch rates. By including the estimated discards in with the reported catches revised catch rates were possible. No standardization was possible using the simple ratio means but a method was devised that attempted to use the standardized catch rates with a multiplier devised from ratio means of total catches (reported catches + discards) divided by total effort.

Table 20.12 Inshore Ocean Perch data for the Alternative TIER 4 calculations using ratio mean catch rates that include discards in the catch rate calculations. Total is the sum of Discards, and other catches. All values in Tonnes. StandCE is the standardized catch rate for Inshore Ocean perch from Zones 10 and 20 in depths $0-200 \mathrm{~m}$ (Haddon, 2012). GeoMean is the geometric mean catch rates (without discards). Discards are estimates from 1998 to present. DiscCE is the standardized catch rates multiplied by [ (Discard/Catch)+1], see Haddon (2011c) for methods.

| Year | Catch | Discards | Total | Effort | (D/C)+1 | StandCE | DiscCE | GeoMean |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 15.239 | 49.930 | 65.169 | 978.4 | 4.276462 | 0.8363 | 0.7531 | 1.2184 |
| 1987 | 12.441 | 34.842 | 47.283 | 1319.8 | 3.800579 | 0.9828 | 0.7865 | 1.0578 |
| 1988 | 16.643 | 49.027 | 65.670 | 1599.5 | 3.945803 | 1.1184 | 0.9292 | 1.2957 |
| 1989 | 16.758 | 50.257 | 67.015 | 1315.2 | 3.998986 | 1.0713 | 0.9021 | 1.4286 |
| 1990 | 17.076 | 88.665 | 105.741 | 1416.9 | 6.192375 | 1.1429 | 1.4903 | 1.3818 |
| 1991 | 26.084 | 106.551 | 132.635 | 1495.5 | 5.084918 | 1.2818 | 1.3725 | 1.4465 |
| 1992 | 16.106 | 106.112 | 122.218 | 742.8 | 7.588352 | 1.696 | 2.7100 | 1.6928 |
| 1993 | 29.267 | 100.307 | 129.574 | 1390.1 | 4.427307 | 1.9063 | 1.7772 | 1.8109 |
| 1994 | 38.765 | 99.192 | 137.957 | 1599.4 | 3.55882 | 1.7357 | 1.3007 | 1.6767 |
| 1995 | 40.881 | 104.606 | 145.487 | 1712.4 | 3.558816 | 1.284 | 0.9622 | 1.5562 |
| 1996 | 51.250 | 131.139 | 182.389 | 2127.5 | 3.558824 | 1.1194 | 0.8389 | 1.2539 |
| 1997 | 34.279 | 87.713 | 121.992 | 1750.3 | 3.558833 | 1.0464 | 0.7842 | 1.0498 |
| 1998 | 39.085 | 124.000 | 163.085 | 1858.4 | 4.17256 | 0.9151 | 0.8040 | 1.0225 |
| 1999 | 25.438 | 78.000 | 103.438 | 2073.3 | 4.066274 | 0.8112 | 0.6946 | 0.8883 |
| 2000 | 47.846 | 100.000 | 147.846 | 4148.9 | 3.090058 | 0.9859 | 0.6415 | 0.8125 |
| 2001 | 37.815 | 89.000 | 126.815 | 3191.9 | 3.353576 | 0.98 | 0.6920 | 0.7479 |
| 2002 | 48.363 | 145.110 | 193.473 | 4661.2 | 4.000439 | 0.6996 | 0.5893 | 0.4651 |
| 2003 | 30.865 | 61.320 | 92.185 | 3742.2 | 2.986715 | 0.5408 | 0.3401 | 0.4112 |
| 2004 | 25.887 | 194.450 | 220.337 | 3285.2 | 8.51161 | 0.5522 | 0.9897 | 0.3989 |
| 2005 | 23.829 | 41.680 | 65.509 | 3103.4 | 2.749095 | 0.625 | 0.3618 | 0.5311 |
| 2006 | 50.439 | 9.760 | 60.199 | 2153.8 | 1.193503 | 0.5206 | 0.1308 | 0.4000 |
| 2007 | 35.923 | 17.195 | 53.117 | 1369.8 | 1.478654 | 0.7329 | 0.2282 | 0.6302 |
| 2008 | 29.746 | 23.433 | 53.180 | 1094.1 | 1.787777 | 0.9001 | 0.3388 | 0.7552 |
| 2009 | 19.480 | 91.350 | 110.830 | 947.4 | 5.689398 | 0.7656 | 0.9172 | 0.7348 |
| 2010 | 21.952 | 132.847 | 154.798 | 1095.3 | 7.051795 | 0.8049 | 1.1952 | 0.6819 |
| 2011 | 16.411 | 269.811 | 286.222 | 1010.5 | 17.44092 | 0.9448 | 3.4698 | 0.6513 |

Discards are calculated now according to the latest ISMP design and this has led to a reassessment of the levels of discards from 2008 onwards; hence the difference between this year's analysis and last year's.

Table 20.13 RBC calculations for Inshore Ocean Perch. $\mathrm{C}_{\text {targ }}$ and CPUE $_{\text {targ }}$ relate to the period $1986-1995, \mathrm{CPUE}_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7).

| Ref_Year | $1986-1995$ |
| ---: | ---: |
| CE_Targ | 1.2984 |
| CE_Lim | 0.5194 |
| CE_Recent | 1.4803 |
| Wt_Discard | 193.067 |
| Scaling | 1.2335 |
| Last Year’s TAC |  |
| C $_{\text {targ }}$ | 101.875 |
| RBC | $\mathbf{1 2 5 . 6 6 1}$ |

InOceanPerchDiscard


Figure 20.5 Alternative InShore Ocean Perch (where catch rates include discards). Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

Table 20.14 RBC calculations for Inshore Ocean Perch. $\mathrm{C}_{\text {targ }}$ and $\mathrm{CPUE}_{\text {targ }}$ relate to the period $1986-1995$, CPUE $_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7). The proxy target is here B40\%.

| Ref_Year | $1986-1995$ |
| ---: | ---: |
| CE_Targ $^{\text {CE_Lim }}$ | 1.2984 |
| CE_Recent | 0.5194 |
| Wt_Discard | 1.4803 |
| Scaling | 193.067 |
| Last Year’s TAC | 1.7079 |
| C $_{\text {targ }}$ |  |
| RBC | 101.875 |
|  | $\mathbf{1 7 3 . 9 9 3}$ |

## InOceanPerchDiscard



Figure 20.6 Alternative InShore Ocean Perch (where catch rates include discards). Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 20.5.6 Offshore Ocean Perch (REG - 37287001 - H percoides) 48\% Target Proxy

The RAG agreed, this year, to attempt to estimate the RBC for Ocean Perch by separately estimating the RBCs for Offshore and Inshore Ocean Perch and combining the result. Offshore Ocean Perch were defined as those records that were reported as being from 200-700 metres depth; Inshore Ocean Perch were defined as those records from depths of $0-200$ metres (A decision of the RAG in 2010, reversing a different decision made in 2009). In addition, the data series of reported catches differ from those previously used as they have been recently reviewed and revised, splitting the landings between Offshore and Inshore Ocean Perch relative to the Commonwealth log book catches for the two depth ranges. This increased the total catches reported, but these data are now the best available information on Ocean Perch catches.

Table 20.15 Offshore Ocean Perch data for the TIER 4 calculations. Total is the sum of Discards, State, Non Trawl and SEF2 catches. All values in Tonnes. CE is the standardized catch rate for Offshore Ocean perch from Zones 10 and 20 in depths $200-700 \mathrm{~m}$ (Haddon, 2012). GeoMean is the geometric mean catch rates. Discards are estimates from 1998 to present. The ratio of discards to catch over the 1998 2006 period was used to estimate the discards between 1986 and 1997. Landings before 1994 were subdivided according to the ratio of inshore to offshore in the Commonwealth logbook data.

| Year | Catch | Discards | Total | State | Non-T | Pdiscard | CE | GeoMean |
| ---: | ---: | ---: | ---: | :--- | ---: | ---: | ---: | ---: |
| 1986 | 218.366 | 31.876 | 250.242 |  |  | 12.74 | 1.0298 | 12.1440 |
| 1987 | 179.087 | 26.142 | 205.230 |  |  | 12.74 | 0.9538 | 8.9237 |
| 1988 | 178.089 | 25.997 | 204.086 |  |  | 12.74 | 1.0669 | 10.5074 |
| 1989 | 207.462 | 30.284 | 237.746 |  |  | 12.74 | 1.0257 | 10.6494 |
| 1990 | 176.918 | 25.826 | 202.744 |  |  | 12.74 | 1.3644 | 12.0207 |
| 1991 | 234.031 | 34.163 | 268.193 |  |  | 12.74 | 1.4423 | 13.4339 |
| 1992 | 349.336 | 50.994 | 400.330 |  |  | 12.74 | 1.2143 | 11.9264 |
| 1993 | 314.476 | 45.906 | 360.382 |  |  | 12.74 | 1.2142 | 12.9555 |
| 1994 | 294.313 | 42.962 | 337.276 | 35.478 | 0.000 | 12.74 | 1.1325 | 11.8001 |
| 1995 | 320.654 | 46.807 | 367.461 | 35.712 | 0.000 | 12.74 | 1.0249 | 10.4874 |
| 1996 | 363.621 | 53.080 | 416.701 | 35.992 | 0.000 | 12.74 | 0.9240 | 9.8364 |
| 1997 | 440.479 | 64.299 | 504.777 | 37.041 | 5.312 | 12.74 | 0.9739 | 9.7119 |
| 1998 | 372.254 | 174.000 | 546.254 | 35.974 | 6.250 | 31.85 | 0.8662 | 9.4285 |
| 1999 | 395.062 | 64.000 | 459.062 | 39.250 | 7.018 | 13.94 | 0.9802 | 9.7566 |
| 2000 | 344.156 | 34.000 | 378.156 | 36.369 | 9.086 | 8.99 | 0.7702 | 7.5464 |
| 2001 | 356.183 | 46.000 | 402.183 | 29.725 | 8.597 | 11.44 | 0.8632 | 8.3956 |
| 2002 | 322.376 | 22.470 | 344.846 | 36.660 | 18.885 | 6.52 | 0.8206 | 7.3709 |
| 2003 | 373.003 | 27.800 | 400.803 | 28.965 | 30.940 | 6.94 | 0.8719 | 7.6242 |
| 2004 | 362.369 | 42.440 | 404.809 | 19.579 | 66.129 | 10.48 | 0.8707 | 8.0648 |
| 2005 | 322.617 | 17.100 | 339.717 | 15.404 | 34.518 | 5.03 | 0.9783 | 9.3641 |
| 2006 | 226.413 | 20.980 | 247.393 | 15.835 | 46.229 | 8.48 | 0.8351 | 7.8433 |
| 2007 | 186.607 | 100.727 | 287.334 | 13.362 | 28.638 | 35.06 | 1.0332 | 9.9183 |
| 2008 | 208.930 | 22.187 | 231.117 | 13.489 | 37.801 | 9.60 | 0.9554 | 9.1917 |
| 2009 | 218.732 | 28.233 | 246.965 | 18.551 | 32.967 | 11.43 | 0.9499 | 9.0355 |
| 2010 | 238.512 | 81.596 | 320.108 | 27.782 | 28.977 | 25.49 | 0.9792 | 9.8647 |
| 2011 | 223.984 | 18.569 | 242.553 | 10.842 | 24.104 | 7.66 | 0.8592 | 9.0998 |

Discards make up approximately $12.68 \%$ \% of the catch over the 1998-2006 period. The catch rates used were for Offshore Ocean Perch from 200 to 700 metres depth. State catches from 1994 to 1997 were compromised through including some

Commonwealth catches. As an agreed upon better estimates, the State catches in these years were replaced with the average State catch from the years 1998 to 2003.

Table 20.16 RBC calculations for Offshore Ocean Perch. $\mathrm{C}_{\operatorname{targ}}$ and CPUE $_{\operatorname{targ}}$ relate to the period $1986-1995, \mathrm{CPUE}_{\mathrm{Lim}}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7).

| Ref_Year | $1986-1995$ |
| ---: | ---: |
| CE_Targ | 1.1469 |
| CE_Lim | 0.4588 |
| CE_Recent | 0.9359 |
| Wt_Discard | 36.906 |
| Scaling | 0.6934 |
| Last Year's TAC |  |
| Ctarg $^{\text {RBC }}$ | 283.369 |

## OffOceanPerch



Figure 20.7 OffShore Ocean Perch. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

Table 20.17 RBC calculations for Offshore Ocean Perch. $\mathrm{C}_{\text {targ }}$ and CPUE $_{\text {targ }}$ relate to the period $1986-1995, \mathrm{CPUE}_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7). The proxy target is here B40\%.

| Ref_Year | $1986-1995$ |
| ---: | ---: |
| CE_Targ $^{\text {CE_Lim }}$ | 0.9557 |
| CE_Recent | 0.4588 |
| Wt_Discard | 0.9359 |
| Scaling | 36.906 |
| Last Year's TAC | 0.9602 |
| C $_{\text {targ }}$ |  |
| RBC | 283.369 |
|  | $\mathbf{2 7 2 . 0 7 7}$ |



OffOceanPerch

Figure 20.8 OffShore Ocean Perch. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

Table 20.18 RBC calculations for Offshore Ocean Perch. $\mathrm{C}_{\text {targ }}$ and $\mathrm{CPUE}_{\text {targ }}$ relate to the period $1986-1995, \mathrm{CPUE}_{\mathrm{Lim}}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7). Includes Discards

| Ref_Year | $1986-1995$ |
| ---: | ---: |
| CE_Targ | 1.1299 |
| CE_Lim | 0.4519 |
| CE_Recent | 0.9401 |
| Wt_Discard | 36.906 |
| Scaling | 0.72 |
| Last Year’s TAC |  |
| C $_{\text {targ }}$ | 283.369 |
| RBC | $\mathbf{2 0 4 . 0 2 6}$ |

OffOceanPerchDiscard


Figure 20.9 OffShore Ocean Perch. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

Table 20.19 RBC calculations for Offshore Ocean Perch. $\mathrm{C}_{\text {targ }}$ and CPUE $_{\text {targ }}$ relate to the period $1986-1995$, CPUE $_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7). Discards Included. The proxy target is here B40\%.

| Ref_Year | $1986-1995$ |
| ---: | ---: |
| CE_Targ | 0.9416 |
| CE_Lim | 0.4519 |
| CE_Recent | 0.9401 |
| Wt_Discard | 36.906 |
| Scaling | 0.9969 |
| Year’s TAC |  |
| Ctarg | 283.369 |
| RBC | $\mathbf{2 8 2 . 5 0 0}$ |

OffOceanPerchDiscard





Figure 20.10 OffShore Ocean Perch. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 20.5.7 Royal Red Prawn (PRR - 28714005 - Haliporoides sibogae)

Table 20.20 Royal Red Prawn data for the TIER 4 calculations. Total is the sum of Discards, State, Non Trawl and SEF2 catches. All values in Tonnes. CE is the standardized catch rate for Zone 10 in depths 0 - 400m (Haddon, 2012). GeoMean is the geometric mean catch rates. Discards are estimates from 1998 to present. The ratio of discards to catch over the 1998 - 2006 period was used to estimate the discards between 1986 and 1997, the proportion of which is the PDiscard.

| Year | Catch | Discards | Total | State | Non-T PDiscard | CE | GeoMea |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| n |  |  |  |  |  |  |  |

Discards make up approximately 4.3 \% of the catch over the 1998-2006 period.

Table 20.21 RBC calculations for Royal Red Prawn. $\mathrm{C}_{\text {targ }}$ and $\mathrm{CPUE}_{\text {targ }}$ relate to the period 1986-1995, $\mathrm{CPUE}_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7).

| Ref_Year | $1986-1995$ |
| ---: | ---: |
| CE_Targ | 1.0611 |
| CE_Lim | 0.4244 |
| CE_Recent | 0.9738 |
| Wt_Discard | 6.672 |
| Scaling | 0.8629 |
| Year's TAC |  |
| C $_{\text {targ }}$ | 408.075 |
| RBC | $\mathbf{3 5 2 . 1 2 3}$ |

RRP


Figure 20.11 Royal Red Prawn. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 20.5.7.1 Royal Red Prawn Taken with Different Mesh Sizes

Royal Red Prawns are principally taken in SESSF zone 10 and just north of the northern border of the SESSF along the NSW coastline (including relatively small amounts north of Barrenjoey). When they are specifically targeted it is standard practice to change the net to one with a much smaller mesh. However, in the standard analysis of catch rates, because the information on mesh size is only available for a limited number of years (2002-2011) no attention has been paid to which mesh is in use for each shot despite there being higher catch rates with the smaller meshed nets (Figure 20.12). It has been requested that the effect of mesh size on catch rates be examined for the limited years data was available.


Figure 20.12. Log book data for Royal Red Prawn since 2003 (about $81 \%$ of data has mesh data provided each year; in 2002 only about $12 \%$ mesh data was provided). The two modes in the log transformed catch rate data illustrates, crudely, the difference between the smaller meshed nets (higher catch rates) and the larger meshed nets (lower catch rates).
In the log books there are 23 different mesh sizes recorded. When the log-transformed catch rates are plotted against the mesh size used to make each individual catch there are three clusters apparent (Figure 20.13), with the mean catch rate of each cluster increasing as the mesh size increases.


Figure 20.13. The catch rates of Royal Red Prawn obtained by different meshed nets. The mesh size has been jittered (a small random number added) so as to make clusters of observations apparent. The three red lines represent the mean catch rate of each of the three groups of observations.

Given that there is a clear difference between the smaller meshes and the larger meshes and there were only few records around $70-75 \mathrm{~mm}$ (Table 20.22), the data were grouped into those vessels with meshes $<60 \mathrm{~mm}$, those with meshes $>80 \mathrm{~mm}$, and those in between. Subsequent analyses were then conducted only on the smaller and larger meshed groups.

Table 20.22. The relative catches reported in the log books by mesh categories. In 2002 only $12 \%$ of catches had mesh size recorded.

| Year | No Mesh | $<60 \mathrm{~mm}$ | $>=60 \&<80 \mathrm{~mm}$ | $>=80 \mathrm{~mm}$ |
| ---: | ---: | ---: | ---: | ---: |
| 2003 | 41.475 | 80.250 | 20.840 | 20.619 |
| 2004 | 48.150 | 84.858 | 29.951 | 7.722 |
| 2005 | 50.046 | 80.135 | 6.770 | 22.854 |
| 2006 | 69.820 | 50.185 |  | 58.574 |
| 2007 | 18.445 | 50.410 |  | 47.575 |
| 2008 |  | 51.215 | 0.700 | 18.690 |
| 2009 | 8.625 | 54.248 |  | 4.734 |
| 2010 | 28.158 | 45.550 |  | 9.113 |
| 2011 | 40.970 | 56.200 |  | 11.790 |

Table 20.23. The standardized catch rates for the eight different models fitted to the Royal Red Prawn data (including the smaller and larger mesh classes). The optimum model was $\mathrm{LnCE}=$ Year + Vessel + Mesh + DepCat + Month + DayNight.

| Year | Year | Vessel | Mesh | DepCat | Month | DayNight | Month:DepCat | DN:Month |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2002 | 0.9923 | 1.1450 | 0.9687 | 1.0375 | 1.0299 | $\mathbf{1 . 0 3 0 8}$ | 1.0133 | 1.0275 |
| 2003 | 0.7535 | 1.2447 | 0.9751 | 1.0464 | 1.0022 | $\mathbf{0 . 9 8 5 2}$ | 0.9970 | 0.9866 |
| 2004 | 0.8767 | 1.1113 | 1.0311 | 1.0725 | 1.0516 | $\mathbf{1 . 0 3 9 0}$ | 1.0842 | 1.0524 |
| 2005 | 0.8780 | 1.0594 | 0.9697 | 0.9914 | 0.9527 | $\mathbf{0 . 9 4 0 0}$ | 0.9376 | 0.9370 |
| 2006 | 0.8688 | 1.1939 | 1.3374 | 1.2319 | 1.1756 | $\mathbf{1 . 1 5 9 8}$ | 1.1819 | 1.1558 |
| 2007 | 1.0624 | 0.7905 | 1.0345 | 1.0031 | 1.0074 | $\mathbf{1 . 0 2 9 6}$ | 0.9995 | 1.0057 |
| 2008 | 0.9007 | 0.6595 | 0.7815 | 0.7959 | 0.8098 | $\mathbf{0 . 8 2 3 9}$ | 0.8019 | 0.8191 |
| 2009 | 1.4024 | 0.8534 | 0.7784 | 0.7550 | 0.7978 | $\mathbf{0 . 8 0 2 9}$ | 0.8080 | 0.8072 |
| 2010 | 0.7655 | 0.7555 | 0.8487 | 0.8340 | 0.8747 | $\mathbf{0 . 8 8 3 7}$ | 0.8622 | 0.8829 |
| 2011 | 1.4998 | 1.1869 | 1.2749 | 1.2324 | 1.2982 | $\mathbf{1 . 3 0 5 0}$ | 1.3145 | 1.3259 |



Figure 20.14. Standardized catch rates for Royal Red Prawn including Mesh size in the standardization. The geometric mean is represented by the dashed line and the vertical bars are two times the standard errors.


Figure 20.15. The relative impact of the different factors on the trend in catch rates. The blue bars indicate where the addition of a factor leads to the trend rising above the previous model while a red bar indiciats where the trend drops below the previous model's prediction. The effect of mesh appears to relate to the reduced catches taken by the bigger meshes and increased catches with no mesh. The impact of vessel reflects the completion of the structural adjustment.

While there are some differences when a comparison is made between standardizations using all available data, the larger and smaller mesh only data, and the smaller mesh data only (Figure 20.16) the general trends over the period 2002-2011 are basically the same. In all cases the catch rates in 2011 have all increased markedly.


Figure 20.16. a comparison of the standardization based on all data with the standardizations that relate to the data where mesh size information was available for both large and small meshes, and also for a separate analysis where only the small mesh data were standardized.

### 20.5.7.2 The Effect of the Endeavour Dogfish Closure

Catches in the final version of the Endeavour Dogfish closure reached between 15 $21 \%$ between $1998-2001$ but have always been less than that in other years (Table 20.24).

Table 20.24. Catches of Royal Red Prawn in the Endeavour dogfish closure and elsewhere.

| Year | Open | Endeavour | Year | Open | Endeavour |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 228.150 | 3.694 | 1999 | 283.239 | 65.565 |
| 1987 | 320.209 | 4.507 | 2000 | 340.739 | 57.735 |
| 1988 | 340.567 | 3.890 | 2001 | 180.289 | 48.410 |
| 1989 | 303.417 | 7.343 | 2002 | 406.385 | 10.985 |
| 1990 | 311.118 |  | 2003 | 156.969 | 6.215 |
| 1991 | 299.370 |  | 2004 | 167.451 | 3.230 |
| 1992 | 145.291 | 0.790 | 2005 | 159.605 | 0.200 |
| 1993 | 232.774 |  | 2006 | 177.629 | 0.950 |
| 1994 | 240.363 |  | 2007 | 116.430 |  |
| 1995 | 237.595 | 15.310 | 2008 | 70.605 |  |
| 1996 | 258.345 | 14.330 | 2009 | 67.587 | 0.020 |
| 1997 | 152.173 | 14.530 | 2010 | 82.221 | 0.600 |
| 1998 | 152.960 | 37.772 | 2011 | 108.960 |  |

Catches within what has become the Endeavour dogfish closure have been less than 4 tonnes since 2004. Once all data from this area are removed from the Royal Red Prawn data a standardization demonstrated no appreciable difference from the trend exhibited by using all data.


Figure 20.17. The standardization of all Royal Red Prawn
20.5.8 Silver Trevally (TRE - 37337062 - Pseudocaranx dentex)

Table 20.25 Silver Trevally data for the TIER 4 calculations. Total is the sum of Discards, State, Non Trawl and SEF2 catches. All values in Tonnes. CE is the standardized catch rate for Zones 10 and 20 from depths 0 to 200 m (Haddon, 2012) with records from the Bateman's Bay MPA removed. GeoMean is the geometric mean catch rates. Discards are estimates from the ISMP from 1998 to present. The ratio of discards to catch over the 1998 - 2006 period was used to estimate the discards between 1986 and 1997, the proportion of which is the PDiscard.

| Year | Catch Discards | Total | State | Non-T PDiscard | CE | GeoM |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 1166.400 | 5.413 | 1171.813 |  |  | 0.46 | 1.1373 | 17.0086 |
| 1987 | 1142.400 | 5.301 | 1147.701 |  |  | 0.46 | 1.3565 | 17.5072 |
| 1988 | 1226.400 | 5.691 | 1232.091 |  |  | 0.46 | 1.7779 | 23.7642 |
| 1989 | 1394.400 | 6.471 | 1400.871 |  |  | 0.46 | 1.8873 | 23.0657 |
| 1990 | 1587.600 | 7.367 | 1594.967 |  |  | 0.46 | 2.2336 | 23.2975 |
| 1991 | 990.000 | 4.594 | 994.594 |  |  | 0.46 | 2.0038 | 18.1137 |
| 1992 | 949.200 | 4.405 | 953.605 |  |  | 0.46 | 1.1420 | 12.0774 |
| 1993 | 1030.800 | 4.783 | 1035.583 |  |  | 0.46 | 1.2467 | 13.4863 |
| 1994 | 842.815 | 3.911 | 846.726 | 711.358 | 0.000 | 0.46 | 0.9557 | 9.4912 |
| 1995 | 1001.628 | 4.648 | 1006.276 | 799.748 | 0.000 | 0.46 | 1.0853 | 10.2789 |
| 1996 | 1025.880 | 4.761 | 1030.640 | 810.673 | 0.000 | 0.46 | 0.8718 | 7.5806 |
| 1997 | 794.220 | 3.686 | 797.905 | 626.612 | 0.526 | 0.46 | 0.8265 | 6.2012 |
| 1998 | 648.496 | 0.000 | 648.496 | 536.581 | 12.215 | 0.00 | 0.6064 | 5.2414 |
| 1999 | 492.585 | 2.000 | 494.585 | 412.781 | 7.275 | 0.40 | 0.6057 | 4.9696 |
| 2000 | 500.297 | 0.000 | 500.297 | 405.277 | 2.707 | 0.00 | 0.4529 | 3.6777 |
| 2001 | 646.433 | 9.000 | 655.433 | 490.555 | 2.170 | 1.37 | 0.5290 | 4.1345 |
| 2002 | 521.838 | 1.100 | 522.938 | 361.519 | 2.444 | 0.21 | 0.4299 | 3.0864 |
| 2003 | 528.815 | 1.510 | 530.325 | 402.604 | 2.452 | 0.28 | 0.4218 | 3.3755 |
| 2004 | 659.720 | 7.400 | 667.120 | 519.086 | 2.036 | 1.11 | 0.5836 | 4.5401 |
| 2005 | 513.373 | 0.100 | 513.473 | 416.717 | 0.640 | 0.02 | 0.5154 | 4.7971 |
| 2006 | 429.737 | 1.820 | 431.557 | 358.778 | 2.045 | 0.42 | 0.7212 | 5.7178 |
| 2007 | 369.851 | 3.065 | 372.916 | 303.373 | 2.070 | 0.82 | 0.8211 | 7.4274 |
| 2008 | 296.810 | 2.460 | 299.270 | 185.746 | 0.319 | 0.82 | 0.8476 | 8.0833 |
| 2009 | 324.382 | 0.000 | 324.382 | 167.808 | 0.740 | 0.00 | 0.8553 | 9.2632 |
| 2010 | 386.444 | 0.160 | 375.400 | 164.161 | 0.302 | 0.04 | 1.0954 | 11.7000 |
| 2011 | 331.176 | 11.955 | 375.400 | 125.817 | 0.122 | 3.18 | 0.9902 | 11.0945 |

Discards make up approximately $0.16 \%$ of the catch over the 1998-2006 period.
Silver Trevally exhibited a period of high catch rates during 1989-1991 which were the result of a set of highly efficient vessels entering the fishery. These catch rates were considered not to represent a sustainable fishery and are not expected to be repeated. Therefore 1992-2001 was selected by the RAG as being a more representative reference period. In addition, the coastal waters within the Bateman's Bay MPA were removed from consideration during the catch rate standardization; the catches were deemed possible as fish could move from the MPA, but catch rates are not expected to be so high outside the MPA.

Table 20.26 RBC calculations for Silver Trevally. $\mathrm{C}_{\text {targ }}$ and CPUE $_{\text {targ }}$ relate to the period 1992-2001, $\mathrm{CPUE}_{\mathrm{Lim}}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7).

| Ref_Year | $1992-2001$ |
| ---: | ---: |
| CE_Targ | 0.8322 |
| CE_Lim | 0.3329 |
| CE_Recent | 0.9471 |
| Wt_Discard (t) | 6.582 |
| Scaling | 1.2302 |
| Last Year's TAC | 540.000 |
| Ctarg | 796.955 |
| RBC | $\mathbf{9 8 0 . 3 8 4}$ |

## SilverTrevally






Figure 20.18 Silver Trevally. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates (with records within the Bateman's Bay MPA removed; Haddon, 2012) with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 20.5.9 Ribaldo (RBD - 37224002 - Mora moro)

It was decided that this year the option of treating Ribaldo as one of the primary target species would be examined. This entailed changing the implied target reference point from $48 \%$ of the unfished state to $40 \%$ of the unfished state. Because the target catch rate is taken as a proxy for $48 \%$ unfished biomass, to make it equivalent to $40 \%$ means the average catch rate over the reference period should be multiplied by 0.8333 (thus $0.83334 \times 48=40$ ).

Table 20.27 Ribaldo data for the TIER 4 calculations. Total is the sum of Discards, State, Non Trawl and SEF2 catches. All values in Tonnes. CE is the standardized catch rate for Zones 10 to 50 in depths $0-$ 1000 m (Haddon, 2012). GeoMean is the geometric mean catch rates. Discards are estimates from 1998 to present. The ratio of discards to catch over the 1998 - 2006 period was used to estimate the discards between 1986 and 1997, the proportion of which is the PDiscard.

| Year | Catch | Discards | Total | State | Non-T | PDiscard | CE | GeoMean |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 4.800 | 0.723 | 5.523 |  |  | 13.09 | 2.2797 | 14.6630 |
| 1987 | 8.400 | 1.265 | 9.665 |  |  | 13.09 | 1.2772 | 10.2593 |
| 1988 | 8.400 | 1.265 | 9.665 |  |  | 13.09 | 2.0037 | 16.5570 |
| 1989 | 8.400 | 1.265 | 9.665 |  |  | 13.09 | 1.8029 | 18.2556 |
| 1990 | 2.400 | 0.362 | 2.762 |  |  | 13.09 | 1.4196 | 8.9113 |
| 1991 | 7.200 | 1.085 | 8.285 |  |  | 13.09 | 1.3647 | 7.9930 |
| 1992 | 15.600 | 2.350 | 17.950 |  |  | 13.09 | 1.3480 | 9.7616 |
| 1993 | 36.000 | 5.423 | 41.423 |  |  | 13.09 | 1.1172 | 11.2449 |
| 1994 | 28.021 | 0.063 | 28.021 | 0.418 | 0.000 | 13.09 | 1.2565 | 11.8156 |
| 1995 | 95.719 | 0.814 | 95.719 | 5.401 | 0.000 | 13.09 | 1.3011 | 12.3128 |
| 1996 | 85.154 | 0.529 | 85.154 | 3.510 | 0.000 | 13.09 | 1.0009 | 10.1757 |
| 1997 | 103.704 | 0.907 | 103.704 | 4.057 | 1.962 | 13.09 | 0.8776 | 9.8023 |
| 1998 | 95.427 | 23.766 | 119.193 | 0.102 | 2.431 | 90.37 | 0.8530 | 9.6696 |
| 1999 | 64.076 | 6.555 | 70.631 | 0.031 | 3.335 | 66.07 | 0.7871 | 8.7093 |
| 2000 | 63.117 | 8.284 | 71.401 | 0.022 | 8.736 | 48.61 | 0.7152 | 7.4217 |
| 2001 | 75.565 | 4.468 | 80.033 | 0.303 | 21.161 | 17.23 | 0.6655 | 6.7639 |
| 2002 | 171.727 | 7.305 | 179.033 | 0.000 | 95.820 | 4.08 | 0.6229 | 6.7944 |
| 2003 | 205.908 | 26.457 | 232.365 | 0.037 | 103.460 | 11.39 | 0.6103 | 6.7153 |
| 2004 | 199.188 | 16.087 | 215.275 | 0.061 | 102.509 | 7.47 | 0.6613 | 7.2233 |
| 2005 | 105.471 | 21.800 | 127.271 | 0.118 | 52.297 | 29.37 | 0.5715 | 6.3488 |
| 2006 | 116.822 | 3.100 | 119.921 | 0.000 | 73.324 | 2.58 | 0.6153 | 6.3304 |
| 2007 | 61.126 | 0.451 | 61.577 | 0.000 | 36.371 | 0.73 | 0.4015 | 3.2493 |
| 2008 | 97.215 | 2.629 | 99.843 | 0.000 | 70.985 | 2.63 | 0.5556 | 4.7326 |
| 2009 | 134.086 | 3.626 | 137.712 | 0.000 | 86.624 | 2.63 | 0.6149 | 5.6978 |
| 2010 | 111.395 | 1.955 | 113.350 | 0.000 | 65.348 | 1.72 | 0.6321 | 5.5851 |
| 2011 | 116.712 | 7.076 | 123.789 | 0.030 | 56.931 | 5.72 | 0.6447 | 5.8331 |

Discards make up approximately $13.1 \%$ of the catch over the 1998-2006 period.
There was no significant effect on the catch rate standardization of whether a shot was within or outside of one of the current closures (Haddon, 2011). As the standardized catch rate trend was indistinguishable from the series without the spatial factor it was not included.

Table 20.28 RBC calculations for Ribaldo. $\mathrm{C}_{\text {targ }}$ and $\mathrm{CPUE}_{\text {targ }}$ relate to the period 1995-2004, $\mathrm{CPUE}_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7).

| Ref_Year | $1995-2004$ |
| ---: | ---: |
| CE_Targ $^{\text {CE_Lim }}$ | 0.4047 |
| CE_Recent | 0.1619 |
| Wt_Discard | 0.6118 |
| Scaling | 4.954 |
| Last Year's TAC | 1.8527 |
| C $_{\text {targ }}$ | 168 |
| RBC | 125.251 |
|  | $\mathbf{2 3 2 . 0 5 4}$ |

## Ribaldo



Figure 20.19 Ribaldo. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

Table 20.29 RBC calculations for Ribaldo. $\mathrm{C}_{\text {targ }}$ and CPUE $_{\text {targ }}$ relate to $83.33 \%$ of the average over 1995-2004, CPUE $_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7). The proxy target is here $\mathrm{B} 40 \%$. The $50 \%$ rule may be required if this RBC is used.

| Ref_Year | $1995-2004$ |
| ---: | ---: |
| CE_Targ | 0.3373 |
| CE_Lim | 0.1619 |
| CE_Recent | 0.6118 |
| Wt_Discard | 4.954 |
| Scaling | 2.5653 |
| Year's TAC | 168 |
| C | 125.251 |
| RBC | $\mathbf{3 2 1 . 3 0 9}$ |

Ribaldo


Figure 20.20 Ribaldo. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine purple line representing the target catch rate $(83.33 \%$ of the average over the reference period) and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 20.6 Deep-Water

### 20.6.1 Summary Results

Table 20.30. Summary statistics from the TIER 4 analyses for each fishery.

| Fishery | Ref Years | Target Catch | RBC | $\mathbf{5 0 \%}$ Meta |
| :--- | :---: | ---: | ---: | ---: |
| Smooth Oreo Cascade | $1996-2005$ | 199.260 | 710.823 | Yes |
| Smooth Oreo Non- | $1989-1998(-1992)$ | 23.264 | 20.638 | Yes |
| Cascade | $1993-2001$ | 151.822 | 120.412 | Yes |
| Mixed Oreo | $1995-2004$ | 124.207 | 374.417 | Yes |
| Western Deepwater Shark | 105.307 | 89.511 |  |  |
| Eastern Deepwater Shark | $1995-2004$ | 127.620 | 189.548 | Yes |
| Alfonsino | $2001-2005$ |  |  |  |

### 20.6.2 Oreos General

Table 20.31. The catch of all species of Oreos in tonnes reported in each fishery. GAB is the Great Australian Bight, SET is the South East Fishery, and HST is High Seas Trawl STR is South Tasman Rise fishery and the WDW is the Western Deep Water Trawl fishery.

| Year | GAB | GHT | HSN | HST | SEN | SET | SPF | SSF | STR | VIT | WDW |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 |  |  |  |  |  | 56.636 |  |  |  |  |  |
| 1987 | 0.581 |  |  |  |  | 89.630 |  |  |  |  |  |
| 1988 | 67.935 |  |  |  |  | 89.242 |  |  |  |  |  |
| 1989 | 215.481 |  |  |  |  | 533.720 |  |  |  |  |  |
| 1990 | 10.178 |  |  |  |  | 1090.260 |  |  |  |  |  |
| 1991 | 6.982 |  |  |  |  | 1129.201 |  |  |  |  |  |
| 1992 | 94.219 |  |  |  |  | 3201.806 |  |  |  |  | 58.000 |
| 1993 | 2.780 |  |  |  |  | 1036.616 |  |  |  |  | 58.030 |
| 1994 | 48.184 |  |  |  |  | 1043.359 |  |  |  |  | 20.795 |
| 1995 | 0.730 |  |  |  |  | 1025.771 |  |  |  |  | 1.186 |
| 1996 | 5.264 |  |  |  |  | 771.783 |  |  |  |  | 8.268 |
| 1997 | 39.757 |  |  |  |  | 2050.730 |  |  |  |  | 0.635 |
| 1998 | 20.916 |  |  |  | 0.009 | 2021.332 |  |  |  |  |  |
| 1999 | 20.437 |  |  | 2.896 | 0.019 | 882.455 |  |  |  |  |  |
| 2000 | 49.187 |  |  |  | 0.001 | 1010.255 |  | 0.100 |  |  | 0.111 |
| 2001 | 12.647 |  |  |  | 0.007 | 1079.123 |  |  | 25.450 |  | 4.314 |
| 2002 | 0.580 |  | 0.007 | 24.389 | 0.137 | 828.422 |  |  | 2.500 |  |  |
| 2003 | 5.678 | 0.527 |  | 129.630 |  | 750.909 |  |  |  | 0.070 |  |
| 2004 | 8.782 | 0.702 |  | 168.647 |  | 432.483 |  |  | 32.683 |  | 0.633 |
| 2005 | 24.215 | 0.807 |  | 92.576 |  | 233.887 |  |  | 151.600 |  |  |
| 2006 | 16.621 | 1.168 |  | 0.246 |  | 173.732 | 0.034 |  | 22.520 |  |  |
| 2007 | 3.447 | 0.823 |  | 1.224 |  | 129.664 |  |  |  |  |  |
| 2008 | 0.275 | 0.685 |  |  |  | 77.386 |  |  |  |  | 0.020 |
| 2009 | 1.796 | 1.958 |  | 101.491 |  | 85.975 |  |  |  |  |  |
| 2010 | 1.180 | 1.047 |  | 146.562 |  | 89.314 |  |  |  |  |  |
| 2011 | 0.080 | 0.400 |  | 4.579 |  | 101.976 |  |  |  |  |  |

Table 20.32. The catch of each recognized species of Oreos in tonnes reported in the GenLog (SEF1) database. Smooth and Spiky Oreos are the most commonly reported.

|  | Oreo | Spiky | Oxeye | Smooth | Warty | Black | Oreo <br> Dory |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | 37266000 | 37266001 | 37266002 | 37266003 | 37266004 | 37266005 | 37266902 |

### 20.6.3 Smooth Oreo (Cascade) (DOO - 37266003 - Smooth Oreo Pseudocyttus maculatus and DOE 37266902 - Oreo Dory)

After examination of the depth distribution of records, only data from OR Zone 40 in depths $650-1250 \mathrm{~m}$ were used. All vessels recording smooth oreos in orange roughy zone 40 were included in the analysis. The discard rate estimated in 2007 was $12.3 \%$. Catch rates as Kg /Tow.

Table 20.33. Number of records where Smooth Oreos or Oreos (CAAB codes 37266003 , and 37266902
$=$ Smooth Oreo, and Oreo Dory) on the Cascade are reported by trawling in OR Zone 40, in depths 650 to 1250 m . Used are the number of records excluding those reported as being in the 700 m closure. Vessels represent the count of vessels reporting oreos. Effort H and CatchT are the reported effort and catch of Smooth Oreos from the used records. The geometric mean CE is the raw unstandardized catch rate in $\mathrm{Kg} /$ tow. StandCE is the standardized catch rates and StErrCE is the standard error of the standardized catch rates (Figure 20.24).

| Year | Records | Vessels | Effort <br> H | CatchT | Geo Mean CE | StandCE | StErrCE |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1989 | 211 | 5 | 120.80 | 127.768 | 267.387 |  |  |
| 1990 | 296 | 7 | 126.30 | 91.494 | 146.934 |  |  |
| 1991 | 7 | 1 | 2.70 | 1.060 | 86.926 |  |  |
| 1992 | 13 | 4 | 7.55 | 11.320 | 426.816 |  |  |
| 1993 | 19 | 1 | 7.96 | 2.098 | 50.017 |  |  |
| 1994 | 24 | 4 | 140.02 | 94.474 | 142.348 | 0.5044 | 0 |
| 1995 | 94 | 6 | 88.44 | 14.288 | 49.713 | 0.3827 | 0.3282 |
| 1996 | 457 | 8 | 311.20 | 142.244 | 64.177 | 0.4944 | 0.3574 |
| 1997 | 305 | 7 | 185.87 | 281.722 | 99.386 | 0.5764 | 0.3784 |
| 1998 | 166 | 8 | 126.66 | 103.366 | 128.204 | 0.6806 | 0.3833 |
| 1999 | 94 | 9 | 52.75 | 98.568 | 191.733 | 0.9863 | 0.4000 |
| 2000 | 358 | 10 | 240.07 | 295.843 | 195.144 | 0.9112 | 0.3678 |
| 2001 | 216 | 9 | 109.39 | 276.287 | 234.844 | 1.2054 | 0.3770 |
| 2002 | 354 | 9 | 118.38 | 284.595 | 110.842 | 0.5629 | 0.3635 |
| 2003 | 161 | 7 | 63.81 | 104.069 | 139.562 | 0.6631 | 0.3877 |
| 2004 | 116 | 5 | 27.73 | 100.785 | 375.609 | 2.2969 | 0.3912 |
| 2005 | 88 | 5 | 35.19 | 60.033 | 149.794 | 1.1473 | 0.4051 |
| 2006 | 46 | 3 | 10.94 | 61.300 | 288.216 | 1.3624 | 0.4395 |
| 2007 | 53 | 2 | 28.49 | 45.408 | 168.150 | 1.1275 | 0.4328 |
| 2008 | 85 | 3 | 50.72 | 16.245 | 44.721 | 0.8945 | 0.4395 |
| 2009 | 35 | 2 | 18.85 | 2.485 | 41.907 | 0.6613 | 0.4780 |
| 2010 | 29 | 2 | 27.13 | 7.315 | 144.194 | 2.1550 | 0.5002 |
| 2011 | 10 | 2 | 7.99 | 1.320 | 73.602 | 1.3877 | 0.6754 |


| Table 20.34. Catches and numbers of records for Smooth Oreo (CAAB 37266003) and, from 2006, for |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| both Smooth Oreo and the new category Oreo Dory (CAAB 37266902). |  |  |  |  |  |  |  |
| Year | Smooth | Smooth | Year | Smooth | Smooth | OreoDory | OreoDory |
| 1989 | 127.768 | 211 | 2006 | 60.910 | 34 | 0.390 | 8 |
| 1990 | 91.494 | 296 | 2007 | 43.698 | 32 | 1.710 | 16 |
| 1991 | 1.060 | 7 | 2008 | 12.365 | 14 | 3.880 | 71 |
| 1992 | 11.320 | 13 | 2009 | 0.060 | 3 | 2.425 | 32 |
| 1993 | 2.098 | 19 | 2010 | 3.200 | 5 | 4.115 | 24 |
| 1994 | 94.474 | 241 | 2011 |  |  | 1.320 | 10 |
| 1995 | 14.288 | 94 |  |  |  |  |  |
| 1996 | 142.244 | 457 |  |  |  |  |  |
| 1997 | 281.722 | 305 |  |  |  |  |  |
| 1998 | 103.366 | 166 |  |  |  |  |  |
| 1999 | 98.568 | 94 |  |  |  |  |  |
| 2000 | 295.843 | 358 |  |  |  |  |  |
| 2001 | 276.287 | 216 |  |  |  |  |  |
| 2002 | 284.595 | 354 |  |  |  |  |  |
| 2003 | 104.069 | 161 |  |  |  |  |  |
| 2004 | 100.785 | 116 |  |  |  |  |  |
| 2005 | 60.033 | 88 |  |  |  |  |  |

Table 20.35. Statistical model structures used in this analysis. DepCat is a series of 50 metre depth categories.

```
Model Year
1
Model Year + Vessel
2
Model Year + Vessel + DepCat
3
Model Year + Vessel + DepCat + Month
4
Model Year + Vessel + DepCat + Month + DayNight
5
Model Year + Vessel + DepCat + Month + DayNight + Vessel:Month
6
Model Year + Vessel + DepCat + Month + DayNight + DepCat:Month
7
```

Table 20.36. Model selection criteria, including the AIC and other diagnostics. Smooth Oreos (Cascade). The model with the smallest AIC and largest Adjusted $r^{2}$ is accepted as best. RSS is residual sum of squares, MSS is Model sum of squares.

|  | Year | Vessel | DepCat | Month | DayNight | Vessel:Mth | DepCat:Mth |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 3298 | 3193 | 2809 | 2800 | $\mathbf{2 7 9 7}$ | 2895 | 2914 |
| RSS | 8929 | 8529 | 7411 | 7331 | $\mathbf{7 3 1 0}$ | 6797 | 6946 |
| MSS | 799 | 1198 | 2317 | 2397 | $\mathbf{2 4 1 8}$ | 2931 | 2782 |
| Nobs | 2908 | 2908 | 2895 | 2895 | $\mathbf{2 8 9 5}$ | 2895 | 2895 |
| Npars | 18 | 32 | 44 | 55 | $\mathbf{5 8}$ | 212 | 190 |
| Adj_r2 | 7.672 | 11.375 | 22.666 | 23.208 | $\mathbf{2 3 . 3 4 8}$ | 24.635 | 23.606 |
| $\Delta \mathrm{r}^{2}$ | 0.000 | 3.703 | 11.291 | 0.541 | $\mathbf{0 . 1 4 1}$ | 1.287 | -1.029 |



Figure 20.21. Smooth Oreo (Cascade) are reported from trawling in OR Zone 40, in depths 650 to 1250 m . The top left is the depth distribution of all records reporting Smooth Oreo (not just Cascade), the top right graph depicts the depth distribution of shots containing Smooth Oreo in OR Zone 40 and depths $650-1250 \mathrm{~m}$. The middle left diagram depicts the distribution of catch across all years by depth within OR zone 40 , the right hand middle graph depicts the number of vessels reporting smooth oreos through time. The bottom left reflects the number of records used in analysis, and bottom right are the Smooth Oreo catches used in the analysis.


Figure 20.22. The catch by month for each year of smooth oreos on the Cascade from 1989-2011. Each axis is identical.


Figure 20.23. Catch in tonnes by depth category (in metres) for smooth oreos on the cascade from 1989 2011.


Figure 20.24. Standardized catch rates for the Smooth Oreo on the Cascade. The geometric mean catch rates are depicted as a dashed line, while the standardized catch rates are the solid line. Numbers of data points before 1994 were too few for standardization. The error bars are two times the Standard Errors.

### 20.6.3.1 TIER 4 Smooth Oreo (Cascade)

It is very doubtful whether the catch rate values for 2009 - 2011 are valid as there were so few data points, especially in 2011. In addition, the extremely rapid changes in apparent catch rates indicates that the observed catch rates are unlikely to be representative of the stock size, so the validity of applying even a TIER 4 needs to be questioned. The error bars illustrated in Figure 20.24 are expected to overestimate the certainty with which the mean catch rates are estimated, which suggests that the catch rates have not deviated significantly from each other since 1994 (despite the large changes in the apparent mean catch rate). Catches were so small because the deepwater fishery is barely being pursued anywhere. Because the catches were so small it would not have been valid to update the TIER 4 analysis, which is in-line with a RAG decision to only update the Tier 4 assessment if there were more than 10 t of catch taken. Despite the lack of assessment there were no signs of stress in these fishery data in terms of the distribution of catches or the catch rates of those catches that were reported.

### 20.6.4 Smooth Oreo (non-Cascade) (DOO - 37266003 - Pseudocyttus maculates)

After examination of the depth distribution of records, only data from OR Zones 10, 20, 21, 30, and 50, taken by trawl in the SET fishery in depths $400-1200 \mathrm{~m}$ were used. All vessels recording smooth oreos were included in the analysis. The Cascade, GAB and zone 70 Smooth Oreos were excluded. The discard rate estimated in 2007 was $12.3 \%$ and this was assumed for other years. The ratio of catches inside relative to outside the current closures is $84.9 \%$ versus $15.7 \%$ out of a total of 7236 t considered in the analysis.

Table 20.37. Number of records where Smooth Oreos not on the Cascade are reported from trawling in OR Zones 10, 20, 21, 30, 50, in depths 400 to 1200 m . Vessels represents the count of vessels reporting smooth oreos. CatchT is the reported catch of Smooth Oreos. The geometric mean CE is the raw unstandardized catch rate in $\mathrm{Kg} /$ tow. The left hand five columns represent data, in both the closed and currently open areas the right hand five columns (post-fixed O) represent the areas left open following the 700 m closure.

| Year | Records Vessels | Effort | Yield | Geom RecordsO VesselsO | EffortO | YieldO | GeomO |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1987 | 33 | 3 | 74.3 | 6.250 | 118.343 | 27 | 2 | 61.8 | 4.660 | 112.932 |
| 1988 | 41 | 9 | 71.8 | 39.363 | 232.252 | 15 | 6 | 21.0 | 5.218 | 144.408 |
| 1989 | 247 | 22 | 151.9 | 177.234 | 209.771 | 25 | 8 | 33.2 | 8.855 | 136.437 |
| 1990 | 648 | 38 | 478.5 | 715.045 | 302.562 | 54 | 12 | 35.9 | 62.269 | 382.833 |
| 1991 | 667 | 34 | 689.3 | 904.830 | 242.766 | 134 | 20 | 236.2 | 119.733 | 129.486 |
| 1992 | 1327 | 30 | 1062.7 | 2216.456 | 396.338 | 231 | 21 | 289.5 | 384.811 | 206.685 |
| 1993 | 999 | 31 | 691.2 | 605.649 | 136.366 | 95 | 19 | 140.4 | 68.926 | 97.532 |
| 1994 | 1068 | 26 | 743.7 | 574.904 | 93.488 | 109 | 18 | 171.7 | 43.981 | 91.736 |
| 1995 | 667 | 21 | 1175.5 | 493.353 | 114.545 | 76 | 11 | 260.6 | 34.425 | 105.413 |
| 1996 | 498 | 18 | 810.0 | 171.377 | 72.869 | 77 | 15 | 178.3 | 13.503 | 54.227 |
| 1997 | 407 | 20 | 774.9 | 153.412 | 108.713 | 77 | 16 | 223.8 | 21.482 | 107.409 |
| 1998 | 342 | 19 | 900.8 | 134.877 | 114.236 | 59 | 16 | 200.4 | 28.092 | 116.670 |
| 1999 | 278 | 21 | 1043.9 | 61.895 | 101.167 | 51 | 13 | 253.1 | 5.444 | 60.900 |
| 2000 | 314 | 23 | 1133.2 | 91.490 | 94.029 | 80 | 16 | 375.5 | 19.153 | 71.681 |
| 2001 | 520 | 23 | 2017.3 | 282.152 | 175.312 | 194 | 22 | 844.2 | 86.807 | 159.792 |
| 2002 | 516 | 22 | 2538.8 | 222.806 | 132.965 | 163 | 19 | 876.1 | 56.186 | 109.442 |
| 2003 | 444 | 17 | 2008.6 | 166.908 | 114.728 | 141 | 14 | 788.4 | 40.513 | 90.968 |
| 2004 | 404 | 18 | 1987.7 | 110.666 | 95.065 | 126 | 16 | 655.9 | 32.213 | 101.907 |
| 2005 | 191 | 10 | 762.7 | 53.557 | 89.466 | 60 | 9 | 295.9 | 12.648 | 69.210 |
| 2006 | 26 | 7 | 49.7 | 15.019 | 113.430 | 11 | 4 | 44.2 | 0.589 | 13.588 |
| 2007 | 8 | 2 | 3.5 | 0.886 | 73.216 | 3 | 2 | 2.7 | 0.156 | 49.716 |
| 2008 | 3 | 2 | 19.3 | 0.910 | 125.992 | 3 | 2 | 19.3 | 0.910 | 125.992 |
| 2009 | 15 | 8 | 49.0 | 1.295 | 47.042 | 14 | 7 | 43.0 | 1.265 | 48.579 |
| 2010 | 11 | 4 | 48.9 | 0.579 | 32.832 | 11 | 4 | 48.9 | 0.579 | 32.832 |
| 2011 | 17 | 7 | 104.7 | 4.727 | 92.224 | 17 | 7 | 104.7 | 4.727 | 92.224 |



Figure 20.25. Smooth Oreo (Non-Cascade) are reported from trawling in OR Zones 10, 20, 21, 30, and 50 , in depths 600 to 1200 m . The top left is the depth distribution of all records reporting Smooth Oreo (not just Cascade), the top right graph depicts the depth distribution of shots containing Smooth Oreo (non-Cascade) in OR Zones 10, 20, 21, 30, and 50, in depths 600 to 1200 m . The middle left diagram depicts the distribution of catch across all years by depth within OR zones, the right hand middle graph depicts the number of vessels reporting smooth oreos through time. The bottom left reflects the number of records for the non-Cascade, and bottom right are the Smooth Oreo catches used in the analysis.


Figure 20.26. Catch by month for smooth oreo from non-Cascade areas (except GAB and Zone 70) from 1989-2011. Catches since 2006 have been very low.


Figure 20.27. Standardized catch rates for the Smooth Oreo not on the Cascade. The geometric mean catch rates are depicted as a dashed line, while the standardized catch rates are the solid line. The error bars are two times the Standard Errors. The times series are scaled to the mean of each series for visual comparison. Data since 2007 has been minimal.

Table 20.38. Model selection criteria, including the AIC, the residual sum of squares, the Model sum of squares, the number of usable observations, the number of parameters, the adjusted $r^{2}$ and the increment in adjusted $\mathrm{r}^{2}$. The complete model was optimal. The effect of being in the open or closed areas (Closure) was only minor.

|  | Year | Vessel | DepCat | Month | ORZone |  | Closure | DayNight | DepCat:Month |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 10568 | 9783 | 9381 | 9355 | 9340 | 9325 | 9316 | $\mathbf{9 3 1 3}$ |  |
| RSS | 28689 | 25956 | 24828 | 24704 | 24645 | 24602 | 24563 | $\mathbf{2 3 8 9 0}$ |  |
| MSS | 2741 | 5474 | 6601 | 6726 | 6784 | 6827 | 6866 | $\mathbf{7 5 4 0}$ |  |
| Nobs | 9691 | 9691 | 9618 | 9618 | 9618 | 9618 | 9618 | $\mathbf{9 6 1 8}$ |  |
| Npars | 25 | 118 | 130 | 141 | 145 | 146 | 149 | $\mathbf{2 8 1}$ |  |
| adj_r $^{2}$ | 8.494 | 16.407 | 19.929 | 20.238 | 20.394 | 20.525 | 20.625 | $\mathbf{2 1 . 7 1 0}$ |  |
| $\Delta \mathrm{r}^{2}$ | 0.000 | 7.913 | 3.522 | 0.309 | 0.156 | 0.131 | 0.100 | $\mathbf{1 . 0 8 6}$ |  |

Table 20.39. The standardized catch rates for the alternative statistical models for Smooth Oreos in OR zones $10,20,21,30$, and 50 , in depths 600 to 1200 m . The optimal model was included all factors to DayNight (not Closure). St Err is the estimate of standard error for the optimum model. Values are relative to the mean of the standardized catch rates. Month is omitted for brevity. Note the relatively large standard errors, which imply the trend does not differ from 1.0 since 1993.

| Year | Year | Vessel | DepCat | ORZone | Closure | DayNight | DepCat:Month | StErr |
| ---: | ---: | ---: | :--- | :--- | :--- | :--- | ---: | ---: | ---: |
| 1987 | 0.7942 | 0.9492 | 0.7319 | 0.7634 | 0.8381 | 0.8269 | $\mathbf{0 . 9 3 0 0}$ | 0.0000 |
| 1988 | 1.6905 | 1.3412 | 1.2717 | 1.3302 | 1.3950 | 1.3535 | $\mathbf{1 . 2 8 0 3}$ | 0.4417 |
| 1989 | 1.4815 | 1.3067 | 1.4426 | 1.4952 | 1.4619 | 1.4319 | $\mathbf{1 . 3 9 8 1}$ | 0.3418 |
| 1990 | 2.1288 | 2.1056 | 2.0817 | 2.0305 | 1.9741 | 1.9463 | $\mathbf{2 . 0 4 5 2}$ | 0.3275 |
| 1991 | 1.7080 | 2.0953 | 1.9875 | 1.8502 | 1.8148 | 1.7909 | $\mathbf{1 . 9 2 7 1}$ | 0.3283 |
| 1992 | 2.7854 | 2.8512 | 2.9614 | 2.8050 | 2.7374 | 2.7579 | $\mathbf{2 . 7 5 4 6}$ | 0.3254 |
| 1993 | 0.9587 | 1.1058 | 1.2181 | 1.1800 | 1.1479 | 1.1588 | $\mathbf{1 . 1 7 7 7}$ | 0.3271 |
| 1994 | 0.6572 | 0.6891 | 0.7926 | 0.7952 | 0.7763 | 0.7850 | $\mathbf{0 . 8 0 9 6}$ | 0.3274 |
| 1995 | 0.8059 | 0.7798 | 0.8533 | 0.8514 | 0.8297 | 0.8366 | $\mathbf{0 . 8 4 6 0}$ | 0.3293 |
| 1996 | 0.5131 | 0.4681 | 0.5161 | 0.5246 | 0.5137 | 0.5171 | $\mathbf{0 . 5 2 5 3}$ | 0.3324 |
| 1997 | 0.7659 | 0.8202 | 0.8835 | 0.8849 | 0.8687 | 0.8788 | $\mathbf{0 . 8 8 5 8}$ | 0.3344 |
| 1998 | 0.8054 | 0.6994 | 0.7016 | 0.7282 | 0.7093 | 0.7157 | $\mathbf{0 . 7 2 8 1}$ | 0.3372 |
| 1999 | 0.7140 | 0.6585 | 0.6555 | 0.6809 | 0.6641 | 0.6694 | $\mathbf{0 . 7 0 2 2}$ | 0.3402 |
| 2000 | 0.6632 | 0.7823 | 0.8235 | 0.8130 | 0.8015 | 0.8053 | $\mathbf{0 . 8 1 7 9}$ | 0.3383 |
| 2001 | 1.2342 | 1.3053 | 1.2592 | 1.2565 | 1.2691 | 1.2826 | $\mathbf{1 . 3 0 9 6}$ | 0.3326 |
| 2002 | 0.9361 | 0.8766 | 0.8607 | 0.9083 | 0.8998 | 0.9115 | $\mathbf{0 . 9 1 7 8}$ | 0.3338 |
| 2003 | 0.8081 | 0.8486 | 0.8383 | 0.8882 | 0.8823 | 0.8914 | $\mathbf{0 . 9 0 6 8}$ | 0.3346 |
| 2004 | 0.6698 | 0.7806 | 0.7793 | 0.8285 | 0.8182 | 0.8237 | $\mathbf{0 . 8 4 1 4}$ | 0.3365 |
| 2005 | 0.6329 | 0.7299 | 0.6706 | 0.6874 | 0.6834 | 0.6849 | $\mathbf{0 . 7 3 1 0}$ | 0.3461 |
| 2006 | 0.8430 | 0.9402 | 0.7735 | 0.7804 | 0.7879 | 0.7893 | $\mathbf{0 . 5 8 0 6}$ | 0.4724 |
| 2007 | 0.6187 | 0.7805 | 0.6681 | 0.5989 | 0.6113 | 0.6215 | $\mathbf{0 . 6 5 0 5}$ | 0.6635 |
| 2008 | 1.4505 | 0.6834 | 0.6930 | 0.6918 | 0.7247 | 0.7249 | $\mathbf{0 . 5 1 9 1}$ | 1.1207 |
| 2009 | 0.3646 | 0.4831 | 0.3459 | 0.3991 | 0.4276 | 0.4342 | $\mathbf{0 . 4 5 8 7}$ | 0.5561 |
| 2010 | 0.2638 | 0.2557 | 0.2270 | 0.2466 | 0.2745 | 0.2715 | $\mathbf{0 . 2 6 5 4}$ | 0.5968 |
| 2011 | 0.7065 | 0.6640 | 0.9635 | 0.9816 | 1.0889 | 1.0905 | $\mathbf{0 . 9 9 1 1}$ | 0.5315 |

### 20.6.4.1 TIER 4 Smooth Oreo (Non-Cascade)

As with the Cascade smooth oreos assessment it is doubtful whether the catch rate value for 2011 is valid as there were less than 5 t of data that met the reporting requirements. It remains unknown whether catch rates reflect the stock status but there are so few records it appears highly unlikely. Certainly since 1993 the standard error estimates for the standardized catch rates are so large that the mean catch rates cannot be claimed to have differed significantly from 1.0 since 1993.

Again, as for Cascade smooth oreo, catches were so small because the deepwater fishery is barely being pursued anywhere. Because the catches were so small it would not have been valid to update the TIER 4 analysis, which is in-line with a RAG decision to only update the Tier 4 assessment if there were more than 10 t of catch taken. Despite the lack of assessment there were no signs of stress in this fishery in terms of the distribution of catches or the catch rates of those catches that were reported.

### 20.6.5 Mixed Oreo Basket (warty, spikey, rough, black, \& Oreo Dory)

Allocyttus verrucosus (warty), Neocyttus rhomboidalis (spiky), Neocyttus psilorhynchus (rough), Allocyttus niger black). CAAB codes: 37266004, 37266001, 37266006, 37266005, 37266901,37266902 (group code). Estimated discard rate in 2007 was $16.2 \%$ and there is no update on that figure. $97.01 \%$ of the reported catch is given as spikey oreo (Neocyttus rhomboidalis), $2.98 \%$ as warty oreo (Allocyttus verrucosus), and $0.01 \%$ as black oreo (Allocyttus niger)(Table 20.32). In the last five years $80-91 \%$ has been reported as Oreo Dory and the remainder as Spiky oreos. Only data from OR Zones 10, 20, 21, 30, 50, in depths $500-$ 1200 m were used, in particular only the data from outside the closures are used. All vessels recording mixed oreos were included in the analysis. Orange Roughy zones 40, 60, 70 and unknown were removed.

Table 20.40. Number of records where Mixed Oreos are reported from trawling in OR Zones 10, 20, 21, 30, and 50 , in depths 500 to 1200 m . Vessels represents the count of vessels reporting mixed oreos. Yield is the reported catch of mixed Oreos. The geometric mean CE is the raw unstandardized catch rate in $\mathrm{Kg} /$ tow. The left hand five columns represent all data, the right hand five columns represent the areas left open following the 700 m closure.

| Year | Recor <br> ds | Vessel | sffort | Yield | Geom |  | Record | Vessel | Effort | YieldO |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | SO | O |  |  |  |  |  |  |  |  |
| 1986 | 166 | 9 | 367 | 50.966 | 114.224 | 138 | 9 | 329 | 47.586 | 128.028 |
| 1987 | 145 | 16 | 353 | 59.909 | 133.794 | 84 | 12 | 217 | 17.390 | 108.044 |
| 1988 | 161 | 12 | 372 | 33.809 | 82.647 | 68 | 7 | 192 | 12.228 | 77.821 |
| 1989 | 352 | 18 | 497 | 189.239 | 137.647 | 114 | 10 | 263 | 25.771 | 103.141 |
| 1990 | 257 | 22 | 172 | 257.178 | 292.016 | 23 | 11 | 61 | 7.335 | 107.273 |
| 1991 | 215 | 22 | 532 | 86.887 | 85.155 | 113 | 16 | 389 | 18.421 | 72.479 |
| 1992 | 577 | 31 | 848 | 607.582 | 227.389 | 174 | 22 | 499 | 76.258 | 111.068 |
| 1993 | 832 | 38 | 1621 | 281.255 | 94.969 | 337 | 29 | 1144 | 80.648 | 111.752 |
| 1994 | 1077 | 34 | 2494 | 284.569 | 75.354 | 419 | 32 | 1543 | 97.882 | 86.332 |
| 1995 | 1766 | 30 | 6060 | 482.242 | 92.167 | 953 | 23 | 3835 | 311.961 | 128.068 |
| 1996 | 2107 | 33 | 6898 | 420.967 | 69.658 | 1237 | 32 | 4824 | 284.955 | 91.185 |
| 1997 | 2274 | 34 | 9607 | 572.827 | 103.523 | 1502 | 31 | 6813 | 387.711 | 115.469 |
| 1998 | 2348 | 33 | 9873 | 666.856 | 121.631 | 1455 | 30 | 6170 | 448.279 | 132.626 |
| 1999 | 1912 | 33 | 7905 | 441.017 | 105.804 | 1191 | 31 | 4968 | 313.340 | 120.753 |
| 2000 | 1726 | 38 | 7739 | 376.494 | 97.319 | 1033 | 36 | 4541 | 253.999 | 114.544 |
| 2001 | 1926 | 37 | 8622 | 399.034 | 98.900 | 1262 | 36 | 5714 | 247.178 | 101.183 |
| 2002 | 1457 | 36 | 7174 | 212.546 | 70.372 | 931 | 33 | 4597 | 145.658 | 75.006 |
| 2003 | 1462 | 30 | 7411 | 229.224 | 75.450 | 915 | 28 | 4685 | 145.208 | 77.220 |
| 2004 | 1445 | 30 | 7502 | 181.402 | 66.947 | 912 | 28 | 4802 | 121.256 | 72.045 |
| 2005 | 813 | 22 | 4271 | 101.266 | 64.123 | 553 | 20 | 2882 | 72.176 | 67.852 |
| 2006 | 643 | 23 | 3230 | 80.260 | 50.683 | 422 | 22 | 2168 | 53.096 | 53.582 |
| 2007 | 388 | 17 | 2026 | 58.754 | 55.456 | 340 | 17 | 1831 | 52.028 | 54.586 |
| 2008 | 305 | 16 | 1751 | 48.564 | 72.522 | 280 | 16 | 1602 | 42.937 | 70.213 |
| 2009 | 500 | 17 | 2743 | 73.639 | 65.057 | 455 | 17 | 2482 | 65.576 | 62.511 |
| 2010 | 508 | 15 | 2900 | 76.137 | 65.407 | 467 | 15 | 2683 | 62.542 | 60.014 |
| 2011 | 571 | 17 | 3514 | 78.262 | 76.354 | 529 | 17 | 3244 | 70.490 | 74.866 |


| Year | Total | 10 | 20 | 21 | 30 | 50 | Open | Closed |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 50.966 | 0.160 | 30.520 |  | 20.278 | 0.008 | 47.586 | 3.380 |
| 1987 | 59.909 | 0.130 | 6.470 |  | 53.309 |  | 17.390 | 42.519 |
| 1988 | 33.809 | 0.020 | 0.150 |  | 33.549 | 0.090 | 12.228 | 21.581 |
| 1989 | 189.239 | 0.030 | 98.650 | 37.090 | 53.409 | 0.060 | 25.771 | 163.468 |
| 1990 | 257.178 | 4.340 | 183.043 | 62.965 | 6.700 | 0.130 | 7.335 | 249.843 |
| 1991 | 86.887 | 3.191 | 47.720 | 17.251 | 18.340 | 0.385 | 18.421 | 68.466 |
| 1992 | 607.582 | 31.646 | 352.204 | 190.614 | 31.622 | 1.496 | 76.258 | 531.324 |
| 1993 | 281.255 | 1.392 | 106.148 | 36.651 | 107.769 | 29.295 | 80.648 | 200.607 |
| 1994 | 284.569 | 0.882 | 90.447 | 34.734 | 136.647 | 21.859 | 97.882 | 186.687 |
| 1995 | 482.242 | 1.388 | 64.172 | 8.076 | 402.359 | 6.247 | 311.961 | 170.281 |
| 1996 | 420.967 | 8.539 | 92.953 | 3.451 | 278.999 | 37.025 | 284.955 | 136.012 |
| 1997 | 572.827 | 43.955 | 129.864 | 1.390 | 377.317 | 20.301 | 387.711 | 185.116 |
| 1998 | 666.856 | 33.724 | 130.862 | 1.492 | 379.621 | 121.157 | 448.279 | 218.577 |
| 1999 | 441.017 | 13.860 | 126.159 | 1.295 | 241.554 | 58.149 | 313.340 | 127.677 |
| 2000 | 376.494 | 26.075 | 111.417 | 0.775 | 213.565 | 24.662 | 253.999 | 122.495 |
| 2001 | 399.034 | 17.880 | 134.639 | 7.785 | 218.687 | 20.043 | 247.178 | 151.856 |
| 2002 | 212.546 | 36.018 | 59.214 | 1.025 | 105.532 | 10.757 | 145.658 | 66.888 |
| 2003 | 229.224 | 33.272 | 57.005 | 7.550 | 118.164 | 13.233 | 145.208 | 84.016 |
| 2004 | 181.402 | 12.011 | 40.705 | 1.820 | 115.255 | 11.612 | 121.256 | 60.145 |
| 2005 | 101.266 | 5.967 | 22.182 | 1.500 | 62.499 | 9.118 | 72.176 | 29.090 |
| 2006 | 80.260 | 8.581 | 12.259 | 0.270 | 56.955 | 2.195 | 53.096 | 27.164 |
| 2007 | 58.754 | 2.340 | 18.565 | 1.194 | 35.345 | 1.310 | 52.028 | 6.726 |
| 2008 | 48.564 | 2.262 | 17.114 |  | 26.527 | 2.661 | 42.937 | 5.627 |
| 2009 | 73.639 | 4.105 | 17.271 | 0.058 | 48.027 | 4.178 | 65.576 | 8.063 |
| 2010 | 76.137 | 5.344 | 25.346 | 5.860 | 37.301 | 2.286 | 62.542 | 13.595 |
| 2011 | 78.262 | 3.643 | 20.661 | 1.990 | 48.064 | 3.904 | 70.490 | 7.772 |
| Total | 6350.885 | 300.755 | 1995.740 | 424.836 | 3227.394 | 402.161 | 3461.910 | 2888.975 |



Figure 20.28. Mixed Oreo are reported from trawling in OR Zones $10,20,21,30$, and 50 , in depths 500 to 1200 m . The top left is the depth distribution of all records reporting Mixed Oreo, the top right graph depicts the depth distribution of shots containing Mixed Oreo in OR Zones 10, 20, 21, 30, 50, in depths 500 to 1200 m . The middle left diagram depicts the distribution of catch across all years by depth within separate OR zones, the right hand middle graph depicts the number of vessels reporting mixed oreos through time. The bottom left reflects the number of records for mixed oreos, and bottom right are the Mixed Oreo catches used in the analysis.


Figure 20.29. Mixed Oreo are reported from trawling in OR Zones 10, 20, 21, 30, and 50, in depths 500 to 1200 m . The top left show the geometric mean catch rate. The top right graph depicts the catch from the open and closed areas. The bottom left graph depicts the number of hours effort recorded for the open and closed areas. Finally, the bottom right hand graph depicts the number of records/shots of effort for the open and closed areas.


Figure 20.30. The standardized catch rates showing the optimum model (solid black line) and the geometric mean catch rate (dashed line) each scaled to the mean of each time series. The error bars are two times the standard errors.


Figure 20.31. Relative impact of each factor on the final trend. Blue bars indicate the standardization is above the previous model, red bars indicate it is below. Closures appear to have only a very small effect.


Figure 20.32. A comparison of last year's standardization with this year's.

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Table 20.42. Statistical model structures used with Mixed Oreos. DepCat is a series of 50 metre depth
categories. Closure relates to whether the area is open or closed.
Model 1 Year
Model 2 Year + Vessel
Model 3 Year + Vessel + DepCat
Model 4 Year + Vessel + DepCat + Month
Model 5 Year + Vessel + DepCat + Month + ORZone
Model 6 Year + Vessel + DepCat + Month + ORZone + DayNight
Model 7 Year + Vessel + DepCat + Month + ORZone + DayNight + Closure
Model 8 Year + Vessel + DepCat + Month + ORZone + DayNight + Closure +
    Vessel:Month
Model 9 Year + Vessel + DepCat + Month + ORZone + DayNight + Closure +
    DepCat:Month
```

Table 20.43. Model selection criteria, including the AIC, the residual sum of squares, the Model sum of squares, the number of usable observations, the number of parameters, the adjusted $r^{2}$ and the increment in adjusted $r^{2}$. The DepCat:Month model (model 9) was optimal. The effect of being in the open or closed areas (Closed) was minor (Figure 20.31).

|  | Year | Vessel | DepCat | Month | ORZone | DayNight | Closed | Vess: | Dep:Mth |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AIC | 16486 | 11838 | 10083 | 9143 | 8532 | 8354 | 8317 | 8382 | 8022 |
| RSS | 48873 | 40521 | 37665 | 36284 | 35421 | 35169 | 35117 | 32158 | 34304 |
| MSS | 2248 | 10599 | 13456 | 14836 | 15699 | 15952 | 16004 | 18963 | 16817 |
| Nobs | 25933 | 25933 | 25752 | 25752 | 25752 | 25752 | 25752 | 25752 | 25752 |
| Npars | 26 | 132 | 146 | 157 | 161 | 164 | 165 | 1331 | 319 |
| adj_r ${ }^{2}$ | 4.304 | 20.332 | 25.904 | 28.590 | 30.277 | 30.766 | 30.866 | 33.669 | 32.057 |
| $\Delta \mathrm{r}^{2}$ | 0.000 | 16.027 | 5.573 | 2.686 | 1.687 | 0.489 | 0.100 | 2.803 | 1.191 |

Table 20.44. Reported catches by CAAB code for the data analysed. In 2010 the group code Oreo Dory, 37266902, was previously omitted from the analysis because of confusion with Black Oreo (37266901). The 37266902 reporting code (Oreo Dories) appears only to have been introduced in 2005 when quotas were first applied to Mixed Oreos.

|  | Spiky | Warty | OreoDory |  | Spiky | Warty | OreoDory |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | 37266001 | 37266004 | 37266902 | Year | 37266001 | 37266004 | 37266902 |
| 1986 | 19.269 | 31.697 |  | 1999 | 429.802 | 11.215 |  |
| 1987 | 40.834 | 19.075 |  | 2000 | 345.507 | 30.987 |  |
| 1988 | 13.860 | 19.949 | 2001 | 392.974 | 6.060 |  |  |
| 1989 | 175.798 | 13.441 |  | 2002 | 210.951 | 1.595 |  |
| 1990 | 254.921 | 2.257 | 2003 | 228.924 | 0.300 |  |  |
| 1991 | 86.359 | 0.528 | 2004 | 179.862 | 1.540 |  |  |
| 1992 | 606.532 | 1.050 | 2005 | 93.756 |  | 7.510 |  |
| 1993 | 278.224 | 3.031 | 2006 | 38.109 |  | 42.151 |  |
| 1994 | 265.949 | 18.620 | 2007 | 11.771 |  | 46.983 |  |
| 1995 | 468.212 | 14.030 | 2008 | 6.983 |  | 41.581 |  |
| 1996 | 405.361 | 15.606 | 2009 | 6.851 |  | 66.788 |  |
| 1997 | 552.637 | 20.190 | 2010 | 8.061 |  | 68.076 |  |
| 1998 | 642.05 | 24.806 | 2011 | 6.802 |  | 71.460 |  |

Table 20.45. The standardized catch rates for the alternative statistical models for Mixed Oreos in OR zones $10,20,21,30$, and 50 , in depths 500 to 1200 m . The optimal model was DepCat:Month. St Err is the estimate of standard error for the optimum model. Values are relative to the mean of the standardized catch rates. The Month and closure factors column was omitted for clarity; their relative effect can be seen in Figure 20.31

| Year | Year | Vessel | DepCat | ORZone | DayNight | Vess:Mth | Dep:Mth | StErr |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 1.1362 | 1.1326 | 1.3198 | 1.2399 | 1.3170 | 1.1242 | $\mathbf{1 . 2 6 5 4}$ | 0.0000 |
| 1987 | 1.3472 | 1.3873 | 1.4599 | 1.5452 | 1.5435 | 1.4421 | $\mathbf{1 . 4 9 9 7}$ | 0.1560 |
| 1988 | 0.8317 | 0.8665 | 1.0074 | 1.0935 | 1.0860 | 1.0109 | $\mathbf{0 . 9 8 1 1}$ | 0.1643 |
| 1989 | 1.3807 | 1.6139 | 1.8223 | 1.9028 | 1.9166 | 1.9174 | $\mathbf{1 . 9 5 8 0}$ | 0.1408 |
| 1990 | 2.9320 | 3.3330 | 3.9427 | 3.5164 | 3.4706 | 4.1547 | $\mathbf{3 . 5 3 5 9}$ | 0.1444 |
| 1991 | 0.8556 | 1.1797 | 1.3570 | 1.3128 | 1.3121 | 1.2199 | $\mathbf{1 . 3 7 9 3}$ | 0.1469 |
| 1992 | 2.2785 | 2.7537 | 2.7093 | 2.6352 | 2.5811 | 2.4923 | $\mathbf{2 . 6 1 2 6}$ | 0.1281 |
| 1993 | 0.9511 | 1.1773 | 1.1666 | 1.1977 | 1.1736 | 1.1961 | $\mathbf{1 . 2 1 2 8}$ | 0.1283 |
| 1994 | 0.7545 | 0.8761 | 0.8639 | 0.9444 | 0.9315 | 1.0303 | $\mathbf{0 . 9 7 1 5}$ | 0.1266 |
| 1995 | 0.9225 | 1.0227 | 0.8766 | 0.9586 | 0.9668 | 0.9782 | $\mathbf{1 . 0 0 3 0}$ | 0.1242 |
| 1996 | 0.6972 | 0.8124 | 0.6864 | 0.7325 | 0.7340 | 0.7445 | $\mathbf{0 . 7 1 5 0}$ | 0.1246 |
| 1997 | 1.0361 | 1.0099 | 0.8712 | 0.8903 | 0.8908 | 0.8686 | $\mathbf{0 . 8 7 1 4}$ | 0.1245 |
| 1998 | 1.2173 | 1.0642 | 0.9611 | 0.9941 | 0.9931 | 0.9904 | $\mathbf{0 . 9 9 9 0}$ | 0.1245 |
| 1999 | 1.0590 | 0.9422 | 0.8640 | 0.8627 | 0.8630 | 0.8493 | $\mathbf{0 . 8 5 4 3}$ | 0.1249 |
| 2000 | 0.9741 | 0.8273 | 0.7697 | 0.7613 | 0.7628 | 0.7395 | $\mathbf{0 . 7 5 5 7}$ | 0.1253 |
| 2001 | 0.9899 | 0.9074 | 0.8480 | 0.8031 | 0.8043 | 0.7830 | $\mathbf{0 . 8 0 4 8}$ | 0.1251 |
| 2002 | 0.7044 | 0.6062 | 0.5737 | 0.5806 | 0.5836 | 0.5632 | $\mathbf{0 . 5 7 6 5}$ | 0.1261 |
| 2003 | 0.7553 | 0.6236 | 0.5988 | 0.6056 | 0.6083 | 0.5804 | $\mathbf{0 . 6 0 6 8}$ | 0.1262 |
| 2004 | 0.6702 | 0.5399 | 0.5213 | 0.5305 | 0.5328 | 0.5178 | $\mathbf{0 . 5 2 5 0}$ | 0.1264 |
| 2005 | 0.6422 | 0.4846 | 0.4485 | 0.4484 | 0.4519 | 0.4368 | $\mathbf{0 . 4 4 5 1}$ | 0.1291 |
| 2006 | 0.5078 | 0.4245 | 0.3847 | 0.4168 | 0.4211 | 0.4143 | $\mathbf{0 . 4 1 4 6}$ | 0.1311 |
| 2007 | 0.5561 | 0.4602 | 0.3883 | 0.4055 | 0.4084 | 0.3977 | $\mathbf{0 . 3 8 7 3}$ | 0.1368 |
| 2008 | 0.7278 | 0.5009 | 0.3826 | 0.3824 | 0.3864 | 0.3784 | $\mathbf{0 . 3 8 2 4}$ | 0.1408 |
| 2009 | 0.6520 | 0.4862 | 0.3895 | 0.4192 | 0.4257 | 0.3899 | $\mathbf{0 . 4 2 8 4}$ | 0.1335 |
| 2010 | 0.6555 | 0.4520 | 0.3765 | 0.3998 | 0.4049 | 0.3741 | $\mathbf{0 . 3 8 9 4}$ | 0.1327 |
| 2011 | 0.7651 | 0.5158 | 0.4103 | 0.4208 | 0.4302 | 0.4061 | $\mathbf{0 . 4 2 5 0}$ | 0.1316 |

### 20.6.5.1 TIER 4 Mixed Oreo Target Proxy 48\%

Using the standardized catch rates and the updated catches for 2010, which now include the Oreo Dory (CAAB code 37266902) previously omitted, the TIER 4 analysis shows the recent catch rates to be not far from the target ( $\sim 82 \%$ of the target) so the RBC calculation is restrained.

The RAG, in Oct 2011, recommended the reference period be moved from 1992-2001 to become 1993-2001. The reasoning behind this move was that 1992 was the last year of the Orange Roughy fishery in which mixed oreos were a significant discard component, while from 1993 onwards Oreos were landed much more often.

Table 20.46. CE are the standardized catch rates. GeoCE is the geometric mean catch rate from the raw data. Total is the total catch in the open areas, including discards (estimated at 16.2\%). The target catch rate and target catch are both halved to allow for an assumed lack of exploitation prior to the reference period.

| Year | Catch | Total | CE | GeoCE |  |  |  |
| ---: | ---: | ---: | ---: | ---: | :--- | ---: | ---: |
| 1986 | 47.586 | 56.785 | 1.2654 | 128.0280 | Ref_Year | 1993 |  |
| 1987 | 17.390 | 20.752 | 1.4997 | 108.0437 | Ref_Year | 2001 |  |
| 1988 | 12.228 | 14.592 | 0.9811 | 77.8207 | lecept Yr |  |  |
| 1989 | 25.771 | 30.753 | 1.9580 | 103.1413 | CE_Targ | 0.4549 |  |
| 1990 | 7.335 | 8.753 | 3.5359 | 107.2726 | CE_Lim | 0.1819 |  |
| 1991 | 18.421 | 21.982 | 1.3793 | 72.4795 | CE_Recent | 0.4063 |  |
| 1992 | 76.258 | 91.000 | 2.6126 | 111.0680 | Wt_Discard | 12.735 |  |
| 1993 | 80.648 | 96.239 | 1.2128 | 111.7519 |  |  |  |
| 1994 | 97.882 | 116.804 | 0.9715 | 86.3323 |  |  | 0.8221 |
| 1995 | 311.961 | 372.268 | 1.0030 | 128.0685 | Ccaling |  |  |
| 1996 | 284.955 | 340.042 | 0.7150 | 91.1849 | RBC |  |  |
| 1997 | 387.711 | 462.662 | 0.8714 | 115.4691 |  | 160.830 |  |
| 1998 | 448.279 | 534.939 | 0.9990 | 132.6256 |  | 132.213 |  |
| 1999 | 313.340 | 373.914 | 0.8543 | 120.7534 |  |  |  |
| 2000 | 253.999 | 303.101 | 0.7557 | 114.5441 |  |  |  |
| 2001 | 247.178 | 294.962 | 0.8048 | 101.1833 |  |  |  |
| 2002 | 145.658 | 173.816 | 0.5765 | 75.0058 |  | Years | TAC |
| 2003 | 145.208 | 173.279 | 0.6068 | 77.2203 |  | 2005 | 200 |
| 2004 | 121.256 | 144.697 | 0.5250 | 72.0455 |  | 2006 | 200 |
| 2005 | 72.176 | 86.129 | 0.4451 | 67.8523 |  | $2007 / 08$ | 190 |
| 2006 | 53.096 | 63.360 | 0.4146 | 53.5816 |  | $2008 / 09$ | 150 |
| 2007 | 52.028 | 62.086 | 0.3873 | 54.5861 |  | $2009 / 10$ | 188 |
| 2008 | 42.937 | 51.237 | 0.3824 | 70.2134 |  | $2010 / 11$ | 188 |
| 2009 | 65.576 | 78.253 | 0.4284 | 62.5112 |  | $2011 / 12$ | 113 |
| 2010 | 62.542 | 74.632 | 0.3894 | 60.0141 |  | $2012 / 13$ |  |
| 2011 | 70.490 | 84.117 | 0.4250 | 74.8662 |  |  |  |



Figure 20.33. Tier 4 analysis for mixed oreos. Top left is the total catch in the open areas with the target catch indicated by the horizontal line. The target period is indicated by the thickened section of the line. Top right, illustrates the standardized catch rates plus both the target and limit catch rates, as well as the recent average catch rate, again with the target period identified with a thickened line. The distance of the mean of the last four points from the target indicates the potential scaling used to produce the RBC. Bottom left is total removals. Bottom left is the geometric mean catch rate compared to the standardized catch rates, scaled to the mean of the unstandardized rates.

MixedOreo


Figure 20.34. An expanded version of the Tier 4 analysis of catch rates to improve the illustration of the reference period and the recent mean catch rates.

### 20.6.6 Eastern Deepwater Sharks

Table 20.47. The names of the various species identified in the catch and effort database.

| CAAB Code | Common Name | Scientific Name |
| ---: | :--- | :--- |
| 37020000 | Dogfish | Squalidae |
| 37020002 | Black | Dalatias licha |
| 37020003 | Brier | Deania calcea |
| 37020004 | Platypus | Deania quadrispinosa |
| 37020013 | Plunket's Dogfish | Centroscymnus plunketi |
| 37020904 | Roughskin | Centroscymnus \& Deania sps. |
| 37020905 | Pearl | Deania calcea \& D. quadrispinosa |
| 37020906 | Black (roughskin) | Centroscymnus sps. |
| 37990003 | Other Sharks | Other Sharks |

The estimated discard rate is 2.8\% (Wayte \& Fuller (2009).
This basket quota group is made up of many recognized species but only ten have any records, and only eight of these have any significant catches. Dogfish and Other Sharks dominate catches until about 2000. The Black Shark is possibly confounded with two group categories, the Roughskin and the Black Shark - Roughskin. Plunket's Dogfish is possibly confounded with the Roughskin Shark group. Similarly, the Pearl Shark group is a combination of the Brier and Platypus Sharks. The reported distributions of the Brier shark, the Roughskin Shark, and especially the Plunket's Dogfish categories are much less widespread than the others.


Figure 20.35. Eastern Deepwater Sharks catches broken down by species taken by trawling in OR Zones 10, 20, and 50 (catches in 21 and 40 were trivial), in depths 600 to 1250 m .

Table 20.48. Number of records where Eastern Deepwater Sharks are reported from trawling in OR Zones $10,20,21$, and 50 , in depths 600 to 1250 m . Recs is the number of records used. Vess represents the count of vessels reporting Deepwater Sharks. Yield is the total reported catch in tonnes. The geometric mean CE is the raw unstandardized catch rate in $\mathrm{Kg} /$ tow. The left hand five columns represent all data, the right hand five columns represent the areas left open following the 700 m closure.

|  | Year | Yield | Records | Effort | Vessels | Geom | YieldO |  |  | Record <br> SO |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 28.926 | 254 | 1052 | 25 | 45.111 | 25.898 | 209 | 874 | 24 | 46.096 |
| 1987 | 6.061 | 105 | 327 | 28 | 26.456 | 4.821 | 89 | 272 | 24 | 26.085 |
| 1988 | 5.746 | 47 | 137 | 22 | 45.312 | 4.919 | 37 | 107 | 17 | 45.225 |
| 1989 | 5.561 | 85 | 220 | 21 | 37.910 | 5.080 | 76 | 191 | 19 | 37.505 |
| 1990 | 7.228 | 69 | 125 | 23 | 42.032 | 3.189 | 42 | 67 | 19 | 23.441 |
| 1991 | 20.213 | 129 | 316 | 24 | 62.171 | 10.119 | 87 | 208 | 21 | 54.265 |
| 1992 | 64.054 | 115 | 463 | 25 | 120.583 | 5.527 | 49 | 206 | 20 | 48.652 |
| 1993 | 95.237 | 295 | 968 | 26 | 132.886 | 17.922 | 118 | 322 | 22 | 48.635 |
| 1994 | 112.086 | 434 | 1605 | 30 | 130.137 | 38.050 | 215 | 780 | 27 | 96.916 |
| 1995 | 115.605 | 368 | 1453 | 22 | 179.615 | 61.899 | 220 | 804 | 22 | 163.944 |
| 1996 | 327.383 | 966 | 3712 | 30 | 191.197 | 260.404 | 777 | 2949 | 26 | 183.367 |
| 1997 | 194.243 | 907 | 4091 | 26 | 131.258 | 135.947 | 684 | 3062 | 24 | 122.844 |
| 1998 | 206.076 | 1105 | 4989 | 24 | 117.628 | 170.931 | 927 | 4093 | 23 | 114.465 |
| 1999 | 156.977 | 1013 | 4667 | 28 | 95.560 | 128.817 | 842 | 3829 | 26 | 91.905 |
| 2000 | 187.075 | 889 | 4252 | 28 | 124.127 | 150.371 | 707 | 3326 | 24 | 121.916 |
| 2001 | 140.954 | 893 | 4097 | 28 | 86.377 | 113.107 | 724 | 3224 | 26 | 90.318 |
| 2002 | 161.446 | 898 | 4230 | 29 | 102.917 | 130.026 | 752 | 3450 | 28 | 97.882 |
| 2003 | 130.839 | 974 | 4769 | 25 | 76.461 | 93.895 | 749 | 3534 | 22 | 73.496 |
| 2004 | 104.208 | 724 | 3459 | 29 | 79.814 | 78.429 | 587 | 2773 | 27 | 79.701 |
| 2005 | 61.426 | 480 | 2470 | 17 | 74.410 | 48.427 | 377 | 1949 | 15 | 75.336 |
| 2006 | 43.617 | 410 | 1960 | 21 | 51.361 | 33.066 | 279 | 1274 | 20 | 63.563 |
| 2007 | 8.418 | 106 | 494 | 17 | 43.938 | 8.378 | 104 | 484 | 17 | 44.636 |
| 2008 | 12.904 | 100 | 658 | 10 | 65.755 | 11.859 | 96 | 628 | 10 | 62.155 |
| 2009 | 39.137 | 232 | 1227 | 14 | 81.789 | 38.692 | 229 | 1208 | 14 | 81.183 |
| 2010 | 25.529 | 251 | 1264 | 13 | 48.906 | 24.302 | 241 | 1198 | 13 | 49.139 |
| 2011 | 4.154 | 36 | 151 | 8 | 51.408 | 4.154 | 36 | 151 | 8 | 51.408 |

```
Table 20.49. Statistical model structures used with Deepwater Sharks. DepCat is a series of 20 metre
depth categories. Deep relates to whether the area is open or closed. DayNight reduced the quality of fit..
Model 1 Year
Model 2 Year + Vessel
Model 3 Year + Vessel + DepCat
Model 4 Year + Vessel + DepCat + Month
Model 5 Year + Vessel + DepCat + Month + ORZone
Model 6 Year + Vessel + DepCat + Month + ORZone + Deep
Model 7 Year + Vessel + DepCat + Month + ORZone + Deep + ORZone:Month
Model 8 Year + Vessel + DepCat + Month + ORZone + Deep + Vessel:Month
```



Figure 20.36. Eastern Deepwater Sharks reported from trawling in OR Zones 10, 20, 21, 50, in depths 600 to 1250 m . The top left is the depth distribution of all records reporting Deepwater Sharks, the top right graph depicts the depth distribution of shots containing Deepwater Sharks in OR Zones 10, 20, 21, 50, in depths 600 to 1250 m . The middle left diagram depicts the distribution of catch across all years by depth within separate OR zones (most catch is in zones 10,20 , and 50 ), the right hand middle graph depicts the number of vessels reporting Eastern Deepwater Sharks through time. The bottom left reflects the number of records for Deepwater Sharks, and bottom right are the Deepwater Shark catches used in the analysis.


Figure 20.37. Eastern Deepwater Sharks catches taken by trawling in OR Zones 10, 20, 21, and 50, in depths 600 to 1250 m . Less than 7.0 t was reported in OR Zone 70 across all years.


Figure 20.38. Depth distribution of the eight main species of Eastern Deepwater Sharks catches taken by trawling in OR Zones 10, 20, 21, and 50, in depths 600 to 1250 m. 37020000: Dogfish, 37020002: Black Shark, 37020003: Brier Shark, 37020004: Platypus Shark, 37020904:Roughskin Shark, 37020905: Pearl Shark, and 37020906: Black Shark - Roughskin category, 37990003: Other Shark. Data updated to 2010.

Table 20.50. Model selection criteria, including the AIC, the residual sum of squares, the Model sum of squares, the number of usable observations, the number of parameters, the adjusted $r^{2}$ and the increment in adjusted $\mathrm{r}^{2}$. The model including the ORZone:Month interaction term (model 7) was optimal. There was a trivial effect of being in the open or closed areas (Deep) on the statistical model fit. Year, Vessel, and DepCat dominated the analysis. The DayNight factor was omitted because it detracted from the fit.

|  | Year | Vessel | DepCat | Month | ORzone | deep | ORzone:Mth | Vessel:Month |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 3077 | 1602 | 751 | 740 | 736 | 737 | $\mathbf{7 0 4}$ | 1743 |
| RSS | 13890 | 11870 | 10631 | 10596 | 10583 | 10582 | $\mathbf{1 0 4 5 7}$ | 9908 |
| MSS | 1265 | 3286 | 4524 | 4559 | 4572 | 4573 | $\mathbf{4 6 9 8}$ | 5247 |
| Nobs | 10352 | 10352 | 10115 | 10115 | 10115 | 10115 | $\mathbf{1 0 1 1 5}$ | 10115 |
| Npars | 17 | 93 | 124 | 135 | 139 | 140 | $\mathbf{1 8 4}$ | 976 |
| adj_r $^{2}$ | 8.204 | 20.977 | 28.990 | 29.145 | 29.200 | 29.201 | $\mathbf{2 9 . 7 2 8}$ | 27.648 |
| $\Delta \mathbf{r}^{2}$ | 0.000 | 12.773 | 8.013 | 0.155 | 0.055 | 0.001 | $\mathbf{0 . 5 2 7}$ | -1.553 |

Table 20.51. The standardized catch rates for the alternative statistical models for Eastern Deepwater Sharks in OR zones10, 20, 21, and 50, in depths 600 to 1250 m . The optimal model was Model 7. St Err is the estimate of standard error for the optimum model. Values are relative to the mean of the standardized catch rates. The models for Deep and Vessel:Month were omitted for brevity.

| Year | Year | Vessel | DepCat | Month | ORzone | Deep | ORzone:Mth | StErr |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1995 | 1.8988 | 1.8206 | 1.6859 | 1.7180 | 1.7287 | 1.7191 | $\mathbf{1 . 6 8 6 3}$ | 0.0000 |
| 1996 | 2.0263 | 2.0835 | 2.0721 | 2.0806 | 2.0162 | 2.0128 | $\mathbf{2 . 0 0 8 4}$ | 0.0726 |
| 1997 | 1.3912 | 1.3797 | 1.2422 | 1.2467 | 1.2311 | 1.2284 | $\mathbf{1 . 2 4 4 5}$ | 0.0704 |
| 1998 | 1.2465 | 1.1467 | 1.0327 | 1.0378 | 1.0393 | 1.0401 | $\mathbf{1 . 0 5 1 8}$ | 0.0695 |
| 1999 | 1.0127 | 1.0171 | 0.8896 | 0.8929 | 0.8972 | 0.8978 | $\mathbf{0 . 8 8 6 8}$ | 0.0697 |
| 2000 | 1.3156 | 1.3093 | 1.1287 | 1.1260 | 1.1280 | 1.1297 | $\mathbf{1 . 1 1 5 1}$ | 0.0715 |
| 2001 | 0.9155 | 1.0355 | 0.9330 | 0.9330 | 0.9439 | 0.9457 | $\mathbf{0 . 9 4 7 3}$ | 0.0721 |
| 2002 | 1.0908 | 1.1143 | 1.0451 | 1.0577 | 1.0666 | 1.0681 | $\mathbf{1 . 0 6 3 9}$ | 0.0719 |
| 2003 | 0.8103 | 0.8230 | 0.7431 | 0.7431 | 0.7474 | 0.7469 | $\mathbf{0 . 7 5 1 8}$ | 0.0718 |
| 2004 | 0.8461 | 0.8128 | 0.7616 | 0.7560 | 0.7652 | 0.7659 | $\mathbf{0 . 7 7 2 3}$ | 0.0740 |
| 2005 | 0.7892 | 0.7595 | 0.7301 | 0.7293 | 0.7355 | 0.7353 | $\mathbf{0 . 7 3 5 7}$ | 0.0797 |
| 2006 | 0.5448 | 0.5307 | 0.6440 | 0.6391 | 0.6405 | 0.6398 | $\mathbf{0 . 6 4 6 6}$ | 0.0825 |
| 2007 | 0.4683 | 0.4716 | 0.7287 | 0.7289 | 0.7315 | 0.7336 | $\mathbf{0 . 7 3 1 0}$ | 0.1299 |
| 2008 | 0.7011 | 0.6604 | 1.0293 | 1.0340 | 1.0424 | 1.0445 | $\mathbf{1 . 0 3 8 9}$ | 0.1271 |
| 2009 | 0.8687 | 0.9349 | 1.1736 | 1.1806 | 1.1882 | 1.1905 | $\mathbf{1 . 2 0 8 2}$ | 0.0963 |
| 2010 | 0.5193 | 0.5312 | 0.5793 | 0.5811 | 0.5849 | 0.5864 | $\mathbf{0 . 5 8 7 7}$ | 0.0939 |
| 2011 | 0.5547 | 0.5692 | 0.5810 | 0.5150 | 0.5135 | 0.5154 | $\mathbf{0 . 5 2 3 5}$ | 0.1962 |



Figure 20.39. Eastern Deepwater Sharks reported from trawling in OR Zones 10, 20, 21, and 50, in depths 600 to 1250 m . The black dashed line from 86-11 represents the geometric mean catch rate and the solid black line the optimum standardized catch rates (Model 7). The graph scales the catch rates relative to the mean of the standardized catch rates (depicted by the horizontal grey line at 1.0).

### 20.6.6.1 TIER 4 Eastern Deepwater Sharks

As with the western deepwater sharks it is doubtful whether the catch rate value for 2011 is valid as there were less than 5 t of data that met the reporting requirements. It remains unknown whether catch rates reflect the stock status but there are so few records it appears highly unlikely. Certainly the standard error estimates in 2011 is relatively large. The RAG decided last year that when catches were less than 10 t no update of the Tier 4 would be made.

The low catches are in fact an artefact of trying to identify which shots are in closures and which are out. Many trawl shots are made immediately next to the closure boundary because the catch rates there are the best that are available. However, the precision with which a vessel's position is recorded is less than the precision with which we can define the closure boundaries. Western deep water sharks were used as an example in the RAG to demonstrate this but the same phenomenon occurs in the eastern fishery. Actual reported catches were approximately 33 t in the east but to conduct an analysis of catch rates it remains necessary to identify those shots that were definitely outside the reserves. It may be necessary to simply assume all shots are now outside the reserves and use all available data. Because the TAC was rolled over this year it was decided to analyse this in more detail in next year's assessment.

The closures have undoubtedly had a great impact.. A consideration of Figure 20.38 indicates that many earlier catches were taken in water deeper than 700 m . The closures introduced in the deepwater have thus removed large areas where catches of deepwater catches were taken.

### 20.6.7 Western Deepwater Sharks

There are numerous species grouped together into the Deepwater Sharks (Table 20.52) but only some have data and even fewer have significant catches reported.

| Table 20.52. The names of the various species identified in the catch and effort database. |  |  |
| ---: | :--- | :--- |
| CAAB Code | Common Name | Scientific Name |
| 37020000 | Dogfish | Squalidae |
| 37020002 | Black | Dalatias licha |
| 37020003 | Brier | Deania calcea |
| 37020004 | Platypus | Deania quadrispinosa |
| 37020904 | Roughskin | Centroscymnus \& Deania sps. |
| 37020905 | Pearl | Deania calcea \& D. quadrispinosa |
| 37020906 | Black (roughskin) | Centroscymnus sps. |
| 37990003 | Other Sharks | Other Sharks |

The estimated discard rate is $2.8 \%$ (Wayte \& Fuller (2009).
This basket quota group is made up of many recognized species but only seven have any records, and only four have any significant catches reported recently. The Black Shark is possibly confounded with two group categories, the Roughskin and the Black Shark - Roughskin. Similarly, the Pearl Shark is a combination of the Brier and Platypus Sharks.


Figure 20.40. The depth distribution of the six main species with catches reported in the western deepwater shark fishery. The vertical line at 600 m illustrates the cut-off used in data selection.


Figure 20.41. Catch by depth for the six main species and species groups. 37020000: Dogfish, 37020002: Black Shark, 37020003: Brier Shark, 37020004: Platypus Shark, 37020905: Pearl Shark, and 37020906: Black Shark - Roughskin category, 37990003: Other Shark.


Figure 20.42. Western Deepwater Sharks catches broken down by species taken by trawling in OR Zone 30, in depths 600 to 1100 m for the years $1995-2010$.

Table 20.53. Number of records where Western Deepwater Sharks are reported from trawling in OR Zone 30, in depths 600 to 1100 m . Vess represents the count of vessels reporting Deepwater Sharks. Yield is the total reported catch. The geometric mean CE is the raw unstandardized catch rate in $\mathrm{Kg} /$ tow. The left hand five columns represent all data, the right hand five columns represent the areas left open following the 700 m closure. There appear to be captures in the closed areas because many vessels track the edge of the closures and the software is making category errors.

| Year | Yield | Records | Effort Vessels | Geom | YieldO | RecordsO | EffortO | VesselsO | GeomO |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 1.030 | 14 | 56 | 3 | 54.016 | 0.290 | 5 | 19 | 3 | 45.731 |
| 1987 | 0.603 | 21 | 62 | 5 | 22.509 | 0.498 | 17 | 48 | 4 | 22.239 |
| 1988 | 0.525 | 4 | 11 | 2 | 122.474 | 0.100 | 1 | 2 | 1 | 100.000 |
| 1989 | 1.238 | 15 | 40 | 3 | 65.597 | 0.428 | 7 | 16 | 3 | 44.501 |
| 1990 | 0.314 | 5 | 13 | 4 | 34.822 | 0.010 | 1 | 2 | 1 | 10.000 |
| 1991 | 0.315 | 5 | 18 | 3 | 42.929 | 0.195 | 2 | 5 | 2 | 51.962 |
| 1992 | 3.600 | 21 | 94 | 4 | 128.049 | 3.460 | 19 | 86 | 4 | 137.919 |
| 1993 | 2.025 | 18 | 61 | 3 | 79.840 | 1.725 | 13 | 43 | 2 | 91.106 |
| 1994 | 1.612 | 23 | 128 | 4 | 55.626 | 0.572 | 9 | 43 | 3 | 57.241 |
| 1995 | 95.106 | 593 | 2929 | 10 | 93.596 | 43.007 | 256 | 1221 | 8 | 97.550 |
| 1996 | 186.252 | 956 | 4491 | 23 | 105.541 | 96.030 | 526 | 2365 | 17 | 108.016 |
| 1997 | 325.955 | 1975 | 10102 | 19 | 95.986 | 157.971 | 1054 | 5142 | 19 | 92.683 |
| 1998 | 396.667 | 2905 | 16202 | 18 | 88.170 | 147.941 | 1180 | 6124 | 18 | 82.990 |
| 1999 | 312.960 | 2212 | 12544 | 19 | 89.926 | 117.333 | 970 | 5227 | 18 | 81.342 |
| 2000 | 311.679 | 1872 | 10454 | 17 | 111.018 | 124.522 | 818 | 4155 | 17 | 103.720 |
| 2001 | 242.052 | 1832 | 10384 | 19 | 84.155 | 99.095 | 834 | 4490 | 19 | 79.789 |
| 2002 | 251.392 | 1625 | 10161 | 17 | 98.832 | 115.775 | 734 | 4399 | 17 | 100.324 |
| 2003 | 166.630 | 1431 | 9008 | 16 | 73.359 | 75.445 | 665 | 4024 | 16 | 74.507 |
| 2004 | 209.774 | 1733 | 10870 | 15 | 78.244 | 99.719 | 798 | 4778 | 14 | 80.257 |
| 2005 | 82.725 | 818 | 4816 | 13 | 61.230 | 40.390 | 396 | 2246 | 12 | 62.709 |
| 2006 | 72.064 | 617 | 3806 | 12 | 70.529 | 38.211 | 312 | 1832 | 12 | 75.480 |
| 2007 | 8.612 | 112 | 682 | 9 | 38.108 | 6.041 | 73 | 406 | 8 | 46.414 |
| 2008 | 15.625 | 121 | 784 | 8 | 76.979 | 10.211 | 83 | 538 | 8 | 71.433 |
| 2009 | 34.072 | 233 | 1487 | 10 | 79.505 | 25.736 | 168 | 1085 | 9 | 87.506 |
| 2010 | 35.775 | 268 | 1620 | 10 | 68.800 | 28.171 | 214 | 1271 | 10 | 69.635 |
| 2011 | 4.004 | 46 | 338 | 5 | 61.421 | 2.729 | 33 | 240 | 4 | 59.548 |

Table 20.54. Statistical model structures used with Deepwater Sharks. DepCat is a series of 20 metre depth categories. Deep relates to whether the area is open or closed.

```
Model Year
1
Model Year + Vessel
2
Model Year + Vessel + DepCat
3
Model Year + Vessel + DepCat + Month
4
Model Year + Vessel + DepCat + Month + DayNight
5
Model Year + Vessel + DepCat + Month + DayNight + Deep
6
Model Year + Vessel + DepCat + Month + DayNight + Deep + Vessel:Month
7
```








Figure 20.43. Western Deepwater Sharks reported from trawling in OR Zone 30, in depths 600 to 1100 m . The top left is the depth distribution of all records reporting Deepwater Sharks, the top right graph depicts the depth distribution of shots containing Deepwater Sharks in OR Zone 30, in depths 600 to 1100 m . The middle left diagram depicts the distribution of catch across all years by depth within separate OR zones (only catch from zone 30), the right hand middle graph depicts the number of vessels reporting Western Deepwater Sharks through time. The bottom left reflects the number of records for Deepwater Sharks, and bottom right are the Deepwater Shark catches used in the analysis.

Table 20.55. Model selection criteria, including the AIC, the residual sum of squares, the Model sum of squares, the number of usable observations, the number of parameters, the adjusted $r^{2}$ and the increment in adjusted $\mathrm{r}^{2}$. Model 6 was optimal. The effect of being in the open or closed areas (Deep) was minor.

|  | Year | Vessel | DepCat | Month | DayNight | Deep | Vessel:Month |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AIC | 1364 | 71 | -2507 | -2680 | -2694 | -2699 | -2399 |
| RSS | 20725 | 19304 | 16782 | 16613 | 16596 | 16590 | 16080 |
| MSS | 490 | 1911 | 4433 | 4602 | 4619 | 4625 | 5135 |
| Nobs | 19349 | 19349 | 19278 | 19278 | 19278 | 19278 | 19278 |
| Npars | 17 | 58 | 83 | 94 | 97 | 98 | 549 |
| adj_r $^{2}$ | 2.228 | 8.740 | 20.558 | 21.312 | 21.381 | 21.406 | 21.988 |
| $\Delta r^{2}$ | 0.000 | 6.512 | 11.817 | 0.755 | 0.069 | 0.025 | 0.582 |

Table 20.56. The standardized catch rates for the alternative statistical models for Western Deepwater Sharks in OR zone 30, in depths 600 to 1100 m . The optimal model was Model 6. St Err is the estimate of standard error for the optimum model. Values are relative to the mean of the standardized catch rates.

| Year | Year | DepCa <br> t | Vessel |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | Month | DayNig |
| ---: |
| ht |$\quad$| Deep |
| ---: | Vessel:Month | StErr |
| ---: | :--- |



Figure 20.44. Western Deepwater Sharks reported from trawling in OR Zone 30, in depths 600 to 1100 m . The black dashed line from 95-10 represents the geometric mean catch rate and the solid black line the optimum standardized catch rates (Model 6). The graph standardizes catch rates relative to the mean of the standardized catch rates, represented by the horizontal fine grey line.


Figure 20.45. The relative impact of the different factors on the changes in the standardized trend. The major effects of both the structural adjustment, with its change of vessels, and the deepwater closures is clear.

### 20.6.7.1 TIER 4 Western Deepwater Sharks

It is doubtful whether the catch rate value for 2011 is valid as there were less than 5 t of data that met the reporting requirements. It remains unknown whether catch rates reflect the stock status but there are so few records it appears highly unlikely. Certainly the standard error estimates in 2011 is relatively large. The RAG decided last year that when catches were less than 10 t no update of the Tier 4 would be made.

The low catches are in fact an artefact of trying to identify which shots are in closures and which are out. Many trawl shots are made immediately next to the closure boundary because the catch rates there are the best that are available. However, the precision with which a vessel's position is recorded is less than the precision with which we can define the closure boundaries. Western deep water sharks were used as an example in the RAG to demonstrate this but the same phenomenon occurs in the eastern fishery. Actual reported catches were approximately 53 t in the west but to conduct an analysis of catch rates it remains necessary to identify those shots that were definitely outside the reserves. It may be necessary to simply assume all shots are now outside the reserves
and use all available data. Because the TAC was rolled over this year it was decided to analyse this in more detail in next year's assessment.

### 20.6.8 Alfonsino (ALF - 37258002)

There were no reported catches of Alfonsino in the East Coast Deepwater fishery in 2010 so the analysis conducted in 2011 (Haddon, 2012b) still stands. This year some summary information is given instead of simply reiterating the same information.

The SESSF is made up of the Commonwealth trawl sector, the Great Australian Bight Sector, the East Coast Deepwater Trawl sector, and the Gillnet, Hook and Trap sector. Currently the Tier 4 analysis focuses on the East Coast Deepwater trawl fishery but it should include the South east Trawl fishery and the GAB. Currently there are only intermittent reported catches of Alfonsino in the ECD, so no analyses can proceed, but the TAC set (via a Tier 3 analysis) is applicable to the SET and the ECD. If a Tier 4 analysis were to be used, strictly it should include catches taken in each of these jurisdictions.

Table 20.57. Reported catches of Alfonsino by method. AL - autoline, BL - Bottom Line, DL - Drop Line, DS - Danish Seine, FP - , GN - Gill net, LL - Long Line, RR - , TL - Trot Line, and TW - trawl.

| Year | Unknown | AL | BL | DL | DS | FP | GN | HL | RR | TL |
| :--- | ---: | :--- | ---: | :--- | :--- | :--- | :--- | ---: | ---: | ---: | TW

While the obvious hotspots are in the ECD, there are catches taken in the SET. There are very low catches spread widely there appear to be relative hot spots, the same order of magnitude as in the ECD) one off Eddystone point on the East Coast of Tasmania, one south of the Tasman Peninsula. There is another, somewhat smaller spot, off Macquarie Harbour on the west coast of Tasmania and another spot off Robe or Cape Jaffa in western Victoria.

Table 20.58. Catch of Alfonsino taken by trawl in the different fisheries. CSF - Coral Sea Fishery, ECD - East Coast Deepwater, GAB - Great Australian Bight, HST - High Seas Trawl, NFO - Norfolk Island Offshore Demersal Finfish, SET - South East Trawl, and WDW Western Deepwater. Currently no attention is paid to the catches in the SET

|  | CSF | ECD | GAB | HST | NFO | SET | WDW |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1988 |  |  |  |  | 0.538 |  |  |
| 1989 |  |  | 0.276 |  | 2.302 |  |  |
| 1990 |  |  | 0.010 |  | 3.634 |  |  |
| 1991 |  |  |  |  | 5.652 |  |  |
| 1992 |  |  |  |  | 17.787 |  |  |
| 1993 |  |  |  |  | 5.231 |  |  |
| 1994 |  |  |  |  | 7.842 | 7.760 |  |
| 1995 |  |  |  |  | 8.423 | 0.166 |  |
| 1996 |  |  |  |  | 12.427 |  |  |
| 1997 |  |  |  |  | 8.290 | 0.016 |  |
| 1998 |  |  | 4.094 | 8.836 |  | 4.762 |  |
| 1999 |  |  |  |  | 3.736 |  |  |
| 2000 |  | 66.950 | 384.332 |  | 40.620 | 2.710 |  |
| 2001 |  | 307.271 | 0.827 | 6.720 | 0.010 | 6.774 | 4.342 |
| 2002 | 63.560 | 42.036 | 0.270 | 2508.871 | 0.245 | 25.181 |  |
| 2003 | 58.640 | 140.771 | 0.025 | 1611.061 |  | 6.180 |  |
| 2004 | 14.163 | 509.466 | 0.042 | 867.980 |  | 16.062 |  |
| 2005 |  | 136.050 | 0.039 | 296.964 |  | 8.525 |  |
| 2006 | 14.091 |  | 0.320 | 429.620 |  | 11.239 |  |
| 2007 | 55.582 | 85.397 | 0.124 | 372.904 |  | 12.283 |  |
| 2008 |  |  | 0.052 | 206.851 |  | 50.348 |  |
| 2009 |  | 14.156 | 0.042 | 62.275 |  | 17.870 |  |
| 2010 |  |  | 0.028 | 46.999 |  | 8.914 |  |
| 2011 |  | 147.500 |  | 460.950 |  | 4.374 |  |



Figure 20.46. The relative catch in four of the fisheries listed in Table 20.58.Table 20.58 Note the different scales in the different fisheries. To indicate isolated years of reported catches points are added to the graphs.

### 20.7 Non-Tier 4 Species

### 20.7.1 Blue Grenadier (GRE - 37227001 - Macruronus novaezelandiae)

Table 20.59 Blue Grenadier data for the TIER 4 calculations. Total is the sum of Discards, State, Non Trawl and SEF2 catches. All values in Tonnes. CE is the standardized catch rate the non-spawning fishery (Haddon, 2012). GeoMean is the geometric mean catch rates. Discards are estimates from 1998 to present. The ratio of discards to catch over the 1998 - 2006 period was used to estimate the discards between 1986 and 1997, the proportion of which is the PDiscard.

| Year | Catch | Discards | Total | State | Non-T | PDisca <br> rd | CE | GeoMea <br> $n$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 1408.800 | 87.889 | 1496.689 |  |  | 5.87 | 1.4764 | 36.7375 |
| 1987 | 2197.200 | 137.074 | 2334.274 |  |  | 5.87 | 1.9414 | 37.3307 |
| 1988 | 1760.400 | 109.824 | 1870.224 |  |  | 5.87 | 2.1038 | 36.6778 |
| 1989 | 1798.800 | 112.220 | 1911.020 |  |  | 5.87 | 2.1808 | 45.3866 |
| 1990 | 2433.600 | 151.822 | 2585.422 |  |  | 5.87 | 2.1567 | 47.9497 |
| 1991 | 3812.400 | 237.840 | 4050.240 |  |  | 5.87 | 1.5519 | 48.2874 |
| 1992 | 3338.400 | 208.269 | 3546.669 |  |  | 5.87 | 1.2755 | 40.5408 |
| 1993 | 3412.800 | 212.910 | 3625.710 |  |  | 5.87 | 0.9613 | 33.2638 |
| 1994 | 3282.175 | 204.761 | 3486.936 | 126.682 | 0.000 | 5.87 | 0.8636 | 29.5414 |
| 1995 | 2812.359 | 175.451 | 2987.811 | 51.541 | 0.000 | 5.87 | 0.5978 | 19.4025 |
| 1996 | 3078.789 | 192.073 | 3270.861 | 40.338 | 0.000 | 5.87 | 0.5459 | 15.8910 |
| 1997 | 4550.755 | 283.902 | 4834.658 | 17.700 | 0.000 | 5.87 | 0.5664 | 13.3293 |
| 1998 | 5745.683 | 2959.00 | 8704.683 | 12.824 | 0.000 | 33.99 | 0.9286 | 18.8682 |
| 1999 | 9333.962 | 140.000 | 9473.962 | 8.359 | 0.000 | 1.48 | 0.9810 | 22.7820 |
| 2000 | 8655.402 | 129.000 | 8784.402 | 0.599 | 0.000 | 1.47 | 0.6986 | 16.8751 |
| 2001 | 9128.199 | 1.000 | 9129.199 | 0.469 | 3.684 | 0.01 | 0.3992 | 11.4735 |
| 2002 | 9164.727 | 5.270 | 9169.997 | 0.011 | 3.808 | 0.06 | 0.4006 | 13.3454 |
| 2003 | 8482.833 | 9.810 | 8492.643 | 0.057 | 8.925 | 0.12 | 0.3344 | 10.1345 |
| 2004 | 6401.449 | 27.190 | 6428.639 | 0.042 | 9.878 | 0.42 | 0.5633 | 16.9690 |
| 2005 | 4293.080 | 526.640 | 4819.720 | 0.075 | 10.222 | 10.93 | 0.6724 | 19.8341 |
| 2006 | 3624.811 | 246.570 | 3871.381 | 0.076 | 11.436 | 6.37 | 0.8930 | 26.9839 |
| 2007 | 3183.767 | 63.140 | 3246.907 | 4.584 | 8.015 | 1.94 | 0.7950 | 25.1832 |
| 2008 | 3937.055 | 41.982 | 3979.037 | 0.033 | 6.285 | 1.06 | 0.8654 | 28.8353 |
| 2009 | 3269.003 | 66.605 | 3335.609 | 0.075 | 9.655 | 2.00 | 0.8035 | 25.9256 |
| 2010 | 4194.794 | 20.010 | 4214.803 | 0.147 | 9.545 | 0.47 | 0.7879 | 25.9279 |
| 2011 | 2820.365 | 684.326 | 3504.691 | 0.147 | 5.913 | 19.53 | 0.6557 | 19.3008 |

Discards make up approximately 5.9 \% of the catch over the 1998-2006 period. NOTE: Actual landings in 2011 were in fact 4201.400 t rather than 2820 t , the source of this error is still being investigated. However, had this been used in a full Tier 4 it would not have influenced the result because the RBC depends on catch rates and the target catch not the current catch.

Table 20.60 RBC calculations for Blue Grenadier. $\mathrm{C}_{\text {targ }}$ and $\mathrm{CPUE}_{\text {targ }}$ relate to the period 1986-1995, $\mathrm{CPUE}_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7).

| Ref_Year | $1986-1995$ |
| ---: | ---: |
| CE_Targ | 1.5109 |
| CE_Lim | 0.6044 |
| CE_Recent | 0.7781 |
| Wt_Discard | 381.989 |
| Scaling | 0.1917 |
| Last Year's TAC |  |
| Ctarg $^{2789.499}$ |  |

BlueGrenadier


Figure 20.47 Blue Grenadier. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 20.7.2 Flathead (FLT - 37296001 - Neoplatycephalus richardsoni)

Table 20.61 Tiger Flathead data for the TIER 4 calculations. Total is the sum of Discards, State, Non Trawl and SEF2 catches. All values in Tonnes. CE is the standardized catch rate for otter trawl for Zones 10 and 20 in depths $0-400 \mathrm{~m}$ (Haddon, 2012).GeoMean is the geometric mean catch rates. Discards are estimates from 1998 to present. The ratio of discards to catch over the 1998 - 2006 period was used to estimate the discards between 1986 and 1997, the proportion of which is the PDiscard.

| Year | Catch | Discards | Total | State | Non- <br> T | PDiscard | CE | GeoMea <br> n |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 2133.600 | 158.670 | 2292.270 |  |  | 6.92 | 0.8010 | 16.7357 |
| 1987 | 2496.000 | 185.620 | 2681.620 |  |  | 6.92 | 1.0713 | 20.4621 |
| 1988 | 2444.400 | 181.783 | 2626.183 |  |  | 6.92 | 1.1721 | 23.7988 |
| 1989 | 2623.200 | 195.080 | 2818.280 |  |  | 6.92 | 1.1672 | 23.9908 |
| 1990 | 2188.800 | 162.775 | 2351.575 |  |  | 6.92 | 1.3861 | 30.1854 |
| 1991 | 2620.800 | 194.901 | 2815.701 |  |  | 6.92 | 1.3155 | 28.7154 |
| 1992 | 3564.000 | 265.044 | 3829.044 |  |  | 6.92 | 1.0282 | 23.8898 |
| 1993 | 3132.000 | 232.918 | 3364.918 |  |  | 6.92 | 1.0500 | 23.8001 |
| 1994 | 2786.959 | 207.258 | 2994.217 | 1290.69 | 0.000 | 6.92 | 0.7610 | 17.9798 |
| 1995 | 2735.929 | 203.463 | 2939.392 | 1023.89 | 0.000 | 6.92 | 0.8067 | 18.0790 |
| 1996 | 2725.609 | 202.696 | 2928.305 | 832.370 | 0.000 | 6.92 | 0.7138 | 16.4549 |
| 1997 | 3093.299 | 230.040 | 3323.339 | 586.530 | 0.000 | 6.92 | 0.7166 | 16.8264 |
| 1998 | 2933.991 | 291.000 | 3224.991 | 391.360 | 0.000 | 9.02 | 0.7588 | 17.7430 |
| 1999 | 3729.333 | 267.000 | 3996.333 | 272.155 | 0.000 | 6.68 | 0.9100 | 20.4344 |
| 2000 | 3427.408 | 511.000 | 3938.408 | 205.714 | 0.000 | 12.97 | 1.0130 | 24.4338 |
| 2001 | 2992.436 | 160.000 | 3152.436 | 147.950 | 0.281 | 5.08 | 0.9759 | 22.3118 |
| 2002 | 3272.572 | 193.970 | 3466.542 | 128.764 | 0.337 | 5.60 | 1.0657 | 22.8273 |
| 2003 | 3670.170 | 178.030 | 3848.200 | 175.179 | 0.809 | 4.63 | 1.0530 | 22.5536 |
| 2004 | 3596.871 | 228.380 | 3825.251 | 214.094 | 0.858 | 5.97 | 0.9091 | 19.7879 |
| 2005 | 3295.823 | 195.140 | 3490.963 | 293.570 | 1.145 | 5.59 | 0.7770 | 17.7159 |
| 2006 | 3017.332 | 201.730 | 3219.062 | 318.879 | 0.607 | 6.27 | 0.9429 | 22.2550 |
| 2007 | 3052.284 | 278.562 | 3330.847 | 204.789 | 0.486 | 8.36 | 1.1537 | 31.3544 |
| 2008 | 3446.847 | 43.736 | 3490.582 | 249.130 | 0.362 | 1.25 | 1.2088 | 31.6602 |
| 2009 | 2925.235 | 155.881 | 3081.116 | 246.308 | 0.403 | 5.06 | 1.1115 | 30.0219 |
| 2010 | 2991.840 | 251.039 | 3242.878 | 265.559 | 0.297 | 7.74 | 1.0698 | 29.4565 |
| 2011 | 2876.701 | 492.160 | 3368.861 | 205.907 | 0.686 | 14.61 | 1.0611 | 28.3798 |

Discards make up approximately 6.9 \% of the catch over the 1998-2006 period.

The catch rate trend used was from trawl caught flathead in zones 10 and 20. The fishery was well developed before 1986, the start of our data series.

Table 20.62 RBC for Flathead. $\mathrm{C}_{\text {targ }}$ and $\mathrm{CPUE}_{\text {targ }}$ relate to the period 1986-1995, $\mathrm{CPUE}_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7).

| Ref_Year | $1986-1995$ |
| ---: | ---: |
| CE_Targ | 1.0559 |
| CE_Lim | 0.4224 |
| CE_Recent | 1.1128 |
| Wt_Discard | 353.129 |
| Scaling | 1.0898 |
| Year's TAC |  |
| C $_{\text {targ }}$ | 2871.32 |



Figure 20.48 Tiger Flathead. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 20.7.3 Eastern Gemfish (GEM - 37439002 - Rexea solandri)

Table 20.63 Eastern Gemfish data for the TIER 4 calculations. Total is the sum of Discards, State, Non Trawl and SEF2 catches. All values in Tonnes. CE is the standardized catch rate for Zones 10, 20, and 30 in depths $300-500 \mathrm{~m}$ from June to September (Haddon, 2012). GeoMean is the geometric mean catch rate. Discards are estimates from 1998 to present. The ratio of discards to catch over the 1998 2006 period was used to estimate the discards between 1986 and 1997, the proportion of which is the PDiscard.

| Year | Catch | Discards | Total | State | Non-T | PDiscard | CE | GeoMean |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 1945.200 | 218.268 | 2163.468 |  |  | 10.09 | 2.3896 | 15.3241 |
| 1987 | 2208.000 | 247.757 | 2455.757 |  |  | 10.09 | 2.9835 | 25.2674 |
| 1988 | 1148.400 | 128.860 | 1277.260 |  |  | 10.09 | 2.6781 | 20.3738 |
| 1989 | 980.400 | 110.009 | 1090.409 |  |  | 10.09 | 1.9458 | 12.8697 |
| 1990 | 979.200 | 109.875 | 1089.075 |  |  | 10.09 | 1.7081 | 12.0080 |
| 1991 | 301.200 | 33.797 | 334.997 |  |  | 10.09 | 1.1499 | 8.4919 |
| 1992 | 1028.400 | 115.395 | 1143.795 |  |  | 10.09 | 1.5743 | 10.6133 |
| 1993 | 457.200 | 51.302 | 508.502 |  |  | 10.09 | 1.2954 | 8.9852 |
| 1994 | 266.110 | 29.860 | 295.970 | 131.931 | 0.000 | 10.09 | 0.8826 | 6.2854 |
| 1995 | 251.022 | 28.167 | 279.189 | 157.756 | 0.000 | 10.09 | 0.8009 | 5.4906 |
| 1996 | 315.471 | 35.399 | 350.869 | 204.700 | 0.000 | 10.09 | 0.5858 | 3.9966 |
| 1997 | 529.152 | 59.375 | 588.527 | 136.395 | 0.000 | 10.09 | 0.5934 | 4.1253 |
| 1998 | 373.133 | 23.000 | 396.133 | 127.144 | 0.000 | 5.81 | 0.5742 | 4.0091 |
| 1999 | 247.201 | 31.000 | 278.201 | 88.664 | 0.000 | 11.14 | 0.4234 | 2.7336 |
| 2000 | 123.746 | 29.000 | 152.746 | 30.747 | 0.000 | 18.99 | 0.3905 | 2.5299 |
| 2001 | 110.245 | 8.000 | 118.245 | 23.859 | 2.702 | 6.77 | 0.3369 | 1.9996 |
| 2002 | 77.867 | 13.600 | 91.467 | 16.174 | 3.564 | 14.87 | 0.2622 | 1.5421 |
| 2003 | 82.841 | 115.170 | 198.011 | 7.781 | 2.697 | 58.16 | 0.2954 | 1.6954 |
| 2004 | 97.542 | 83.210 | 180.752 | 17.731 | 2.683 | 46.04 | 0.4142 | 2.5873 |
| 2005 | 112.493 | 77.650 | 190.143 | 15.751 | 8.598 | 40.84 | 0.4370 | 2.7875 |
| 2006 | 101.951 | 46.350 | 148.301 | 15.153 | 6.564 | 31.25 | 0.4652 | 2.8952 |
| 2007 | 93.213 | 128.758 | 221.971 | 14.091 | 10.096 | 58.01 | 0.6460 | 4.0265 |
| 2008 | 118.957 | 164.319 | 283.276 | 11.607 | 20.277 | 58.01 | 0.8828 | 5.5997 |
| 2009 | 101.999 | 171.228 | 273.227 | 16.294 | 11.688 | 62.67 | 0.9137 | 5.4510 |
| 2010 | 112.643 | 190.964 | 303.607 | 20.128 | 16.264 | 62.90 | 0.7197 | 3.8269 |
| 2011 | 85.035 | 106.586 | 191.620 | 14.579 | 10.492 | 55.62 | 0.6515 | 3.5366 |

Discards make up approximately 10.08 \% of the catch over the 1998-2002 period. The reduced period, relative to other species, reflects the bycatch nature of the fishery in recent years.

Table 20.64 RBC calculations for Eastern Gemfish. $\mathrm{C}_{\text {targ }}$ and CPUE $_{\text {targ }}$ relate to the period 1993-2002, $\mathrm{CPUE}_{\mathrm{Lim}}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7).

| Ref_Year | $1993-2002$ |
| ---: | ---: |
| CE_Targ | 0.6145 |
| CE_Lim | 0.2458 |
| CE_Recent | 0.7919 |
| Wt_Discard | 141.554 |
| Scaling | 1.4811 |
| Year’s TAC |  |
| C $_{\text {targ }}$ | 305.985 |



Figure 20.49 Eastern Gemfish. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 20.7.4 Western Gemfish (GEM - 37439002 - Rexea solandri)

This relates solely to the SESSF zones 40 and 50 ; specifically it does not include the GAB, either in the catch rate standardization or the catches.

[^8]| Year | Catch | Discards | Total | State | Non-T | PDiscard | CE | GeoMea |
| ---: | ---: | ---: | ---: | :--- | :--- | ---: | ---: | ---: |
| $n$ |  |  |  |  |  |  |  |  |

Discards make up approximately $4.6 \%$ of the catch over the 1998-2006 period.

Table 20.66 RBC calculations for Western Gemfish. $\mathrm{C}_{\text {targ }}$ and CPUE $_{\text {targ }}$ relate to the period 1992-2001, $\mathrm{CPUE}_{\mathrm{Lim}}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Only catches from zones 40 and 50 included. Wt_Discard is the weighted average discards from the last four years, as with Equ (7).

| Ref_Year | $1992-2001$ |
| ---: | ---: |
| CE_Targ | 0.8678 |
| CE_Lim | 0.3471 |
| CE_Recent | 0.6635 |
| Wt_Discard | 80.108 |
| Scaling | 0.6077 |
| Last Year’s TAC |  |
| C $_{\text {targ }}$ | 203.063 |

## WesternGemfish



Figure 20.50 Western Gemfish. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 20.7.5 Jackass Morwong (MOR - 37377003 -Nemadactylus macropterus)

Table 20.67 Jackass Morwong data for the TIER 4 calculations. Total is the sum of Discards, State, Non Trawl and SEF2 catches. All values in Tonnes. CE is the standardized catch rate for Zones 10 to 50 in depths $70-360 \mathrm{~m}$ (Haddon, 2012). GeoMean is the geometric mean catch rates. Discards are estimates from 1998 to present. The ratio of discards to catch over the 1998-2006 period was used to estimate the discards between 1986 and 1997, the proportion of which is the PDiscard.

| Year | Catch Discards | Total | State | Non-T | PDisca <br> rd | CE | GeoMea <br> $n$ |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 1390.800 | 67.232 | 1458.032 |  |  | 4.61 | 1.8326 | 22.5592 |
| 1987 | 1459.200 | 70.538 | 1529.738 |  |  | 4.61 | 2.0862 | 26.1917 |
| 1988 | 1742.400 | 84.228 | 1826.628 |  |  | 4.61 | 2.0565 | 29.1554 |
| 1989 | 1971.600 | 95.308 | 2066.908 |  |  | 4.61 | 1.9994 | 33.9001 |
| 1990 | 1129.200 | 54.586 | 1183.786 |  |  | 4.61 | 1.6572 | 24.2137 |
| 1991 | 1406.400 | 67.986 | 1474.386 |  |  | 4.61 | 1.4684 | 21.1181 |
| 1992 | 888.000 | 42.926 | 930.926 |  |  | 4.61 | 1.2157 | 19.1937 |
| 1993 | 1132.800 | 54.760 | 1187.560 |  |  | 4.61 | 1.2320 | 21.3530 |
| 1994 | 1034.932 | 50.029 | 1084.961 | 243.396 | 0.000 | 4.61 | 1.0502 | 18.0744 |
| 1995 | 981.801 | 47.461 | 1029.261 | 160.992 | 0.000 | 4.61 | 0.9875 | 16.3623 |
| 1996 | 972.505 | 47.011 | 1019.517 | 89.072 | 0.211 | 4.61 | 0.9246 | 13.8607 |
| 1997 | 1213.726 | 58.672 | 1272.398 | 95.060 | 3.192 | 4.61 | 0.9910 | 16.1581 |
| 1998 | 942.082 | 34.000 | 976.082 | 59.783 | 4.519 | 3.48 | 0.8476 | 13.4363 |
| 1999 | 992.195 | 45.000 | 1037.195 | 41.481 | 17.667 | 4.34 | 0.8793 | 14.1587 |
| 2000 | 950.483 | 27.000 | 977.483 | 41.087 | 29.294 | 2.76 | 0.7263 | 10.1983 |
| 2001 | 866.752 | 12.000 | 878.752 | 50.298 | 2.263 | 1.37 | 0.5431 | 8.3295 |
| 2002 | 879.234 | 25.440 | 904.674 | 29.445 | 1.874 | 2.81 | 0.5693 | 8.3275 |
| 2003 | 776.411 | 71.850 | 848.261 | 28.583 | 3.311 | 8.47 | 0.4899 | 7.9077 |
| 2004 | 797.330 | 47.380 | 844.710 | 37.380 | 4.593 | 5.61 | 0.4905 | 8.6153 |
| 2005 | 840.172 | 38.610 | 878.782 | 42.118 | 5.979 | 4.39 | 0.5259 | 8.9785 |
| 2006 | 812.736 | 78.550 | 891.286 | 34.415 | 5.306 | 8.81 | 0.6003 | 11.5427 |
| 2007 | 586.065 | 70.704 | 656.769 | 18.299 | 4.507 | 10.77 | 0.5987 | 12.2504 |
| 2008 | 715.142 | 86.276 | 801.418 | 12.108 | 5.740 | 10.77 | 0.7042 | 13.7889 |
| 2009 | 465.638 | 56.176 | 521.814 | 11.506 | 2.812 | 10.77 | 0.6212 | 11.4713 |
| 2010 | 376.393 | 21.121 | 397.515 | 8.435 | 3.007 | 5.31 | 0.4635 | 8.5497 |
| 2011 | 411.007 | 46.510 | 457.517 | 5.001 | 2.399 | 10.17 | 0.4389 | 8.5254 |

Discards make up approximately 4.6 \% of the catch over the 1998-2006 period.

Table 20.68. RBC calculations for Jackass Morwong. $\mathrm{C}_{\text {targ }}$ and $\mathrm{CPUE}_{\text {targ }}$ relate to the period 1986-1995, $\mathrm{CPUE}_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. The Wt_Discard is the weighted discards from 2008-2011, as in Equ (7) .

| Ref_Year | $1986-1995$ |
| ---: | ---: |
| CE_Targ | 1.5586 |
| CE_Lim | 0.6234 |
| CE_Recent | 0.557 |
| Wt_Discard | 43.680 |
| Scaling | 0 |
| Year's TAC |  |
| C $_{\text {targ }}$ | 1377.219 |

JackassMorwong





Figure 20.51 Jackass Morwong. Top left is the total removals with the line illustrating the target catch. Top right represents the standardized catch rates with the upper line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 20.7.6 John Dory (DOJ - 37264004 - Zeus faber)

It was decided that this year the option of treating John Dory as a non-target species would be examined. This entailed changing the implied target reference point from $48 \%$ of the unfished state to $40 \%$ of the unfished state. Because the target catch rate is taken as a proxy for $48 \%$ unfished biomass, to make it equivalent to $40 \%$ means the average catch rate over the reference period should be multiplied by 0.8333 (thus $0.83334 \times 48=40$ ).

Table 20.69 John Dory data for the TIER 4 calculations. Total is the sum of Discards, State, Non Trawl and SEF2 catches. All values in Tonnes. CE is the standardized catch rate for Zones 10 and 20 in depths $0-200 \mathrm{~m}$ (Haddon, 2012). GeoMean is the geometric mean catch rates. Discards are estimates from 1998 to present. The ratio of discards to catch over the 1998 - 2006 period was used to estimate the discards between 1986 and 1997, the proportion of which is the PDiscard.

| Year | Catch | Discards | Total | State | Non-T | PDiscard | CE | GeoMea <br> $n$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 301.200 | 7.987 | 309.187 |  |  | 2.58 | 1.5483 | 7.6948 |
| 1987 | 240.000 | 6.364 | 246.364 |  |  | 2.58 | 1.7713 | 8.5155 |
| 1988 | 226.800 | 6.014 | 232.814 |  |  | 2.58 | 1.6647 | 8.3856 |
| 1989 | 252.000 | 6.683 | 258.683 |  |  | 2.58 | 1.8250 | 9.5319 |
| 1990 | 212.400 | 5.633 | 218.033 |  |  | 2.58 | 1.6491 | 8.7451 |
| 1991 | 236.400 | 6.269 | 242.669 |  |  | 2.58 | 1.3564 | 7.1954 |
| 1992 | 240.000 | 6.364 | 246.364 |  |  | 2.58 | 1.1209 | 5.6282 |
| 1993 | 400.800 | 10.629 | 411.429 |  |  | 2.58 | 1.4522 | 7.0963 |
| 1994 | 289.728 | 7.683 | 297.411 | 176.767 | 0.000 | 2.58 | 1.3690 | 6.7516 |
| 1995 | 243.673 | 6.462 | 250.135 | 129.268 | 0.000 | 2.58 | 1.1594 | 5.9610 |
| 1996 | 137.004 | 3.633 | 140.637 | 2.107 | 0.000 | 2.58 | 0.8950 | 4.5279 |
| 1997 | 178.118 | 4.723 | 182.841 | 88.373 | 0.000 | 2.58 | 0.7008 | 3.3776 |
| 1998 | 138.811 | 3.000 | 141.811 | 23.993 | 0.000 | 2.12 | 0.7274 | 3.6350 |
| 1999 | 178.334 | 3.000 | 181.334 | 40.806 | 0.000 | 1.65 | 0.8410 | 3.9411 |
| 2000 | 209.229 | 17.000 | 226.229 | 39.601 | 0.000 | 7.51 | 0.7920 | 3.5716 |
| 2001 | 164.643 | 6.000 | 170.643 | 29.821 | 0.051 | 3.52 | 0.6650 | 2.9450 |
| 2002 | 182.316 | 1.660 | 183.976 | 19.794 | 0.014 | 0.90 | 0.6556 | 3.1506 |
| 2003 | 193.130 | 3.190 | 196.320 | 28.348 | 0.084 | 1.62 | 0.6388 | 3.1538 |
| 2004 | 193.824 | 1.740 | 195.564 | 27.679 | 0.113 | 0.89 | 0.6745 | 3.4191 |
| 2005 | 132.030 | 3.530 | 135.560 | 29.319 | 0.060 | 2.60 | 0.5638 | 2.6772 |
| 2006 | 107.020 | 0.640 | 107.660 | 23.481 | 0.011 | 0.59 | 0.6354 | 2.8463 |
| 2007 | 82.383 | 1.355 | 83.738 | 13.849 | 0.016 | 1.62 | 0.5790 | 2.8023 |
| 2008 | 177.122 | 0.596 | 177.718 | 41.012 | 0.011 | 0.34 | 0.8620 | 4.3014 |
| 2009 | 127.476 | 4.252 | 131.728 | 19.671 | 0.012 | 3.23 | 0.7984 | 4.1921 |
| 2010 | 86.573 | 2.942 | 89.515 | 14.267 | 0.025 | 3.29 | 0.5177 | 2.6414 |
| 2011 | 119.316 | 39.760 | 159.076 | 27.282 | 0.013 | 24.99 | 0.5373 | 2.7474 |

Discards make up approximately $2.6 \%$ of the catch over the 1998-2006 period.

Table 20.70 RBC calculations for John Dory. $\mathrm{C}_{\text {targ }}$ and $\mathrm{CPUE}_{\text {targ }}$ relate to the period 1986-1995, $\mathrm{CPUE}_{\mathrm{Lim}}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7).

| Ref_Year | $1986-1995$ |
| ---: | ---: |
| CE_Targ | 1.4916 |
| CE_Lim | 0.5967 |
| CE_Recent | 0.6788 |
| Wt_Discard | 5.686 |
| Scaling | 0.0918 |
| Last Year's TAC | 221 |
| C $_{\text {targ }}$ | 271.309 |

JohnDory


Figure 20.52. John Dory. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

Table 20.71 RBC calculations for John Dory. $\mathrm{C}_{\text {targ }}$ and CPUE $_{\text {targ }}$ relate to $83.33 \%$ of the average over 1986-1995, $\mathrm{CPUE}_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7). The proxy target is here B40\%.

| Ref_Year | $1986-1995$ |
| ---: | ---: |
| CE_Targ | 1.2430 |
| CE_Lim | 0.5967 |
| CE_Recent | 0.6788 |
| Wt_Discard | 5.686 |
| Scaling | 0.1272 |
| Last Year's TAC | 221 |
| C $_{\text {targ }}$ | 271.309 |

## JohnDory



Figure 20.53. John Dory. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine purple line representing the target catch rate $(83.33 \%$ of the average over the reference period) and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 20.7.7 Mirror Dory (DOM - 37264003 - Zenopsis nebulosus)

Table 20.72 Mirror Dory data for the TIER 4 calculations. Total is the sum of Discards, State, Non
Trawl and SEF2 catches. All values in Tonnes. CE is the standardized catch rate for Zones 10 to 50 in
depths $0-600 \mathrm{~m}$ (Haddon, 2012). GeoMean is the geometric mean catch rates. Discards are estimates
from 1998 to present. The ratio of discards to catch over the $1998-2006$ period was used to estimate the
discards between 1986 and 1997, the proportion of which is the PDiscard.

| Year | Catch | Discard <br> s | Total | State | Non-T | PDiscar <br> d | CE | GeoMea <br> $n$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 336.000 | 80.920 | 416.920 |  |  | 19.41 | 1.2117 | 18.6423 |
| 1987 | 340.800 | 82.076 | 422.876 |  |  | 19.41 | 1.2147 | 19.7476 |
| 1988 | 373.200 | 89.879 | 463.079 |  |  | 19.41 | 1.1875 | 16.9455 |
| 1989 | 542.400 | 130.628 | 673.028 |  |  | 19.41 | 1.4715 | 23.1957 |
| 1990 | 267.600 | 64.447 | 332.047 |  |  | 19.41 | 1.3530 | 20.6077 |
| 1991 | 277.200 | 66.759 | 343.959 |  |  | 19.41 | 1.1550 | 13.9567 |
| 1992 | 357.600 | 86.122 | 443.722 |  |  | 19.41 | 0.9990 | 11.3487 |
| 1993 | 537.600 | 129.472 | 667.072 |  |  | 19.41 | 1.0955 | 13.7999 |
| 1994 | 324.664 | 78.190 | 402.854 | 21.816 | 0.000 | 19.41 | 0.9802 | 11.4667 |
| 1995 | 289.953 | 69.830 | 359.783 | 22.320 | 0.000 | 19.41 | 0.9047 | 10.0782 |
| 1996 | 404.725 | 97.471 | 502.196 | 21.715 | 0.000 | 19.41 | 0.8782 | 8.9039 |
| 1997 | 547.416 | 131.836 | 679.252 | 22.021 | 0.000 | 19.41 | 0.9362 | 9.6820 |
| 1998 | 439.374 | 115.000 | 554.374 | 26.988 | 0.000 | 20.74 | 0.8485 | 9.0983 |
| 1999 | 382.139 | 52.000 | 434.139 | 36.911 | 0.000 | 11.98 | 0.7039 | 8.0995 |
| 2000 | 217.405 | 93.000 | 310.405 | 11.121 | 0.000 | 29.96 | 0.4847 | 4.6519 |
| 2001 | 306.752 | 292.000 | 598.752 | 10.600 | 0.096 | 48.77 | 0.5653 | 5.1157 |
| 2002 | 545.156 | 96.920 | 642.076 | 21.650 | 0.029 | 15.09 | 0.7542 | 7.1647 |
| 2003 | 737.989 | 163.710 | 901.699 | 68.468 | 0.000 | 18.16 | 0.9179 | 8.6661 |
| 2004 | 628.392 | 170.310 | 798.702 | 106.386 | 0.505 | 21.32 | 0.8819 | 8.2044 |
| 2005 | 663.887 | 52.720 | 716.607 | 73.442 | 0.008 | 7.36 | 0.9762 | 9.3924 |
| 2006 | 490.852 | 26.880 | 517.732 | 85.434 | 0.058 | 5.19 | 0.9631 | 9.7517 |
| 2007 | 335.763 | 64.522 | 400.284 | 29.067 | 0.060 | 16.12 | 0.9292 | 9.5152 |
| 2008 | 463.424 | 89.595 | 553.019 | 22.103 | 0.002 | 16.20 | 1.1150 | 12.2034 |
| 2009 | 561.250 | 369.419 | 930.669 | 35.112 | 0.000 | 39.69 | 1.2256 | 13.1797 |
| 2010 | 629.323 | 275.697 | 905.020 | 12.028 | 0.037 | 30.46 | 1.1642 | 12.8612 |
| 2011 | 571.704 | 247.571 | 819.275 | 6.077 | 3.492 | 30.22 | 1.0831 | 10.8311 |

Discards make up approximately 19.41 \% of the catch over the 1998-2006 years.

Table 20.73 RBC calculations for Mirror Dory. $\mathrm{C}_{\text {targ }}$ and $\mathrm{CPUE}_{\text {targ }}$ relate to the period 1986-1995, $\mathrm{CPUE}_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7).

| Ref_Year | $1992-1997 \& 2003-2006$ |
| ---: | ---: |
| CE_Targ | 0.9533 |
| CE_Lim | 0.3813 |
| CE_Recent | 1.147 |
| Wt_Discard | 260.786 |
| Scaling | 1.3386 |
| Last Year’s TAC |  |
| C $_{\text {targ }}$ | 598.962 |

MirrorDory





Figure 20.54 Mirror Dory. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 20.7.8 Mirror Dory East (DOM - 37264003 - Z. nebulosus)

Table 20.74 Mirror Dory data for the TIER 4 calculations. Total is the sum of Discards, State, Non
Trawl and SEF2 catches. All values in Tonnes. CE is the standardized catch rate for Zones 10 to 30 in
depths $0-600 \mathrm{~m}$ (Haddon, 2012). GeoMean is the geometric mean catch rates. Discards are estimates
from 1998 to present. The ratio of discards to catch over the $1998-2006$ period was used to estimate the
discards between 1986 and 1997, the proportion of which is the PDiscard.

| Year | Catch | Discard <br> s | Total | State | Non-T | PDiscar <br> d | CE | GeoMea <br> $n$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 329.399 | 79.330 | 408.729 |  |  | 19.41 | 1.1585 | 18.7487 |
| 1987 | 328.474 | 79.107 | 407.581 |  |  | 19.41 | 1.1556 | 19.9429 |
| 1988 | 356.164 | 85.776 | 441.939 |  |  | 19.41 | 1.1336 | 16.8882 |
| 1989 | 530.901 | 127.858 | 658.759 |  |  | 19.41 | 1.3791 | 23.1617 |
| 1990 | 257.511 | 62.017 | 319.528 |  |  | 19.41 | 1.2896 | 20.5538 |
| 1991 | 257.915 | 62.114 | 320.029 |  |  | 19.41 | 1.1339 | 14.2052 |
| 1992 | 337.458 | 81.271 | 418.728 |  |  | 19.41 | 0.9845 | 11.7312 |
| 1993 | 503.639 | 121.293 | 624.932 |  |  | 19.41 | 1.0792 | 14.1976 |
| 1994 | 303.620 | 73.121 | 376.741 | 20.402 | 0.000 | 19.41 | 0.9448 | 11.6924 |
| 1995 | 242.777 | 58.469 | 301.245 | 18.688 | 0.000 | 19.41 | 0.8577 | 10.2913 |
| 1996 | 262.435 | 63.203 | 325.638 | 14.081 | 0.000 | 19.41 | 0.7617 | 7.7998 |
| 1997 | 361.397 | 87.036 | 448.433 | 14.538 | 0.000 | 19.41 | 0.8100 | 8.6425 |
| 1998 | 292.102 | 76.454 | 368.556 | 17.942 | 0.000 | 20.74 | 0.7297 | 8.0944 |
| 1999 | 301.020 | 40.962 | 341.981 | 29.076 | 0.000 | 11.98 | 0.6626 | 7.8713 |
| 2000 | 187.852 | 80.358 | 268.209 | 9.610 | 0.000 | 29.96 | 0.4995 | 4.7885 |
| 2001 | 168.695 | 160.582 | 329.277 | 5.829 | 0.053 | 48.77 | 0.4995 | 4.0443 |
| 2002 | 243.846 | 43.352 | 287.198 | 9.684 | 0.013 | 15.09 | 0.6208 | 5.2594 |
| 2003 | 534.068 | 118.474 | 652.541 | 49.549 | 0.000 | 18.16 | 0.9105 | 7.7688 |
| 2004 | 406.456 | 110.160 | 516.616 | 68.813 | 0.327 | 21.32 | 0.8636 | 7.2635 |
| 2005 | 537.095 | 42.651 | 579.747 | 59.416 | 0.006 | 7.36 | 1.1053 | 9.9946 |
| 2006 | 402.462 | 22.040 | 424.502 | 70.049 | 0.048 | 5.19 | 1.1074 | 10.3893 |
| 2007 | 254.433 | 48.893 | 303.326 | 22.026 | 0.046 | 16.12 | 1.1963 | 11.4463 |
| 2008 | 391.327 | 75.656 | 466.983 | 18.664 | 0.002 | 16.20 | 1.3389 | 14.4563 |
| 2009 | 411.441 | 270.813 | 682.255 | 25.740 | 0.000 | 39.69 | 1.4196 | 15.8458 |
| 2010 | 430.160 | 188.446 | 618.607 | 8.221 | 0.025 | 30.46 | 1.1806 | 14.3976 |
| 2011 | 393.757 | 170.513 | 564.270 | 4.185 | 2.405 | 30.22 | 1.1774 | 12.7502 |

Discards make up approximately 19.41 \% of the catch over the 1998-2006 period.

Table 20.75 RBC calculations for Mirror Dory East. $\mathrm{C}_{\text {targ }}$ and CPUE $_{\text {targ }}$ relate to the period 1986-1995, $\mathrm{CPUE}_{\mathrm{Lim}}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7).

| Ref_Year | $1986-1995$ |
| ---: | ---: |
| CE_Targ | 1.1117 |
| CE_Lim | 0.4447 |
| CE_Recent | 1.2791 |
| Wt_Discard | 182.345 |
| Scaling | 1.2511 |
| Last Year's TAC |  |
| C $_{\text {targ }}$ | 427.821 |

## MirrorDoryE






Figure 20.55 Mirror Dory. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.
20.7.9 Mirror Dory West (DOM - 37264003 - Z. nebulosus)

Table 20.76 Mirror Dory data for the TIER 4 calculations. Total is the sum of Discards, State, Non
Trawl and SEF2 catches. All values in Tonnes. CE is the standardized catch rate for Zones 40 to 50 in
depths $0-600 \mathrm{~m}$ (Haddon, 2012). GeoMean is the geometric mean catch rates. Discards are estimates
from 1998 to present. The ratio of discards to catch over the $1998-2006$ period was used to estimate the
discards between 1986 and 1997, the proportion of which is the PDiscard.

| Year | Catch | Discard <br> s | Total | State | Non-T | PDiscar <br> d | CE | GeoMea <br> $n$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 6.601 | 1.590 | 8.190 |  |  | 19.41 | 2.3666 | 13.7130 |
| 1987 | 12.326 | 2.968 | 15.294 |  |  | 19.41 | 1.5740 | 16.0832 |
| 1988 | 17.036 | 4.103 | 21.139 |  |  | 19.41 | 1.2966 | 18.4525 |
| 1989 | 11.499 | 2.769 | 14.268 |  |  | 19.41 | 1.6475 | 24.6757 |
| 1990 | 10.089 | 2.430 | 12.518 |  |  | 19.41 | 1.1110 | 21.6631 |
| 1991 | 19.285 | 4.645 | 23.930 |  |  | 19.41 | 0.7909 | 11.7670 |
| 1992 | 20.142 | 4.851 | 24.993 |  |  | 19.41 | 0.6541 | 8.1608 |
| 1993 | 33.961 | 8.179 | 42.139 |  |  | 19.41 | 0.7732 | 10.1017 |
| 1994 | 21.044 | 5.068 | 26.113 | 1.414 | 0.000 | 19.41 | 0.6758 | 9.3264 |
| 1995 | 47.176 | 11.362 | 58.538 | 3.632 | 0.000 | 19.41 | 0.8594 | 9.0896 |
| 1996 | 142.290 | 34.268 | 176.558 | 7.634 | 0.000 | 19.41 | 1.2433 | 13.3473 |
| 1997 | 186.019 | 44.799 | 230.818 | 7.483 | 0.000 | 19.41 | 1.2670 | 12.8686 |
| 1998 | 147.272 | 38.546 | 185.818 | 9.046 | 0.000 | 20.74 | 1.2404 | 12.6121 |
| 1999 | 81.119 | 11.038 | 92.158 | 7.835 | 0.000 | 11.98 | 0.8103 | 8.8763 |
| 2000 | 29.554 | 12.642 | 42.196 | 1.512 | 0.000 | 29.96 | 0.4281 | 4.0569 |
| 2001 | 138.057 | 131.418 | 269.475 | 4.771 | 0.043 | 48.77 | 0.7384 | 7.9361 |
| 2002 | 301.310 | 53.568 | 354.878 | 11.966 | 0.016 | 15.09 | 1.0837 | 11.7181 |
| 2003 | 203.921 | 45.236 | 249.158 | 18.919 | 0.000 | 18.16 | 0.9361 | 11.0165 |
| 2004 | 221.936 | 60.150 | 282.086 | 37.573 | 0.178 | 21.32 | 0.9388 | 10.3786 |
| 2005 | 126.791 | 10.069 | 136.860 | 14.026 | 0.002 | 7.36 | 0.7379 | 8.0456 |
| 2006 | 88.390 | 4.840 | 93.231 | 15.385 | 0.010 | 5.19 | 0.6304 | 8.0395 |
| 2007 | 81.330 | 15.629 | 96.959 | 7.041 | 0.015 | 16.12 | 0.5653 | 6.7120 |
| 2008 | 72.097 | 13.939 | 86.036 | 3.439 | 0.000 | 16.20 | 0.6335 | 7.5767 |
| 2009 | 149.809 | 98.605 | 248.414 | 9.372 | 0.000 | 39.69 | 0.9617 | 9.7010 |
| 2010 | 199.163 | 87.250 | 286.413 | 3.807 | 0.012 | 30.46 | 1.1448 | 11.0745 |
| 2011 | 177.946 | 77.058 | 255.004 | 1.891 | 1.087 | 30.22 | 0.8911 | 8.6540 |

Discards make up approximately 19.41 \% of the catch over the 1998-2006 period.

Table 20.77 RBC calculations for Mirror Dory. $\mathrm{C}_{\text {targ }}$ and $\mathrm{CPUE}_{\text {targ }}$ relate to the period 1996-2005, $\mathrm{CPUE}_{\mathrm{Lim}}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7).

| Ref_Year | $1996-2005$ |
| ---: | ---: |
| CE_Targ | 0.9424 |
| CE_Lim | 0.377 |
| CE_Recent | 0.9078 |
| Wt_Discard | 78.441 |
| Scaling | 0.9388 |
| Year's TAC |  |
| C $_{\text {targ }}$ | 202.001 |



Figure 20.56 Mirror Dory. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 20.7.10 Pink Ling (LIG - 37228002 - Genypterus blacodes)

Table 20.78 Pink Ling data for the TIER 4 calculations. Total is the sum of Discards, State, Non Trawl and SEF2 catches. All values in Tonnes. CE is the standardized catch rate for Zones 10, 20 and 30 in depths $0-1000 \mathrm{~m}$ (Haddon, 2012). GeoMean is the geometric mean catch rates. Discards are estimates from 1998 to present. The ratio of discards to catch over the 1998 - 2006 period was used to estimate the discards between 1986 and 1997, the proportion of which is the PDiscard.

| Year | Catch | Discards | Total | State | Non-T | PDisc <br> ard | CE | LnCE |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 650.400 | 3.627 | 654.027 |  |  | 0.55 | 1.1020 | 20.6650 |
| 1987 | 802.800 | 4.477 | 807.277 |  |  | 0.55 | 1.1680 | 19.4240 |
| 1988 | 621.600 | 3.467 | 625.067 |  |  | 0.55 | 1.1140 | 20.2590 |
| 1989 | 744.000 | 4.149 | 748.149 |  |  | 0.55 | 0.9590 | 19.1570 |
| 1990 | 776.400 | 4.330 | 780.730 |  |  | 0.55 | 1.4110 | 26.8200 |
| 1991 | 910.800 | 5.080 | 915.880 |  |  | 0.55 | 1.4120 | 26.3050 |
| 1992 | 1081.200 | 6.030 | 1087.230 |  |  | 0.55 | 1.0890 | 24.8500 |
| 1993 | 1657.200 | 9.243 | 1666.443 |  |  | 0.55 | 1.0230 | 25.3070 |
| 1994 | 1463.324 | 8.161 | 1471.485 | 538.219 | 0.000 | 0.55 | 1.0370 | 23.5160 |
| 1995 | 1944.501 | 10.845 | 1955.346 | 672.495 | 0.000 | 0.55 | 1.3070 | 25.8110 |
| 1996 | 2244.320 | 12.517 | 2256.837 | 811.461 | 0.000 | 0.55 | 1.3000 | 27.6570 |
| 1997 | 2128.990 | 11.874 | 2140.864 | 393.906 | 0.000 | 0.55 | 1.3230 | 27.9370 |
| 1998 | 1933.870 | 41.000 | 1974.870 | 52.110 | 202.385 | 2.08 | 1.3230 | 26.0160 |
| 1999 | 2022.297 | 12.000 | 2034.297 | 50.847 | 270.504 | 0.59 | 1.2090 | 25.2290 |
| 2000 | 1860.795 | 11.000 | 1871.795 | 19.036 | 251.991 | 0.59 | 1.0770 | 22.4050 |
| 2001 | 1733.968 | 5.000 | 1738.968 | 9.879 | 376.583 | 0.29 | 0.8390 | 19.0620 |
| 2002 | 1610.520 | 6.640 | 1617.160 | 15.634 | 522.209 | 0.41 | 0.7330 | 15.8660 |
| 2003 | 1617.638 | 1.390 | 1619.028 | 8.277 | 477.475 | 0.09 | 0.7520 | 18.2930 |
| 2004 | 1766.179 | 1.390 | 1767.569 | 12.201 | 850.448 | 0.08 | 0.6720 | 16.7980 |
| 2005 | 1421.688 | 3.330 | 1425.018 | 20.897 | 644.493 | 0.23 | 0.6280 | 16.3340 |
| 2006 | 1200.188 | 2.840 | 1203.028 | 15.646 | 455.183 | 0.24 | 0.7500 | 21.3190 |
| 2007 | 1010.801 | 21.554 | 1032.355 | 23.812 | 339.055 | 2.09 | 0.7360 | 20.5010 |
| 2008 | 1182.085 | 16.542 | 1198.627 | 32.212 | 443.663 | 1.38 | 0.8520 | 25.1510 |
| 2009 | 900.949 | 50.088 | 951.036 | 16.474 | 298.114 | 5.27 | 0.6220 | 18.2950 |
| 2010 | 1098.359 | 57.616 | 1155.975 | 54.356 | 388.518 | 4.98 | 0.7610 | 20.7210 |
| 2011 | 1244.238 | 14.446 | 1258.684 | 25.420 | 429.517 | 1.15 | 0.8030 | 23.4440 |

Discards make up approximately 0.54 \% of the catch over the 1998-2006 period. The standardized catch rate series used was from Zones 10,20 and 30 as taken by trawl.

Table 20.79 RBC calculations for Pink Ling. $\mathrm{C}_{\text {targ }}$ and $\mathrm{CPUE}_{\text {targ }}$ relate to the period 1986-1995, $\mathrm{CPUE}_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7).

| Ref_Year | $1986-1995$ |
| ---: | ---: |
| CE_Targ | 1.1622 |
| CE_Lim | 0.4649 |
| CE_Recent | 0.7595 |
| Wt_Discard | 30.85 |
| Scaling | 0.4225 |
| Last Year's TAC |  |
| C $_{\text {targ }}$ | 1071.163 |

## PinkLingE



Figure 20.57 Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 20.7.11 RedFish (RED - 37258003 - Centroberyx affinis)

The period of the redfish fishery between 1991 to 1998 appears to have been during an era of heightened availability for redfish. This period is no longer considered to be representative of the fishery as it normally runs and has been running for the last few years.

Table 20.80 Redfish data for the TIER 4 calculations. Total is the sum of Discards, State, Non Trawl and
SEF2 catches. All values in Tonnes. CE is the standardized catch rate for Zone 10 in depths $0-400 \mathrm{~m}$
(Haddon, 2012) relative to the catch rate in 1986. GeoMean is the geometric mean catch rates. Discards are
estimates from 1998 to present. The ratio of discards to catch over the $1998-2006$ period was used to
estimate the discards between 1986 and 1997, the proportion of which is the PDiscard.

| Year | Catch | Discards | Total | State | Non- <br> T | PDisc <br> ard | CE | GeoMean |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 1426.800 | 904.992 | 2331.792 |  |  | 38.81 | 1.5835 | 38.3044 |
| 1987 | 986.400 | 625.655 | 1612.055 |  |  | 38.81 | 1.2675 | 35.9993 |
| 1988 | 961.200 | 609.671 | 1570.871 |  |  | 38.81 | 1.3164 | 37.3114 |
| 1989 | 649.200 | 411.775 | 1060.975 |  |  | 38.81 | 1.0928 | 29.4122 |
| 1990 | 792.000 | 502.350 | 1294.350 |  |  | 38.81 | 1.4854 | 37.2522 |
| 1991 | 1737.600 | 1102.126 | 2839.726 |  |  | 38.81 | 1.5398 | 39.9367 |
| 1992 | 2443.200 | 1549.675 | 3992.875 |  |  | 38.81 | 1.9882 | 50.0990 |
| 1993 | 2114.400 | 1341.123 | 3455.523 |  |  | 38.81 | 2.5202 | 56.0385 |
| 1994 | 1957.210 | 1241.421 | 3198.631 | 1345.606 | 0.000 | 38.81 | 1.7668 | 35.8972 |
| 1995 | 1999.572 | 1268.290 | 3267.862 | 789.249 | 0.000 | 38.81 | 1.1458 | 27.8589 |
| 1996 | 2219.833 | 1407.997 | 3627.831 | 784.092 | 0.000 | 38.81 | 0.9344 | 26.2588 |
| 1997 | 1840.798 | 1167.583 | 3008.380 | 304.137 | 0.000 | 38.81 | 1.0959 | 33.5183 |
| 1998 | 1835.469 | 2324.000 | 4159.469 | 83.849 | 0.000 | 55.87 | 1.3762 | 43.1196 |
| 1999 | 1346.976 | 69.000 | 1415.976 | 94.939 | 0.000 | 4.87 | 1.0812 | 32.7876 |
| 2000 | 859.909 | 233.000 | 1092.909 | 27.446 | 0.000 | 21.32 | 0.7288 | 22.7760 |
| 2001 | 846.662 | 738.000 | 1584.662 | 52.093 | 0.545 | 46.57 | 0.7123 | 17.8301 |
| 2002 | 926.928 | 894.850 | 1821.778 | 46.951 | 0.155 | 49.12 | 0.6089 | 16.4201 |
| 2003 | 726.661 | 347.500 | 1074.161 | 48.604 | 0.828 | 32.35 | 0.5826 | 17.0122 |
| 2004 | 557.603 | 377.440 | 935.043 | 58.124 | 1.005 | 40.37 | 0.4941 | 15.2541 |
| 2005 | 579.526 | 126.180 | 705.706 | 46.690 | 0.568 | 17.88 | 0.5057 | 16.1484 |
| 2006 | 397.194 | 13.070 | 410.264 | 75.690 | 0.541 | 3.19 | 0.4745 | 15.6812 |
| 2007 | 283.332 | 2.681 | 286.013 | 53.689 | 0.089 | 0.94 | 0.4237 | 15.4678 |
| 2008 | 230.566 | 2.182 | 232.748 | 29.369 | 0.163 | 0.94 | 0.3993 | 13.9780 |
| 2009 | 207.440 | 231.285 | 438.726 | 25.489 | 0.076 | 52.72 | 0.3253 | 11.3207 |
| 2010 | 187.992 | 27.086 | 215.079 | 22.340 | 0.019 | 12.59 | 0.3059 | 10.4815 |
| 2011 | 111.956 | 26.064 | 138.020 | 13.438 | 0.247 | 18.88 | 0.2447 | 8.5118 |
|  |  |  |  |  |  |  |  |  |

Discards make up approximately $38.8 \%$ of the catch over the $1998-2006$ period. The standardized catch rate series is from Zone 10.

Table 20.81 RBC calculations for Redfish. $\mathrm{C}_{\text {targ }}$ and CPUE $_{\text {targ }}$ relate to the period 1986-1990 and 1999-2003 (omitting a period of enhanced availability during the 1990 s). $\mathrm{CPUE}_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7).

| Ref_Year | 1986-1990\&1999-2003 |
| ---: | ---: |
| CE_Targ | 1.0459 |
| CE_Lim | 0.4184 |
| CE_Recent | 0.3188 |
| Wt_Discard | 52.107 |
| Scaling | 0 |
| Last Year’s TAC | 276 |
| C $_{\text {targ }}$ | 1485.953 |

## Redfish



Figure 20.58 Redfish. Left is the total removals with the fine line illustrating the target catch. Right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 20.7.12 RedFish plus Discards (RED - 37258003 - C. affinis)

The period of the redfish fishery between 1991 to 1998 appears to have been during an era of heightened availability for redfish. This period is no longer considered to be representative of the fishery as it normally runs and has been running for the last few years.

Table 20.82 Redfish data for the Alternative TIER 4 calculations using ratio mean catch rates that include
discards in the catch rate calculations. Total is the sum of Discards, and other catches. All values in
Tonnes. StandCE is the standardized catch rate for redfish from Zone 10 in depths $0-400 \mathrm{~m}$ (Haddon,
2012). GeoMean is the geometric mean catch rates (without discards). Discards are estimates from 1998
to present. DiscCE is the standardized catch rates multiplied by [ (Discard/Catch)+1], see Haddon
(2011c) for methods.
Year Catch Discards Total Effort (D/C)+1 StandCE DiscCE GeoMean

| 1986 | 1426.800 | 904.992 | 2331.792 | 15230 | 1.6343 | 1.5835 | 1.6149 | 1.4133 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1987 | 986.400 | 625.655 | 1612.055 | 11696 | 1.6343 | 1.2675 | 1.2927 | 1.3282 |
| 1988 | 961.200 | 609.671 | 1570.871 | 12009 | 1.6343 | 1.3164 | 1.3425 | 1.3767 |
| 1989 | 649.200 | 411.775 | 1060.975 | 8184 | 1.6343 | 1.0928 | 1.1145 | 1.0852 |
| 1990 | 792.000 | 502.350 | 1294.350 | 8217 | 1.6343 | 1.4854 | 1.5149 | 1.3745 |
| 1991 | 1737.600 | 1102.126 | 2839.726 | 10584 | 1.6343 | 1.5398 | 1.5704 | 1.4735 |
| 1992 | 2443.200 | 1549.675 | 3992.875 | 9440 | 1.6343 | 1.9882 | 2.0277 | 1.8485 |
| 1993 | 2114.400 | 1341.123 | 3455.523 | 11231 | 1.6343 | 2.5202 | 2.5702 | 2.0676 |
| 1994 | 1957.210 | 1241.421 | 3198.631 | 16182 | 1.6343 | 1.7668 | 1.8019 | 1.3245 |
| 1995 | 1999.572 | 1268.290 | 3267.862 | 16829 | 1.6343 | 1.1458 | 1.1685 | 1.0279 |
| 1996 | 2219.833 | 1407.997 | 3627.831 | 17763 | 1.6343 | 0.9344 | 0.9529 | 0.9689 |
| 1997 | 1840.798 | 1167.583 | 3008.380 | 13300 | 1.6343 | 1.0959 | 1.1177 | 1.2367 |
| 1998 | 1835.469 | 2324.000 | 4159.469 | 12958 | 2.2662 | 1.3762 | 1.9462 | 1.5910 |
| 1999 | 1346.976 | 69.000 | 1415.976 | 12048 | 1.0512 | 1.0812 | 0.7093 | 1.2097 |
| 2000 | 859.909 | 233.000 | 1092.909 | 14534 | 1.2710 | 0.7288 | 0.5780 | 0.8404 |
| 2001 | 846.662 | 738.000 | 1584.662 | 14186 | 1.8717 | 0.7123 | 0.8320 | 0.6579 |
| 2002 | 926.928 | 894.850 | 1821.778 | 16725 | 1.9654 | 0.6089 | 0.7468 | 0.6058 |
| 2003 | 726.661 | 347.500 | 1074.161 | 13389 | 1.4782 | 0.5826 | 0.5374 | 0.6277 |
| 2004 | 557.603 | 377.440 | 935.043 | 13137 | 1.6769 | 0.4941 | 0.5170 | 0.5628 |
| 2005 | 579.526 | 126.180 | 705.706 | 12939 | 1.2177 | 0.5057 | 0.3843 | 0.5958 |
| 2006 | 397.194 | 13.070 | 410.264 | 8826 | 1.0329 | 0.4745 | 0.3058 | 0.5786 |
| 2007 | 283.332 | 2.681 | 286.013 | 6104 | 1.0095 | 0.4237 | 0.2669 | 0.5707 |
| 2008 | 230.566 | 2.182 | 232.748 | 6407 | 1.0095 | 0.3993 | 0.2515 | 0.5157 |
| 2009 | 207.440 | 231.285 | 438.726 | 5457 | 2.1149 | 0.3253 | 0.4293 | 0.4177 |
| 2010 | 187.992 | 27.086 | 215.079 | 6456 | 1.1441 | 0.3059 | 0.2184 | 0.3867 |
| 2011 | 111.956 | 26.064 | 138.020 | 4760 | 1.2328 | 0.2447 | 0.1883 | 0.3141 |

Discards make up approximately $38.8 \%$ of the catch over the 1998-2006 period. The standardized catch rate series is from Zone 10.

Table 20.83 RBC calculations for Redfish. $\mathrm{C}_{\text {targ }}$ and $\mathrm{CPUE}_{\text {targ }}$ relate to the period 1986-1990 and 1999-2003 (omitting a period of enhanced availability during the 1990 s). $\mathrm{CPUE}_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7).

| Ref_Year | 1986-1990\&1999-2003 |
| ---: | ---: |
| CE_Targ | 1.0283 |
| CE_Lim | 0.4113 |
| CE_Recent | 0.2719 |
| Wt_Discard | 52.107 |
| Scaling | 0 |
| Last Year’s TAC |  |
| C targ | 1485.953 |

## Redfishdis



Figure 20.59 Redfish. Left is the total removals with the fine line illustrating the target catch. Right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 20.7.13 School Whiting (WHS - 37330014 - Sillago flindersi)

Table 20.84 School Whiting data for the TIER 4 calculations. Total is the sum of Discards, State, Non Trawl and SEF2 catches. All values in Tonnes. CE is the standardized catch rate for Zone 60 from depths 0 to 100 m (Haddon, 2012). GeoMean is the geometric mean catch rates. Discards are estimates from 1998 to present. The ratio of discards to catch over the $1998-2006$ period was used to estimate the discards between 1986 and 1997, the proportion of which is the PDiscard.

| Year | Catch | Discards | Total | State | $\begin{array}{r} \text { Non } \\ -T \end{array}$ | PDiscard | CE | GeoMean |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 1903.200 | 29.333 | 1932.533 |  |  | 1.518 | 1.1573 | 112.3054 |
| 1987 | 1320.000 | 20.344 | 1340.344 |  |  | 1.518 | 1.2735 | 131.1624 |
| 1988 | 1549.200 | 23.877 | 1573.077 |  |  | 1.518 | 1.6496 | 168.5490 |
| 1989 | 1220.400 | 18.809 | 1239.209 |  |  | 1.518 | 1.0919 | 127.0438 |
| 1990 | 2007.600 | 30.942 | 2038.542 |  |  | 1.518 | 1.6853 | 165.2959 |
| 1991 | 1866.000 | 28.760 | 1894.760 |  |  | 1.518 | 1.4283 | 164.1905 |
| 1992 | 1219.200 | 18.791 | 1237.991 |  |  | 1.518 | 1.0180 | 124.7066 |
| 1993 | 2007.600 | 30.942 | 2038.542 |  |  | 1.518 | 1.4500 | 152.4819 |
| 1994 | 1647.018 | 25.385 | 1672.403 | 766.818 | 0 | 1.518 | 0.8477 | 93.9314 |
| 1995 | 1990.79 | 30.683 | 2021.473 | 910.4204 | 0 | 1.518 | 1.0666 | 122.4731 |
| 1996 | 1695.105 | 26.126 | 1721.231 | 1038.743 | 0 | 1.518 | 0.6978 | 81.4339 |
| 1997 | 1556.38 | 23.988 | 1580.367 | 1169.811 | 0 | 1.518 | 0.5411 | 64.5619 |
| 1998 | 1813.848 | 48.000 | 1861.848 | 1396.053 | 0 | 2.578 | 0.5219 | 66.0158 |
| 1999 | 1448.81 | 5.000 | 1453.810 | 1011.862 | 0 | 0.344 | 0.5966 | 84.3634 |
| 2000 | 1289.46 | 9.000 | 1298.460 | 797.749 | 0 | 0.693 | 0.6006 | 65.1233 |
| 2001 | 1719.332 | 28.000 | 1747.332 | 1217.815 | 0 | 1.602 | 0.8457 | 93.2089 |
| 2002 | 1577.598 | 9.760 | 1587.358 | 1052.245 | 0 | 0.615 | 0.8697 | 90.8874 |
| 2003 | 1490.494 | 46.340 | 1536.834 | 926.2246 | 0 | 3.015 | 0.8898 | 87.1013 |
| 2004 | 1463.803 | 26.360 | 1490.163 | 1040.482 | 0 | 1.769 | 0.8435 | 79.7648 |
| 2005 | 1468.64 | 37.500 | 1506.140 | 1013.53 | 0 | 2.490 | 0.9513 | 77.2502 |
| 2006 | 1551.224 | 3.090 | 1554.314 | 1095.517 | $\begin{array}{r} 0.00 \\ 11 \end{array}$ | 0.199 | 0.8220 | 76.2250 |
| 2007 | 1636.456 | 3.260 | 1639.716 | 1197.766 | 0 | 0.199 | 1.0799 | 89.2381 |
| 2008 | 1369.947 | 2.729 | 1372.676 | 959.3665 | 0 | 0.199 | 1.0860 | 92.3448 |
| 2009 | 1227.521 | 2.445 | 1229.966 | 814.915 | 0 | 0.199 | 1.1350 | 93.6200 |
| 2010 | 1226.626 | 18.316 | 1244.942 | 846.88 | 0 | 1.471 | 1.0156 | 88.7190 |
| 2011 | 1240.111 | 58.467 | 1298.578 | 880.657 | 0 | 4.502 | 0.8354 | 72.0269 |

Discards make up approximately $1.5 \%$ of the catch over the 1998-2006 period.

Table 20.85 RBC calculations for School Whiting. $\mathrm{C}_{\text {targ }}$ and CPUE $_{\text {targ }}$ relate to the period 1986-1995, $\mathrm{CPUE}_{\mathrm{Lim}}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7).

| Ref_Year | $1986-1995$ |
| ---: | ---: |
| CE_Targ $^{\text {CE_Lim }}$ | 1.2668 |
| CE_Recent | 0.5067 |
| Wt_Discard | 1.018 |
| Scaling | 36.575 |
| Year's TAC | 0.6726 |
| C $_{\text {targ }}$ |  |
|  | 1698.887 |

## SchoolWhiting



Figure 20.60 School Whiting. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 20.7.14 Silver Warehou (TRS - 37445006 - Seriolella punctata)

Table 20.86 Spotted/Silver Warehou data for the TIER 4 calculations. Total is the sum of Discards, State, Non Trawl and SEF2 catches. All values in Tonnes. CE is the standardized catch rate for Zones 10 to 50 in depths $0-1000 \mathrm{~m}$ (Haddon, 2012). GeoMean is the geometric mean catch rates. Discards are estimates from 1998 to present. The ratio of discards to catch over the $1998-2006$ period was used to estimate the discards between 1986 and 1997, the proportion of which is the PDiscard.

| Year | Catch | Discards | Total | State | Non-T PDiscard | CE |  | GeoMean |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 1142.927 | 242.793 | 1385.720 |  |  | 17.52 | 1.4228 | 32.2897 |
| 1987 | 779.270 | 165.541 | 944.811 |  |  | 17.52 | 1.5121 | 35.5040 |
| 1988 | 1637.312 | 347.815 | 1985.127 |  |  | 17.52 | 1.9169 | 42.9346 |
| 1989 | 916.714 | 194.738 | 1111.452 |  |  | 17.52 | 1.5678 | 30.7291 |
| 1990 | 1319.413 | 280.284 | 1599.697 |  |  | 17.52 | 1.6590 | 40.6488 |
| 1991 | 1421.943 | 302.064 | 1724.007 |  |  | 17.52 | 1.1675 | 25.6848 |
| 1992 | 709.181 | 150.652 | 859.833 |  |  | 17.52 | 1.0123 | 27.9469 |
| 1993 | 1775.414 | 377.152 | 2152.566 |  |  | 17.52 | 1.1448 | 33.2988 |
| 1994 | 2054.296 | 436.396 | 2490.692 | 188.226 |  | 17.52 | 1.2249 | 34.7142 |
| 1995 | 2213.896 | 470.299 | 2684.196 | 148.791 |  | 17.52 | 1.1080 | 29.7825 |
| 1996 | 2735.681 | 581.143 | 3316.824 | 181.480 |  | 17.52 | 1.0502 | 22.7319 |
| 1997 | 2807.462 | 596.391 | 3403.853 | 37.925 |  | 17.52 | 1.0783 | 25.3481 |
| 1998 | 2433.954 | 2150.000 | 4583.954 | 24.112 |  | 46.90 | 1.0390 | 26.6416 |
| 1999 | 3255.217 | 45.000 | 3300.217 | 1.746 |  | 1.36 | 0.8951 | 31.2330 |
| 2000 | 3726.592 | 123.000 | 3849.592 | 0.464 |  | 3.20 | 0.8113 | 26.0708 |
| 2001 | 3295.454 | 695.000 | 3990.454 | 0.324 | 0.923 | 17.42 | 0.6811 | 21.7853 |
| 2002 | 4101.870 | 552.470 | 4654.340 | 0.487 | 0.701 | 11.87 | 0.7357 | 22.9919 |
| 2003 | 3060.003 | 769.760 | 3829.763 | 1.007 | 12.642 | 20.10 | 0.7411 | 20.4815 |
| 2004 | 3315.032 | 1183.280 | 4498.312 | 3.774 | 0.251 | 26.30 | 0.8231 | 23.3323 |
| 2005 | 2912.725 | 434.830 | 3347.555 | 4.996 | 0.139 | 12.99 | 0.8088 | 20.0277 |
| 2006 | 2374.182 | 95.630 | 2469.812 | 2.494 | 0.086 | 3.87 | 0.7116 | 18.2160 |
| 2007 | 1987.060 | 82.453 | 2069.513 | 4.373 | 0.056 | 3.98 | 0.6722 | 20.1239 |
| 2008 | 1522.999 | 49.718 | 1572.717 | 0.541 | 0.063 | 3.16 | 0.6018 | 16.1202 |
| 2009 | 1379.268 | 33.280 | 1412.548 | 1.240 | 0.002 | 2.36 | 0.6203 | 15.8837 |
| 2010 | 1288.673 | 17.155 | 1305.827 | 0.561 | 1.285 | 1.31 | 0.5126 | 13.2653 |
| 2011 | 1229.278 | 428.738 | 1658.015 | 0.547 | 0.1116 | 25.858 | 0.4816 | 12.5782 |

Discards make up approximately 17.52 \% of the catch over the 1998-2006 period. The standardization is an annual analysis conducted for the TIER 4 analysis.

Table 20.87 RBC calculations for Silver Warehou. $\mathrm{C}_{\text {targ }}$ and $\mathrm{CPUE}_{\text {targ }}$ relate to the period 1986-1995, $\mathrm{CPUE}_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. Wt_Discard is the weighted average discards from the last four years, as with Equ (7).

| Ref_Year | $1986-1995$ |
| ---: | ---: |
| CE_Targ | 0.6868 |
| CE_Lim | 0.2747 |
| CE_Recent $^{\text {Wt_Discard }}$ | 0.5541 |
| Scaling | 240.987 |
| Last Year's TAC | 0.6779 |
| C $_{\text {targ }}$ | 1693.81 |

## SilverWarehou



Figure 20.61 Spotted/Silver Warehou. Top left is the total removals with the fine line illustrating the target catch. Top right represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

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# 21. Saw Shark and Elephant Fish TIER 4 Analyses (Data 1980 2011) 

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### 21.1 Summary

The stock assessments that feed into the management control rules that reflect the harvest strategy adopted in the SESSF are arranged in a tiered system ranging from fully quantified modelled stock assessments (Tier 1) down to empirical rules based only on catch and catch rates (Tier 4). For those species where biological and fisheries data are limited an examination of trends in catch rates is used to modify allowable catches with the objective of managing the particular fishery towards a target that represents a desirable state for the fishery that also acts as a proxy for the general Harvest Strategy Policy target of $48 \% B_{0}$.

The Tier 4 control rule is used to calculate Recommended Biological Catches (RBCs) for saw sharks and elephant fish from the southern shark fishery. Standardized catch rates for both species (Haddon, 2012) were used along with total catches of the respective species in a standard analysis. This year's analysis varied from previous analyses by comparing the outcome of treating the catch rate target as a proxy for $48 \% B_{0}$ versus with a proxy for $40 \% B_{0}$ as an alternative target for these non-target species. For saw sharks the reported catches by trawl are now approaching the level of gill net catches so an additional analysis was conducted where the standardized catch rate for trawl saw shark catches was used instead of the gillnet catch rates.

The gillnet catch rates for saw sharks in 2011 barely differed from that in 2010 but owing to the initial drop in catch rates in 2010 the tier 4 analysis, which considers the average catch rate over the last four years, generates a RBC for saw sharks at the $48 \%$ target that has now declined to about $64 \%$ of the target catch. Whether the decline in the gillnet catch rates constitute a reasonable reflection of the stock status remains questionable due to the level of avoidance that occurs in the fishery (due to low and reducing value of saw sharks in the market). Importantly, when the trawl catch rates for saw sharks are standardized a different trend is apparent; the catches by trawl are almost a the same level as that taken by gill net.

The decline in catch rates in elephant fish seen in 2010 continued in 2011 and this implies a decrease in the RBC (Table 20.1). However, these values relate to the target catch rate being a proxy for $48 \%$ of unfished biomass. Neither saw sharks or elephant fish are targeted in the fishery (when using any method) and so the analyses were repeated except using a proxy target of $40 \% B_{0}$ which, given the control rule, will always increase the RBC if it is above zero.

Table 21.1. TIER 4 outcomes by species. The RBC in tonnes; this has not had discards, State catches, or recreational catches removed. The 2010 and 2011 values came from Haddon $(2010 ; 2011)$ and the 2009 values came from Rodriguez and McLoughlin (2009a,b).

| Species | RBC09 | RBC10 | RBC11 | RBC12 |
| :--- | :---: | :---: | :---: | :---: |
| SawSharks @ 48\% | 369.610 | 339.756 | 268.186 | $\mathbf{2 3 4 . 0 1 0}$ |
| Saw Sharks Trawl @ 48\% | Zones 20,60,50 |  |  | $\mathbf{5 1 3 . 6 1 2}$ |
| Saw Sharks Trawl @ 48\% | Zones 20,60,50,83,82 |  |  | $\mathbf{4 7 7 . 0 0 9}$ |
| Elephant Fish @ 48\% | 122.81 | 135.499 | 208.263 | $\mathbf{1 8 6 . 4 2 8}$ |
|  |  |  |  |  |
| SawSharks @ 40\% |  |  |  | $\mathbf{3 2 4 . 0 1 6}$ |
| Saw Sharks Trawl @ 40\% |  |  |  | $\mathbf{7 1 1 . 1 6 0}$ |
| Elephant Fish @ 40\% |  |  |  | $\mathbf{2 5 8 . 1 3 2}$ |

### 21.2 Introduction

The TIER 4 harvest control rule is the default procedure applied to species for which only limited information is available; specifically, if no reliable information is available relating to either current biomass levels or current exploitation rates. Ideally, in line with the notion of being more precautionary in the absence of information, the outcome from these analyses should be more conservative than those available from higher TIER analyses. In essence TIER 4 analyses require as a minimum, knowledge of the time series of total catches and of standardized catch rates.

Initially a control rule was implemented that was based around using any trend in recent catch rates to scale average recent catches. However, in 2008, an alternative was proposed that would not be prone to a declining ratchet effect on catches, and, in line with the Harvest Strategy Policy, could manage each fishery towards a target catch rate and away from a limit catch rate (Little, et al., 2008) The current TIER 4 analysis and control rule underwent Management Strategy Evaluation (Wayte, 2009; Little et al, 2011), which demonstrated its advantages over the original implementation.

The Tier 4 assessment requires the definition of a reference period for catches and catch rates which are to constitute the effective target for the fishery. This reference period is intended to act as a proxy for the fishery in a desirable state; ideally close to the stock size that leads to the maximum economic yield, and so in practice this target is also taken as a proxy for $B_{M E Y}$. In practice, in TIER 4 analyses, all that is really known about the reference period is that the RAG considers this period to be when the fishery was in a desirable state both biologically and economically. The Harvest Strategy Policy does not require that all species in a multispecies fishery aim to achieve the maximum economic yield, and this is especially the case with bycatch species. Nevertheless, the objective of avoiding the limit reference point remains. Within the current Tier 4 methodology the limit reference point is defined as $40 \%$ of the target catch rate. In addition, the Harvest Strategy Policy also states that:

Consideration should also be given to:

- Demonstrating that economic modelling and other advice clearly supports such action;
- No cost effective, alternative management options (e.g. gear modifications or spatial management) are available; and
- The associated ecosystem risks have been considered in full.
(DAFF, 2007, p 25)

If the average catch rate over the last four years drops below this limit the RBC is automatically zero.

### 21.3 Methods

### 21.3.1 TIER 4 Methods

The data required are time series of catches and catch rates. The analyses have been conducted on total catches across the entire SESSF (including State catches, SEF2 landing records, any discards, and any recreational catches (for elephant fish). Despite the fishery now operating from May through to April each year, the fishery data was collated in calendar years for consistency with the earlier fishery.

The fishery for both saw sharks and elephant fish was established before the catch rate standardization period selected by the RAG (i.e. significant catches were taken in the 1970s). Thus, although the Shark RAG did not consider the stocks of saw sharks and elephant fish to be seriously depleted by 1980, the stock was not pristine. In previous TIER 4 analyses (Rodriguez \& McLoughlin, 2009a, b) two reference periods were examined for saw sharks, 1986-2001 and 2002-2008, and two for elephant fish, 1980 - 1992 and 1998-2004. The earlier period had an extra source of uncertainty because the estimates of trawl bycatch and discards were likely under-estimated. To avoid these uncertainties and focus on a period when the total catches are known with most certainty the Shark RAG has selected 2002-2008 as the reference period for saw sharks and 1996-2007 for elephant fish.

All data to the end of 2010 relating to catches and discards, from both State waters and SEF2 data sets were provided by John Garvey of AFMA, with initial processing by Dr Neil Klaer and Mike Fuller of CSIRO. For saw sharks the species codes used in the landings database were SAW (Pristiophorus cirratus or Common Saw Shark), SHN (Pristiophorus nudipinnis or Southern Saw Shark), and SHW (Pristiophoridae or saw sharks). For elephant fish the species code in the landings database was SHE (Callorhinchus milii or Elephantfish). All catch rate data from the GHT fishery for both species were derived from the CANDE11.csv data files and analysed in Haddon (2012). All analyses of trawl caught fish used data straight from the AFMA Log Book database following pre-processing by Mike Fuller and Neil Klaer of CSIRO.

Standard analyses were set up in the statistical software, R, which provided the tables and graphs required for the TIER 4 analyses. The data and results for each analysis are presented for clarity. The TIER 4 harvest control rule formulation essentially uses a ratio of current catch rates with respect to the selected limit and target reference points to calculate a scaling factor. This scaling factor is applied to the target catch to generate an RBC:

$$
\begin{equation*}
\text { Scaling Factor }=S F=\max \left(0, \frac{\overline{C P U E}-C P U E_{\mathrm{lim}}}{C P U E_{t \mathrm{arg}}-C P U E_{\mathrm{lim}}}\right) \tag{1}
\end{equation*}
$$

$$
\begin{equation*}
R B C=C^{*} \times S F \tag{2}
\end{equation*}
$$

where
CPUE $_{\text {targ }}$ is the target CPUE for the species (half the average CPUE for the reference period). CPUE $_{\text {lim }}$ is the limit CPUE for the species; which is $40 \%$ CPUE $_{\text {targ }}$
$\overline{C P U E}$ the average CPUE over the past $m$ years
$C^{*} \quad$ is a catch target derived from a period of historical catch that has been identified as a desirable target in terms of CPUE, catches and status of the fishery (e.g. 1996 2007, as for Elephant fish). This is an average of the total removals for the selected reference period, including any discards.

$$
\begin{equation*}
C P U E_{\mathrm{targ}}=\frac{\sum_{y=y r 1}^{y r 2} C P U E_{y}}{(y r 2-y r 1+1)} \tag{3}
\end{equation*}
$$

where $C P U E_{y}$ is the catch rate in year $y, y r 2$ and $y r 1$ represent the last and the first years in the reference period respectively. The catch target is the mean of the total catch across the reference years.

$$
\begin{equation*}
C^{*}=\frac{\sum_{y=y r 1}^{y r 2} L_{y}}{(y r 2-y r 1+1)} \tag{4}
\end{equation*}
$$

where $L_{y}$ represents the total catch (landings plus discards) in year $y$.
Usually there are three rules used to select/estimate the CPUE/catch target:

1. The CPUE target for stocks fully exploited at or prior to 1986 is based on the average CPUE from 1986-1995.
2. Where fishing exploitation up to 1986 is thought to be minimal, the CPUE determined in step 1 is halved (to provide a catch rate proxy for $B_{M E Y}$ ).
3. Where fishing exploitation after 1986 is low, the first year in which catches are above 100t signifies the start of the 10 year period from which the target CPUE and catch targeted are calculated.

With bycatch shark species these rules are not always applicable (for example, with elephant fish the total catch rarely reaches 100 tonnes. Instead periods were chosen during which the fishery was considered to be well developed but in a good and relatively stable condition. For elephant fish the reference period chosen was 1996 - 2007 and for saw sharks the reference period chosen was 2002-2008.

Once the average CPUE for the reference period has been selected as the target CPUE (assumed a proxy for $B_{48 \%}$ which is assumed to be a proxy for $B_{M E Y}$ ) then the limit CPUE is defined as $40 \%$ of the that target. The maximum of the terms in the brackets, that is either zero or the ratio of CPUE values, is a scaling factor which is multiplied by the catch target $\left(\mathrm{C}^{*}\right)$ to determine the expected total catch. If the $\overline{C P U E}$ is less than the CPUE ${ }_{\text {lim }}$ this will automatically set the scaling factor to be negative, which means that the scaling factor will be set to zero and the consequent RBC will be zero.

To estimate the expected discards in the coming year a weighted average is used:

$$
\mathrm{D}_{\mathrm{CUR}}=\left(1.0 \mathrm{D}_{i-1}+0.5 \mathrm{D}_{i-2}+0.25 \mathrm{D}_{i-3}+0.125 \mathrm{D}_{i-4}\right) / 1.875
$$

where $D_{i}$ is the discards rate in year $i$, the discard rate in year $i$ is the ratio of discards to the sum of landed catches plus discards:

$$
D_{i}=\frac{\text { Discard }_{i}}{\left(\text { Catches }_{i}+\text { Discard }_{i}\right)}
$$

Plots are given of the total removals illustrating the target catch level. In addition, the standardized catch rates are illustrated with the target catch rate and the limit catch rate.

There are a number of meta-rules that are used when translating the RBCs into TACs. Two that relate to all species are:

- No TAC will change by more than $50 \%$ (either increase or decrease)
- Only changes greater than $10 \%$ (up or down) will be implemented.


### 21.3.2 Catches

The discard data for both saw sharks and elephant fish have been included in the most recent SESSF data summaries (Klaer and Upston, 2012) and this has led to some changes in the histories. Fortunately, the changes to the tier4 targets have barely changed as a result so this aspect of the change should have little effect. On the other hand the discard rate for elephant fish appears to have increased dramatically in 2011 from a base level of about 30 t up to about 132 t . This change calls into question the previous discard estimates. There have been no updates of information concerning State or recreational catches and these have been assumed to be equal to the last available estimates. This is unfortunate because there are anecdotal reports that recreational catches of elephant fish has been larger recently. Commonwealth landings were derived from the Quota landings database.

### 21.3.3 Treatment of Non-Target Species

In 2012, the SESSF RAG determined that the assessments of those species which do not constitute the economic drivers for a fishery might use the proxy for $\mathrm{B}_{\text {MSY }}$ as the target instead of $\mathrm{B}_{\text {MEY }}$. In practice this means that the target is assumed to be a proxy for $\mathrm{B}_{40}$ rather than $\mathrm{B}_{48}$. For the Tier 4, this means modifying the control rule used to estimate the RBC by multiplying the original target catch rate by $5 / 6$. If the original target was a proxy for $B_{48 \%}$ and $B_{M E Y}$, then $5 / 6^{\text {th }}$ or 0.83333 of this target would be a proxy for $B_{40 \%}$. The graphs illustrate this by a line below the original target. The limit reference point is not altered so as to maintain the same level of low risk to the stock.

The key requirement for non-key commercial species is to avoid allowing the stock to fall below the proxy for $B_{L I M}$ and there does not appear to be anything in the current Harvest Strategy Policy that recommends an alternative target. However, for the Tier 4 (and Tier 3) control rules to operate a specified target reference point is required. Clark (1993) used simulations to demonstrate that fishing at $F_{40 \%}$ instead of $F_{35 \%}$ didn't change the predicted yield by much but reduced the number of times the stock approached a limit of $B_{20 \%}$ (which Clark used as a threshold to indicate overfishing). Setting a target of $B_{40 \%}$ was thus considered to be a reasonable solution for setting a different target from $B_{48 \%}$ for use in the Tier 4 control rule, see Equ. 1, whilst retaining some robustness to falling below $B_{20 \%}$.

### 21.4 Results

### 21.4.1 Saw Sharks

Table 21.2. Saw Sharks. Data used in the Tier 4 analysis of saw sharks (full details of the available data are given in the Tables appendix (see Table 21.11). See the methods for a description of how the discards are calculated. The standardized catch rates (CE) are derived from Haddon (2012). The greyed cells reflect the reference period.

| Year | Catch | Discards | Total | CE |
| ---: | ---: | ---: | ---: | ---: |
| 1986 | 300.007 | 31.407 | 331.414 | 0.7981 |
| 1987 | 343.811 | 31.937 | 375.748 | 0.7692 |
| 1988 | 279.727 | 37.755 | 317.482 | 0.7489 |
| 1989 | 234.846 | 26.428 | 261.274 | 0.6761 |
| 1990 | 207.187 | 23.874 | 231.061 | 0.6530 |
| 1991 | 246.785 | 28.213 | 274.998 | 0.9637 |
| 1992 | 259.68 | 31.399 | 291.079 |  |
| 1993 | 340.195 | 40.162 | 380.357 |  |
| 1994 | 387.141 | 51.517 | 438.658 |  |
| 1995 | 447.775 | 47.723 | 495.498 |  |
| 1996 | 378.107 | 49.728 | 427.835 |  |
| 1997 | 296.93 | 38.773 | 335.703 | 1.0433 |
| 1998 | 278.413 | 39.659 | 318.072 | 1.3982 |
| 1999 | 223.661 | 34.922 | 258.583 | 1.4052 |
| 2000 | 195.973 | 32.211 | 228.184 | 1.2633 |
| 2001 | 264.441 | 30.699 | 295.140 | 1.1532 |
| 2002 | 315.372 | 30.592 | 345.964 | 1.0579 |
| 2003 | 367.676 | 32.486 | 400.162 | 1.1094 |
| 2004 | 376.150 | 32.981 | 409.131 | 1.1856 |
| 2005 | 353.910 | 31.671 | 385.581 | 1.1021 |
| 2006 | 373.515 | 30.656 | 404.171 | 1.2245 |
| 2007 | 269.940 | 41.977 | 311.917 | 0.9862 |
| 2008 | 273.382 | 42.512 | 315.894 | 1.0821 |
| 2009 | 259.743 | 40.392 | 300.135 | 0.9363 |
| 2010 | 245.475 | 38.173 | 283.648 | 0.7200 |
| 2011 | 251.408 | 39.095 | 290.503 | 0.7238 |

### 21.4.1.1 Proxy Target 48\% Gillnet

Table 21.3. Saw Sharks RBC calculations. C* and $\mathrm{CPUE}_{\text {targ }}$ relate to the period 2002-2008, CPUE $_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. The Wt_discards is the expected weight of discards. Implied proxy target $=48 \% B_{0}$.
$1^{\text {st }}$ Reference Year 2002
$2^{\text {nd }}$ Reference Year 2008
C*
367.546

CPUE $_{\text {targ }}$
1.107
$\mathrm{CPUE}_{\text {Lim }}$
0.4427
$\overline{C P U E}$
0.8656

Scaling Factor
0.6367

Wt_Discard
39.250

RBC
234.010



Figure 21.1 Saw Sharks. Top panel is the total removals with the fine line illustrating the target catch. Bottom panel represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 21.4.1.2 Proxy Target $40 \%$ Gillnet

Table 21.4. Saw Sharks RBC calculations. C* and CPUE $_{\text {targ }}$ relate to the period $2002-2008$, CPUE $_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. The Wt_discards is the expected weight of discards. Implied proxy target $=40 \% B_{0}$.

| $1^{\text {st }}$ Reference Year | 2002 |
| ---: | ---: |
| $2^{\text {nd }}$ Reference Year | 2008 |
| $\mathrm{C}^{*}$ | 367.546 |
| CPUE $_{\text {targ }}$ | 0.9220 |
| CPUE $_{\text {Lim }}$ | 0.4427 |
| $\overline{C P U E}$ | 0.8656 |
| Scaling Factor | 8816 |
| Wt_Discard | 39.250 |
|  |  |
| RBC | $\mathbf{3 2 4 . 0 1 6}$ |



Figure 21.2 Saw Sharks. Top panel is the total removals with the fine line illustrating the target catch. Bottom panel represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate. The fine purple line below the target CPUE target is the revised target based on a $40 \%$ $B_{0}$ proxy target for non-target species in a mixed fishery. The limit reference point (the red line) does not change.

### 21.4.1.3 Proxy Target 48\% Trawl SESSF Zones 20,60,50

Table 21.5. TRAWL: Saw Sharks RBC calculations. C* and CPUE $_{\text {targ }}$ relate to the period 2002-2008, CPUE $_{\text {Lim }}$ is $48 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. The Wt_discards is the expected weight of discards. Implied proxy target $=48 \% B_{0}$.
$1^{\text {st }}$ Reference Year 2002
$2^{\text {nd }}$ Reference Year 2008
C* 367.546
CPUE $_{\text {targ }} \quad 0.794$
$\begin{array}{cr}\text { CPUE }_{\text {Lim }} & 0.3177 \\ \overline{C P U E} & 0.9838\end{array}$
Scaling Factor 1.3974
Wt_Discard 39.250
RBC 513.612


Figure 21.3 Saw Sharks taken by Trawl in Zones 20, 60, and 50. Top panel is the total removals with the fine line illustrating the target catch. Bottom panel represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate. The fine purple line below the target CPUE target is the revised target based on a $48 \% B_{0}$ proxy target for non-target species in a mixed fishery. The limit reference point is represented by the red line.

### 21.4.1.4 Proxy Target 40\% Trawl SESSF Zones 20,60,50

Table 21.6. TRAWL: Saw Sharks RBC calculations. C* and CPUE $_{\text {targ }}$ relate to the period $2002-2008, \mathrm{CPUE}_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. The Wt discards is the expected weight of discards. Implied proxy target $=40 \% B_{0}$.

| $1^{\text {st }}$ Reference Year | 2002 |
| ---: | ---: |
| $2^{\text {nd }}$ Reference Year | 2008 |
| C $^{*}$ | 367.546 |
| CPUE $_{\text {targ }}$ | 0.662 |
| CPUE $_{\text {Lim }}$ | 0.3177 |
| $\overline{C P U E}$ | 0.9838 |
| Scaling Factor | 1.9349 |
| Wt_Discard | 39.250 |
|  |  |
| RBC | $\mathbf{7 1 1 . 1 6 0}$ |




Figure 21.4 Saw Sharks taken by Trawl in Zones 20, 60, and 50. Top panel is the total removals with the fine line illustrating the target catch. Bottom panel represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate. The fine purple line below the original target CPUE is the revised target based on a $40 \% B_{0}$ proxy target for non-target species in a mixed fishery. The limit reference point does not change.
21.4.1.5 Proxy Target 48\% Trawl SESSF Zones 20,60,50,83 \& 82

Table 21.7. TRAWL ( $20,60,50,83,82$ ): Saw Sharks RBC calculations. C* and CPUE $_{\text {targ }}$ relate to the period $2002-2008$, CPUE $_{\text {Lim }}$ is $48 \%$ of the target, and $\overline{C P U E}$ is average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. The Wt discards is the expected weight of discards. Implied proxy target $=$ $48 \% B_{0}$.

| $1^{\text {st }}$ Reference Year | 2002 |
| ---: | ---: |
| $2^{\text {nd }}$ Reference Year | 2008 |
| C* $^{*}$ | 367.546 |
| CPUE $_{\text {targ }}$ | 0.852 |
| CPUE $_{\text {Lim }}$ | 0.3408 |
| $\overline{C P U E}$ | 1.0043 |
| Scaling Factor | 1.2978 |
| Wt_Discard | 39.250 |

RBC
477.009


Figure 21.5 Saw Sharks taken by Trawl in Zones 20, 60, 50, 83 and 82 . Top panel is the total removals with the fine line illustrating the target catch. Bottom panel represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate.

### 21.4.2 Elephant Fish

Table 21.8. Elephant Fish. Data used in the Tier 4 analysis of saw sharks (full details of the available data are given in the Tables appendix (see Table 21.12). See the methods for a description of how the discards are calculated. The standardized catch rates (CE) are derived from Haddon (2012). The greyed cells relate to the reference period. Catch from 2002 onwards is the reported catches from the CDRs plus 29t of recreational fishing.

| Year | Catch | Discards | Total | CE |
| ---: | ---: | ---: | ---: | ---: |
| 1986 | 70.522 | 6.537 | 77.059 | 0.5961 |
| 1987 | 65.209 | 6.336 | 71.545 | 0.5178 |
| 1988 | 79.400 | 6.710 | 86.110 | 0.6909 |
| 1989 | 65.460 | 6.211 | 71.671 | 0.8412 |
| 1990 | 57.729 | 5.579 | 63.308 | 0.9046 |
| 1991 | 74.617 | 6.920 | 81.537 | 1.3399 |
| 1992 | 76.829 | 7.107 | 83.936 | 1.3325 |
| 1993 | 57.060 | 5.434 | 62.494 | 0.6588 |
| 1994 | 64.199 | 5.950 | 70.149 | 0.7727 |
| 1995 | 54.694 | 5.184 | 59.878 | 1.1228 |
| 1996 | 111.796 | 12.524 | 124.320 | 1.1090 |
| 1997 | 94.550 | 9.573 | 104.123 | 0.9266 |
| 1998 | 89.802 | 8.539 | 98.341 | 0.9030 |
| 1999 | 111.624 | 9.448 | 121.072 | 1.0194 |
| 2000 | 95.801 | 8.189 | 103.990 | 1.0827 |
| 2001 | 87.880 | 7.533 | 95.413 | 1.2929 |
| 2002 | 88.744 | 18.782 | 107.526 | 1.0046 |
| 2003 | 105.582 | 18.500 | 124.082 | 0.7764 |
| 2004 | 109.548 | 0.176 | 109.724 | 0.6278 |
| 2005 | 114.461 | 4.150 | 118.611 | 0.7377 |
| 2006 | 104.498 | 0.306 | 104.804 | 0.9117 |
| 2007 | 96.642 | 18.628 | 115.270 | 1.2929 |
| 2008 | 100.291 | 27.523 | 127.814 | 1.6919 |
| 2009 | 114.555 | 22.190 | 136.745 | 1.7726 |
| 2010 | 100.035 | 22.940 | 122.975 | 1.1630 |
| 2011 | 94.227 | 132.325 | 226.552 | 0.9103 |

### 21.4.2.1 Proxy Target 48\% Gillnet

Table 21.9. Elephant Fish RBC calculations. C* and CPUE $_{\text {targ }}$ relate to the period 1996 - 2007, $\mathrm{CPUE}_{\text {Lim }}$ is $48 \%$ of the original target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. The Wt_discards is the expected weight of discards. Implied proxy target $=48 \% B_{0}$.

| $1^{\text {st }}$ Reference Year | 1996 |
| ---: | ---: |
| $2^{\text {nd }}$ Reference Year | 2007 |
| C* $^{*}$ | 109.467 |
| CPUE $_{\text {targ }}$ | 0.974 |
| CPUE $_{\text {Lim }}$ | 0.3895 |
| $\overline{C P U E}$ | 1.3845 |
| Scaling Factor | 1.703 |
| Wt_Discard | 81.484 |
|  |  |
| RBC | $\mathbf{1 8 6 . 4 2 8}$ |



Figure 21.6 Elephant Fish. Top panel is the total removals with the fine line illustrating the target catch. Bottom panel represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates (1996 2007), and the recent average catch rate (last four years).

### 21.4.2.2 Proxy Target 40\% Gillnet

Table 21.10. Elephant Fish RBC calculations. C* and CPUE $_{\text {targ }}$ relate to the period 1996 - 2007, CPUE $_{\text {Lim }}$ is $40 \%$ of the target, and $\overline{C P U E}$ is the average catch rate over the last four years. The RBC calculation does not account for predicted discards of predicted State catches. The Wt_discards is the expected weight of discards. Implied proxy target $=40 \% B_{0}$.

| $1^{\text {st }}$ Reference Year | 1996 |
| ---: | ---: |
| $2^{\text {nd }}$ Reference Year | 2007 |
| $\mathrm{C}^{*}$ | 109.467 |
| CPUE $_{\text {targ }}$ | 0.811 |
| CPUE $_{\text {Lim }}$ | 0.3895 |
| $\overline{C P U E}$ | 1.3845 |
| Scaling Factor | 2.3581 |
| Wt_Discard | 81.484 |
|  |  |
| RBC | $\mathbf{2 5 8 . 1 3 2}$ |



Figure 21.7 Elephant Fish. Top panel is the total removals with the fine line illustrating the target catch. Bottom panel represents the standardized catch rates with the upper fine line representing the target catch rate and the lower line the limit catch rate. Thickened lines represents the reference period for catches, catch rates, and the recent average catch rate (1996-2007). The fine purple line below the target CPUE target is the revised target based on a $40 \% B_{0}$ proxy target for non-target species in a mixed fishery. The limit reference point does not change.

### 21.5 Discussion

The Recommended Biological Catches recommended by the original TIER 4 analyses recommend a slight drop for saw sharks and a slight increase for elephant fish. The RBCs usually have the State and recreational catches, and the discards removed as well as a $15 \%$ discount factor applied to estimate the Total Allowable Catch.

In the case of Saw sharks the reference years overlap the last four years used to generate an estimate of the recent catches, which will tend to keep the reference estimates and the current rate estimates similar. Fortunately, this now only includes 2008 in both series and any effects of this will decline as the years advance but it does mean there is an element of uncertainty about these estimates that doesn't normally enter into the calculations. This year, this problem no longer applies to elephant fish.

The capture of Elephant fish by recreational fishers is not insignificant but the estimates of catch are uncertain. In the analysis these have been held constant at 29 t since 1996. Braccini et al (2009) derive an estimated catch of Elephant fish of 13.931 t in 2008 inside Western Port (of which they estimated $70 \%$ were females). If this were included rather than the default 29 t it would not influence the Tier 4 calculation of the RBC but it might influence the removals taken from the RBC to form the TAC, although that would depend on whether such an adjustment to the total catches were made across the reference period as well as more recently. However, this may not represent all recreational catches of Elephant fish around Victoria and so the analysis retained the default value for recreational catches. Clearly a new estimate of total recreational catch would have value. It does suggest that the catch rate dynamics are likely being influenced by larger catches than believed, which in terms of the commercial fishery implies that the resulting RBC will be relatively conservative, as long as recreational catches are now stable, which is unknown.

As expected the use of a proxy CPUE target for $B_{40 \%}$ led to increases in the RBC in all cases. This is simply because the ratio of the target to the current average is bound to increase as long as the average is above the limit.

Not as expected, the standardized catch rates for trawl caught saw sharks behave differently to those from the gill net fishery, so much so that the analysis of trawl caught catch rates recommends an increase in the RBC (Table 20.1). Catches of saw sharks by trawl are now almost as high as those taken by gill net so this finding illustrates the uncertainty in this analysis, which provides some evidence that there may be an element of avoidance by gill net fishers. This avoidance would, in turn, lead to a reduction in gill net catch rates.

### 21.6 Tables

Table 21.11. Saw Sharks. Total catches and discards by fishery and Standardized catch rates, ready for the TIER 4 analysis (only the total catches and Standardized catch rates are used). Columns starting with Disc relate to discards. Only the Catch T and Std CE columns are used in the Tier 4 analysis, the first four columns derive from log-book data and under-estimate the landings and leave out the discards.

| Year | GHAT | SET | GAB | State | Discard | Catch T | Std CE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1976 | 248.65 |  |  |  |  | 263.569 |  |
| 1977 | 230.377 |  |  |  |  | 244.200 |  |
| 1978 | 269.2 |  |  |  |  | 285.352 |  |
| 1979 | 236.76 |  |  |  |  | 250.966 |  |
| 1980 | 227.969 |  |  |  |  | 241.647 | 1.6438 |
| 1981 | 193.592 |  |  |  |  | 205.208 | 1.0880 |
| 1982 | 244.047 |  |  |  |  | 258.690 | 0.9073 |
| 1983 | 234.673 |  |  |  |  | 248.753 | 1.0027 |
| 1984 | 230.465 |  |  |  |  | 244.293 | 1.0074 |
| 1985 | 262.913 | 4.11 |  |  | 3.075 | 285.873 | 1.0121 |
| 1986 | 280.529 | 19.478 |  |  | 14.575 | 331.414 | 0.7729 |
| 1987 | 327.365 | 16.431 | 0.015 |  | 12.295 | 375.748 | 0.7450 |
| 1988 | 248.708 | 30.514 | 0.505 |  | 22.833 | 317.482 | 0.7253 |
| 1989 | 212.59 | 18.608 | 3.983 |  | 13.673 | 261.274 | 0.6548 |
| 1990 | 180.123 | 17.598 | 9.601 |  | 13.067 | 231.061 | 0.6325 |
| 1991 | 211.606 | 23.931 | 14.442 |  | 15.517 | 274.998 | 0.9334 |
| 1992 | 209.242 | 25.541 | 25.265 |  | 18.844 | 291.079 |  |
| 1993 | 289.205 | 31.782 | 20.506 |  | 22.810 | 380.357 |  |
| 1994 | 327.406 | 43.078 | 17.149 |  | 31.873 | 438.658 |  |
| 1995 | 390.983 | 32.762 | 24.375 |  | 24.264 | 495.498 |  |
| 1996 | 310.827 | 37.963 | 29.537 |  | 31.078 | 427.835 |  |
| 1997 | 158.440 | 36.176 | 27.611 | 17.528 | 24.773 | 335.703 | 1.0105 |
| 1998 | 249.497 | 29.418 | 25.726 | 10.444 | 25.010 | 318.072 | 1.3542 |
| 1999 | 242.185 | 35.155 | 23.123 | 14.33 | 22.156 | 258.583 | 1.3609 |
| 2000 | 274.919 | 53.421 | 23.645 | 15.24 | 20.150 | 228.184 | 1.2236 |
| 2001 | 262.689 | 41.698 | 33.684 | 8.387 | 20.150 | 295.140 | 1.1168 |
| 2002 | 158.250 | 75.473 | 20.355 | 17.106 | 20.150 | 345.964 | 1.0245 |
| 2003 | 190.996 | 78.034 | 47.541 | 26.31 | 20.150 | 400.162 | 1.0745 |
| 2004 | 193.424 | 87.501 | 33.488 | 28.953 | 20.150 | 409.131 | 1.1482 |
| 2005 | 172.616 | 85.607 | 38.071 | 33.949 | 20.150 | 385.581 | 1.0674 |
| 2006 | 158.713 | 112.938 | 45.982 | 36.352 | 20.150 | 404.171 | 1.1859 |
| 2007 | 107.878 | 77.417 | 28.719 | 34.602 | 41.977 | 311.917 | 0.9551 |
| 2008 | 115.421 | 75.926 | 19.648 | 24.718 | 42.512 | 315.894 | 1.0480 |
| 2009 | 89.441 | 79.631 | 22.344 | 33.357 | 40.392 | 300.135 | 0.9068 |
| 2010 | 92.732 | 67.389 | 32.260 | 32.371 | 38.173 | 283.648 | 0.6973 |
| 2011 | 102.973 | 72.867 | 17.637 | 22.527 | 39.095 | 290.503 | 0.7010 |

Table 21.12. Elephant Fish. Total catches and discards by fishery and Standardized catch rates, ready for the TIER 4 analysis (only the total catches and Standardized catch rates are used). Columns starting with Disc relate to discards. Recr is the recreational catch.

| Year | GHAT | SET | GAB | State | Recr | DiscSSF | DiscS_G | CatchT | Std CE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1976 | 42.188 |  |  |  |  | 4.219 |  | 46.407 |  |
| 1977 | 68.334 |  |  |  |  | 6.833 |  | 75.167 |  |
| 1978 | 65.575 |  |  |  |  | 6.558 |  | 72.133 |  |
| 1979 | 100.581 |  |  |  |  | 10.058 |  | 110.639 |  |
| 1980 | 82.283 |  |  |  |  | 8.228 |  | 90.511 | 1.2954 |
| 1981 | 82.065 |  |  |  |  | 8.207 |  | 90.272 | 1.4169 |
| 1982 | 58.663 |  |  |  |  | 5.866 |  | 64.529 | 1.1052 |
| 1983 | 80.478 |  |  |  |  | 8.048 |  | 88.526 | 1.1028 |
| 1984 | 78.195 |  |  |  |  | 7.82 |  | 86.015 | 0.7679 |
| 1985 | 108.987 | 0.911 |  |  |  | 10.899 |  | 120.797 | 0.7279 |
| 1986 | 65.368 | 5.154 |  |  |  | 6.537 |  | 77.059 | 0.5961 |
| 1987 | 63.363 | 1.846 |  |  |  | 6.336 |  | 71.545 | 0.5178 |
| 1988 | 67.1 | 12.2 | 0.1 |  |  | 6.71 |  | 86.11 | 0.6909 |
| 1989 | 62.109 | 3.207 | 0.144 |  |  | 6.211 |  | 71.671 | 0.8412 |
| 1990 | 55.792 | 1.892 | 0.045 |  |  | 5.579 |  | 63.308 | 0.9046 |
| 1991 | 69.2 | 5.385 | 0.032 |  |  | 6.92 |  | 81.537 | 1.3399 |
| 1992 | 71.071 | 5.698 | 0.06 |  |  | 7.107 |  | 83.936 | 1.3325 |
| 1993 | 54.335 | 2.725 | 0 |  |  | 5.434 |  | 62.494 | 0.6588 |
| 1994 | 59.502 | 3.987 | 0.71 |  |  | 5.95 |  | 70.149 | 0.7727 |
| 1995 | 51.836 | 2.819 | 0.039 |  |  | 5.184 |  | 59.878 | 1.1228 |
| 1996 | 77.111 | 5.41 | 0.275 |  | 29 | 7.711 | 4.813 | 124.32 | 1.1090 |
| 1997 | 59.857 | 5.598 | 0.095 |  | 29 | 5.986 | 3.587 | 104.123 | 0.9266 |
| 1998 | 52.832 | 7.9 | 0.07 |  | 29 | 5.283 | 3.256 | 98.341 | 0.9030 |
| 1999 | 59.199 | 7.46 | 0.965 | 0.384 | 29 | 5.92 | 3.528 | 121.072 | 1.0194 |
| 2000 | 53.888 | 8.913 | 0 | 0.699 | 29 | 5.389 | 2.8 | 103.99 | 1.0827 |
| 2001 | 47.33 | 8.444 | 0.106 | 0.420 | 29 | 4.733 | 2.8 | 95.413 | 1.2929 |
| 2002 | 24.659 | 17.888 | 0.191 | 0.472 | 29 | 2.466 | 2.8 | 107.526 | 1.0046 |
| 2003 | 42.763 | 20.4088 | 2.032 | 0.439 | 29 | 4.879 | 2.8 | 124.082 | 0.7764 |
| 2004 | 29.088 | 27.2915 | 1.619 | 0.731 | 29 | 3.523 | 2.8 | 109.724 | 0.6278 |
| 2005 | 34.853 | 27.2535 | 1.878 | 0.663 | 29 | 4.052 | 2.8 | 118.611 | 0.7377 |
| 2006 | 36.061 | 17.865 | 1.426 | 3.933 | 29 | 4.014 | 2.8 | 104.804 | 0.9117 |
| 2007 | 36.206 | 14.093 | 1.701 | 11.954 | 29 | 21.845 | 2.8 | 115.270 | 1.2929 |
| 2008 | 40.471 | 19.297 | 0.834 | 2.092 | 29 | 23.023 | 2.8 | 127.814 | 1.6919 |
| 2009 | 44.136 | 20.2703 | 0.520 | 3.848 | 29 | 27.630 | 2.8 | 136.745 | 1.7726 |
| 2010 | 34.754 | 20.7817 | 0.310 | 3.553 | 29 | 22.941 | 2.8 | 122.975 | 1.1630 |
| 2011 | 33.906 | 15.7776 | 0.285 | 7.150 | 29 | 132.395 | 2.8 | 226.552 | 0.9103 |

### 21.7 Acknowledgements

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# 22. Projecting the School Shark Model into the Future: Rebuilding Timeframes and Auto-Longlining in South Australia ${ }^{1}$ 

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### 22.1 Summary

The current version of the school shark model predicts that catches of up to 250 t allow recovery of the stock, but $275 t$ will not.

Rebuilding to the limit reference point $\left(\mathrm{B}_{20}\right)$ cannot be achieved in a generation time plus time 10 years ( 32 years) given current levels of catch (176t). Rebuilding in three generation times ( 66 years) can be achieved with future catches of up to 225 t . If the limit reference point is moved from $\mathrm{B}_{20}$ to half $\mathrm{B}_{\mathrm{MSY}}$ (i.e. $\mathrm{B}_{25}$ ), then rebuilding within 32 years would require catches of close to zero; future catches would need to be of the order of 200 t in order to achieve rebuilding in 66 years.

Recovery times are only slightly lengthened by higher levels of auto-line fishing in South Australia (SA), however, this lowers $\mathrm{B}_{\text {MSY }}$ so the impact of an auto-line fishery would be felt when the school shark stock has recovered to levels where the overall catch can be increased to levels closer to $\mathrm{B}_{\mathrm{MSY}}$. If the auto-line fishery in SA is allowed to take a substantial portion of the catches in that state, the overall maximum sustainable catch for school shark will be lower than it would be if the auto-line fishery remained small relative to the gillnet fishery.

The results shown here are valid for a fishery whose seasonality and regional distribution are similar to that of the 2011 school shark fishery. Substantial (or perhaps even subtle) deviations from this pattern could alter these findings by altering the size and sex composition of the commercial school shark catch.

### 22.2 Background

The School Shark Recovery Plan (DEWHA, 2008) states that school shark recovery to a limit and a target reference point should occur within a biologically reasonable timeframe, and suggests that one generation time plus 10 years is such a timeframe. The Commonwealth Fisheries Harvest Strategy Policy (HSP) (DAFF 2007) also gives 3 generation times as a possible time frame to use for recovery. This policy states that the target reference point should be $\mathrm{B}_{\text {MSY }}$ and the limit reference point should be half $\mathrm{B}_{\text {MSY }}$. If $\mathrm{B}_{\text {MSY }}$ is unknown, $\mathrm{B}_{40}$ is used as a proxy, along with $\mathrm{B}_{20}$ for the limit. At the time of writing of the School Shark Recovery Plan, $\mathrm{B}_{\text {MSY }}$ for school shark had not been calculated so $\mathrm{B}_{40}$ was recommended. However, $\mathrm{B}_{\text {MSY }}$ has since been calculated to be

[^9]approximately $50 \%$ of pristine (Thomson and Punt, 2009), giving $\mathrm{B}_{50}$ as a target and $\mathrm{B}_{25}$ as a limit reference point.

The current rebuilding timeframe of one generation time plus 10 years is 32 years (Thomson and Punt, 2009). The ability of the stock to recover in this timeframe is explored here.

Management of the southern shark fishery has included restriction to fishing with smaller gillnet mesh sizes in order to protect large, more fecund, females. Figure 22.1 shows the impact on a pristine school shark stock of fishing heavily (taking 900t p.a. until 2011) with just one gear type. Line gear, and large mesh sizes, deplete the stock more rapidly and to lower levels than smaller mesh sizes. For this reason the planned introduction of an auto-line fishery for sharks in South Australia in response to exclusion of mesh gear from certain areas (to protect Australian sea lions and dolphins) needs to be carefully considered.

This paper explores recovery times for the school shark stock given a range of future fishing scenarios, both in terms of catch and gear composition.

### 22.3 Methods

The 2006 stock assessment update (Punt et al 2006) used an MSYR value of $3.5 \%$, which is essentially a measure of the productivity of the stock. Unlike previous school shark stock assessments, this parameter was fixed at $3.5 \%$ instead of being estimated by the model. The reason for this may have been that the estimated value was $3.5 \%$ and that fixing the parameter at its estimated value would have greatly speeded subsequent calculations. The MSYR parameter has remained fixed at a value of $3.5 \%$ in all subsequent calculations with the school shark model, possibly leading to lower perceived productivity than might otherwise have been the case. The parameter has been freed in the calculations shown in this paper, resulting in an estimated value of 4.4\%.

The most recent school shark assessment update used data to 2008 (Thomson and Punt, 2009). Catches taken between 2009 and 2011 were extracted from the GENLOG database (see a summary in Table 22.1). Catches were entered into the model by year, gear, month and shark region. The database contained only partial information for 2012 so the catches for 2011 were used again to represent those for 2012.

The 2009 base case model was used, except that the small SAV region which spans the South Australian - Victorian border which has traditionally been used in school shark stock assessments, but was excluded from the 2009 Base Case, was include here. The 2009 base case model dissolved that region into the two adjacent regions, but as this is not possible to do for all data sources, and has created technical problems, that change has not been made here. The MSYR parameter was freely estimated. This model:

- uses ISMP data for West only (not WSA, not East) (see Thomson and Punt 2009 for
explanation)
- uses commercial gillnet CPUE to 1996
- uses commercial gillnet CPUE for WBAS
- uses survey CPUE data
- uses recent CPUE series (1998-2008) separate to the CPUE up to 1996


### 22.3.1 Future auto longline fishery for South Australia

We assume that future catches are split among months, regions and gears in the same proportions seen in recent catches - with one exception: that a proportion of the catch currently taken by the gillnet fishery in South Australia (shark regions 1 WSA, 2 CSA and 3 SAV) is taken by a new auto-line fishery. This fishery has a knife-edged size at first selectivity of 536 mm (the smallest school shark caught in a recent auto-line survey, Figure 22.2). In the past a size threshold of approximately 630 mm for males and 640 mm for females (calculated from an age-based threshold of 2 years) has been assumed for a combined line and trawl fishery/fleet.

We consider a recovery scenario in which all future catches are 150 t, or 200 t or 250 t p.a., for a range of future gear configurations. Alternatively, to mimic the harvest strategy, we fixed catches at 200 t or 250 t p.a. for 50 years and after that allowed catches to increase by 2 t p.a.

### 22.4 Results and conclusions

### 22.4.1 Rebuilding times for status quo fishing

Catches of up to 250 t allow recovery of the school shark stock, but 275 t is not sustainable (Figure 22.3, Table 22.2a).

Rebuilding to the limit reference point $\left(\mathrm{B}_{20}\right)$ cannot be achieved in a generation time plus time 10 years ( 32 years) given current levels of catch (150t). Rebuilding in three generation times ( 66 years) can be achieved with future catches of up to 225 t . If the limit reference point is moved from $\mathrm{B}_{20}$ to half $\mathrm{B}_{\mathrm{MSY}}$ (i.e. $\mathrm{B}_{25}$ ), then rebuilding within 32 years would require catches of close to zero; future catches would need to be of the order of 200 t in order to achieve rebuilding in 66 years.

### 22.4.2 Split of future catches

Note that the future projections used here assume that the split of the catch between months, and regions is the same as that observed during 2011. If the split observed during 2008 is maintained instead (as was done for projections previously show to the sharkRAG), somewhat more pessimistic results are obtained (Table 22.2). This is due to interannual differences in the seasonality and spatial distribution of catches, highlighting that the results shown here pertain only to a fishery similar to that operating in 2011. Any substantial shifts in the timing or location of the bulk of the catches would require another investigation into the sustainability, and rebuilding timeframe of the new catch regime.

### 22.4.3 Future auto longline fishery for South Australia

For interest sake, we present the model prediction of the status of the future school shark population if the whole fishery moved to auto-line gear, as compared with using only 6 inch or 6.5 inch mesh gillnets in future.

As the size of any future auto-line fishery is unknown, we consider 3 possible futures ranging from no future auto-line fishery ( $0 \%$ ) to $100 \%$ auto-line fishing (in South Australia alone).

Compared with 6 inch gillnets, auto-line fishing takes both younger fish and the larger, more fecund females (Figure 22.2). It therefore impacts the stock more heavily than does fishing with 6 or 6.5 inch gillnets (Figure 22.1).

However, if catches remain at the present level (150t p.a.) the impact on the stock of an auto-line fishery in South Australia is forecast to be negligible (Figure 22.4, Table 22.3). However, more realistic scenarios in which larger catches are taken, particularly in the future, reveal that a large line fishery in South Australia would reduce the overall sustainable catch for the stock (Figure 22.5). This is because $\mathrm{B}_{\mathrm{MSY}}$ is lower for a line only fishery than for a 6.5 inch mesh, only, fishery.

Recovery times are only slightly lengthened by higher levels of auto-line fishing in South Australia (SA) (Table 22.3), however, this lowers B BSY so the impact of an autoline fishery would be felt when the school shark stock has recovered to levels where the overall catch can be increased to levels closer to $\mathrm{B}_{\text {MSY }}$. If the auto-line fishery in SA is allowed to take a substantial portion of the catches in that state, the overall maximum sustainable catch for the school shark fishery will be lower than it would be if the autoline fishery remained small relative to the gillnet fishery.

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### 22.6 Figures



Figure 22.1. Using a greatly simplified version of the school shark Tier 1 model, catches of 900 t p.a. were taken every year from 1927 to 2011 after which all fishing stopped. All of the catch was taking using a single gear type.

ISMP onbd


Figure 22.2. Length frequencies (grey bars) for commercially (gillnet) caught school shark in South Australia during summer, measured onboard by the ISMP Observer Program. The length frequency for the school shark caught during the summer auto-line survey in South Australia is shown (black line). The gear selectivity for 6 inch (red) and 6.5 inch (blue) mesh gillnets are shown.


Figure 22.3. Projected future depletion (pup production divided by pristine pup production) for the school shark stock for the Tier 12009 base case assessment model. Projections are shown for 9 future catch scenarios. Catches between 2008 (marked by a vertical line) and 2011 are the actual catches taken by the fishery.


Figure 22.4. Projected future depletion (pup production divided by pristine pup production) for the school shark stock for the Tier 12009 base case assessment model. Projections are shown for various future catch combinations. The lower plot shows results for just the years 2013 to 2020, in more detail.


Figure 22.5. Projected future depletion (pup production divided by pristine pup production) for the school shark stock for the Tier 12009 base case assessment model. Projections are shown for future catch scenarios in which 200t p.a. (ALL 200t) or 250 t (ALL 250t) are taken for 50 years from 2013 after which catches are increased by 2 t p.a. results are shown for scenarios in which an auto-line fleet in South Australia takes $0 \%, 50 \%$ or $100 \%$ of the catches previously taken by gillnet in that state.

### 22.7 Tables

Table 22.1. Catches of school shark between 2009 and 2011 by gear (taken from GENLOG database). Catches for 2012 are assumed to be the same as those for 2011.

| Year | Line and <br> Trawl | 6 inch | 6.5 inch | Total |
| :---: | :---: | :---: | :---: | :---: |
| 2009 | 19.1 | 140.6 | 85.6 | 245.4 |
| 2010 | 23.0 | 108.2 | 38.8 | 169.9 |
| 2011 | $7.8^{*}$ | 113.8 | 32.7 | 154.3 |

* this figure appears to be missing the trawl component of the catch, but this does not noticeably affect the conclusions of this paper.

Table 22.2. Number of years after 2008 when the school shark stock is predicted to achieve limit $\left(B_{20}\right.$, $\left.B_{25}\right)$ or target reference points $\left(B_{40}, B_{50}\right)$ under future catches ranging between 0 and 275 t. Future projections assume either that catches are distributed according to 2011 proportions, or 2008 proportions. A generation time plus 10 years is 32 years.

|  | 0t | 100 t | 125t | 150 t | 175t | 200t | 225t | 250 t | 275 t |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 Base Case - 2011 proportions |  |  |  |  |  |  |  |  |  |
| $\mathrm{B}_{20}$ | 23 | 30 | 32 | 36 | 40 | 47 | 58 | 80 | - |
| $\mathrm{B}_{25}$ | 30 | 38 | 42 | 46 | 51 | 59 | 71 | 95 | - |
| $\mathrm{B}_{40}$ | 45 | 57 | 62 | 67 | 74 | 83 | 97 | 124 | - |
| $\mathrm{B}_{50}$ | 50 | 62 | 67 | 73 | 80 | 89 | 104 | 132 | - |
| 2009 Base Case - 2008 proportions |  |  |  |  |  |  |  |  |  |
| $\mathrm{B}_{20}$ | 23 | 30 | 33 | 37 | 42 | 50 | 64 | 99 | - |
| $\mathrm{B}_{25}$ | 30 | 39 | 42 | 47 | 53 | 63 | 78 | 117 | - |
| $\mathrm{B}_{40}$ | 45 | 58 | 63 | 69 | 76 | 87 | 105 | 150 | - |
| $\mathrm{B}_{50}$ | 50 | 63 | 68 | 74 | 82 | 93 | 111 | 159 | - |

Table 22.3. Number of years after 2008 when the school shark stock is predicted to achieve limit $\left(B_{20}\right.$, $B_{25}$ ) or target reference points ( $B_{40}, B_{50}$ ) under different future gear combinations, taking a constant catch over all future years. Three scenarios in which a single gear is used across all regions are included for interest sake only. Three projections in which a proportion ( 0,50 or $100 \%$ ) of the gillnet catch in South Australia is transferred to a new auto-line fishery are shown (note that the $0 \%$ column resembles the 175 t and matches the 200 t or 250 t columns in Table 2). A generation time plus 10 years is 32 years.

|  | Global <br> line | Global <br> 6.5inch | Global 6 <br> inch | $0 \%$ ALL | $50 \%$ <br> ALL | 100\% ALL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 150t p.a. |  |  |  |  |  |  |
| $\mathrm{B}_{20}$ | 42 | 40 | 41 | 41 | 41 | 41 |
| $\mathrm{~B}_{25}$ | 53 | 51 | 52 | 52 | 52 | 52 |
| $\mathrm{~B}_{40}$ | 75 | 73 | 75 | 74 | 75 | 75 |
| $\mathrm{~B}_{50}$ | 81 | 79 | 81 | 80 | 80 | 81 |
| $200 t$ p.a. |  |  |  |  |  |  |
| $\mathrm{B}_{20}$ | 49 | 46 | 47 | 47 | 47 | 48 |
| $\mathrm{~B}_{25}$ | 61 | 58 | 60 | 59 | 60 | 60 |
| $\mathrm{~B}_{40}$ | 85 | 82 | 84 | 83 | 84 | 84 |
| $\mathrm{~B}_{50}$ | 91 | 88 | 90 | 89 | 90 | 90 |
| $250 t p . a$. |  |  |  |  |  |  |
| $\mathrm{B}_{20}$ | 87 | 75 | 83 | 80 | 82 | 83 |
| $\mathrm{~B}_{25}$ | 103 | 90 | 99 | 95 | 97 | 99 |
| $\mathrm{~B}_{40}$ | 130 | 117 | 129 | 124 | 126 | 128 |
| $\mathrm{~B}_{50}$ | 137 | 124 | 137 | 132 | 134 | 136 |

# 23. Incidental bycatch ratios of school shark (Galeorhinus galeus) to gummy shark (Mustelus antarcticus) off South Australia when using automatic longlines compared with gillnets 

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### 23.1 Summary

This chapter compares relative incidental bycatch ratios of school shark (Galeorhinus galeus) and gummy shark (Mustelus antarcticus) off South Australia using automatic longlines and gillnets. Data on catches of school and gummy sharks collected during scientific fishing trials using automatic longlines across South Australia were compared with reported catches using gillnets during the same period and broad area from the Commonwealth logbook database. We used a variety of methods for averaging and calculating the ratio of school shark to gummy shark, including or excluding zero catches and discards. Overall, these results provide strong evidence in favour of the conclusion that the bycatch of school shark is not greater when using automatic longlines as compared with gillnets. However, sample sizes from the automatic longline trials are relatively small, seasonal coverage is lacking (being confined to just summer months) and deliberate avoidance of school shark during the trial may have been greater than that practiced by gillnet fishers not participating in the trial. Consequently, we recommend proceeding with great caution should fishing using automatic longlines in South Australia commence. The ratio of unavoidable bycatch of school shark to gummy shark based on ordinary fishing operations (i.e. non-trials) should also be closely monitored.

### 23.2 Introduction

The gummy shark (Mustelus antarcticus) comprises the main species targeted using demersal gillnets in the Southern and Eastern Scalefish and Shark Fishery (SESSF). School shark (Galeorhinus galeus), common sawshark (Pristiophorus cirratus), southern sawshark (Pristiophorus nudipinnis) and elephant fish (Callorhinchidae and Rhinochimaeridae) are considered by-product species (Klaer et al. 2012; Woodhams and Vieira, 2012).

Concerns regarding incidental bycatch of TEP species (i.e. Australian sea lion) by gillnets, resulted in area closures in waters off South Australia (SA). This has prompted calls to allow the use of automatic longline gear in SA waters shallower than 183 m . There is, however, a concern that such a shift in gear type could result in high levels of unavoidable bycatch of school shark, which is classified as conservation-dependent under the EPBC Act, and is currently undergoing a rebuilding strategy (Woodhams et al. 2012). To address these concerns, trials using automatic longlines targeting gummy shark were conducted between November 2011 and January 2012 in SESSF waters off SA to determine the potential impact if fishers who currently employ demersal gillnets to target gummy shark shift to automatic longline gear (Knuckey et al. 2012).

This chapter compares incidental bycatch ratios of school shark to gummy shark when using automatic longlines compared to gillnets in three management zones in waters off SA.

### 23.3 Methods

## Trials and Commonwealth logbook data

Catch data, including retained and/or discarded catches, arising from four automatic longline trials targeting gummy shark in SESSF waters off SA between November 2011 and January 2012 (Knuckey et al. 2012) were used to estimate incidental bycatch ratios of school shark to gummy shark (SS-GS). These SS-GS ratios were also estimated using gillnet catch data extracted from the Commonwealth logbook database (GENLOG) corresponding to the same period. Relative ratios were then calculated as the quotient of these two ratios.

Two estimation methods were examined. The "Ratio [mean]" method estimates the ratio of the average school shark catch across all sets to that of gummy shark. By contrast, the "Mean[ratio]" method calculates the ratio of school shark to gummy shark in each set and then estimates the average across all sets. The latter tends to over-estimate the true ratio (Ye, 2002).

Relative SS-GS ratios were estimated for three management zones in waters off SA: western (WSA), central (CSA) and eastern SA (ESA) (Figure 23.1). Separate ratios were estimated using (i) retained catches and (ii) retained and discarded catches as well as including or excluding zero catches (Table 23.1). Variance estimates were not calculated since (i) the trials data contained small sample sizes and (ii) estimation methods were available to us.


Figure 23.1. Gummy shark management zones in waters off SA used in SS-GS ratio analyses.

### 23.4 Results

Relative SS-GS ratios of $<1$ suggest that school shark are less likely to be caught incidentally compared to gummy shark using automatic longlines (Table 23.1).

Relative ratios using the Ratio[mean] method ranged between $0.24-0.41$ based on retained and discarded catches (including zero catches) across the three zones (Table 23.1). These ratios ranged between $0.15-1.24$ when zeros were excluded from the analyses.

Relative ratios using the Mean[ratio] method were $>1$ in four instances, two when excluding zero observations in the CSA and ESA, based on both retained and discarded catches (Table 23.1). The ratio of 11.355 was based on one automatic longline set where 13 kg of school shark were retained and only 0.2 kg of gummy shark were discarded.

Relative incidental bycatch ratios were similar using the Ratio[mean] method when either retained or both retained and discarded catches were combined in analyses, and also similar when zero catches were either included or excluded from analyses (Table 23.1). However, care should be taken when interpreting these results as the automatic longline trials employed in the analyses were based on small sample sizes when aggregated by management zone.

Table 23.1. Incidental relative school shark to gummy shark ratio (SS-GS) using automatic longlines compared to gillnets $\times$ management zone (WSA, CSA and ESA) $\times$ Data type $\times$ Method $\times$ use of retained catch and/or discards. Data type: Include refers to the inclusion of zero catches; exclude: exclude zero catches. Retained: retained catches included only; Retained + discarded: retained and discarded catches included.

|  |  | Retained |  | Retained + discarded |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Data <br> type | Zone | Ratio [mean] | Mean [ratio] | Ratio [mean] | Mean [ratio] |
| include | WSA | 0.063 | 0.044 | $\mathbf{0 . 2 4 4}$ | 0.326 |
|  | CSA | 0.383 | 0.192 | $\mathbf{0 . 4 0 8}$ | $9.449^{\mathrm{b}}$ |
|  | ESA | 0.291 | 0.501 | $\mathbf{0 . 3 8 2}$ | 0.479 |
| exclude | WSA | 0.070 | 0.060 | 0.146 | 0.195 |
|  | CSA | 0.497 | 0.281 | 0.486 | $11.355^{\mathrm{b}}$ |
|  | ESA | $1.176^{\mathrm{a}}$ | $1.663^{\mathrm{a}}$ | $1.237^{\mathrm{c}}$ | $1.549^{\mathrm{c}}$ |

Note: Overall relative ratio exceeds 1 in a-c above.
${ }^{\text {a }}$ : Ratios for school and gummy sharks are low and similar for auto-longline and logbook-gillnets respectively. Overall ratio exceeds 1.
${ }^{\mathrm{b}}$ : One set (auto-longline shots) dominates this ratio (school shark: retained ( 13 kg ) and gummy shark: discarded ( 0.2 kg ) ).
${ }^{\text {c }}$ : Ratios for school and gummy sharks are low and similar for auto-longline and logbook-gillnets.
Zone: west South Australia (WSA), central South Australia (CSA) and east South Australia (ESA).

### 23.5 Conclusions

Overall, these results suggest that the incidental catch of school shark when targeting gummy shark is not higher using automatic longlines gear than gillnets. However, fishing trials were conducted during summer (November to January) only and it has been suggested that large female school shark are abundant off King Island in winter (SharkRAG 2012). Trials were conducted during winter 2011 but unfortunately these were confined to a small area off the Coorong region (Knuckey et al. 2012).

In addition, sample sizes were relatively small and gillnet fishing wasn't conducted alongside ALL operations. Great care was taken during this trial to avoid areas of known school shark abundance (Ian Knuckey pers. comm.) but these results had to be compared with general gillnet fishing recorded in the Commonwealth logbook database. This may be an unfair comparison, hence biasing these results towards lower apparent school shark bycatch using ALL.

Given the relative paucity of data on which these results are based, the lack of seasonal coverage, and the possibility of a (perhaps grossly) biased comparison, we recommend proceeding with great caution if longline fishing in South Australia is to be commenced. The ratio of school to gummy shark in non-trial catches should be closely monitored.

### 23.6 Acknowledgements

Matt Koopman (Fishwell Consulting) is thanked for providing data corresponding to the four automatic longline trials conducted between November 2011 and January 2012, and discussions during data error checks. Thanks also go to Russell Hudson (Fishwell consulting) for his useful insights into vessel operations during the trials and Ian Knuckey (Fishwell consulting) for useful discussions. Both Neil Klaer (CSIRO) and Mike Fuller (CSIRO) are thanked for providing the Commonwealth logbook data used in the analyses. Members of SharkRAG are also thanked for useful discussions.

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## 24. Predicted pup production of Gummy Shark (Mustelus antartictcus) assuming an automatic longline fishery off South Australia

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### 24.1 Summary

This chapter estimates relative changes in the pup production of gummy shark (Mustelus antarcticus) in South Australia under different levels of fishing effort using automatic longline (ALL) gear. Results indicate that to achieve the same equilibrium level of future pup production based on a 300 t future annual catch and using an existing combination of 6 and 6.5 inch gillnets and small line gears, catches would need to drop to (i) 280-285 t if half of the catch currently taken by gillnets shifts to ALL gear, and to (ii) 260-270 $t$ if all of the catch is obtained by ALL gear. If behavioral factors which cause larger, older gummy shark to be unavailable to gillnet gear do not apply to ALL gear, to achieve the same equilibrium levels of pup production, a 300 t catch would need to be reduced to $250-280 \mathrm{t}$ or 210-250 t based on a $50 \%$ or $100 \%$ transfer of catch from gillnet to ALL gear respectively. Higher future catches widen the gap between sustainable catch levels for gillnet versus mixed gillnet and ALL sectors. Although the gummy shark stock in SA would be able to sustain an introduction of a future ALL fishery, sustainable catch levels would be lower than they would be in the absence of a large line fishery. These calculations were based on an update of the gummy shark stock assessment in 2010 (which employed data to end of 2009), and assumed that all (available) fish greater than 76 cm were caught by ALL gear. These results are based on an introduction of an ALL sector in SA only, despite potential for related changes to the fishery in Victoria.

### 24.2 Introduction

Gummy shark (Mustelus antarcticus) are mostly caught using demersal gillnets and comprise the main species targeted in the Southern and Eastern Scalefish and Shark Fishery (SESSF). Gummy shark are also caught using otter trawls, Danish seines and hooks, although captures are significantly lower compared to gillnets (Klaer et al. 2012). However, due to existing area closures and potential incidental captures of TEP species by gillnets in waters off South Australia (SA), calls have been made to legalize the use of automatic longline (ALL) gear in waters shallower than 183 m . In addition, a relatively narrow size-range of sub-adult gummy shark are caught using gillnets, whereas a much wider size range (including adults), are likely to be vulnerable to capture using ALL gear.

Trials using ALLs to target gummy shark were conducted between November 2011 and January 2012 within the SESSF in South Australia to determine the potential impact on fishers who currently employ demersal gillnet gear (Knuckey et al. 2012).

Using information gained from the ALL trials, this chapter examines the likely impact of a change in selectivity on (i) future equilibrium pup production of gummy shark and (ii) sustainable yields in SA. The existing stock assessment methodology for gummy shark approved by sharkRAG is based on three reproductively separate populations: SA, Bass Strait (Victoria), and Tasmania. We used the existing gummy shark stock assessment developed for SA (Punt and Thomson 2011) and employed three model configurations to simulate future pup production assuming (i) different levels of harvesting based on ALL and gillnet gear and (ii) whether or not large sharks are available to ALL gear.

We also examine the future annual catch required to provide the same equilibrium pup production obtained under current harvesting, when half or all of the catch is obtained from ALL gear.

### 24.3 Methods

The most recent gummy shark stock assessment conducted in 2010 (using data to end of 2009) does not employ one base case model. Instead it employs six alternative model configurations assuming the same weight of belief (see Punt and Thomson 2011). Although, a stock assessment update was not performed, recorded Commonwealth logbook catches for 2010-11 inclusive, obtained from the AFMA logbook database were included in the analyses (Table 24.1). Each of the six model configurations are detailed in Table 24.2.

Table 24.1. Yearly Commonwealth logbook catches ( $t$ ) for the period 2009-2011 and gear type for South Australia (shark regions WSA, CSA and ESA). Note that "Line" indicates gummy shark catches recorded as caught by lines, but precedes the start of ALL fishing.

| Year | Gear |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Line | 6 " mesh | 6.5 " mesh | $7{ }^{\prime \prime}$ mesh | 8 " mesh |
| 2009 | 43.6 | 74.68 | 319.38 | 0.01 | 0 |
| 2010 | 39.57 | 97.03 | 286.17 | 0 | 0 |
| 2011 | 60.01 | 76.78 | 152.62 | 0 | 0 |

Initial analyses based on six different model configurations (B, D, F, G, I, J; Table 24.2) and reported as the final set of models assessed in 2010 (Punt and Thomson 2011) were projected to 2112 . Model configurations $\mathrm{F}, \mathrm{G}$ and J were excluded from further analyses since they produced unstable oscillations in projections (Table 24.2). Model configurations (B, D, I) were projected for 100 years (to 2112) to ensure that equilibrium conditions were achieved.

Table 24.2. Model configurations used in 2010 gummy shark stock assessment. Model configurations B, D and I (bold) were used to obtain projections due to the instability of the other configurations ( $\mathrm{F}, \mathrm{G}$ and J). $\mathrm{M}=$ natural mortality; $\mathrm{DD}=$ density dependence.

| Model <br> configuration | Description |
| :---: | :--- |
| $\mathbf{B}^{\wedge}$ <br> (reference case) | - DD is a function of total (1+) biomass <br> - DD impacts rate of M for animals 0-30 years <br> -gear competition modeled on Equation 1a (Punt and Thomson <br> 2010) |
| D | Model B; DD on M for ages 0-15 (B1+) |
| F | Model B; DD on M for ages 0-30 (B_mat) |
| G | Model B; DD on M for ages 0-15 (B_mat) |
| I | Model B; DD on M for ages 0-2 (B1+) |
| J | Model B; DD on M for ages 0-2 (B_mat) |

^ closest model configuration to that used in July 2010 preliminary gummy shark assessment from Thomson and Punt (2010).

The mean proportion of total annual catch from the most recent year (2011) was estimated for each gear type. Future proportions used in the scenarios through the transfer of catches from gillnet to ALL gear were based on these estimated proportions. Given that the TACs have been greatly reduced in recent years, it was assumed that the most recent catches (and their proportional catch by gear type) are more representative of the fishery than those in the past. Transfers of $0 \%, 50 \%$ or $100 \%$ of gummy shark catches from gillnet to ALL gear types off SA were considered. Note that a $0 \%$ transfer does not correspond to $0 t$ catch corresponding to line gear since line catches are nonzero. Instead, a $0 \%$ transfer corresponds to no additional transfer of catch resulting from gillnets to ALLs (Table 24.3). Also, a $50 \%$ transfer corresponds to $50 \%$ of the catch obtained from gillnets is transferred to ALLs in same relative proportions among the four mesh sizes (Table 24.3).

Table 24.3. Line and gillnet gear proportions employed in analyses.

|  | Gear |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Transfer to ALL <br> $(\%)$ | Line | 6 " mesh | 6.5 " mesh | 7 " mesh | 8 " mesh |
| 0 | 0.207 | 0.265 | 0.528 | 0 | 0 |
| 50 | 0.604 | 0.132 | 0.264 | 0 | 0 |

Two future catch levels were also considered i.e., based on the total catch across all gear types in $2011(\sim 300 \mathrm{t})$ and higher future catches ( $\sim 450 \mathrm{t}$ ), which corresponds to the highest catch recorded over the 1927-2011 period. Higher future catches are plausible because the gummy shark fishery is currently constrained by the need to keep the bycatch of school shark (Galeorhinus galeus) at low levels. As school shark populations recover, higher gummy shark catches may be permitted.

### 24.4 Results and discussion

Examination of the size range of gummy sharks caught and retained using ALL gear (Figure 24.1, see also Knucky et al. 2012) supports the assumption made in previous gummy shark stock assessments (e.g. Punt and Thomson, 2011) i.e., that sharks greater than 76 cm are caught and that there is no upper size limit. However, some smaller sharks are caught and discarded using ALL gear (Figure 24.1). While, existing gummy shark stock assessment models do not include discarding, future calculations should consider the effects of discard mortality. Since Commonwealth logbook catches based on gillnets were used to calculate gummy shark to school shark catch ratios (Sporcic and Thomson 2013), the 76 cm minimum size limit was employed here.


Figure 24.1. Length distribution of gummy sharks retained (dark green bars) and discarded (red bars) during ALL trials (left axis). Measurements are total length (cm) binned in 1 cm length intervals. Right axis: Selectivity of (i) 6.5 inch mesh (solid blue line) and (ii) line gear (solid black line). Model estimated availability from Models B, D and I (solid red, green and orange lines respectively).

In addition to gear selectivity, the stock assessment models of Punt and Thomson (2011) assume that larger sharks are unavailable to the fishery based on an estimated availability function. It is unknown whether or not this would apply to ALL gear. However, opinion was expressed during presentation of this work to sharkRAG members that it would not apply to ALL gear (sharkRAG, November 2012).

### 24.4.1 Gear selectivity; availability applied to line

Predicted equilibrium pup production (relative to pristine; 1927, hereafter referred to as "pup production") ranged from $42 \%$ to $80 \%$ (Model B, D, I;Table 24.4). This was based on the assumption that large gummy sharks are unavailable to the fishery. A predicted $42 \%$ pup production estimate was obtained when all catches in SA were taken using

ALLs and future annual landed catches of 450 t were maintained (Model I; Table 24.4). Minimal changes in pup production occurred if future annual catches of 300 t were maintained across the three percentage transfers to ALLs ( $0-100 \%$ ) under three model configurations (Model B, D, I). If the total annual future catch increased to 450 t , pup production ranged from $42 \%$ to $67 \%$ (Model B, D, I;Table 24.4). A $62 \%$ pup production estimate resulted in SA if all catches were obtained from ALLs (Model B), $58 \%$ (Model D) and $42 \%$ (Model I), respectively. Estimated annual pup production is based on model runs to 2112, but equilibrium levels were obtained by approximately 2030.

If $50 \%$ of the catch transferred to ALLs in SA, and future annual catches were 300 t , i.e., similar to current levels, future equilibrium pup production ranged from $64 \%$ to $78 \%$ (Model B, D, I; Table 24.4). Minimal changes to these production estimates occurred if the entire SA fishery shifted to using ALLs under the same future catch scenario (Table 24.4).

Table 24.4. Large sharks unavailable to automatic longline (ALL) gear. Equilibrium pup production levels for model configurations B, D and I, under three transfer (\%) to ALL gear scenarios and two annual future catch scenarios ( $300 \mathrm{t}, 450 \mathrm{t}$ ). $\mathrm{P}=$ pup production.

| Model | Transfer to ALL <br> $(\%)$ | Pup production <br> $\left(\mathrm{P}_{2112} / \mathrm{P}_{1927}\right)$ | Pup production <br> $\left(\mathrm{P}_{2112} / \mathrm{P}_{1927}\right)$ |
| :---: | :---: | :---: | :---: |
|  | $(300 \mathrm{t})$ | $(450 \mathrm{t})$ |  |
|  | 0 | 0.80 | 0.67 |
|  | 50 | 0.78 | 0.64 |
|  | 100 | 0.77 | 0.62 |
| D | 0 | 0.76 | 0.61 |
|  | 50 | 0.75 | 0.60 |
|  | 100 | 0.74 | 0.58 |
|  | 0 | 0.67 | 0.48 |
|  | 50 | 0.64 | 0.45 |
|  | 100 | 0.63 | 0.42 |

If $50 \%$ of the SA catch transferred to the ALLs, then future annual catches of 280-285 t would yield the same future equilibrium pup production that would be obtained if the fishery continued taking 300 t annually with no transfer from gillnets to ALLs (Table 24.5). If $100 \%$ of the catch transferred to ALLs, the equivalent future annual catches ranged between 260 t and 270 t . These results assumed that an availability function applied to ALL gear (Table 24.5).

Table 24.5. Large sharks unavailable to automatic longline (ALL) gear. Changes in future catch levels required to provide pup production at the same equilibrium levels obtained under current harvesting when transferring $50 \%$ and $100 \%$ of the total annual catch to ALLs.

| Model | Transfer to ALL <br> $(\%)$ | Current catch <br> $(\mathrm{t})$ | Equivalent catch <br> $(\mathrm{t})$ |
| :---: | :---: | :---: | :---: |
| B | 50 |  | 280 |
| D |  | 300 | 285 |
| I |  |  | 285 |
| B | 100 | 300 | 260 |
| D |  |  | 270 |
| I |  | 270 |  |

### 24.4.2 Gear selectivity; availability not applied to line

Future equilibrium pup production ranged from $34 \%$ to $78 \%$ (Table 24.6). The lowest value was based on the scenario where $100 \%$ of the SA gummy shark fishery converted its catch to ALLs, large gummy shark were available to the fishery and annual future catches were high (Model I; 450 t). The only other pup production estimate below $40 \%$ occurred when half of the SA fishery converted its catch to ALLs and future catches were high (Model I; 450 t ). Pup production estimates below $40 \%$, occurred when both future annual catches were high ( 450 t ) and at least $50 \%$ of the total annual catch was obtained by ALL gear. If the total annual catch remained at approximately 300 t , pup production was estimated to be above $40 \%$ across the three percentage transfers to ALLs (Models B, D and I; Table 24.6).

Table 24.6. Large sharks available to automatic longline (ALL) gear. Equilibrium pup production levels for model configurations $\mathrm{B}, \mathrm{D}$ and I , under three transfer (\%) to ALL gear scenarios and two future annual catch scenarios ( $300 \mathrm{t}, 450 \mathrm{t}$ ). $\mathrm{P}=$ pup production.

| Model | Transfer to ALL <br> $(\%)$ | Pup production <br> $\left(\mathrm{P}_{2112} / \mathrm{P}_{1927}\right)$ | Pup production <br> $\left(\mathrm{P}_{2112} / \mathrm{P}_{1927}\right)$ |
| :---: | :---: | :---: | :---: |
|  | 0 | $(300 \mathrm{t})$ | $(450 \mathrm{t})$ |
|  | 0.78 | 0.60 |  |
| B | 50 | 0.72 | 0.53 |
|  | 100 | 0.68 | 0.47 |
|  | 0 | 0.74 | 0.50 |
| D | 50 | 0.70 | 0.45 |
|  | 100 | 0.66 | 0.41 |
|  | 0 | 0.63 | 0.45 |
| I | 50 | 0.58 | 0.39 |
|  | 100 | 0.54 | 0.34 |

If $50 \%$ of the SA catch was transferred to the ALL sector, then future annual catches of $250-280 \mathrm{t}$ would give the same future equilibrium pup production that would be obtained if the fishery continued taking 300 t annually with no transfer from gillnets to ALLs (Model B, D, I; Table 24.7). If $100 \%$ of the catch of gummy shark by the SA fishery transferred to ALLs, equivalent future annual catches ranged between 210 t and 250 t . These calculations assumed that an availability function did not apply to ALL gear (Table 24.7).

Table 24.7. Large sharks available to automatic longline (ALL) gear. Changes in future catch levels required to provide pup production at the same equilibrium levels obtained under current harvesting when transferring 50 and $100 \%$ of the total annual catch to ALL gear.

| Model | Transfer to ALL <br> $(\%)$ | Current catch <br> $(\mathrm{t})$ | Equivalent catch <br> $(\mathrm{t})$ |
| :---: | :---: | :---: | :---: |
| B |  |  | 250 |
| D | 50 | 300 | 270 |
| I |  |  | 280 |
| B | 100 | 300 | 210 |
| D |  |  | 220 |
| I |  | 250 |  |

### 24.5 Conclusions

As expected, information gained from the automatic longline trials indicate that larger (and older) gummy sharks are caught by ALLs than gillnets. Consequently, harvesting employing ALLs would require lower TACs compared to employing gillnets alone in order to achieve the same level of sustainability in terms of future pup production (Tables 24.4 to 24.7). If future annual catches remain at current levels ( $\sim 300 \mathrm{t}$ in SA), and if larger sharks are unavailable to the fishery (e.g. due to behavioral reasons), the impact of an ALL fishery can be predicted to be relatively low: future pup production drops from 67-80\% (relative to pristine) to $63-78 \%$ depending on (i) catch levels using ALL gear and (ii) model configuration employed. To achieve the same level of pup production in the absence of an ALL sector, future catches are required to drop from 300 t to 280-285 t if $50 \%$ of former gillnet catch is obtained by ALLs, and 260-270 t if $100 \%$ of former gillnet catch is obtained by ALLs.

The presence of an ALL sector has a greater impact if future annual catches are higher. Catches of 450 t decrease pup production from $48-67 \%$ to $42-64 \%$ following the introduction of an ALL fishery in SA, depending on the percent transfer of catches from gillnets to ALLs and the model configuration used. This assumes that larger sharks are unavailable to the fishery.

If large sharks are available to ALL gear, as suggested by sharkRAG, the effect of an ALL sector is still greater. Future pup production drops from $63-78 \%$ to $54-72 \%$ under a 300 t future catch scenario, and from $45-60 \%$ to $34-53 \%$ under a 450 t catch scenario. Future catches would have to drop from 300 t to $250-280 \mathrm{t}$ or $210-250 \mathrm{t}$ given $50 \%$ or $100 \%$ transfer of catch from gillnet to ALL gear respectively.

This work considers the introduction of an ALL sector in SA alone. Sustainable catch levels in Victoria could also be affected if ALL quotas are issued across the SA-Vic border and between different gears.

The gummy shark population in SA waters can sustain the introduction of a future ALL fishery, though sustainable catch levels will be lower than they would be in the absence of a large ALL fishery. Future work should consider the effect of capture and possible mortality of gummy shark smaller than 76 cm .

### 24.6 Acknowledgements

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### 24.7 References

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Thomson, R., Punt, A.E. (2010). Gummy shark assessment update for July 2010 SharkRAG meeting, using data to the end of 2008. SharkRAG document 2010/03.

## 25. Benefits

The results of this project have had a direct bearing on the management of the Southern and Eastern Scalefish and Shark Fishery. Direct benefits to the commercial fishing industry in the SESSF have arisen from improvements to, or the development of, assessments under the various Tier Rules of the Commonwealth Harvest Strategy Policy for selected quota and non-quota species. Information from the stock assessments has fed directly into the TAC setting process for SESSF quota species. As specific and agreed harvest strategies are being developed for SESSF species (a process required by and agreed to under EPBC approval for the fishery), improvements in the assessments developed under this project have had direct and immediate impacts on quota levels or other fishery management measures (in the case of non-quota species).

Participation by the project's staff on the SESSF Resource Assessment Groups has enabled the production of critical assessment reports and clear communication of the reports' results to a wide audience (including managers, industry). Project staff's scientific advice on quantitative and qualitative matters is also clearly valued.

The stock assessments presented in this report have provided managers and industry greater confidence when making key commercial and sustainability decisions for species in the SESSF. These assessments have provided the most up-to-date information, in terms of data and methods, to facilitate the management of the Southern and Eastern Scalefish and Shark Fishery.

## 26. Conclusion

- Provide quantitative and qualitative species assessments in support of the five SESSFRAG assessment groups, including RBC calculations within the SESSF harvest strategy framework.

The 2012 assessment of the stock status of key Southern and Eastern Scalefish and Shark fishery species is based on the methods presented in this report. Documented are the latest quantitative assessments (Tier 1) for several of the key quota species (pink ling (east and west), silver warehou, and deepwater flathead), as well as catch curve analyses and cpue standardisations for shelf, slope, deepwater and shark species. Typical assessment outputs provided indications of current stock status and an application of the Commonwealth Harvest Strategy framework. This framework is based on a set of assessment methods and harvest control rules, with the decision to apply a particular combination dependent on the type and quality of information available to determine stock status (Tiers 1 to 4).

The assessment outputs from this project are a critical component of the management and TAC setting process for these fisheries. The results from these studies are being used by SESSFRAG, industry and management to help manage the fishery in accordance with agreed sustainability objectives.

## Stock status and Recommended Biological Catch (RBC) conclusions:

The 2012 base-case (aggregated zones model) Tier 1 assessment of pink ling (Genypterus blacodes) concluded that the eastern stock is $0.26 \mathrm{~B}_{0}$ at the start of 2013 and the western stock is $0.43 \mathrm{~B}_{0}$ at this time (under the assumption that the TAC for 2012 of $1,000 \mathrm{t}$ is taken). The Recommended Biological Catches (RBCs) arising from the base-case models are 223 t for the eastern stock and 490 t for the western stock; giving a total RBC of 713 t for the SESSF pink ling stocks. The long term RBC (for the year 2032) is 829 tonnes for the eastern stock and 548 tonnes for the western stock; giving a total long-term RBC of $1,377 \mathrm{t}$. An alternative model was considered in addition to the base-case as a next step towards the development of a model to account for lack of spatial homogeneity in population processes within the eastern and western stocks of pink ling. This alternative model treats the zone-based CPUE indices and the age- and length-compositions by zone as coming from different 'fleets'. Further work on the zone-based model is expected over the coming years.

A quantitative Tier 1 assessment of silver warehou (Seriolella punctata) in the SESSF was conducted using data up to 31 December 2011. The last full quantitative assessment was presented in 2009. The base-case assessment estimates that the projected 2013 spawning stock biomass will be $46.6 \%\left(0.466 \mathrm{~B}_{0}\right)$ of virgin stock biomass. The RBC from the base-case model for 2013 is 2,544t for the 20:35:48 harvest control rule, with a long-term yield of $2,618 \mathrm{t}$. If recent recruitments (2008-2011), which are not currently estimated by the model, are assumed to be poor and at similar levels to recruitment during the period 2002-2005, then depletion in 2013 could fall below $40 \%$. Under this scenario, setting a multi-year TAC could result in depletion levels falling below $30 \%$ by 2015.

While a full quantitative assessment of jackass morwong (Nemadactylus macropterus) was not conducted in 2012, to calculate the 2013 RBC, the 2011 assessments for both eastern and western morwong were projected for one more year, using actual catches from 2011, and estimated catches for 2012. No other data were added and no new parameter estimation was performed. The 'recruitment shift' assessment model accepted as the base-case for the eastern stock in 2011, and the base-case model for the western stock from 2011 were used for the projections. Current spawning biomass in the eastern stock is projected to be $37.7 \%$ of 1988 spawning stock biomass, and the 2013 RBC under the 20:35:48 harvest control rule is 380 t . For the western stock, current spawning biomass is projected to be $66 \%$ of unexploited stock biomass, and the 2013 RBC is 275t.

The current version of the school shark (Galeorhinus galeus) model predicts that catches of up to 250 t allow recovery of the stock, but that 275 t will not. Rebuilding to the limit reference point $\left(\mathrm{B}_{20}\right)$ cannot be achieved in a generation time plus time 10 years ( 32 years) given current levels of catch (176t). Rebuilding in three generation times ( 66 years) can be achieved with future catches of up to 225 t. If the limit reference point is moved from $\mathrm{B}_{20}$ to half $\mathrm{B}_{\text {MSY }}$ (i.e. $\mathrm{B}_{25}$ ), then rebuilding within 32 years would require catches of close to zero; future catches would need to be of the order of 200t in order to achieve rebuilding in 66 years.

An update of the 2010 assessment of deepwater flathead (Neoplatycephalus conatus) was conducted, providing estimates of stock status in the Great Australian Bight at the start of 2013/14. The base-case assessment estimates an unexploited spawning stock biomass $\left(\mathrm{SSB}_{0}\right)$ of $8,921 \mathrm{t}$ and a current depletion of $39 \%$ of $\mathrm{SSB}_{0}$. The 2013/14 RBC under the 20:35:43 harvest control rule is 979 t and the long-term yield (assuming average recruitment in the future) is $1,051 \mathrm{t}$.

Tier 3 calculations use the estimates of total mortality, natural mortality and average recent catches to determine the RBC for the following year. RBCs for alfonsino, John dory, redfish and mirror dory were greater than reference average catches using the Tier 3 rule. Western gemfish, blue grenadier, pink ling, blue-eye trevalla and silver trevally were unable to be assessed using catch curves due to probable dome-shaped selectivity or high recruitment variability.

The Tier 4 harvest control rules are the default procedure applied to species for which only limited information is available; specifically no reliable information on either current biomass levels or current exploitation rates. In 2012 Tier 4 RBCs were only calculated for species that are assessed using the Tier 4 analysis, that is: Blue Eye, Blue Warehou, Inshore Ocean Perch, Offshore Ocean Perch, Redfish, Royal Red Prawns, and Silver Trevally. Among the non-deep water scalefish a total of 18 species with 24 separate Tier 4 analyses were conducted, but these included a number of species for which spatial information was available (blue warehou and mirror dory) leading to analyses for east and west; with an alternative Royal Red Prawn analysis relating catch rates from different mesh sizes. Two fisheries had zero RBCs: blue warehou and redfish.

Among the deep water species the Tier 4 control rule was used to calculate RBCs for the six deepwater fisheries. The target catches were obtained using the total catches
reported outside of the closed areas deeper than 700 m . Reported catches were relatively low in four fisheries so no change could be recommended to the RBC. For mixed oreos the RBC increased slightly from $120-132 \mathrm{t}$. It should be noted that even the standardised catch rates may not reflect changes in stock sizes particularly well. Some of the apparent changes in catch rates exhibited by deep water species are so rapid and so large as to be implausible biologically.

## 27. Appendix: Intellectual Property

No intellectual property has arisen from the project that is likely to lead to significant commercial benefits, patents or licenses.

## 28. Appendix: Project Staff

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[^0]:    Table 13.8. Non-spawning Eastern Gemfish from the SET in depths between $0-600 \mathrm{~m}$, taken by trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr}$ ). Zone:DepCat is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

[^1]:    Table 13.55. BlueEye from the SET and GHT fishery in depths between $200-600 \mathrm{~m}$, taken by Drop Line. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr}$ ). Zone:Month is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

[^2]:    Table 13.69. Silver Warehou from Zones 10 to 50 and depths $0-600 \mathrm{~m}$ by trawl. Statistical model structures used in this analysis. DepCat is a series of 20 metre depth categories.

    | Model 1 | LnCE~Year |
    | :--- | :--- |
    | Model 2 | LnCE~Year+Vessel |
    | Model 3 | LnCE~Year+Vessel+Month |
    | Model 4 | LnCE~Year+Vessel+Month+Zone |
    | Model 5 | LnCE~Year+Vessel+Month+Zone+DepCat |
    | Model 6 | LnCE~Year+Vessel+Month+Zone+DepCat+DayNight |
    | Model 7 | LnCE~Year+Vessel+Month+Zone+DepCat+DayNight+Zone:Month |
    | Model 8 | LnCE $\sim$ Year+Vessel+Month+Zone+DepCat+DayNight+Zone:DepCat |

[^3]:    Table 13.113. Inshore Ocean Perch from zones 10 and 20 in depths $0-200 \mathrm{~m}$ by trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr)}$. Zone:DepCat is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

[^4]:    Table 13.134. Ocean Jackets from Zones 10 to 50 in depths 0 to 300 m by trawl. Total Catch is the total reported in the database, Records is the number of records used in the analysis, CatchT is the reported catch in the area and depth used in the analysis, and Vessels relates to all vessels used in the analysis. Geomean is the geometric mean of catch rates ( $\mathrm{kg} / \mathrm{hr)}$. Zone:DepCat is the optimum model and StDev is the standard deviation relating to the data in the optimum model.

[^5]:    Fspr20 $=0.205$
    Frolen $L=60.995$
    Fspr48 = 0.105
    Hmsy $=64.623$
    Bmsy $=0.228$
    AvLenRef $=45.000$

[^6]:    FO.1 $=0.105$
    Fspr20 $=0.215$
    F2olen $L=49.912$
    Fspr48 $=0.066$
    F48len L $=51.921$
    Fmsy $=0.162$
    Bmsy $=0.256$
    AvLenRef $=44.000$

[^7]:    F0.1 $=0.417$
    Fspr20 $=0.680$
    F2olen $\mathrm{L}=42.947$
    Fspr48 $=0.269$
    F48len $\mathrm{L}=44.427$
    Fmsy
    Bmsy
    $\begin{array}{ll} & 0.286 \\ \text { AvLenRef } & =37.000\end{array}$

[^8]:    Table 20.65 Western Gemfish data for the TIER 4 calculations. Total is the sum of Discards, State, Non Trawl and SEF2 catches. All values in Tonnes. CE is the standardized catch rate for Zones 40 \& 50 in depths $0-600 \mathrm{~m}$, GAB not included (Haddon, 2012). GeoMean is the geometric mean catch rates. Discards are estimates from 1998 to present. The ratio of discards to catch over the 1998 - 2006 period was used to estimate the discards between 1986 and 1997, the proportion of which is the PDiscard.

[^9]:    ${ }^{1}$ Paper presented to Shark RAG November 2012

